

VALIDITY OF THE GENERALIZED LUMPED PARAMETER
HYDROSALINITY MODEL IN PREDICTING IRRIGATION
RETURN FLOW

by

Stephen Gray McLin

Submitted in Partial Fulfillment
of the Requirements for the Degree of
Doctor of Philosophy

NEW MEXICO INSTITUTE OF MINING AND TECHNOLOGY

Socorro, New Mexico

August 1981

ABSTRACT

The generalized lumped parameter model, with a simulated linear or nonlinear subsurface reservoir, is a simplified approach to representing convective transport of a conservative non-reacting solute in a stream connected aquifer system. Since this approach implicitly neglects any hydrodynamic dispersion effects, it is applicable when system-wide temporal responses in groundwater levels and water quality resulting from distributed input stresses are of interest. Although this technique has received relatively little attention in groundwater hydrology, its inherent advantages of reduced structural complexity and input data requirements suggest that it may be a suitable tool in many situations. Furthermore, in both the linear and nonlinear formulations, the requirement for observed or independently computed groundwater pumpage and evapotranspiration rates may be avoided, thereby eliminating those terms that tend to introduce large errors into the typical water balance approach.

Vertical flow effects on the system outflow concentration history resulting from fully or partially penetrating streams and aquifer heterogeneity are examined by utilizing the linearized Dupuit and Laplace distributed models. A hydraulic and mass equivalence between the linear lumped parameter and Dupuit models is established; these analyses produced identical outflow concentration histories for each model. This similitude is extended to include multiple aquifer systems when several lumped models are linked in a parallel configuration. These results indicate that the total mass leaving the Dupuit and lumped multiple systems are identical, while their respective concentration outflow

histories are approximately equivalent. Vertical flow effects on system outflow concentration resulting from varying aquifer geometry are characterized in the Laplace model by using an aspect ratio. This ratio is defined as $n = (L/D)(K_z/K_x)^{1/2}$, where L is the drain half-spacing, D is the aquifer depth, and K_z/K_x is the vertical to horizontal hydraulic conductivity ratio. For values of n greater than about five (and hence, most practical situations), the lumped parameter model approximates the Laplacian concentration break-through curves quite well; however, when n is less than five pronounced vertical flow effects cause significant differences in outflow histories. Vertical flow effects resulting from aquifer stratification are also shown to have an effect on the Laplacian outflow concentration histories. Both the effects of partial stream penetration and aquifer heterogeneity, however, are approximately incorporated into the combined lumped parameter parallel configuration with surprisingly good results by utilizing two empirical correction factors characterizing the aquifer system. These analyses demonstrate that in many situations, pronounced aquifer circulation patterns may not greatly influence system outflow concentrations; furthermore, in those instances where vertical flow effects are important, they may be successfully incorporated into the generalized lumped parameter format.

Finally, the generalized lumped approach is field tested in two separate irrigated valleys in order to demonstrate its general applicability and overall flexibility. The first of these is a linear subsurface reservoir for the Arkansas River Valley of southeastern Colorado over a one year period. Results favorably compare with observed data, and with a more complex convective-dispersive model (Konikow and

Bredehoeft, 1974a, 1974b) previously tested in the same area under similar conditions. A second application with a nonlinear reservoir is tested in the Mesilla Valley of southcentral New Mexico over a ten year period. Results show good to excellent agreement with observed data, and to a model (USBR, 1977) previously tested there (Gelhar and McLin, 1979). Both field applications also demonstrate that the lumped modeling technique can easily be utilized to evaluate the effects of various real and/or hypothetical system stress patterns, including the individual effects of improvements in irrigation efficiency and lining of conveyance canals. These simulations first require a determination of the system's hydraulic and solute response times from known data, and then an estimation of the quantity and quality of applied water from various known sources.

ACKNOWLEDGEMENTS

The author gratefully acknowledges the invaluable suggestions, guidance, and moral support of Professor Lynn W. Gelhar, who served as dissertation advisor throughout this study. He has willingly given both his time and effort to insure the successful completion of this project and for that the author is greatly indebted.

Appreciation is particularly extended to the New Mexico Institute of Mining and Technology for their financial support in the form of research assistantships. Special thanks are due to the dissertation committee members, Dr. Daniel B. Stephens and Dr. Raz Khaleel for their critical review of this research. Further thanks are extended to Dr. Leonard F. Konikow of the USGS Water Resources Division in Reston, Virginia, who graciously supplied field data and simulation results from the USGS Arkansas River Valley study in Colorado.

Above all the author is indebted to his wife, Deborah, and son, Ryan, for making this effort worthwhile.

TABLE OF CONTENTS

	<u>Page</u>
ABSTRACT	ii
ACKNOWLEDGEMENTS	v
LIST OF FIGURES.	ix
LIST OF TABLES	xvi
LIST OF SYMBOLS.	xvii
CHAPTER 1. INTRODUCTION	1
1.1 Background	1
1.2 Review of Previous Work.	3
1.3 Purpose and Scope of Present Research.	11
CHAPTER 2. THEORY OF OUTFLOW CONCENTRATION FROM LINEAR PHREATIC AQUIFERS.	14
2.1 Introduction	14
2.2 Generalized Lumped Parameter Model	15
2.2.1 Water Balance Equation	15
2.2.2 Mass Balance Equatuion	21
2.2.3 Linear Reservoir Models In Combination	26
2.3 Linearized Dupuit Aquifer Model.	34
2.3.1 Dupuit Approximation	34
2.3.2 Relationship Between Lumped Parameter and Linearized Dupuit Models	38
2.3.3 Multi-Layered Dupuit Aquifer System.	44
2.3.4 Relationship Between Combined Lumped Parameter Model and Linearized Multi- Layered Dupuit Aquifer Model	47
2.4 Laplace Aquifer Model.	71
2.4.1 Flow Field Description	71
2.4.2 Convective Outflow Concentration from the Laplace Aquifer Model.	79

TABLE OF CONTENTS (Continued)

		<u>Page</u>
	2.4.3 Effects of Partial Stream Penetration. . . .	86
	2.4.4 Effects of Anisotropic Porous Medium	91
	2.4.5 Impermeable Basement at Infinite Depth . . .	95
	2.4.6 Multiple Aquifer Systems	98
	2.4.7 Relationship Between Laplace and Combined Lumped Models.	131
	2.5 General Conclusions.	149
CHAPTER	3. FIELD APPLICATION OF THE LINEAR RESERVOIR MODEL. . .	153
	3.1 Introduction	153
	3.2 Arkansas River Linear Reservoir Model.	157
	3.3 Parameter Estimation Procedure	165
	3.4 Comparative Study Results.	171
	3.5 Water Management Options with The Linear Model . . .	179
	3.6 Conclusions.	195
CHAPTER	4. FIELD APPLICATION OF THE NONLINEAR RESERVOIR MODEL .	198
	4.1 Introduction	198
	4.2 USBR-EPA Hydrosalinity Model	201
	4.3 Mesilla Valley Nonlinear Reservoir Model	210
	4.4 Comparative Study Results.	227
	4.5 Water Management Options with The Nonlinear Model. .	228
	4.6 Conclusions.	262
CHAPTER	5. SUMMARY, CONCLUSIONS, AND RECOMMENDATIONS.	266
	REFERENCES CITED	272
	APPENDICES	280
	A. Dupuit Model Program	281
	B. Modified U.S. Corps of Engineers Program	292

TABLE OF CONTENTS (Continued)

	<u>Page</u>
C. Linear Reservoir Model Program, Arkansas River Valley	315
D. Data Summary for Arkansas River Valley Study	325
E. Nonlinear Reservoir Model Program, Mesilla Valley. .	370
F. Estimate of Porosity and River Leakage	423
G. Water Management Options, Mesilla Valley	452

LIST OF FIGURES

<u>Figure</u>	<u>Page</u>
2.1 Schematic representation of an idealized stream connected phreatic aquifer system.	17
2.2 Equation (2.2.15) versus time for several values of c_i	25
2.3 Schematic representation of an idealized horizontally stratified aquifer system with a fully penetrating stream, and where $K_u > K_L$	27
2.4 Schematic representation of linear reservoirs in combination.	28
2.5 Schematic vertical cross-section through a stream connected phreatic aquifer system with uniform accretion	35
2.6 Log (c/c_{aq}) versus t/t_{c1} for case (1)	62
2.7 Log (c/c_{aq}) versus t/t_{c1} for case (2)	63
2.8 Log (c/c_{aq}) versus t/t_{c1} for case (3)	64
2.9 Log (c/c_{aq}) versus t/t_{c1} for case (4)	65
2.10 Log (c/c_{aq}) versus t/t_{c1} for case (5)	66
2.11 Log (c/c_{aq}) versus t/t_{c1} for case (6)	67
2.12 Log (c/c_{aq}) versus t/t_{c1} for case (7)	68
2.13 Schematic diagram for an idealized Laplace aquifer showing one possible streamtube.	74
2.14 Idealized relationship indicating the total dimensionless area swept out by a given dimensionless streamline	83
2.15 Idealized breakthrough curve computed from the area versus streamline relationship.	84
2.16 Program grid network showing boundary conditions	87
2.17 $\bar{A} (\bar{\Psi})$ as a function of $\bar{\Psi}$ for different values of n	88
2.18 c/c_0 as a function of t/t_c for different values of n	89
2.19 Log (c/c_0) as a function of t/t_c for different values of n	92
2.20 Schematic Laplace aquifer with impermeable basement at an infinite depth below the drain	96

LIST OF FIGURES (Continued)

<u>Figure</u>	<u>Page</u>
2.21 Schematic diagram showing the Laplace aquifer configuration for geologic layering.	100
2.22 $\bar{A} (\bar{\Psi})$ vs. $\bar{\Psi}$ for different n ; case (1) in Table 2.3 with $\alpha = 20$	103
2.23 c/c_0 vs. t/t_{c1} for different n ; case (1) in Table 2.3 with $\alpha = 20$	104
2.24 $\text{Log } c/c_0$ vs. t/t_{c1} for different n ; case (1) in Table 2.3 with $\alpha = 20$	105
2.25 $\bar{A} (\bar{\Psi})$ vs. $\bar{\Psi}$ for different n ; case (1) in Table 2.3 with $\alpha = 10$	106
2.26 c/c_0 vs. t/t_{c1} for different n ; case (1) in Table 2.3 with $\alpha = 10$	107
2.27 $\text{Log } c/c_0$ vs. t/t_{c1} for different n ; case (1) in Table 2.3 with $\alpha = 10$	108
2.28 $\bar{A} (\bar{\Psi})$ vs. $\bar{\Psi}$ for different n ; case (1) in Table 2.3 with $\alpha = 5$	109
2.29 c/c_0 vs. t/t_{c1} for different n ; case (1) in Table 2.3 with $\alpha = 5$	110
2.30 $\text{Log } c/c_0$ vs. t/t_{c1} for different n ; case (1) in Table 2.3 with $\alpha = 5$	111
2.31 $\bar{A} (\bar{\Psi})$ vs. $\bar{\Psi}$ for different n ; case (1) in Table 2.3 with $\alpha = 2$	112
2.32 c/c_0 vs. t/t_{c1} for different n ; case (1) in Table 2.3 with $\alpha = 2$	113
2.33 $\text{Log } c/c_0$ vs. t/t_{c1} for different n ; case (1) in Table 2.3 with $\alpha = 2$	114
2.34 $\bar{A} (\bar{\Psi})$ vs. $\bar{\Psi}$ for different n ; case (2) in Table 2.3 with $\alpha = 10$	115
2.35 c/c_0 vs. t/t_{c1} for different n ; case (2) in Table 2.3 with $\alpha = 10$	116
2.36 $\text{Log } c/c_0$ vs. t/t_{c1} for different n ; case (2) in Table 2.3 with $\alpha = 10$	117

LIST OF FIGURES (Continued)

<u>Figure</u>	<u>Page</u>
2.37 $\bar{A} (\bar{\Psi})$ vs. $\bar{\Psi}$ for different n ; case (3) in Table 2.3 with $\alpha = 10$	118
2.38 c/c_0 vs. t/t_{c1} for different n ; case (3) in Table 2.3 with $\alpha = 10$	119
2.39 $\text{Log } c/c_0$ vs. t/t_{c1} for different n ; case (3) in Table 2.3 with $\alpha = 10$	120
2.40 $\bar{A} (\bar{\Psi})$ vs. $\bar{\Psi}$ for different n ; case (4) in Table 2.3 with $\alpha = 10$	121
2.41 c/c_0 vs. t/t_{c1} for different n ; case (4) in Table 2.3 with $\alpha = 10$	122
2.42 $\text{Log } c/c_0$ vs. t/t_{c1} for different n ; case (4) in Table 2.3 with $\alpha = 10$	123
2.43 $\bar{A} (\bar{\Psi})$ vs. $\bar{\Psi}$ for different n ; case (5) in Table 2.3 with $\alpha = 10$	124
2.44 c/c_0 vs. t/t_{c1} for different n ; case (5) in Table 2.3 with $\alpha = 10$	125
2.45 $\bar{A} (\bar{\Psi})$ vs. $\bar{\Psi}$ for different n ; case (5) in Table 2.3 with $\alpha = 100$	126
2.46 c/c_0 vs. t/t_{c1} for different n ; case (5) in Table 2.3 with $\alpha = 100$	127
2.47 $\bar{A} (\bar{\Psi})$ vs. $\bar{\Psi}$ for different n ; case (6) in Table 2.3 with $\alpha = 10$	128
2.48 c/c_0 vs. t/t_{c1} for different n ; case (6) in Table 2.3 with $\alpha = 10$	129
2.49 $\text{Log } c/c_0$ vs. t/t_{c1} for different n ; case (6) in Table 2.3 with $\alpha = 10$	130
2.50 Laplace and combined lumped model c/c_{aq} vs. t/t_{c1} curves (see Figures 2.6 and 2.24).	133
2.51 Laplace and combined lumped model c/c_{aq} vs. t/t_{c1} curves (see Figures 2.7 and 2.27).	134
2.52 Laplace and combined lumped model c/c_{aq} vs. t/t_{c1} curves (see Figures 2.8 and 2.30).	135

LIST OF FIGURES (Continued)

<u>Figure</u>	<u>Page</u>
2.53 Laplace and combined lumped model c/c_{aq} vs. t/t_{c1} curves (see Figures 2.9 and 2.33)	136
2.54 Laplace and combined lumped model c/c_{aq} vs. t/t_{c1} curves (see Figures 2.11 and 2.36)	137
2.55 Laplace and combined lumped model c/c_{aq} vs. t/t_{c1} curves (see Figures 2.12 and 2.39)	138
2.56 Lumped (μ/m) vs. Laplace (γ) model slopes for different values of n	141
2.57 Lumped (μ) vs. Laplace (β) model intercepts for different values of n	142
2.58 Corrected lumped vs. Laplace c/c_{aq} vs. t/t_{c1} curves for $n = 10$	143
2.59 Corrected lumped vs. Laplace c/c_{aq} vs. t/t_{c1} curves for $n = 5$	144
2.60 Corrected lumped vs. Laplace c/c_{aq} vs. t/t_{c1} curves for $n = 4$	145
2.61 Corrected lumped vs. Laplace c/c_{aq} vs. t/t_{c1} curves for $n = 3$	146
2.62 Corrected lumped vs. Laplace c/c_{aq} vs. t/t_{c1} curves for $n = 2$	147
2.63 Corrected lumped vs. Laplace c/c_{aq} vs. t/t_{c1} curves for $n = 1$	148
3.1 Lumped parameter representation of the irrigated stream-aquifer system near La Junta, Colorado	159
3.2 Discretized subscript notation for water balance equation.	161
3.3 Discretized subscript notation for mass balance equation	164
3.4 Changes in mean monthly river flow over the study reach.	172
3.5 Total changes in river TDS between the Ft. Lyon canal diversion and the Bent-Otero County line	173
3.6 Study reach changes in river TDS	174
3.7 Changes in average aquifer water levels in the Arkansas River Valley study area.	175

LIST OF FIGURES (Continued)

<u>Figure</u>	<u>Page</u>
3.8	Changes in average aquifer TDS concentration in the study reach. 176
3.9	Study reach changes in streamflow and surface water quality as a result of extending the lumped model simulation period an additional four years 177
3.10	Study reach changes in groundwater storage and water quality as a result of extending the lumped model simulation period an additional four years 178
3.11	Changes in streamflow and surface water quality relative to the base simulation for no surface water diversion for irrigation 182
3.12	Changes in groundwater storage and water quality relative to the base simulation for no surface water diversion for irrigation 183
3.13	Changes in streamflow and surface water quality relative to the base simulation for no groundwater pumpage for irrigation 185
3.14	Changes in groundwater storage and water quality relative to the base simulation for no groundwater pumpage for irrigation 186
3.15	Changes in streamflow and surface water quality relative to the base simulation for a 20 percent improvement in irrigation efficiency. 189
3.16	Changes in groundwater storage and water quality relative to the base simulation for a 20 percent improvement in irrigation efficiency. 190
3.17	Changes in streamflow and surface water quality relative to the base simulation as a result of canal lining to prevent leakage. 193
3.18	Changes in groundwater storage and water quality relative to the base simulation as a result of canal lining to prevent leakage. 194
4.1	Location of the Mesilla Valley and Upper Rio Grande drainage basin, New Mexico. 199
4.2	Schematic nodal representation showing some features simulated in the USBR-EPA model. 202

LIST OF FIGURES (Continued)

<u>Figure</u>	<u>Page</u>
4.3 RMS error in simulated USBR-EPA aquifer water levels and simulated river outflow chemistry as a function of the consumptive use multiplier	207
4.4 USBR-EPA model results using an irrigation efficiency of 50 percent; the RMS error between observed and predicted output is 344 mg/l	208
4.5 USBR-EPA model results using a hypothetical irrigation efficiency of 75 percent; the observed data are shown for reference only	209
4.6 Lumped parameter representation of the irrigated stream-aquifer system in the Mesilla Valley	211
4.7 Monthly observed and simulated average aquifer water levels using the nonlinear lumped model; the RMS error is 0.18 feet	218
4.8 Lumped parameter monthly simulated and observed average valley drain flow; the RMS error is 0.02 feet per month. . .	219
4.9 Lumped parameter monthly observed and simulated Rio Grande TDS at El Paso, Texas; the RMS error is 213 mg/l	221
4.10 Lumped parameter monthly observed and simulated Rio Grande flow at El Paso, Texas; the RMS error is 7324 acre feet per month.	222
4.11 Lumped parameter monthly predicted average aquifer TDS; only a few months of data were observed	223
4.12 Base period simulation of average aquifer water levels and drain flow in the Mesilla Valley	230
4.13 Base period simulation of average aquifer TDS and net average recharge in the Mesilla Valley	231
4.14 Base period simulation of average river TDS at El Paso, Texas, the Mesilla Valley outflow point.	232
4.15 Base period simulation of average river outflow at El Paso, Texas, the Mesilla Valley outflow point.	233
4.16 Case (1) changes from the base simulation in average aquifer water levels and valley drain flow	237
4.17 Case (1) changes from the base simulation in average aquifer TDS and net average aquifer recharge	238

LIST OF FIGURES (Continued)

<u>Figure</u>	<u>Page</u>
4.18 Case (1) changes from the base simulation in average river TDS at El Paso, Texas.	239
4.19 Case (1) changes from the base simulation in average river outflow at El Paso, Texas.	240
4.20 Case (2) changes from the base simulation in average aquifer water levels and average valley drain flow	244
4.21 Case (2) changes from the base simulation in average aquifer TDS and net average aquifer recharge	245
4.22 Case (2) changes from the base simulation in average river TDS at El Paso, Texas.	246
4.23 Case (2) changes from the base simulation in average river outflow at El Paso, Texas.	247
4.24 Case (3) changes from the base simulation in average aquifer water levels and valley drain flow	251
4.25 Case (3) changes from the base simulation in average aquifer TDS at El Paso, Texas.	252
4.26 Case (3) changes from the base simulation in average river TDS at El Paso, Texas.	253
4.27 Case (3) changes from the base simulation in average river outflow at El Paso, Texas.	254
4.28 Case (4) changes from the base simulation in average aquifer water levels and valley drain flow	258
4.29 Case (4) changes from the base simulation in average aquifer TDS and net average aquifer recharge	259
4.30 Case (4) changes from the base simulation in average river TDS at El Paso, Texas.	260
4.31 Case (4) changes from the base simulation in average river outflow at El Paso, Texas.	261
D.1 Arkansas River Valley study area showing well locations and Theissen weighting polygons.	329

LIST OF TABLES

<u>Table</u>	<u>Page</u>
2.1 Comparison between the combined lumped and linearized Dupuit models.	61
2.2 Comparison between the Laplace model and the generalized lumped parameter model	94
2.3 Cases examined for geological layering effects	101
2.4 Areas under c/c_0 vs t/t_c curves.	132
2.5 Values of γ and β from Figures 2.56 and 2.57 with $K_L/K_U = 0.20$, $m=3$, and $\mu = 0.375$	140
3.1 Data summary for the parameter estimation.	170
3.2 Summary comparison between the five year lumped output and the Konikow-Bredehoeft model output.	180
4.1 Summary of major USBR-EPA model features	205
4.2 Comparison of linear and nonlinear aquifer parameters for the Mesilla Valley	217
4.3 Arithmetic means of observed TDS values.	225
4.4 List of wells used for arithmetic means.	226
4.5 Summary of water management options simulated, showing changes from base period simulation.	263

LIST OF SYMBOLS

The following list covers the symbols which are widely used herein. The dimensions of dimensional quantities are given in parenthesis.

<u>Symbol</u>	<u>Description</u>
a	lumped parameter outflow constant (1/T)
a,b,c,l,m	arbitrary constants
A	horizontal surface area of aquifer (L ²)
b _L	thickness of lower aquifer in a multiple aquifer system, (L)
B	equal to b _L /h ₀
c, c ₀ , c _{aq}	average aquifer solute concentration, (M/L ³)
c _i	average aquifer solute input concentration, (M/L ³)
c _s	average solute concentration of surface waters, (M/L ³)
c _a , c _n	average solute concentration of artificial or natural recharge waters, respectively, (M/L ³)
D	Depth to an impermeable basement below the stream or tile drain, (L)
E	lumped time dependent aquifer water inputs minus outputs, (L ³ /T/L ²)
f(x)	constant of integration with respect to z
h, \bar{h}	average saturated thickness of the lumped model aquifer, (L)
h ₀	water elevation above some reference datum at the stream or drain (L)
h _r	average aquifer water level elevation above some reference datum at time t = 0, (L)
h ₁ , h ₂	average saturated thickness in the upper and lower aquifers, respectively, in a multiple aquifer system, (L)
i, j	vector subscript notation corresponding to the x, z unit vector directions, respectively

LIST OF SYMBOLS (Continued)

<u>Symbol</u>	<u>Description</u>
\bar{J}	the hydraulic gradient vector
k	a rate constant in the mass balance equation, (1/T)
K, K_u, K_L	the hydraulic conductivity; upper and lower K values in a multiple aquifer system, (L/T)
L	aquifer length, or drain half spacing, (L)
n	average effective aquifer porosity
q_a, q_n	artificial or natural aquifer volume recharge rate per unit horizontal surface area, ($L^3/T/L^2$)
q_b	net volumetric aquifer boundary inflow per unit horizontal surface area, ($L^3/T/L^2$)
q_l, q_r	net volumetric canal or river leakage rate per unit horizontal surface area lost to the aquifer, ($L^3/T/L^2$)
q_m, q_p	municipal or irrigation pumpage rate per unit horizontal surface area, ($L^3/T/L^2$)
Q	volumetric flow rate, (L^3/T)
r'	volumetric mass source-sink term in the mass balance equation, ($M/L^3/T$)
r	an aquifer shape factor in a multiple aquifer system, and equal to h_2K_L/h_1K_u
S	average aquifer specific yield, or drainable porosity
t	time, (T)
T	aquifer transmissivity, (L^2/T)
t_c	solute response time of the mass balance equation for the lumped parameter and Laplace models, and equal to nh/ϵ , (T)
t_{c1}	solute response time for the combined lumped parameter or multiple layer Laplace model relative to the upper aquifer, and equal to nh_1/ϵ , (T)

LIST OF SYMBOLS (Continued)

<u>Symbol</u>	<u>Description</u>
t_c'	solute response time for the linearized Dupuit model and equal to nh_0/ϵ , (T)
$t_{c'1}$	solute response time for the linearized multiple layer Dupuit aquifer model, and equal to $n(h_0 + \xi/2)/\epsilon$, or $(1+r)t_{c1}$, (T)
t_h	the hydraulic response time of the water balance equation for the lumped parameter model, and equal to S/a , (T)
V	aquifer saturated pore volume, (L^3)
x, y, z	spatial Cartesian coordinates, (L)
\bar{x}, \bar{z}	average horizontal and vertical particle coordinates, (L)
X, Z	equal to x/L and z/L , respectively
α	the ratio K_u/K_L
β	a constant relating a , T , and L
γ	the ratio of the solute response times from the upper and lower aquifers in a multiple aquifer system
ϵ	average aquifer volumetric recharge rate per unit of horizontal surface area, ($L^3/T/L^2$)
η	an aspect ratio defined as, $(L/D)(K_z/K_x)^{1/2}$
μ	the ratio of the aquifer volumetric flow rate
ξ	an aquifer shape factor relating b_L , K_u , and K_L , (L)
$\hat{\xi}$	equal to $\xi/2h_0 = b_L(K_L - K_u)/h_0K_u$
τ	a dummy integration variable, (T)
Ψ	the stream function, ($L^3/T/L$)
Ψ_u, Ψ_L	dimensionless stream function of the upper and lower aquifers, respectively, in a multiple aquifer system

CHAPTER 1

INTRODUCTION

1.1 Background

In the semiarid regions of the western United States, one of the more common sources of shallow groundwater contamination is irrigated agriculture. Typically these irrigated areas are located along alluvial floodplains where relatively inexpensive, high quality surface water supplies are readily available. Elaborate water storage, delivery, and drainage control systems are usually installed to maximize benefits, while attempting to simultaneously control seepage and soil salinity problems. More expensive shallow groundwater of lower quality may also be utilized when surface supplies become limited. The total applied waters tend to increase in solute concentration because of surface evaporation and plant transpiration; furthermore, a potential for upward fluid and solute movement from the underlying saline water table into the plant root zone may also periodically be created. In these irrigated areas it is evident that soluble salts will eventually accumulate in the plant root zone, resulting in lower crop yields, unless preventative measures are taken.

Since the problem solutes in question are primarily transported into or out of the unsaturated plant root zone in the water phase, their temporal and spatial distributions within this zone are controllable through appropriate water and agricultural management techniques. Historically these management approaches have centered around seepage and salinity drainage control techniques, improvements in water conveyance structures, reductions in phreatophyte and crop evapotranspiration rates,

and on farm irrigation efficiency and scheduling improvements. These measures in turn may have associated environmental impacts on hydraulically connected ground and surface water supplies.

In order to evaluate these impacts, many analytical, electric analog, sand tank, and computer based modeling efforts simulating the simultaneous transport of water and solutes through the unsaturated zone have evolved over the past several decades. While these approaches have reached an advanced level of detailed refinement, the saturated flow regime may actually dominate in a systems approach sense. Saturated multidimensional convective-dispersive groundwater models also exist for this situation, but sufficient observed data are seldom adequate to determine the necessary model parameters for site specific applications. Simpler lumped parameter modeling efforts which are more consistent with data availability have been proposed and utilized with some success; these models are not widely accepted because their reliability has not been adequately demonstrated. Hence, there is a need to evaluate the implied limitations and field applicability of lumped models to irrigated stream-aquifer systems. If this reliability can be established, such an approach will offer an important practical tool for management of large scale irrigation systems.

In view of Public Law 92-500 (Federal Water Pollution Control Act Amendments of 1972), U.S. international treaty agreements, and socioeconomic considerations, the lumped modeling approach may prove to be a useful tool in areally extensive river basins where irrigated agriculture is fully developed or still evolving. Hence, the environmental impact of established or newly developed irrigation related water management schemes in one or more subbasins can be evaluated from a

systems oriented approach. For example, if improvements in irrigation efficiency were proposed for an upstream subbasin in the Colorado River or Rio Grande Basins, the resulting shallow ground and surface water alterations at Yuma or El Paso could be simulated before extensive capital expenditures were made. Actually a variety of management options at a number of subbasins within the entire watershed could be easily computed. Furthermore, the technical implications of local, regional, and international legal obligations could be evaluated. Ideally optimal solutions could then be selected so as to maximize basin wide benefits while minimizing undesirable effects.

1.2 Review of Previous Work

We begin this review with a historical discussion on research developments which enable a conceptualization of water and mass transport phenomena in irrigated stream-aquifer systems. The physical basis underlying the various subsurface modeling efforts are briefly highlighted. These basic principles will suggest an appropriate solution technique and determine the required input data for a given approach; furthermore, the intended use of each model type should become apparent. Finally for purposes of specific comparison, we mention several hydrosalinity return flow field simulation efforts utilizing these different methods. The need for and applicability of the generalized lumped parameter model will hopefully become evident in this review.

A great deal of research emphasis in irrigation return flow modeling has focused on the flow and chemical characteristics in the near-surface unsaturated zone above the water table. It is here that the majority of

salinity increases associated with return flow occurs. Evapotranspired waters cause any original contaminants present to further concentrate; these constituents are free to biologically or chemically interact with soil organisms, soil minerals, or interstitial waters, usually resulting in even further degradation. Early physical descriptions of the steady convective leaching, or flushing, phenomenon were developed by Gardner and Brooks (1957), even though dispersion and adsorption effects were already known to be simultaneously operative in laboratory column experiments (Lapidus and Amundson, 1952). Other models developing along this trend continued to emphasize the simultaneous convective flow of solutes and water in one-dimensional unsaturated flow (Bresler, 1967; Bresler and Hanks, 1969; and Freeze, 1969a), but for the unsteady case. Warrick, et al., (1971) improved upon these techniques by including dispersion effects, even though they used a constant dispersion coefficient in the mass transport equation; furthermore, they assumed steady-state conditions would be quickly established around the moving solute front. Later Warrick, et al., (1972) allowed the dispersion coefficient to vary as an empirical function of time in order to account for unsteady flows. Bresler (1973a; 1973b) and Kirda, et al., (1973) extended the analysis to transient unsaturated convective-dispersive transport without these earlier limitations. Bresler's (1973b) numerical approach nicely combined explicit-like efficiency, unconditional stability, and second order accuracy. Wilson and Gelhar (1974) have also overcome some of the previous conceptual and mathematical limitations inherent to earlier unsaturated models by analytically relating the dispersion coefficient to moisture content, seepage velocity, and molecular diffusion.

Solute adsorption and ion exchange reactions in the unsaturated zone were considered separately from convective-dispersive effects at first (Hiester and Vermeuler, 1952; and Dutt, 1962). However comparative studies of several cation adsorption and dispersion models later stressed the coupled importance of these and other possible effects (Biggar and Nielsen, 1963; and Hornsby, 1973). As a result, Dutt, et al., (1972) improved upon the earlier models by including individual soil layer plant water uptake, and chemical and biological effects in their one-dimensional considerations. Other one-dimensional modeling efforts also began to include chemical effects (Bresler, 1973b; and Nimah and Hanks, 1973), but in a convective-dispersive framework. Recent refinements (Bresler, 1977; and Hanks, et al., 1977) include a plant water uptake distribution function in the flow equation, and a source-sink rate term in the mass balance equation. The relative importance between these two terms on aquifer contaminant inputs appears to be site dependent (Hanks, et al., 1977). In addition, the specific form of the plant water uptake function is reported to be more important than variations in either soil hydraulic properties or surface contaminant inputs in causing variations in aquifer contaminant inputs (Jury, et al., 1977).

Without exception input data requirements for these and other unsaturated models grow in direct relation to model complexity; for example, a chemical simulation model might require spatial and temporal distributions of dissolved carbon dioxide, among other data, to realistically represent the unsaturated interval. Without such information, an observer might wonder exactly what a given model is simulating. Furthermore, all of the previously mentioned models stop

short of actually predicting time history concentration fluctuations in hydraulically connected ground or surface waters. Instead these models are designed to simulate solute fluctuations leaving the unsaturated zone and entering a shallow water table.

Not surprisingly, a corresponding amount of research effort has also been directed towards saturated flow through porous media. Early analytical and numerical modeling techniques were developed to simulate steady saturated convective transport (Warrick and Kirkham, 1969; Freeze, 1969b; and Cooley, 1970), but were quickly extended to include the unsteady case as well (Prickett and Lonquist, 1971; Cooley, 1974; and Trescott, et al., 1976). Most recent efforts have concentrated on improving numerical procedures, although Flores W., et al., (1978) have deviated from the more traditional deterministic approach (as summarized by Prickett, 1976; and Gelhar, 1976) in their development of a stochastic water management model for a stream-aquifer system.

Saturated convective-dispersive transport models developed alongside of the saturated hydraulic models (Shamir and Harleman, 1967; and Guymon, 1970), and also initially used the steady flow approximation. This limitation was overcome, however, when Reddell and Sunada (1970) used the method of characteristics in solving the coupled, unsteady convective-dispersive equation, while Guymon, et al., (1970), Huyakorn (1977), and Huyakorn and Nilkula (1979) have suggested applications utilizing the finite element technique. Reactive solute effects were included in early one-dimensional models (Lai and Jainak, 1972; Gupta and Greenkorn, 1973; and Rubin and James, 1973). Various simplifying assumptions made primarily for mathematical or numerical convenience, however, limited many of these efforts. These limitations centered

around the form of the dispersion coefficient, model initial and boundary conditions, various empirical approximations to a cation adsorption function or other fluid-solid disturbances, and the numerical method selected for solution. Schwartz and Domenico (1973) in their refinements, attempted to solve the two-dimensional mass transport problem considering chemical and biological interactive effects for the Upper Kettle Creek, Ontario, Canada, using the steady flow model described by Freeze (1969b).

Some model interfacing between the unsaturated and saturated flow zones has occurred. Freeze (1971) developed a convective three-dimensional transient solution for an entire drainage basin that considers aquifer heterogeneity, compressibility, and soil hysteresis. He has extended this approach to contaminant migration from subsurface waste disposal sites (Freeze, 1972), but does not consider dispersion or hydrochemical reactions. Perez, et al., (1972) presented an overview of a conjunctive use model for a surface-groundwater system. They represented the unsaturated zone by a one-dimensional analytical solution and the groundwater flow by a steady-state solution for each time step. Pikul, et al., (1974) chose to link the solution for one-dimensional vertical unsaturated flow to a one-dimensional horizontal saturated flow condition. Again, none of these models considered simultaneous mass transport by the dispersion process, or any chemical interaction effects. Some of these efforts may be conceptually objectional, but they do attempt to address the problem of surface solute introduction and subsequent convective transport through the groundwater flow regime. Furthermore, recent refinements of this approach (Neuman, et al., 1974) now consider transient, three-dimensional convective-dispersive transport

with slight medium compressibility, plant water uptake in the unsaturated zone, and well withdrawal from the saturated zone. Finally it should be noted that these works suggest the relative importance of the unsaturated zone may sometimes be small in relation to other factors contributing to return flow degradation.

None of the models previously mentioned relates time history fluctuations in contaminant inputs to hydraulically connected surface water outputs. While these spatially distributed convective-dispersive models will directly yield unsaturated and/or saturated concentration profiles for the region of interest, the impact of various water and agricultural management practices has not been directly evaluated from any of these models. Instead these approaches have been used to provide insight on the relative importance of several transport mechanisms involved in irrigation-related water quality degradation.

In irrigated stream-aquifer situations, groundwater contaminant inputs may be regarded as being uniformly distributed throughout the system. In such cases it would appear that hydrodynamic dispersion in saturated media is only a second order effect since adjacent streamlines usually do not show large velocity variations in space or time. Gelhar and Collins (1971) attempted to quantify the relative importance of longitudinal dispersion effects; they used an approximate analytical technique for steady, nonuniform flows with variable dispersion coefficients. Although they did not consider natural nonhomogeneity and anisotropy effects on dispersion, they concluded that accelerating (or converging) flows showed greater dispersion than decelerating (or diverging) flows. One might also logically suspect that transverse dispersion would become dominant over that in the longitudinal direction

when a contaminant front approached a shallow irrigation drain, since convective streamline patterns may begin to move nearly parallel to the front. This may indeed be the case of this idealized example, but Eldor and Dagan (1972) have shown rather convincingly that hydrodynamic dispersion tensor effects (i.e., including both the longitudinal and transverse components) have very little overall effect on contaminant output fluctuations. They examined a two-dimensional flow field with fully penetrating streams under steady uniform recharge. Hence, the relative importance of hydrodynamic dispersion in irrigated stream-aquifer systems appears to be small in comparison to convective effects; this conclusion is obviously of major importance. Field testing of a two-dimensional convective-dispersive model in an irrigated stream-aquifer system in the Arkansas River Valley of Colorado further demonstrated that hydrodynamic dispersion effects were not a major contributor to contaminant mixing in the aquifer (Konikow and Bredehoeft, 1974a, 1974b). These authors concluded that changes in dissolved solid concentrations in shallow groundwaters resulted primarily from convective transport, and by the mixing of recharged waters of different quality. The model developed by them provides flexibility in water planning and management options as demonstrated in the above references, in addition to its application in the San Luis Rey River Basin, California (Helweg and Labadie, 1976; and Labadie, et al., 1976). This model does require extensive input data and access to a digital computer, however.

Recently several workers have begun to re-examine solute travel times along convective stream patterns resulting from typical irrigated stream-aquifer situations. Two-dimensional fluid displacement processes were previously examined analytically for the case of an infinite plane

source to a line sink by Muskat (1946), and for tile drain displacement patterns by Luthin, et al., (1969), and Ortiz and Luthin (1970). More recently, Miyamoto and Warrick (1974a, 1974b) have investigated steady piston displacement of salts into tile drains from surface ponding, using mathematical and sand tank modeling techniques. Jury (1975) numerically integrated the analytical solution of Kirkham (in Kirkham and Powers, 1972, p. 115) in order to calculate relationships between solute travel times and distances from tile drains. Field comparisons of chloride effluent for a ponded leaching experiment were predicted rather well by his model. Raats (1975) developed a conceptually similar technique for unsaturated-saturated solute flow. His method basically involved tracking space-time trajectories of parcels of water for a given flow system, assuming the solute follows this same saturated zone trajectory as the water parcel (i.e., neglecting hydrodynamic dispersion). He has also considered several complicating phenomena in the unsaturated zone; that is, evapotranspiration, hydrodynamic dispersion, adsorption, and dissolution-precipitation reactions for the one-dimensional transient case.

A lumped parameter hydrologic model developed by Riley, et al., (1966) was coupled to a chemical model (described by Tanki and Doneen, 1966; Tanji, et al., 1967; and Dutt, et al., 1970) by Thomas, et al., (1972). An electric analog simulation was used for the hydraulic model, while a digital simulation was utilized for chemical equilibrium reactions within the unsaturated zone. This model was applied to an irrigated area of the Little Bear River Basin of Utah, and showed excellent agreement between observed and predicted monthly TDS outflow history. The model also showed good to excellent histories for water

outflow and the six dominant ions common to western waters (i.e., Ca, Mg, Na, SO_4 , Cl, and HCO_3). This lumped model (i.e., Thomas, et al., 1972), perhaps more than any other, has come closer to satisfactorily relating surface contaminant inputs to resulting downstream surface outputs. The electric analog portion of the model does make routine applications to other sites more difficult, however.

1.3 Purpose and Scope of the Present Research

A general objective of this research is to develop a simple method to predict future changes in downstream water quality resulting from various alternative management practices in irrigated agriculture. Desirable features of this method should include a minimum of input data requirements and overall flexibility and simplicity of operation, while still physically representing the irrigated stream-aquifer system in a realistic manner. The main features of this work are the deterministic representation of the stream-aquifer system by a lumped parameter linear or nonlinear reservoir model, and a comparison of outflow concentration histories between this so called black-box approach and more complex spatially distributed models. The immediate goal of this work is to formulate a basis of analyzing the individual effects of several complicating phenomena on downstream water quality variations that are not explicitly inherent to the generalized lumped parameter groundwater model. The following specific objectives are established:

- (1) To develop solutions to the problem of outflow concentration history from a homogeneous, isotropic aquifer with fully or partially penetrating streams under steady, saturated, convective transport

flow conditions; and with a steady, uniformly distributed conservative contaminant input;

- (2) To extend the approach of step (1) to non-homogeneous, anisotropic aquifers with and without an impermeable basement at some finite depth;
- (3) To develop solutions to the problem of outflow concentration history for the lumped parameter linear reservoir model in single or parallel configurations, and to compare the results to those obtained in steps (1) and (2);
- (4) To examine the predictive capability of the linear lumped parameter model under actual field conditions in the Arkansas River Valley of southeastern Colorado using various management alternatives, and to compare these results to an existing distributed convective-dispersive transport model (Konikow and Bredehoeft, 1974a; 1974b) previously tested in the same area under similar conditions; and
- (5) To examine the predictive capability of the nonlinear lumped parameter model for a second field test in the Mesilla Valley of southcentral New Mexico, and to compare these results to an existing hydrosalinity model (U.S. Bureau of Reclamation, 1977) previously tested in the same area (Gelhar and McLin, 1979) under similar conditions.

The generalized lumped parameter groundwater model is a simplified approach to an overall water and mass balance for a given contaminant transported in a stream-aquifer system. The distinctive feature of this method is that spatial coordinates of variables are not required in the problem formulation and solution. The corresponding problem dimensionality is reduced; variables are considered functions of time

only. When system parameters are expressed in this lumped form, the total system is regarded as a single point in space. These descriptive parameters relate the nature of system input to system output. Parameters are determined from known aquifer properties, or can be estimated from system averaged water levels and water quality changes over time by standard regression methods. Once these parameters are determined, application of the lumped model can provide a technique for future prediction of downstream water quantity and quality changes. Access to a digital computer is not required; calculations can be conveniently made on a programmable pocket calculator. These same parameters can also be used to estimate downstream water quantity and quality changes resulting from various management options in the system operation, if system inputs can be estimated.

CHAPTER 2

THEORY OF OUTFLOW CONCENTRATION FROM LINEAR PHREATIC AQUIFERS

2.1 Introduction

According to the U.S. Soil Conservation Service (1972, p. 18), "the purpose of subsurface drainage is to lower the water table to a point where it will not interfere with plant growth and development. The minimum depth at which the water level should be maintained varies according to both the crop requirement and the soil. One of the principal factors in the height of the water table in arid areas is the control of salinity and alkalinity in the soil and ground water." Over the past several decades considerable research effort has been directed towards developing quantitative drain spacing criteria based on steady and unsteady flow theories (Glover, 1974; and references therein) in order to control the water table elevation in time and space. In more recent years researchers have begun to focus their attention on the simultaneous transport of solutes in and below the plant root zone. Processes that have been used to describe this solute transport include time and/or spatially dependent flow models, often coupled in such a manner so as to include dispersive mixing and hydrochemical effects. Several such recent modeling efforts are classified and compared by Gelhar (1976). In irrigation return flow water quality studies, however, convective transport seems to dominate other solute transport mechanisms in the saturated zone (Eldor and Dagan, 1972; Konikow and Bredehoeft, 1974a), although hydrochemical and dispersive mixing effects may be equally

important in many areas (Schwartz and Domenico, 1973; Shaffer, et al., 1976).

In this chapter the physical and mathematical basis for several saturated, convective transport models are described; these include the generalized lumped parameter model, the linearized Dupuit model, and the Laplace model. Methods of obtaining solute outflow concentration histories from each model are derived. These outflow concentration histories can serve as a basis of model comparison in the problem of water quality prediction of irrigation return flow. Hopefully, advantages and disadvantages of each of these modeling techniques will become apparent, and subsequent potential users will have an additional basis for model selection.

2.2 Generalized Lumped Parameter Model

2.2.1 The Water Balance Equation

The simplest type of model that can be used to describe convective transport in a dynamically connected stream-aquifer system is the lumped parameter model. Structurally it is simple because this type of model treats only system averaged input-output-storage changes over time. Since variables are averaged over space, they are considered independent of spatial coordinates. Thus only the mean temporal fluctuations of these variables are used to describe the physical nature of a given system. This type of modeling approach is applicable when system wide responses in groundwater levels or water quality resulting from distributed input stresses are of interest. This technique has received

relatively little attention in subsurface hydrology (Kraijenhoff van deLeur, 1958; Dooge, 1960; van Schilfgaard, 1965; Eriksson, 1970; and Eliasson, 1971), but renewed interest is evident (Gelhar and Wilson, 1974; Mercado, 1976; Flores W. and Gelhar, 1976; and Updegraff and Gelhar, 1978). Hydrochemical interaction effects between the flowing groundwater and the porous media may also be included (Thomas, et al., 1972), though additional complexity is introduced.

The structural simplicity of the generalized lumped parameter model is inherently related to the systems operation approach it takes; the system is described only to the degree that it relates averaged inputs to averaged outputs (Dooge, 1973). Thus for the stream connected phreatic aquifer system depicted in Figure 2.1, a simple water balance equation can be written as

$$S \frac{dh}{dt} = q_n + q_a - q_o - q_p \quad (2.2.1)$$

where

S	=	average aquifer specific yield (drainable porosity),
$h(t)$	=	average thickness of the saturated zone,
$q_n(t)$	=	natural aquifer volume recharge rate per unit of horizontal aquifer area,
$q_a(t)$	=	artificial aquifer volume recharge rate per unit of horizontal aquifer area,
$q_o(h,t)$	=	natural volume outflow rate from the aquifer per unit of horizontal aquifer area,
$q_p(t)$	=	volume pumping rate per unit of horizontal aquifer area,
t	=	time.

Multiplying each term of (2.2.1) by the horizontal aquifer area, one can see that changes in aquifer storage (left hand side of (2.2.1)) are given by temporal changes in volume inputs minus outputs (right hand side of (2.2.1)). All of the terms in the water balance equation are functions of time only, except for the aquifer outflow term, q_o , which is a

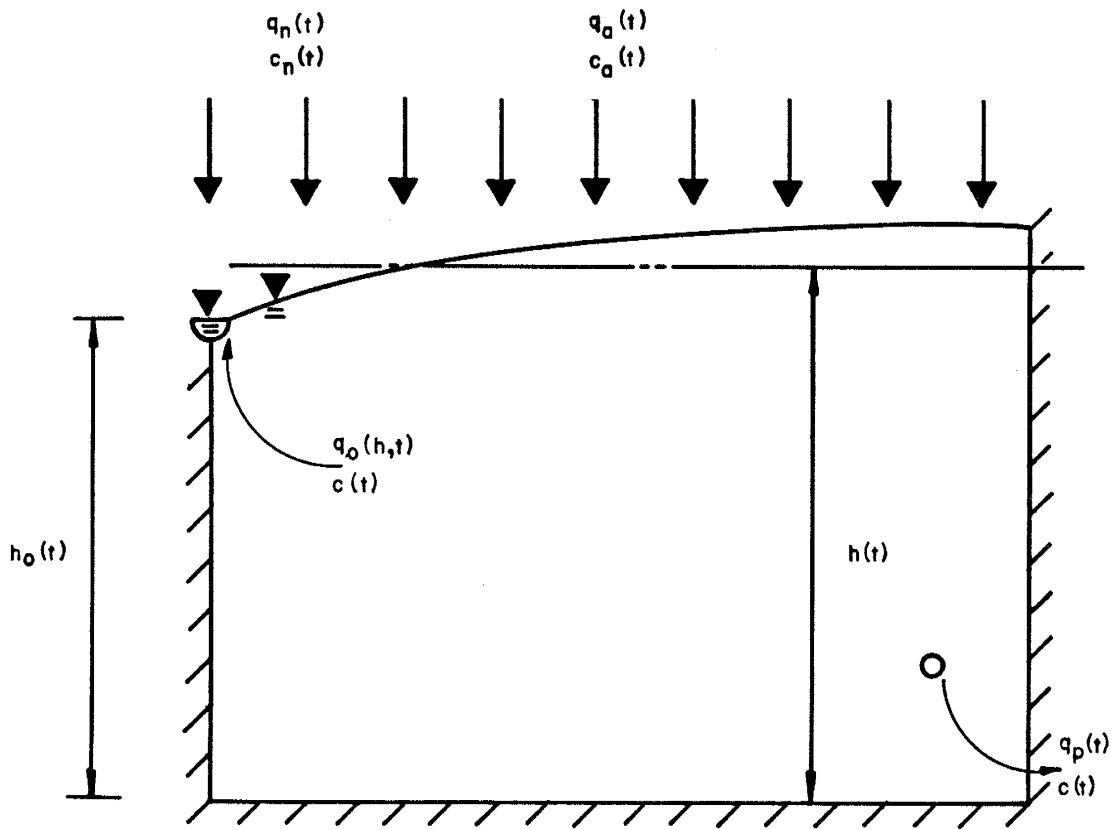


Figure 2.1 Schematic representation of an idealized stream connected phreatic aquifer system.

function of h and t . In general, any number of identifiable input/output terms can be added to the water balance equation without altering its basic form.

The natural outflow rate from the aquifer has been approximated by a linear relationship (Gelhar and Wilson, 1974) given by

$$q_0 = a(h-h_0) \quad (2.2.2)$$

where the parameter a is a lumped outflow constant having units of inverse time and h_0 is the stream reference level. When the average aquifer water level, h , falls below h_0 the flow direction will change according to (2.2.2) and the stream will lose water to the aquifer. Physically, however, when this occurs the stream may no longer have any water to maintain this reversed flow situation. In such cases q_0 must be constrained to values greater than or equal to zero. In general h_0 can be a function of time, but is usually assumed to be constant since fluctuations in h_0 usually are much smaller than those for h . The model represented by (2.2.1) with the natural outflow given by (2.2.2) is a lumped parameter model in the form of a well-mixed linear reservoir. This "well-mixed" assumption will be developed more fully in a later section.

The linear outflow expression given by (2.2.2) is inherently related to a linear form of Darcy's law. As such, the lumped parameter model approximated by a linear reservoir outflow term is intuitively related to some of the simple distributed flow models more frequently encountered in the literature. Quantitative expressions describing this relationship have been developed for several specific cases (Gelhar and Wilson, 1974;

and Flores W. and Gelhar, 1976) which solidify this intuitive connection between the parameter a and certain other aquifer parameters used to characterize distributed systems. One such expression may be written as

$$a = \beta T / L^2 \quad (2.2.3)$$

where T = aquifer transmissivity,
 L = aquifer length,
 β = some constant.

The manner in which β in (2.2.3) is affected by several particular Dupuit type aquifer systems has been explored in some detail (Flores W. and Gelhar, 1976; Flores W., et al., 1978). They considered the steady flow effects on β while varying the recharge distribution and aquifer boundary geometry. They also have considered the unsteady flow effects on β resulting from a declining sinusoidal water table and reported an overall range for β from 1.07 to 8.0; these results essentially verified and expanded upon those of Gelhar, et al., (1974), who used a spectral analysis approach rather than a deterministic one. The physical basis that completes the intuitive coupling of the lumped parameter model and the linearized Dupuit aquifer will be explored in more detail later.

Although the above discussion shows that in principle the parameter a may be estimated analytically, there are problems in its practical application. While both β and T in (2.2.3) may often be estimated without difficulty, the aquifer length L , usually being uncertain, has an important effect on the parameter a . Updegraff and Gelhar (1978) were able to avoid this problem in a field application for the Mesilla Valley of New Mexico. They used a standard least squares linear regression technique to find the parameters a and h_0 in (2.2.2), since values for

q_0 and h were known. Once these lumped aquifer parameters are determined for a given system, (2.2.1) and (2.2.2) can be utilized in several fundamental ways. These are: (1) finding temporal fluctuations in h , knowing system inputs and outputs, (2) finding system inputs, knowing outputs and fluctuations in h , and (3) finding system outputs, knowing inputs and fluctuations in h .

Substituting (2.2.2) into (2.2.1) yields a linear first-order, ordinary differential equation of the form

$$S \frac{dh}{dt} + a(h - h_0) = E \quad (2.2.4)$$

where E represents the lumped time dependent inputs minus outputs. The general solution for (2.2.4) with a constant h_0 is

$$h - h_0 = (h_R - h_0) \exp(-at/S) + \frac{1}{S} \int_0^t E(\tau) \exp[-a(t-\tau)/S] d\tau \quad (2.2.5)$$

where h_R is the average aquifer water level at the starting time $t = 0$, and τ is a dummy variable of integration. The parameter S/a can be referred to as the hydraulic response time (t_h) of the system since it characterizes the average response time of the water balance equation. We can therefore rewrite (2.2.5) in the form

$$h - h_0 = (h_R - h_0) \exp(-t/t_h) + \frac{1}{S} \int_0^t E(\tau) \exp[-(t-\tau)/t_h] d\tau \quad (2.2.6)$$

where $t_h = S/a$.

2.2.2 The Mass Balance Equation

Extending the approach of the previous paragraphs, we can write a mass balance equation for the phreatic aquifer system depicted in Figure 2.1 for some given contaminant of average solute concentration $c(t)$. Thus

$$n \frac{d(hc)}{dt} = q_n c_n + q_a c_a - q_o c - q_p c + nhr' \quad (2.2.7)$$

where n is the effective porosity, c_n and c_a are the average solute concentrations of the natural and artificial recharge waters, respectively, and r' is a volumetric source-sink term that accounts for contaminant additions or degradation within the flow zone; it has units of mass of contaminant per unit volume of water per unit of time. All remaining terms in (2.2.7) have been previously defined. The mass balance equation, like the water balance equation, represents a conservation of mass statement relating changes in mass storage to changes in net mass input. If (2.2.7) is multiplied by the horizontal aquifer area, as was done for the water balance equation, we can see that the mass balance relationship simply becomes a volumetric mass balance statement.

The major underlying assumption in (2.2.7) concerns the concentration of outflow waters; all outflows from the system are assumed to carry the same mass concentration. It is thus implied in (2.2.7) that the aquifer is a well-mixed reservoir and that any system outflow carries this average aquifer concentration. This assumption has been related to the nature of two-dimensional convective transport by Gelhar and Wilson (1974); it will be further discussed in the next section.

It is evident that the source-sink term, r' , in the mass balance equation may be of considerable importance in specific modeling applications. Two widely used forms for r' are discussed by Bear (1972). The first case is that of a contaminant undergoing radioactive decay (or a first order decay). The change in the solute concentration resulting from such a decay can be given by

$$r' = -kc \quad (2.2.8)$$

where k is some rate constant for the contaminant; this form for r' has been used by Przewlocki and Yurtsever (1974) and Rabinowitz, et al. (1976). The second form commonly used for r' is a linear adsorption isotherm which relates the contaminant concentration in the liquid phase to that on the solid matrix. This relation can be written as

$$r' = -\frac{(1-n)}{n} b \frac{\partial c}{\partial t} \quad (2.2.9)$$

where n is the medium porosity and b is some proportionality constant; this form for r' has been used by Robertson (1974), Pickens and Lennox (1976), and Willis (1976). Other more complex forms for r' have occasionally been used; these complex forms generally approach the problem from a chemical equilibrium (Thomas, et al., 1972; Shaffer, et al., 1976) or chemical kinetics (Schwartz and Domenico, 1973; Polciauskas and Domenico, 1976) point of view.

Once all net inputs in the mass balance equation are identified, a solution may be sought. Expanding the left-hand side of (2.2.7) yields

$$nc \frac{dh}{dt} + nh \frac{dc}{dt} = q_n c_n + q_a c_a - q_o c - q_p c + nhr' \quad (2.2.10)$$

Multiplying (2.2.1) by c , substituting the resulting expression for the first term of (2.2.10), and assuming that $S = n$ gives,

$$\frac{dc}{dt} + c \left(\frac{\epsilon}{nh} + k \right) = \frac{\epsilon}{nh} c_i \quad (2.2.11)$$

where

$$\epsilon = q_n + q_a \quad (2.2.12)$$

$$c_i = \frac{q_n c_n + q_a c_a}{\epsilon}$$

and where a first order decay process is used to account for the source-sink effects. The expressions in (2.2.12) are equivalent to a volumetric mixing of inputs and serve to simplify the mass balance equation. Note that the linear outflow relationship given by (2.2.2) need not be used in arriving at (2.2.11). For steady flow, the general solution for (2.2.11) is given by

$$c = c_{aq} \exp [-(\epsilon/nh + k)t] + \int_0^t (\epsilon/nh) c_i(\tau) \exp [-(\epsilon/nh + k)(t-\tau)] d\tau \quad (2.2.13)$$

where c_{aq} is the average aquifer concentration at time $t = 0$. If we assume a constant recharge ($\epsilon = \text{some constant}$) and a step change in input concentration ($c_i = \text{some constant}$) at time $t \geq 0$ (and zero otherwise), the explicit solution for (2.2.13) is

$$c = c_{aq} \exp [-(\epsilon/nh + k)t] + \left(\frac{\epsilon}{\epsilon + nhk} \right) c_i \left\{ 1 - \exp [-(\epsilon/nh + k)t] \right\} \quad (2.2.14)$$

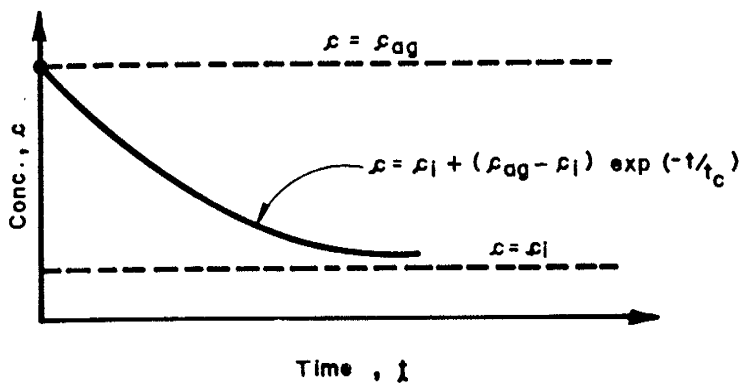
For a conservative solute, k is equal to zero. Thus (2.2.14) reduces to

$$c = c_i + (c_{aq} - c_i) \exp(-t/t_c) \quad (2.2.15)$$

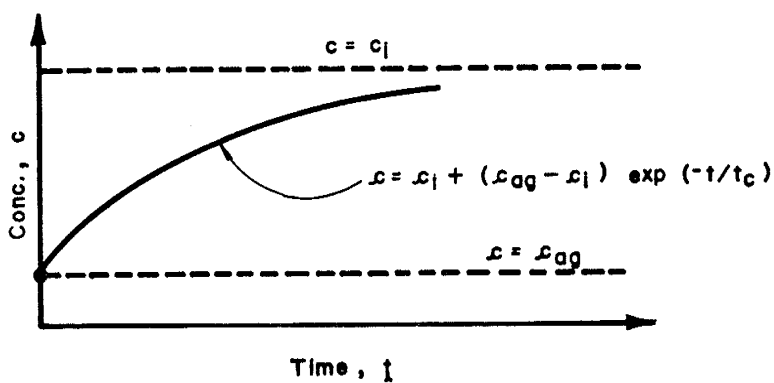
where $t_c = \frac{nh}{\epsilon}$ is the response time for the mass balance system; t_c may be referred to as the solute response time since it characterizes the average residence time of the contaminant $c(t)$ in the aquifer. Figure 2.2 illustrates the general behavior of (2.2.15).

The water and mass balance statements given by (2.2.1) and (2.2.7), together with the linear outflow approximation given by (2.2.2), form a coupled systems description of subsurface water and mass transport in a lumped parameter format. These equations are easily manipulated to supply a variety of explicit analytical solutions for a relatively wide range of specified initial conditions. Furthermore, the system equations are amenable to very simple numerical solution which may be conveniently carried out on a programmable pocket calculator. Examples of both analytical and numerical solutions may be found in Gelhar and Wilson (1974), Flores W., et al., (1978), and Updegraff and Gelhar (1978).

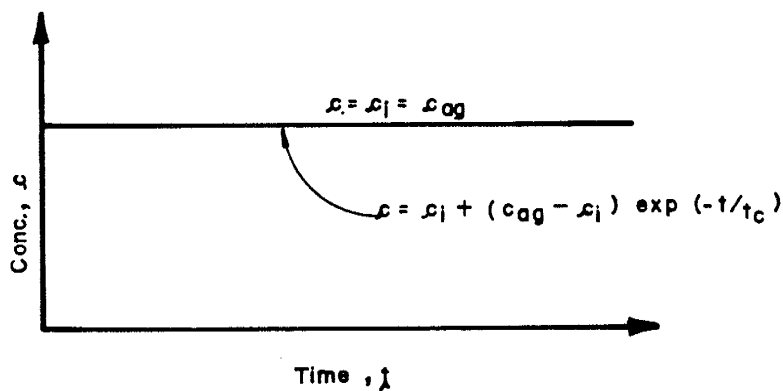
Although the linear outflow assumption made in the water balance equation is the simplest and perhaps most universal approach, the generalized lumped parameter modeling technique certainly is not limited to linear systems. Pinzon (1978) has utilized a nonlinear outflow approximation in a numerical solution of the water balance equation. He obtained recharge estimates for a large irrigated area in New Mexico over the thirty year period 1946-1975. His nonlinear model more accurately simulated time dependent drain flows and average water level fluctuations than did a linear reservoir model tested for the period 1946-1951



(a) For $c_i < c_{ag}$



(b) For $c_i > c_{ag}$



(c) For $c_i = c_{ag}$

Figure 2.2 Equation (2.2.15) versus time for several values of c_i

(Updegraff and Gelhar, 1978). Nonlinearity will be more fully addressed in a later chapter.

Now that some physical basis has been established to describe the lumped parameter model in the form of a well-mixed linear reservoir, we may proceed to expand the lumped parameter modeling technique to multiple reservoir systems.

2.2.3 Linear Reservoir Models in Combination

In certain field situations, the idea of representing physical systems in the generalized lumped format may appear unrealistic, even though net input stresses may be distributed throughout the area of interest. One such example might include the case of horizontally stratified aquifers as depicted in Figure 2.3, where several idealized streamlines could be sketched in as shown. Here K_U is the idealized isotropic, homogenous hydraulic conductivity of the upper layer, while K_L represents some different idealized isotropic, homogenous hydraulic conductivity for the lower layer. The term $\epsilon(t)$ represents the time dependent net input into the aquifer; it is uniformly distributed throughout the system and has units of volume per unit time per unit of surface area.

The simple flow field depicted in Figure 2.3 may be represented by a combination of linear reservoirs as shown in Figure 2.4. The appropriate form of the water balance equation for each linear reservoir shown would be identical to that given by (2.2.4); however, each term in the equation should be subscripted to denote the individual reservoir variables. In the figure all terms except the respective porosities, n_1 , and n_2 , and

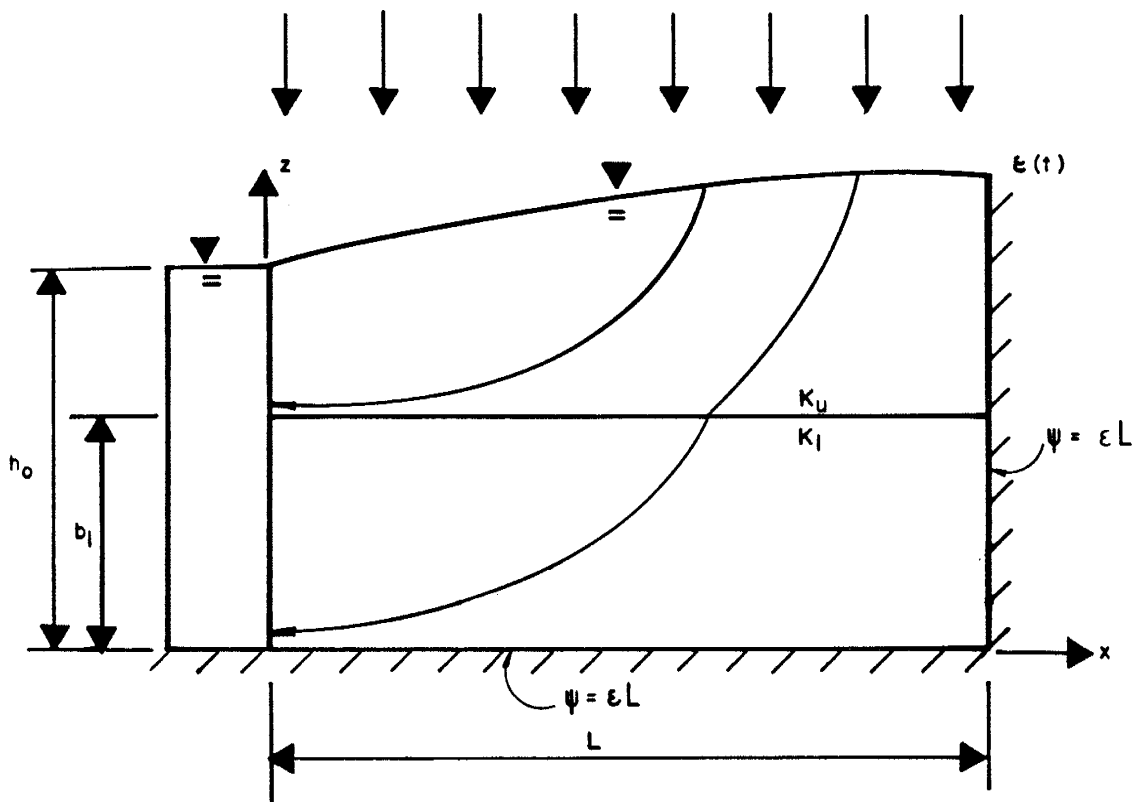


Figure 2.3 Schematic representation of an idealized horizontally stratified aquifer system with a fully penetrating stream, and where $K_U > K_L$.

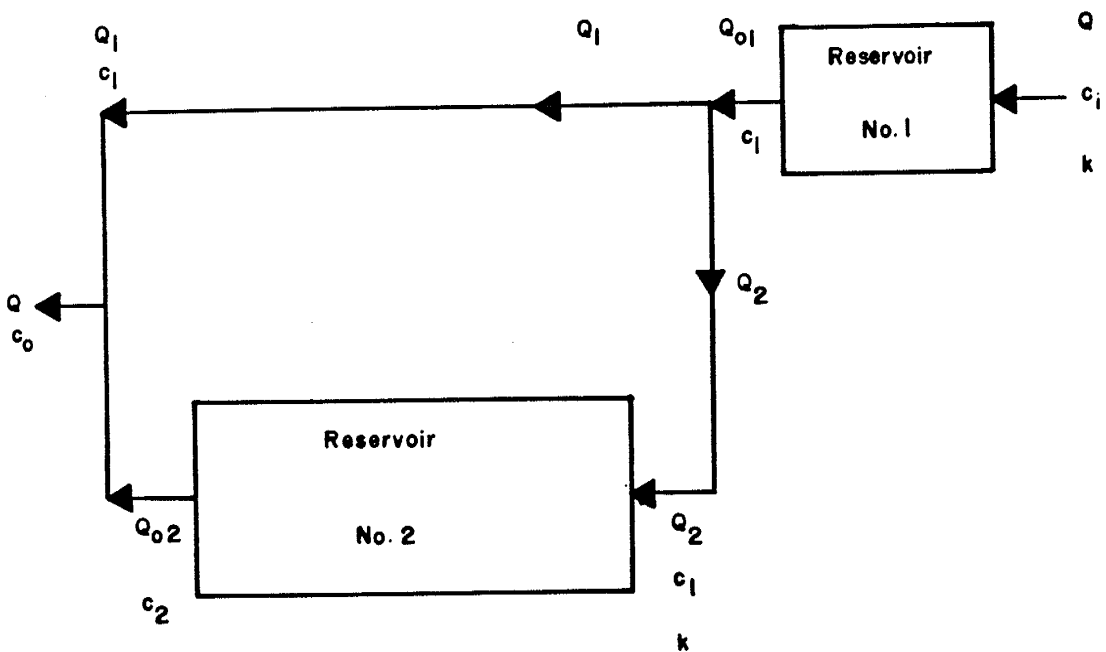


Figure 2.4 Schematic representation of linear reservoirs in combination.

the first order contaminant decay constant, k , are assumed to be functions of time only. As seen in the figure, Q represents the volumetric inflow rate and c_i the contaminant input into the upper aquifer (reservoir no. 1). The outflow from this upper reservoir is Q_{01} , while the corresponding contaminant output is given by c_1 . Only a portion of the outflow from the upper reservoir penetrates into the lower aquifer (reservoir no. 2); the remainder discharges directly into the surface stream. The contaminant input into the lower aquifer is the concentration flowing out of the upper reservoir. The outflow volume and concentration from the lower reservoir are given by Q_{02} and c_2 respectively. Finally, the time history of concentration in the stream can be found by a volumetric mix of concentrations from both reservoirs, noting that the flow rate in the stream is simply the sum of flows from the respective reservoirs.

A mass balance expression can be written for each reservoir depicted in Figure 2.4; the final expression will have the same form as (2.2.11), written below for the upper reservoir only,

$$\frac{dc_1}{dt} + c_1 \left(\frac{\epsilon}{n_1 h_1} + k \right) = \left(\frac{\epsilon}{n_1 h_1} \right) c_i \quad (2.2.16)$$

where $\epsilon = \frac{Q}{A_1}$ and has units of $(L^3/T/L^2)$,

$A_1 =$ horizontal area of the upper aquifer with units of (L^2) .

All other terms have been previously defined. For steady flow, the general solution to (2.2.16) takes the same form as (2.2.13). Thus

$$c_1 = c_{aq_1} \exp [-(\epsilon/n_1 h_1 + k)t] + \int_0^t (\epsilon/n_1 h_1) c_i(\tau) \exp [-(\epsilon/n_1 h_1 + k)(t-\tau)] d\tau \quad (2.2.17)$$

where c_{aq1} is the average aquifer concentration in the upper reservoir at time $t = 0$, and τ is a dummy variable of integration. Likewise for the lower reservoir, the general solution for steady flow would be

$$c_2 = c_{aq2} \exp[-(\epsilon_2/n_2 h_2 + k)t] + \int_0^t (\epsilon_2/n_2 h_2) c_{i2}(\tau) \exp[-(\epsilon_2/n_2 h_2 + k)(t-\tau)] d\tau \quad (2.2.18)$$

where c_{aq2} = average aquifer concentration in the lower reservoir at time $t = 0$,

$c_{i2}(\tau)$ = input concentration into the lower reservoir,

ϵ_2 = $\frac{Q_2}{A_2}$ and has units of $(L^3/T/L^2)$,

A_2 = horizontal area of the lower reservoir with units of (L^2) .

A simple flushing situation may be used to demonstrate the behavior of the combined linear reservoirs in Figure 2.4. Thus for these conditions we would have:

- (1) $c_i(t) = 0$, for $t \geq 0$,
- (2) $c_{aq1} = c_{aq2}$, at $t = 0$,
- (3) $k = 0$,
- (4) $\epsilon_1(t) = \epsilon_1$, a constant for $t \geq 0$,
- (5) $\epsilon_2(t) = \epsilon_2$, a constant for $t \geq 0$.

The first two conditions above state that at $t = 0$ both reservoirs are uniformly mixed with an identical contaminant. All recharge waters introduced into the system at $t \geq 0$ contain no contamination. The resulting time history of concentration in the stream will reflect the time necessary to flush out all pre-existing aquifer contamination that has accumulated up to $t = 0$. The third condition tells us that there is no contaminant degradation over time. Finally, the last two conditions

say there is a steady volume rate of flushing recharge water into both aquifers; these conditions imply that there are no average water table fluctuations over time.

The solution to (2.2.17) is easily found to be

$$c_1 = c_{aq1} \exp(-t/t_{c1}) \quad (2.2.19)$$

where $t_{c1} = \frac{n_1 h_1}{\epsilon}$. Subject to the conditions stated above, we note that in (2.2.18)

$$c_{i2}(t) = c_{aq1} \exp(-t/t_{c1}), \text{ for } t \geq 0.$$

Substituting this input concentration into (2.2.18) and performing the indicated integration, we obtain after simplification

$$C_2 = C_{aq2} \exp(-t/t_{c2}) + \frac{c_{aq1} t_{c1}}{t_{c1} - t_{c2}} [\exp(-t/t_{c1}) - \exp(-t/t_{c2})] \quad (2.2.20)$$

where $t_{c2} = \frac{n_2 h_2}{\epsilon_2}$. Now if we let $\mu = Q_2/Q$ and $h_2 = mh$, where m is some arbitrary constant, then we see that

$$\frac{t_{c1}}{t_{c2}} = \frac{n_1 h_1}{\epsilon} \frac{\epsilon \mu}{n_2 m h_1} = \frac{n_1}{n_2} \frac{\mu}{m} = \gamma$$

Substituting this last relationship into (2.2.20) yields

$$\frac{C_2}{C_{aq1}} = \exp(-\gamma t/t_{c1}) + \frac{\gamma}{\gamma-1} [\exp(-t/t_{c1}) - \exp(-\gamma t/t_{c1})] \quad (2.2.21)$$

The final outflow concentration in the stream can now be calculated since the volumetric flow rate and contaminant output from each reservoir is known. The volumetric mass balance equation takes the form

$$c_0 = (\epsilon_1/\epsilon) c_1 + (\epsilon_2/\epsilon) c_2$$

where $\frac{\epsilon_1}{\epsilon} = \frac{Q_1/A_1}{Q/A_1} = \frac{Q_1}{Q} = \frac{Q-Q_2}{Q} = (1-\mu)$

and $\frac{\epsilon_2}{\epsilon} = \frac{Q_2/A_2}{Q/A_2} = \frac{Q_2}{Q} = \mu$

Thus for $A_1 = A_2$ we see that

$$c_0 = (1-\mu) c_1 + \mu c_2 \quad (2.2.22)$$

Substituting (2.2.19) and (2.2.21) into (2.2.22) and simplifying yields

$$\frac{c_0}{c_{aq1}} = a \exp(-t/t_{c1}) - b \exp(-\gamma t/t_{c1}) \quad (2.2.23)$$

where $a = (1 + \mu/(\gamma - 1))$

$$b = a - 1$$

$$\gamma = \frac{n_1}{n_2} \frac{\mu}{m}$$

$$m = h_2/h_1$$

$$\mu = \epsilon_2/\epsilon = \frac{m K_L/K_U}{1 + mK_L/K_U}$$

This last relationship will be derived in the Section 2.3.4. For the time being, however, we will simply say that the parameter μ can be related to certain physical parameters that characterize the so called Dupuit aquifer. Hence by using this last relationship we can directly compare the behavior of (2.2.23) with the concentration outflow history obtained from a Dupuit aquifer analysis. We therefore establish a basis

in which to explore the relationship between lumped models and spatially distributed models.

Before leaving this section, it would be beneficial to compute the total mass flowing out of the system pictured in Figure 2.4, and that predicted by (2.2.23). For simplicity we will consider only the situation where $n_1 = n_2$, $c_{aq_1} = c_{aq_2}$ at $t < 0$, and for zero aquifer input concentration at $t \geq 0$. Hence the total solute mass convected out of the system is given by

$$\int_0^{\infty} Qc dt = c_{aq} V_{aq} \quad (2.2.24)$$

where V_{aq} is the aquifer fluid volume, and where all other terms have been previously defined. For the steady flow situation Q is constant and (2.2.24) becomes in dimensionless form

$$\int_0^{\infty} \left(\frac{c}{c_{aq}}\right) d\left(\frac{t}{t_{c_1}}\right) = \frac{c_{aq} n (h_1 + h_2) A}{Q} \left(\frac{t_{c_1}}{c_{aq}}\right) \quad (2.2.25)$$

This last equation may be evaluated by inserting (2.2.23) for (c/c_{aq}) , and performing the indicated integration. Thus upon substitution and simplification we obtain

$$\begin{aligned} \int_0^{\infty} \left(\frac{c}{c_{aq}}\right) d\left(\frac{t}{t_{c_1}}\right) &= \int_0^{\infty} [a \exp(-t/t_{c_1}) - b \exp(-\gamma t/t_{c_1})] d(t/t_{c_1}) \\ &= 1 + \frac{\mu m}{\mu - m} - \frac{1}{\gamma} \left(\frac{\mu m}{\mu - m}\right) = 1 + m \end{aligned} \quad (2.2.26)$$

Note that in this evaluation our dimensionless time was normalized with respect to $t_{c_1} = \frac{nh_1}{\epsilon}$.

2.3 Linearized Dupuit Aquifer Model

2.3.1 Dupuit Approximation

The Dupuit approximation is perhaps one of the most powerful and widely applied tools available to the groundwater hydrologist in treating unconfined fluid flow through saturated porous materials. This approximation requires the assumptions that equipotential surfaces are very nearly vertical and the resulting fluid flow practically horizontal (Bear, 1972). As a result the Dupuit approximation is paramount to neglecting the vertical component of fluid flow over some given imaginary section drawn through the aquifer. Thus for a phreatic aquifer system with a fully penetrating stream, as shown in Figure 2.5, the following governing partial differential equation can be derived using this classical approximating technique (Bear, 1972).

$$\frac{\partial}{\partial x} \left(K_x h \frac{\partial h}{\partial x} \right) + \frac{\partial}{\partial y} \left(K_y h \frac{\partial h}{\partial y} \right) + \epsilon = S \frac{\partial h}{\partial t} \quad (2.3.1)$$

where $h(x,y,t)$ = thickness of the saturated flow zone,
 x,y = horizontal coordinates,
 K_x, K_y = directional hydraulic conductivity taken parallel to the (x, y) axes, respectively,
 ϵ = accretion, which is assumed to be uniform over x and y ,
 S = specific yield (drainable porosity),
 t = time.

Equation (2.3.1) is referred to as Boussinesq's equation for unsteady flow in a phreatic aquifer with accretion. Considerable attention has been focused on this and related problems (van Schilfgaarde, 1970; Bear, 1972; Gelhar, et al., 1974; Collins, 1976; and

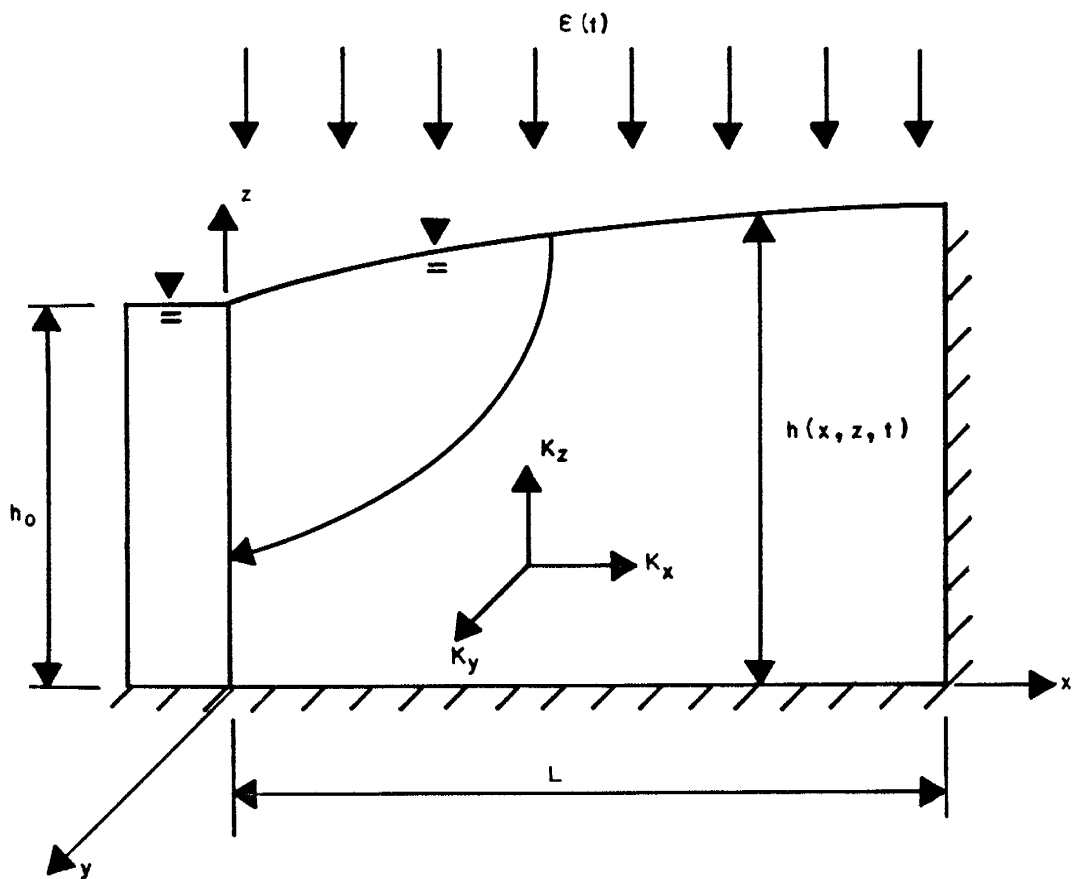


Figure 2.5 Schematic vertical cross-section through a stream connected phreatic aquifer system with uniform accretion.

references therein) in order to describe the temporal and spatial dependence of the free surface elevation. Thus once the necessary initial and boundary conditions are specified, a particular solution to (2.3.1) may be sought. Solution techniques that have been used include: (a) linearizing (2.3.1) in h , (b) linearizing (2.3.1) in h^2 , (c) the method of successive steady states, (d) the method of small perturbations, and (e) numerical methods that are usually related in some fashion to the other four analytical techniques. However, an analysis describing the convective transport of some given contaminant through such systems has been explored only recently (Gelhar and Wilson, 1974).

Consider the case of steady, one dimensional flow in a stream-connected phreatic aquifer system with uniform accretion. The governing flow equation obtained from (2.3.1) is

$$\frac{d}{dx} \left(Kh \frac{dh}{dx} \right) + \epsilon = 0 \quad (2.3.2)$$

where $K = K_x$. Integrating this equation with respect to x yields

$$\frac{K}{2} \frac{d}{dx} (h^2) = -\epsilon x + C$$

The integration constant, C , is easily obtained from the no flux boundary condition $\left(\frac{dh^2}{dx} = 0\right)$ at $x = L$. Thus

$$\frac{K}{2} \frac{d}{dx} (h^2) = \epsilon (L-x) \quad (2.3.3)$$

Noting that $h = h_0$ at $x = 0$, the final solution becomes

$$h^2 - h_0^2 = \frac{\epsilon x}{K} (2L-x) \quad (2.3.4)$$

which reduces to

$$(h - h_0) = \frac{\epsilon x(2L-x)}{K(\bar{h}+h_0)} \quad (2.3.5)$$

If $(h-h_0) \ll h_0$, which is equivalent with our original Dupuit approximation in arriving at (2.3.2), then $h \approx h_0$ and (2.3.5) can be approximately written as

$$(h-h_0) \approx \frac{\epsilon x(2L-x)}{2T} \quad (2.3.6)$$

where $T = K\bar{h}$, the aquifer transmissivity, and \bar{h} represents the average saturated thickness of the aquifer. This last expression gives us the so called Dupuit-Forcheimer parabolically shaped water table (Kirkham and Powers, 1972). At $x = L$ (2.3.6) tells us that the maximum water table elevation is given by

$$h = h_0 + \frac{\epsilon L^2}{2T} \quad (2.3.7)$$

The average saturated thickness of the one-dimensional Dupuit aquifer can be defined by

$$\bar{h} = \frac{1}{L} \int_0^L h \, dx \quad (2.3.8)$$

Inserting (2.3.6) into (2.3.8) and integrating yields

$$\bar{h} = h_0 + \frac{\epsilon L^2}{3T} \quad (2.3.9)$$

Referring back to the lumped parameter model, we recall that the aquifer outflow term in the water balance equation (see (2.2.1)) was

approximated by the linear expression given by (2.2.2). For a steady flow field, the inflow term, ϵ , is equal to the aquifer outflow, q_0 . Thus from (2.2.2) and (2.3.9), we see that

$$a = \frac{\epsilon}{(\bar{h}-h_0)} = \frac{3T}{L^2} \quad (2.3.10)$$

We may therefore conclude that if the linear outflow constant, a , from the lumped parameter model is calculated from (2.3.10), then the lumped model will be hydraulically equivalent to the average steady one-dimensional Dupuit model analysis derived above.

2.3.2 Relationship Between Lumped Parameter and Linearized Dupuit Models

In the previous section we established a hydraulic equivalence between the lumped parameter model described by a linear reservoir and the linearized Dupuit model under steady flow conditions. This equivalence was previously developed by Gelhar and Wilson (1974), and expanded upon by Flores W. and Gelhar (1976) for a variety of steady and unsteady flow conditions. We may now extend the analysis to include a mass transport similitude as well.

The relationship of steady, convective contaminant transport through a stream-connected aquifer system can be derived from an analysis of kinematic particle movement through the aquifer. Following Gelhar and Wilson (1974), we can conceptually visualize a small parcel of contaminated water following some given space-time trajectory through the steady flow field. We can determine the total change in quality of all such parcels at the stream-aquifer boundary where all parcels begin to

mix together. We must assume of course that no chemical interaction effects between adjacent water parcels of different quality or between water parcels and the porous medium occur. Similar conceptual approaches have also been used by Jury (1975) and Raats (1977).

The flow field implied in Figure 2.5 can be quantified by first assuming steady flow and then by using Darcy's law and the stream function Ψ (Bear, 1972). Thus the horizontal component of the specific discharge vector is given by

$$q_x = -K \frac{dh}{dx} = \frac{\partial \Psi}{\partial z} \quad (2.3.11)$$

where Ψ is the stream function that describes the particle trajectories for steady flow through the aquifer. From (2.3.2) and Figure 2.5 we observe that

$$Kh \frac{dh}{dx} = -\epsilon x + \epsilon L$$

and with $h \approx h_0$, (2.3.11) becomes

$$q_x = \frac{\partial \Psi}{\partial z} = -\frac{\epsilon(L-x)}{h_0} \quad (2.3.12)$$

from which we see that

$$\Psi = \Psi_0 - \frac{\epsilon(L-x)z}{h_0} \quad (2.3.13)$$

where $\Psi_0 =$ some constant, since $\Psi = \Psi_0$ at $x = L$. In a similar fashion we can write the vertical component of the specific discharge as

$$q_z = -\frac{\partial \Psi}{\partial x} \quad (2.3.14)$$

Thus

$$q_z = - \frac{\partial}{\partial x} \left[\psi_0 - \frac{\epsilon(L-x)z}{h_0} \right] = - \frac{\epsilon z}{h_0} \quad (2.3.15)$$

The vertical position \bar{z} of some small parcel of water originating at the surface $\bar{z} = h$ can be described by the Darcy flow velocity relation and (2.3.15), so

$$\frac{d\bar{z}}{dt} = \frac{q_z}{n} = - \frac{\epsilon \bar{z}}{nh_0}$$

which can be rewritten as

$$\int_{z=h}^{z=\bar{z}} \frac{dz}{z} = - \int_{\tau=0}^{\tau=t} \frac{\epsilon d\tau}{nh_0}$$

The solution becomes

$$\ln(\bar{z}) - \ln(h) = - \frac{\epsilon t}{nh_0}$$

This last expression may be simplified to yield

$$\bar{z} = h(x) \exp(-t/t_{c1}) \quad (2.3.16)$$

where $t_{c1} = \frac{nh_0}{\epsilon}$. Likewise the horizontal position \bar{x} of the same parcel of water starting at the water surface $\bar{x} = x$ can be described by the Darcy flow velocity and (2.3.12), or

$$\frac{d\bar{x}}{dt} = \frac{q_x}{n} = - \frac{\epsilon(L-\bar{x})}{nh_0}$$

from which we find that

$$(L - \bar{x}) = (L-x) \exp(t/t_{c1}) \quad (2.3.17)$$

where $t_{c1} = \frac{nh_0}{\epsilon}$. When the parcel just arrives at the stream at some time t , $\bar{x} = 0$ and (2.3.17) becomes

$$(1 - x/L) = \exp(-t/t_{c1}) \quad (2.3.18)$$

Substituting (2.3.6) and (2.3.18) into (2.3.16) yields

$$\begin{aligned}
 \bar{z} &= \left[h_0 + \frac{\epsilon x(2L-x)}{2T} \right] \exp(-t/t_{c1}) \\
 &= \left\{ h_0 + \frac{\epsilon L[1-\exp(-t/t_{c1})][2L-L(1-\exp(-t/t_{c1}))]}{2T} \right\} \exp(-t/t_{c1}) \\
 &= \left\{ h_0 + \frac{\epsilon L^2[1-\exp(-t/t_{c1})][1+\exp(-t/t_{c1})]}{2T} \right\} \exp(-t/t_{c1}) \\
 \bar{z} &= h_0 \left\{ 1 + \frac{\epsilon L^2[1-\exp(-2t/t_{c1})]}{2Th_0} \right\} \exp(-t/t_{c1}) \tag{2.3.19}
 \end{aligned}$$

We may observe that the maximum water table elevation given by (2.3.7) can be normalized to

$$\frac{h - h_0}{h_0} = \frac{\epsilon L^2}{2Th_0}$$

But according to the linearization of the Dupuit approximation made earlier, $(h-h_0) \ll h_0$. Therefore in (2.3.19) we have

$$\frac{\epsilon L^2}{2Th_0} \ll 1$$

Thus (2.3.19) may be approximately written as

$$\bar{z} = h_0 \exp(-t/t_{c1}) \tag{2.3.20}$$

Equations (2.3.17) and (2.3.20) define the (\bar{x}, \bar{z}) coordinate positions of some given parcel of water originating on the phreatic surface at (x, z) at time $t = 0$ and flowing through the steady Dupuit aquifer system shown in Figure 2.5. The time required for a water parcel

starting at some x on the phreatic surface to arrive at the stream was given by (2.3.18). This time is the same as that in (2.3.20). Thus

$$\bar{z}/h_0 = 1 - x/L \quad (2.3.21)$$

If the aquifer waters are originally contaminated at some uniform concentration level, c_{aq} , and at time $t=0$ the recharge waters entering the aquifer contain no contaminants, then (2.3.17), (2.3.20), and (2.3.21) can be used to define the uncontaminated front being convected through the aquifer system. For example if some uncontaminated parcel of water enters the phreatic surface at $x/L = 0.40$, it will exit the Dupuit aquifer on the stream boundary $\bar{z}/h_0 = 0.60$ at some dimensionless time $t/t_{c'} = -\ln(0.60) = 0.51$ later. The uncontaminated parcels of water entering the phreatic surface at $x/L < 0.40$ will already have arrived at the stream-aquifer boundary at $\bar{z}/h_0 > 0.60$, while those entering at $x/L > 0.40$ are still in the aquifer at some distance away from this boundary. The concentration in the stream at dimensionless time $t/t_{c'} = 0.51$ will therefore be only 60 percent of that at time $t/t_{c'} = 0$. We can express the entire time history of concentration in the stream as

$$c/c_{aq} = \exp(-t/t_{c'}) = \bar{z}/h_0 = 1 - x/L \quad (2.3.22)$$

Alternately we could have written a simple mass balance statement at the stream-aquifer boundary as

$$q_x h_0 \bar{c} = q_{x_1} (h_0 - \bar{z}) c_i + q_{x_2} \bar{z} c_{aq}$$

where q_x = the specific discharge at the stream boundary,
 h_0 = the stream elevation,
 q_{x_1} , q_{x_2} = uncontaminated and contaminated specific discharges into
the stream, respectively,
 \bar{z} = the elevation of the uncontaminated front at the stream,
 c = the concentration in the stream,
 c_{aq} = the aquifer concentration at $t = 0$,
 c_i = the concentration of the recharge water.

As previously mentioned, the Dupuit approximation neglects any explicit consideration for vertical flow effects so that the horizontal component of specific discharge is implied to be constant over any vertical depth in the aquifer. Therefore $q_x = q_{x_1} = q_{x_2}$ and the average concentration in the stream is given by

$$\bar{c} = \left(1 - \frac{\bar{z}}{h_0}\right) c_i + \left(\frac{\bar{z}}{h_0}\right) c_{aq}$$

But from (2.3.20) we see that $\bar{z}/h_0 = \exp(-t/t_c')$ so the stream concentration may also be written as

$$c = c_i + (c_{aq} - c_i) \exp(-t/t_c') \quad (2.3.23)$$

If $c_i = 0$, then (2.3.23) reduces to the same result obtained in (2.3.22). If t_c' in the above analysis is equal to t_c , then this relationship is also equivalent to that obtained in (2.2.15) for a conservative tracer ($k=0$) with a corresponding step change in input concentration at $t \geq 0$. In (2.2.15) $t_c = nh/\epsilon$, while in (2.3.23) $t_c' = nh_0/\epsilon$. However if the original Dupuit approximation, $(h-h_0) \ll h_0$, is satisfied, then $h \approx h_0$ and $t_c' \approx t_c$.

The average concentration in the Dupuit aquifer can be expressed as

$$c_{aq} = \frac{1}{L} \int_0^L \langle c_{aq} \rangle dx$$

where $\langle c_{aq} \rangle = (1-\bar{z}/h_0)c_i + (\bar{z}/h_0)c_{aq}$ at some \bar{x} . From this relationship and (2.3.20), we conclude that

$$\bar{c}_{aq} = c_i + (c_{aq} - c_i) \exp(-t/t_{c,i})$$

which is identical to the concentration in the stream. These analyses show that the average concentration in the Dupuit aquifer is not only the same as that in the stream, but is also equivalent to the concentration previously found for the well mixed linear aquifer system of the lumped parameter model. Hence we may conclude from the Dupuit analysis presented here, that the well mixed reservoir assumption made for the lumped system can be justified from a distributed systems approach.

2.3.3 Multi-Layered Dupuit Aquifer System

We have previously developed the water and mass balance relationships that can represent the time dependent hydraulic and solute response characteristics of a stream-connected phreatic aquifer system using the generalized lumped parameter format. Furthermore it was shown that a one-dimensional linearized Dupuit model behaves in a similar fashion to this lumped approach when a linear reservoir assumption is employed. At that time it was also suggested that representing certain physical systems by the generalized lumped approach might be unrealistic, even though net input stresses on the system were uniformly distributed throughout it. One such possibility was schematically drawn in Figure 2.3. This physical system was analyzed by utilizing the linear reservoir lumped parameter model in a parallel-type configuration as depicted in Figure 2.4. An obvious question immediately arise: Is this combined

lumped approach capable of accurately representing the physical system shown in Figure 2.3? We might proceed to answer this question by first exploring the hydraulic and solute response characteristics that would result from treating the multi-layered aquifer system of Figure 2.3. as separate Dupuit type reservoirs. Results from this analysis can then be directly compared with those previously obtained from the combined lumped approach.

The governing partial differential equation applicable to the physical situation represented by Figure 2.3 is

$$\frac{\partial}{\partial x_i} [K_u(h-b_L) + K_L b_L] \frac{\partial h}{\partial x_i} + \epsilon = S \frac{\partial h}{\partial t} \quad (2.3.24)$$

where x_i with $i = 1, 2$ are the horizontal coordinates; all other terms have been previously defined. If the flow is steady, then (2.3.24) in one dimension reduces to

$$\frac{d}{dx} [K_u(h-b_L) \frac{dh}{dx}] + \frac{d}{dx} [K_L b_L \frac{dh}{dx}] = -\epsilon \quad (2.3.25)$$

subject to the boundary conditions implied in Figure 2.3:

$$\begin{aligned} (1) \quad h &= h_0 \text{ at } x = 0 \\ (2) \quad dh/dx &= 0 \text{ at } x = L \end{aligned} \quad (2.3.26)$$

Integrating (2.3.25) with respect to x , and applying the second condition from (2.3.26) yields

$$\frac{K_u}{2} \frac{d(h^2)}{dx} - K_u b_L \frac{dh}{dx} + K_L b_L \frac{dh}{dx} = -\epsilon x + \epsilon L$$

which may be simplified to give

$$\frac{d(h^2)}{dx} + \left\{ \frac{d(h)}{dx} \right\} = \frac{2 \epsilon (L-x)}{K_u} \quad (2.3.27)$$

where $\xi = 2b_L(K_L - K_U)/K_U$. Integrating this last equation with respect to x and applying the first boundary condition from (2.3.26) yields

$$(h^2 - h_0^2) + \xi(h - h_0) = \epsilon x(L - x)/K_U \quad (2.3.28)$$

This solution, which describes the steady Dupuit water table shape, is similar in form to that in (2.3.4) for a single Dupuit aquifer. The second term in (2.3.28) alters the water table elevation because of the presence of the lower aquifer. Note that (2.3.4) is actually a special case of (2.3.28); if $K_U = K_L$ then $\xi = 0$ and (2.3.28) reduces to (2.3.4). For $(h - h_0) \ll h_0$ and $h \approx h_0$, equation (2.3.28) may be simplified to

$$(h - h_0) \left(1 + \frac{\xi}{h + h_0}\right) = \frac{\epsilon x(2L - x)}{K_U(h + h_0)}$$

from which we see that $h + h_0 \approx 2h_0$. Hence this last expression may be approximated by

$$h - h_0 = \frac{\epsilon x(2L - x)}{K_U(2h_0 + \xi)} \quad (2.3.29)$$

At $x = L$, the maximum water table elevation is given by

$$h = h_0 + \frac{\epsilon L^2}{K_U(2h_0 + \xi)} \quad (2.3.30)$$

The average saturated thickness of the two layered Dupuit aquifer system can be defined in a similar fashion to (2.3.8), or

$$\bar{h} = \frac{1}{L} \int_0^L h \, dx \quad (2.3.31)$$

Inserting (2.3.29) into (2.3.31), performing the necessary integration, and simplifying yields

$$\bar{h} = h_0 + \frac{2 \epsilon L^2}{3K_u(2h_0 + \xi)} \quad (2.3.32)$$

If we were to choose to represent the multi-aquifer system with a single linear reservoir lumped parameter model for a steady flow field, then the outflow constant given in (2.2.2) would be

$$a = \frac{\epsilon}{(\bar{h} - h_0)} = \frac{3K_u(2h_0 + \xi)}{2L^2} \quad (2.3.33)$$

where $(\bar{h} - h_0)$ in (2.3.33) is given by (2.3.29). Once again, note that if $K_u = K_L$, then $\xi = 0$ and (2.3.33) reduces to (2.3.10) where $T = K_u h_0$.

2.3.4. Relationship Between Combined Lumped Parameter Model and Linearized Multi-Layered Dupuit Aquifer Model

We may proceed in establishing the relationship of steady convective mass transport through the multi-layered Dupuit aquifer system shown in Figure 2.3 using the previously employed conceptual approach of kinematic particle movement. However one would expect the lower aquifer system to either impede or enhance contaminant migration through the system, depending on whether the lower layer hydraulic conductivity was less than or greater than that of the upper layer.

The horizontal component of the specific discharge in the upper layer, q_{xu} , is once again related to Darcy's law and the stream function:

$$q_{xu} = -K_u \frac{dh}{dx} = \left. \frac{\partial \Psi}{\partial z} \right|_u \quad (2.3.34)$$

where Ψ is the stream function describing particle trajectories through the multi-layer system. Differentiating (2.3.29) with respect to x we obtain

$$\frac{dh}{dx} = \frac{2\varepsilon(L-x)}{K_u(2h_0 + \xi)}$$

from which we note

$$q_{xu} = -\frac{2\varepsilon(L-x)}{(2h_0 + \xi)} = \left. \frac{\partial \Psi}{\partial z} \right|_u \quad (2.3.35)$$

Integrating this last equation with respect to z yields

$$\Psi_u = -\frac{2\varepsilon(L-x)z}{(2h_0 + \xi)} + f_1(x) \quad (2.3.36)$$

where $f_1(x)$, the constant of integration with respect to z , is in general some function of x . Equation (2.3.36) describes the stream function, $\Psi_u(x, z)$, in the upper aquifer between ($b_L \leq z \leq h$).

Similarly we note that

$$q_{xL} = -K_L \frac{dh}{dx} = \left. \frac{\partial \Psi}{\partial z} \right|_L$$

from which we immediately see that

$$q_{xL} = -\frac{K_L 2\varepsilon(L-x)}{K_u(2h_0 + \xi)} = \left. \frac{\partial \Psi}{\partial z} \right|_L$$

Thus

$$\Psi_L = -\frac{K_L}{K_u} \frac{2\varepsilon(L-x)z}{(2h_0 + \xi)} + f_2(x) \quad (2.3.37)$$

where $f_2(x)$ is some different function of x . Equation (2.3.37) describes the stream function, $\psi_L(x, z)$, in the lower aquifer between ($0 \leq z \leq b_L$). The implied boundary conditions for the multi-layered Dupuit aquifer system of Figure 2.3. are:

$$\begin{aligned}
 (1) \quad \psi_u &= \epsilon x \text{ at } z = h(x) \approx h_0 \\
 (2) \quad \psi &= \epsilon L \text{ at } x = L \\
 (3) \quad \psi_L &= \epsilon L \text{ at } z = 0 \\
 (4) \quad \psi_u &= \psi_L \text{ at } z = b_L
 \end{aligned}
 \tag{2.3.38}$$

Employing condition (4) from (2.3.38) in (2.3.36) and (2.3.37) yields

$$-\frac{2 \epsilon(L-x)b_L}{(2h_0 + \xi)} + f_1(x) = -\frac{K_L}{K_u} \frac{2 \epsilon(L-x)b_L}{(2h_0 + \xi)} + f_2(x)$$

Solving this last equation for $f_2(x)$ and inserting the result into (2.3.37) gives

$$\psi_L = f_1(x) - \frac{2 \epsilon(L-x)}{(2h_0 + \xi)} \left[\frac{K_L z}{K_u} + b_L \left(1 - \frac{K_L}{K_u}\right) \right]
 \tag{2.3.39}$$

Using condition (1) from (2.3.38) in (2.3.36) yields

$$\psi_u = \epsilon x = f_1(x) - \frac{2 \epsilon(L-x)h_0}{(2h_0 + \xi)}$$

Therefore

$$f_1(x) = \epsilon x + \frac{2 \epsilon(L-x)h_0}{(2h_0 + \xi)}
 \tag{2.3.40}$$

and from (2.3.36) we see that

$$\psi_u = \epsilon x + \frac{\epsilon(L-x)}{(h_0 + \xi/2)} (h_0 - z) \quad (2.3.41)$$

which describes ψ_u in the interval ($b_L \leq z \leq h$). From (2.3.39) and (2.3.40) we observe that

$$\psi_L = \epsilon x + \frac{\epsilon(L-x)}{(h_0 + \xi/2)} \left[h_0 - \frac{K_L z}{K_u} - b_L \left(1 - \frac{K_L}{K_u} \right) \right]$$

which describes ψ_L in the interval ($0 \leq z \leq b_L$). This last equation may be simplified to

$$\psi_L = \epsilon L - \frac{K_L \epsilon(L-x)z}{K_u(h_0 + \xi/2)} \quad (2.3.42)$$

Using the transformations defined below, we may express (2.3.41) and (2.3.42) in a dimensionless form. Thus for

$$\begin{aligned} X &= x/L; \quad 0 \leq X \leq 1 \\ Z &= z/h_0; \quad 0 \leq Z \leq 1 \\ \hat{\xi} &= \xi/2h_0 = b_L (K_L - K_u)/h_0 K_u \\ \bar{\psi}_u &= \psi_u / \epsilon L \\ \bar{\psi}_L &= \psi_L / \epsilon L \\ B &= b_L/h_0 \end{aligned} \quad (2.3.43)$$

we see that

$$\begin{aligned} \bar{\psi}_u &= X - \frac{(1-X)(1-Z)}{(1 + \hat{\xi})}; \quad (B \leq Z \leq 1) \\ \bar{\psi}_L &= 1 - \frac{K_L(1-X)}{K_u(1 + \hat{\xi})}; \quad (0 \leq Z \leq B) \end{aligned} \quad (2.3.44)$$

The vertical position \bar{z} of some small parcel of water originating at the surface $\bar{z} = h_0$ may be described by the Darcy flow velocity, and (2.3.41) and (2.3.42). Thus for the upper aquifer we have

$$\frac{1}{n} q_{zu} = \frac{d\bar{z}}{dt} = -\frac{1}{n} \frac{\partial \psi_u}{\partial x} = -\frac{1}{n} \frac{\partial}{\partial x} \left[\epsilon x + \frac{\epsilon(L-x)(h_0 - \bar{z})}{(h_0 + \xi/2)} \right]$$

or,

$$\frac{1}{n} q_{zu} = -\frac{\epsilon}{n} \left[1 - \frac{(h_0 - \bar{z})}{(h_0 + \xi/2)} \right] = \frac{d\bar{z}}{dt}$$

If we let $V = 1 - (h_0 - \bar{z})/(h_0 + \xi/2)$, then $dV/d\bar{z} = 1/(h_0 + \xi/2)$

and

$$\frac{d\bar{z}}{dt} \frac{dV}{d\bar{z}} = -\frac{\epsilon V}{n(h_0 + \xi/2)}$$

Upon rearranging we note that

$$\int_{V=V_1}^{V=V_2} \frac{dV}{V} = - \int_{t=t_1}^{t=t_2} \frac{\epsilon dt}{n(h_0 + \xi/2)}$$

from which our solution becomes

$$\left[1 - \frac{h_0 - \bar{z}_2}{h_0 + \xi/2} \right] = \left[1 - \frac{h_0 - \bar{z}_1}{h_0 + \xi/2} \right] \exp \left[-\frac{\epsilon(t_2 - t_1)}{n(h_0 + \xi/2)} \right]$$

where $\bar{z}_1 = h_0$ at $t_1 = 0$, and $\bar{z}_2 = \bar{z}_u$ at some later time $t_2 = t$. Thus

$$(\bar{z}_u + \xi/2) = (h_0 + \xi/2) \exp \left[-\frac{\epsilon t}{n(h_0 + \xi/2)} \right] \quad (2.3.45)$$

This last equation gives the vertical position, \bar{z}_u , of some given parcel

of water in the upper aquifer at time t between $(b_L \leq z_u \leq h_0)$ starting at $\bar{z}_u = h_0$ at time $t = 0$.

For the lower aquifer we have from (2.3.42)

$$\frac{1}{n} q_{zL} = \frac{d\bar{z}_L}{dt} = -\frac{1}{n} \frac{d\psi_L}{dx} = -\frac{1}{n} \frac{a}{ax} \left[\epsilon L - \frac{K_L \epsilon (L-x) \bar{z}_L}{(h_0 + \xi/2)} \right]$$

or upon rearranging and integration with respect to t we have

$$\int_{\bar{z}_L = b_L}^{\bar{z}_L = \bar{z}_L} \frac{d\bar{z}_L}{\bar{z}_L} = - \int_{t = t_1}^{t = t_2} \frac{\epsilon K_L dt}{n K_u (h_0 + \xi/2)}$$

where $t = t_1$, represents the travel time of the parcel to reach the lower aquifer. The solution of this last equations is

$$\bar{z}_L = b_L \exp \left[-\frac{\epsilon K_L (t-t_1)}{n K_u (h_0 + \xi/2)} \right] \quad (2.3.46)$$

where t_1 may be obtained from (2.3.45) with $\bar{z}_u = b_L$, or

$$\frac{\epsilon t_1}{n(h_0 + \xi/2)} = -\ln \left[\frac{b_L + \xi/2}{h_0 + \xi/2} \right] \quad (2.2.47)$$

Equation (2.3.46) represents the vertical position of our parcel in the lower aquifer between $(0 \leq \bar{z}_L \leq b_L)$ starting at $\bar{z}_L = b_L$ at some time $t = t_1$. In both (2.3.45) and (2.3.46) we have neglected the travel time of the parcel from h to h_0 , which is assumed small since $(h-h_0) \ll h_0$ in the Dupuit approximation.

In a similar fashion we may find the horizontal position, \bar{x} , of the same parcel. For the upper aquifer we would have

$$\frac{1}{n} q_{xu} = \frac{d\bar{x}}{dt} = \frac{1}{n} \frac{\partial \Psi_u}{\partial z}$$

and from (2.3.41) we obtain

$$\frac{1}{n} q_{xu} = - \frac{\epsilon}{n} \frac{L - \bar{x}_u}{h_0 + \xi/2} = \frac{d\bar{x}_u}{dt}$$

If we let $W = L - \bar{x}_u$, then $dW/d\bar{x}_u = -1$ and

$$\frac{d\bar{x}_u}{dt} \frac{dW}{d\bar{x}_u} = - \frac{\epsilon W}{n(h_0 + \xi/2)} \frac{dW}{d\bar{x}_u}$$

so we see that

$$\int \frac{dW}{W} = \int \frac{\epsilon dt}{n(h_0 + \xi/2)}$$

which yields

$$\ln(W) = \frac{\epsilon t}{n(h_0 + \xi/2)} + C_1$$

where C_1 is some constant of integration. This last equation may be simplified to yield

$$(L - x_u) = C_2 \exp \left[\frac{\epsilon t}{n(h_0 + \xi/2)} \right]$$

where C_2 is some different constant. At $t=0$ the parcel of water is located at the starting point $\bar{x}_u = x$ on h_0 , so $C_2 = (L-x)$ and we obtain

$$(L - \bar{x}_u) = (L - x) \exp \left[\frac{\epsilon t}{n(h_0 + \xi/2)} \right] \quad (2.3.48)$$

Equations (2.3.45) and (2.3.48) together define the (\bar{x}, \bar{z}) particle position in the upper layer at some time t after it enters the system. We may therefore eliminate the term containing this common time t , and arrive at a single expression in (x, z) only. Thus

$$(L - x)/(L - \bar{x}_u) = (\bar{z}_u + \xi/2) / (h_0 + \xi/2) \quad (2.3.49)$$

Using the relationships defined in equation (2.3.43) we may express (2.3.49) in dimensionless form, or

$$(1 - X)/(1 - \bar{X}_u) = (\bar{Z}_u + \hat{\xi}) / (1 + \hat{\xi}) \quad (2.3.50)$$

where the particle position after some time t is given by (\bar{X}_u, \bar{Z}_u) . At the stream boundary $\bar{X}_u = 0$, so that (2.3.50) may be expressed in terms of the original particle starting point, X , and its exit point on the stream, \bar{Z}_u . So

$$(1 - X) = (\bar{Z}_u + \hat{\xi}) / (1 + \hat{\xi}) \quad (2.3.51)$$

which is valid in the interval $(B \leq Z_u \leq 1)$. The maximum value of X may be found when Z_u is a minimum. Thus $X \leq (1 - B)/(1 + \hat{\xi})$ and the maximum dimensionless streamline that remains in the upper aquifer is found to be $\bar{\Psi}_u = (1 - B)/(1 + \hat{\xi})$, since $\bar{\Psi}_u = X$ on $Z = 1$ from (2.3.44).

Likewise for the lower layer we observe that

$$\frac{1}{n} q_{xL} = \frac{d\bar{x}_L}{dt} = \frac{1}{n} \frac{d\psi_L}{dz} = \frac{K_L \epsilon (L - \bar{x}_L)}{nK_u(h_0 + \xi/2)}$$

from (2.3.42); and we note that

$$\frac{d\bar{x}_L}{dt} \frac{dW}{d\bar{x}_L} = - \frac{K_L \epsilon W}{nK_u (h_0 + \xi/2)} \frac{dW}{d\bar{x}_L}$$

where $W = (L - \bar{x}_L)$ and $dW/d\bar{x}_L = -1$. Hence we obtain

$$\ln(W) = \frac{K_L \epsilon t}{nK_u (h_0 + \xi/2)} + C_3$$

We may write this last equation as

$$(L - \bar{x}_L) = C_4 \exp\left[\frac{K_L \epsilon t}{nK_u (h_0 + \xi/2)}\right] \quad (2.3.52)$$

where C_3 and C_4 are different arbitrary constants. Noticing that at $t=t_1$, $\bar{x}_L = \bar{x}_u$, and from (2.3.48) we see that

$$(L - \bar{x}_L) = C_4 \exp\left[\frac{K_L \epsilon t_1}{nK_u (h_0 + \xi/2)}\right] = (L - x) \exp\left[\frac{\epsilon t_1}{n(h_0 + \xi/2)}\right]$$

so that

$$C_4 = (L - x) \exp\left[\frac{\epsilon t_1}{n(h_0 + \xi/2)}\right] (1 - K_L/K_u)$$

and

$$(L - \bar{x}_L) = (L - x) \exp\left[\frac{\epsilon}{n(h_0 + \xi/2)} \left[\frac{K_L}{K_u} (t - t_1) + t_1\right]\right] \quad (2.3.53)$$

where $t > t_1$. Equation (2.3.53) describes the horizontal position, \bar{x}_L , of the parcel of water in the lower aquifer which originally starts at time $t=0$ at the surface position x and intersects the lower boundary

at some later time $t=t_1$. When $\bar{x}_L = 0$ the particle will just arrive at the stream. Thus

$$\frac{L}{L-x} = \exp \left[\frac{K_L \epsilon t}{nK_u(h_0 + \xi/2)} + a(1 - K_L/K_u) \right] \quad (2.3.54)$$

where $a = \frac{\epsilon t_1}{n(h_0 + \xi/2)} = -\ln(b_L + \xi/2)/(h_0 + \xi/2)$, and is obtained from (2.3.47) with $z_u = b_L$ at $t=t_1$. Rearranging (2.3.54) we see that the total travel time of some parcel through both aquifers is given by

$$\frac{\epsilon t}{n(h_0 + \xi/2)} = a(1 - K_L/K_u) - \frac{K_u}{K_L} \ln(1 - x/L) \quad (2.3.55)$$

According to (2.3.43) we may express the above relationship in the dimensionless form

$$t/t_{c'} = A(1 - K_L/K_u) - \frac{K_u}{K_L} \ln(1-X) \quad (2.3.56)$$

where $A = -\ln(B + \hat{\xi})/(1 + \hat{\xi})$, $t_{c'} = n(h_0 + \xi/2)/\epsilon$, and $X > (1-B)/(1 + \hat{\xi})$ according to (2.3.51) with $\bar{z}_u = B$.

We are now in a position to compare the concentration outflow histories between the linearized multi-layered Dupuit aquifer model derived above, and that previously derived for the combined lumped parameter model. Using the simple situation previously employed for a flushing system, we have the following initial conditions:

- (1) $c_{aq} = c_{aq_2}$ at $t = 0$
- (2) $c_i(t) = 0$ for $t \geq 0$
- (3) $\epsilon(t) = \epsilon$, a constant, for $t \geq 0$

To reiterate, these conditions state that our two-layered Dupuit aquifer system of Figure 2.3 has some uniformly mixed contaminant in both reservoirs. The flow is at steady state, but at time $t \geq 0$ the recharge waters no longer carry any contamination into the upper aquifer. Our problem, then, is to define the contaminant outflow concentration in the stream for $t \geq 0$.

We can utilize the expressions given by (2.3.44), (2.3.45), and (2.3.56) to relate the concentration in the stream to time. For example, a particle of water entering the upper aquifer at $t = 0$ at some X (where $Z = 1$) will follow a streamline through this reservoir defined by

$$\bar{\Psi}_u = X + \frac{(1-X)(1-Z)}{(1+\hat{\xi})} \quad (2.3.57)$$

where ($0 \leq X \leq 1$), the horizontal position of the particle,

($0 \leq Z \leq 1$), the vertical position of the particle,

($0 \leq \bar{\Psi}_u \leq 1$), the dimensionless streamline,

and where $\hat{\xi} = b_L (K_L - K_U) / h_o K_U = B(K_L - K_U) / K_U$. According

to (2.3.51) if $X > (1-B)/(1+\hat{\xi})$ at $Z=1$, then the streamline will

penetrate into the lower aquifer; it will be defined by (2.3.57) in the upper reservoir and by

$$\bar{\Psi}_L = 1 - \frac{K_L (1-X)Z}{K_U (1+\hat{\xi})}, \quad 0 \leq Z \leq B \quad (2.3.58)$$

in the lower reservoir. If $X < (1-B)/(1+\hat{\xi})$ at $Z=1$, then the streamline will remain in the upper aquifer and be completely defined by (2.3.57).

Some particle that is following a given $\bar{\psi}_u$ will arrive at the stream in time t , given by (2.3.45), or

$$t/t'_{c_1} = \ln (Z_u + \hat{\xi}) / (1 + \hat{\xi}) \quad (2.3.59)$$

in dimensionless form where $t'_{c_1} = n(h_0 + \xi/2)/\epsilon$; it is called the solute response time since it characterizes the average contaminant residence time in the aquifer system. Note, however, that t'_{c_1} is based on parameters that characterize the system geometry and hydraulic conductivity contrast so that it is actually the average contaminant residence time in the system. If $K_L = K_u$, then $\xi = 0$ and $t'_{c_1} = t_c$, the single Dupuit aquifer solute response time from (2.3.23). The dimensionless travel time (t/t'_c) is therefore based relative to this average solute residence time in the system.

The concentration in the stream at $X = 0$ and at time $t < 0$ will be same as in the system, c_{aq} ; at some later time $t \geq 0$, this concentration, c , will simply be

$$c/c_{aq} = 1 - \bar{\psi}_u = (Z + \hat{\xi}) / (1 + \hat{\xi}) \quad (2.3.60)$$

according to (2.3.51). But from (2.3.45) this last relationship becomes

$$c/c_{aq} = \exp(t/t'_{c_1}) = (Z + \hat{\xi}) / (1 + \hat{\xi}); \quad b \leq Z \leq 1 \quad (2.3.61)$$

where $t/t'_{c_1} \leq -\ln(B + \hat{\xi}) / (1 + \hat{\xi}) = A$, according to (2.3.47) and (2.3.56), and where $\hat{\xi} = b_L(K_L - K_u) / h_0 K_u$. Equation (2.3.61) is similar to the outflow concentration obtained in (2.2.19) for a linear lumped parameter reservoir forming the upper layer of the combined lumped

system previously examined, except in that analysis the solute residence time was defined as $t_{c_1} = n_1 h_1 / \epsilon$.

If the particle starting point on $Z = 1$ is $X > (1-B)/(1 + \hat{\xi})$, then the dimensionless streamline will penetrate into the lower aquifer. Hence from (2.3.46)

$$\bar{z}_L / h_0 = Z_L = \frac{b_L}{h_0} = \exp \left[-\frac{K_L (t-t_1)}{K_u t'_{c_1}} \right] \quad (2.3.62)$$

But from (2.3.44) we note that at the stream $X=0$, and

$$c/c_{aq} = 1 - \bar{\psi}_L = \frac{K_L Z_L}{K_u (1 + \hat{\xi})}, \quad 0 \leq Z \leq B \quad (2.3.63)$$

Combining this last expression with (2.3.62) yields

$$c/c_{aq} = \frac{K_L B}{K_u (1 + \hat{\xi})} \exp \left[-\frac{K_L (t-t_1)}{K_u t'_{c_1}} \right] \quad (2.3.64)$$

where $t \geq t_1$, and where t_1 is given by (2.3.59) with $Z_u = B$.

Equations (2.3.61) and (2.3.64) define the time history solute concentration in the stream. Furthermore, these relationships are identical at $t = t_1$; thus a check at $Z = B$ yields the identity

$$c/c_{aq} = (B + \hat{\xi}) / (1 + \hat{\xi}) = K_L B / K_u (1 + \hat{\xi})$$

Again, notice that the dimensionless travel time is based on the total system thickness, the net recharge rate, ϵ , the upper layer porosity, n , and the parameter $\hat{\xi}$, which characterizes the alteration of the water table shape resulting from the presence of the lower reservoir.

In order to relate the concentration outflow histories of the combined lumped parameter model and the linearized multi-layered Dupuit

aquifer model, we must first relate the parameter μ in equation (2.2.23) to our Dupuit model. From Figure 2.5 we may write a water balance at the stream-aquifer boundary as

$$Q = Q_L + Q_u \quad (2.3.65)$$

where Q represents the total volume outflow rate leaving the aquifer; it is composed of the volume outflows leaving the lower and upper aquifers respectively. For a steady flow field, Q is equal to the net recharge rate ϵ , times the horizontal aquifer area, A . Thus from the definition for μ in (2.2.23) we had

$$\mu = \frac{\epsilon L}{\epsilon} = \frac{Q_L/A}{Q/A} = Q_L/Q$$

But from (2.3.65) we also note that

$$\mu = Q_L/Q = Q_L/(Q_u + Q_L) = Q_2/(Q_1 + Q_2)$$

and from Darcy's Law

$$Q_L = Q_2 = h_L K_L \bar{J}_L \quad \text{and} \quad Q_u = Q_1 = h_u K_u \bar{J}_u$$

where h_L, h_u = lower and upper layer saturated thicknesses,
 K_L, K_u = lower and upper layer hydraulic conductivities,
 \bar{J}_L, \bar{J}_u = lower and upper layer hydraulic gradients.

From the Dupuit approximation we concluded that $\bar{J}_L = \bar{J}_u$ and

$$\mu = (mh_u K_L \bar{J}_L) / (h_u K_u \bar{J}_u + mh_u K_L \bar{J}_L)$$

where $h_L = mh_u$ and m is some constant. Thus

$$\mu = (mK_L/K_U)/(1 + mK_L/K_U) \quad (2.3.66)$$

By selecting μ in (2.2.3) according to (2.3.66) the combined lumped parameter model should approximate the linearized Dupuit results of (2.3.63) and (2.3.64).

Seven specific cases for a comparison between the combined lumped parameter model and the linearized Dupuit model were made. Table 2.1 summarizes the hydraulic features of each case examined.

Table 2.1 Comparison between the combined lumped and linearized Dupuit models

<u>Number</u>	<u>K_U/K_L</u>	<u>m</u>	<u>$1+m$</u> ⁽¹⁾	<u>Remarks</u>
1	20	3	4	Fig. 2.6
2	10	3	4	Fig. 2.7
3	5	3	4	Fig. 2.8
4	2	3	4	Fig. 2.9
5	1	0	1	Fig. 2.10
6	10	9	10	Fig. 2.11
7	10	19	20	Fig. 2.12

(1) Theoretical area under (c/c_{aq}) versus (t/t_{c_1}) curve

The concentration outflow histories from both models are shown in Figures 2.6 through 2.12; each case examined shows good to excellent agreement between the models. In these figures the dimensionless concentration, c/c_{aq} , is plotted against the dimensionless time t/t_{c_1} from the lumped model. Here c is the concentration in the stream, while c_{aq} is the initial average aquifer concentration in the system. Likewise t is the time required for the concentration c to appear in the stream, and t_{c_1}

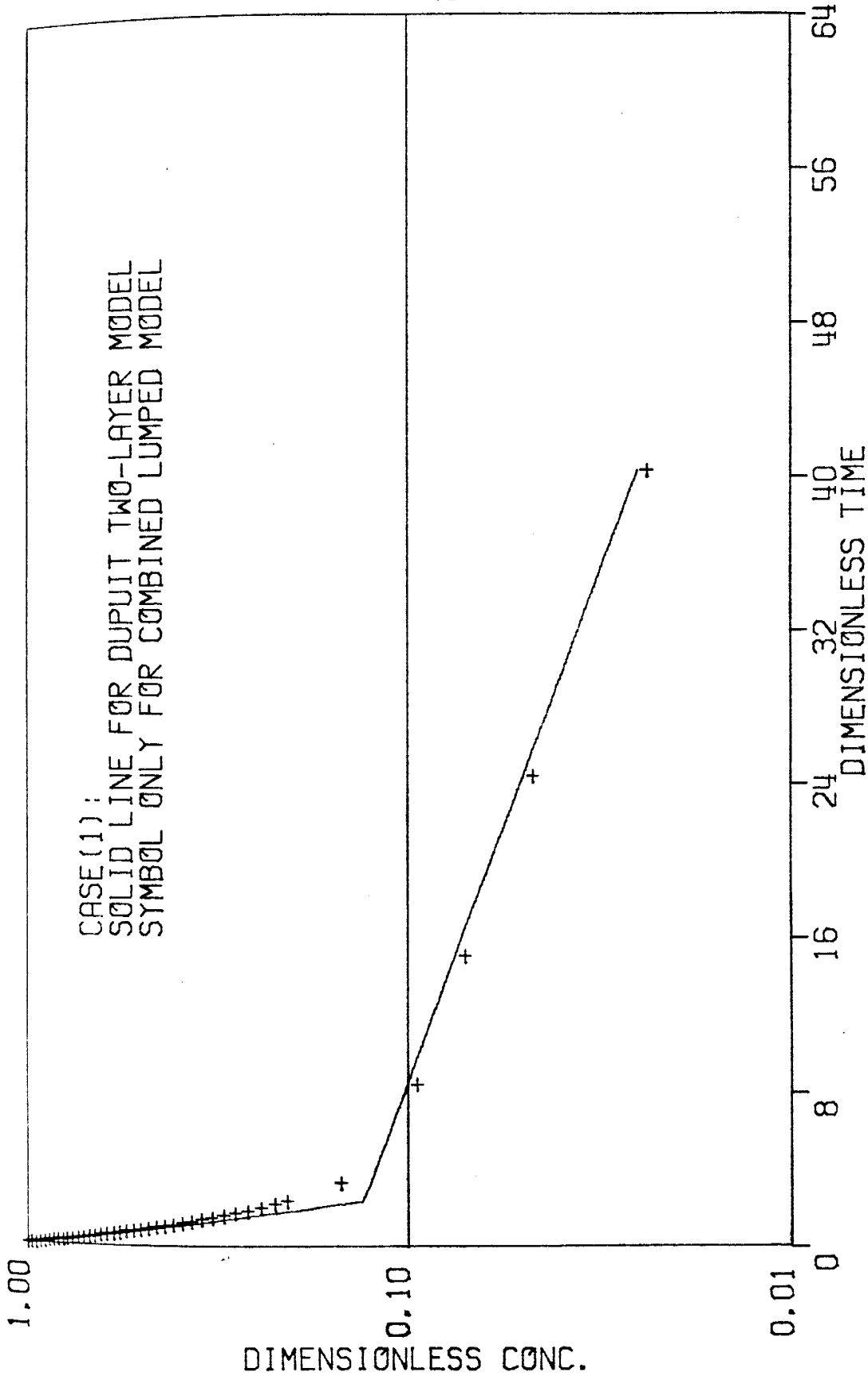


Figure 2.6: Log (c/c_{aq}) versus t/t_{c1} for case (1).

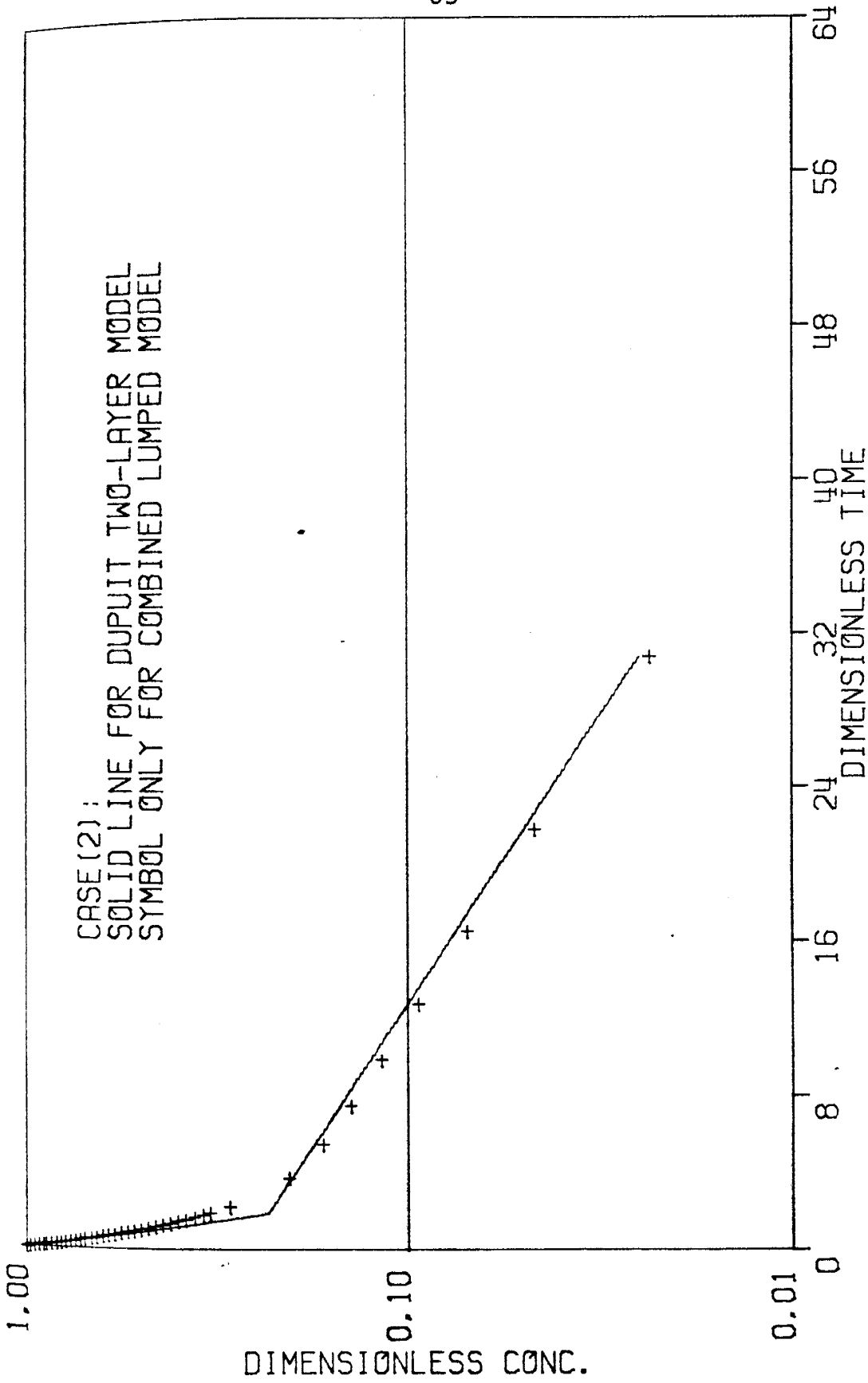


Figure 2.7: $\log(c/c_{aq})$ versus t/t_{c1} for case (2).

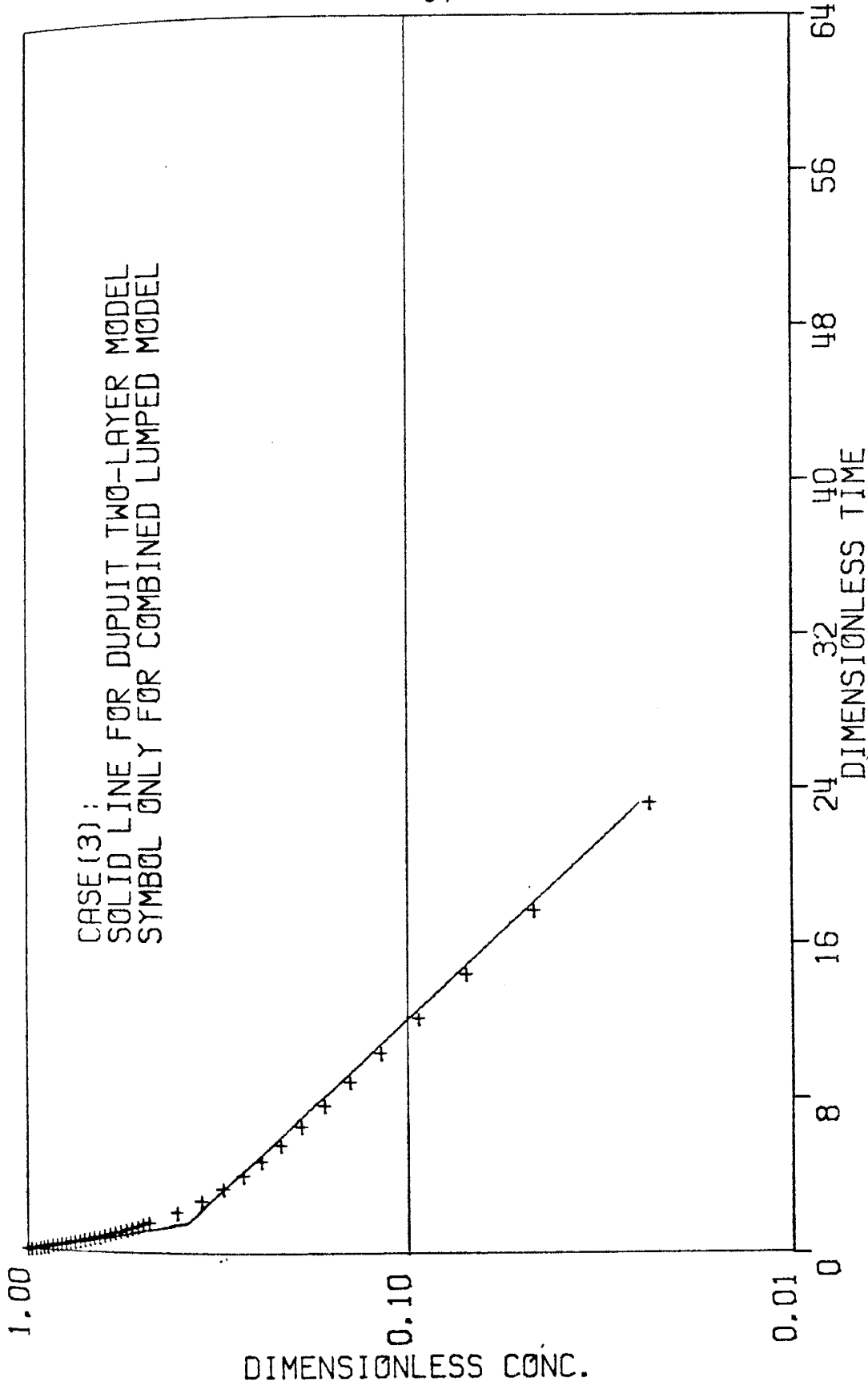


Figure 2.8: $\log (c/c_{aq})$ versus t/t_{c1} for case (3).

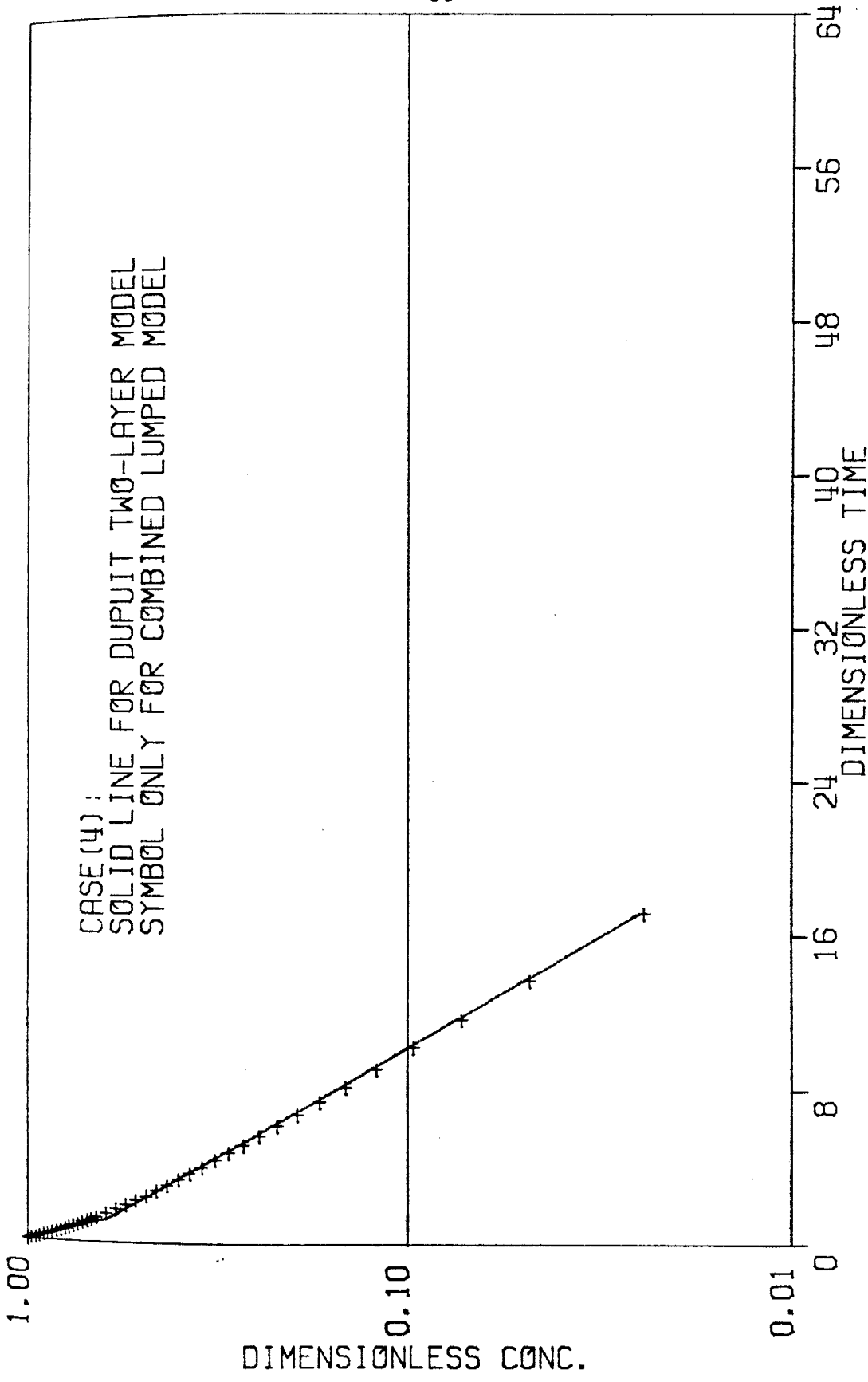


Figure 2.9: Log (c/c_{aq}) versus t/t_{c1} for case (4).

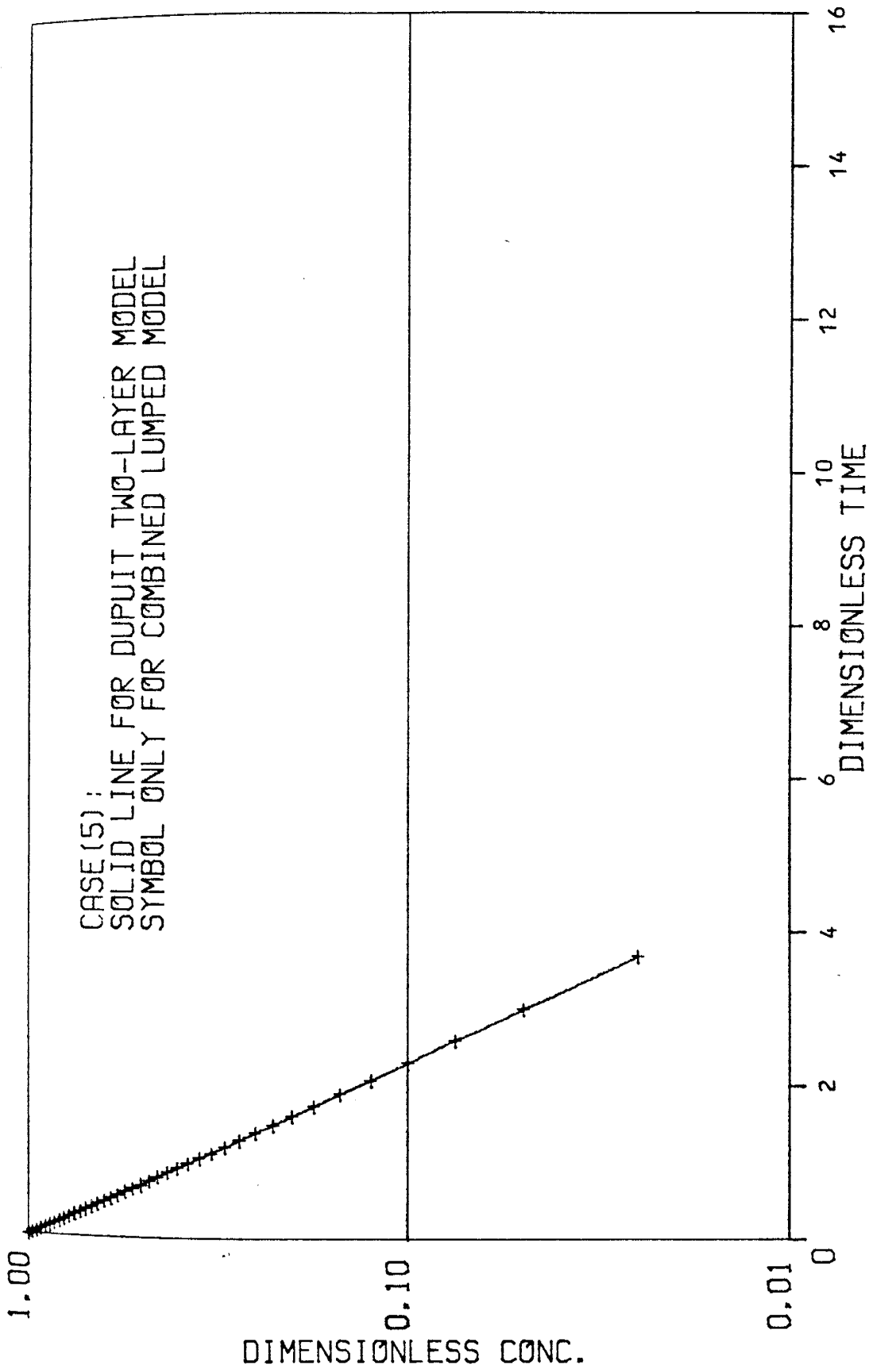


Figure 2.10: Log (c/c_{aq}) versus t/t_{c1} for case (5).

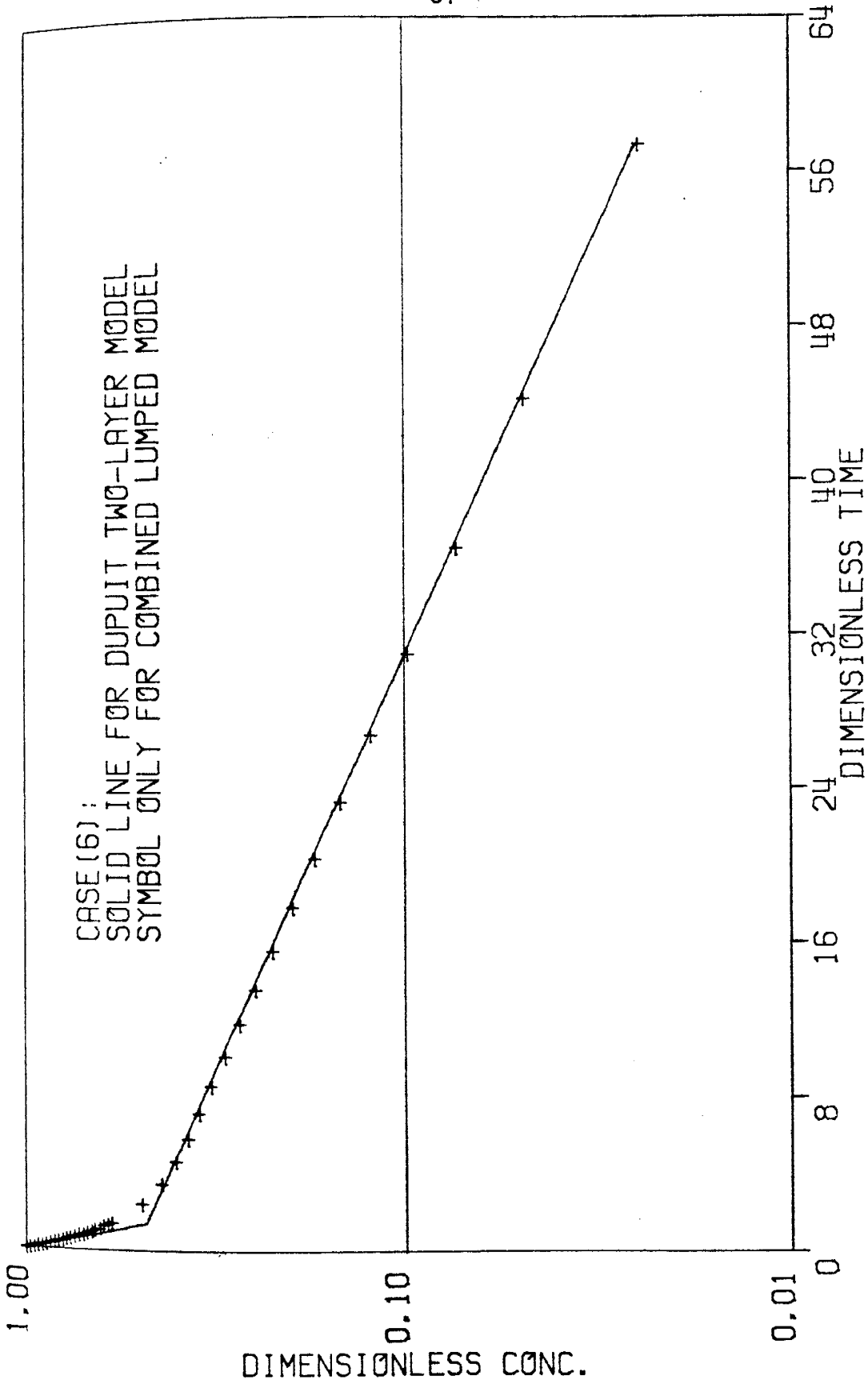


Figure 2.11: $\log (c/c_{aq})$ versus t/t_{c1} for case (6).

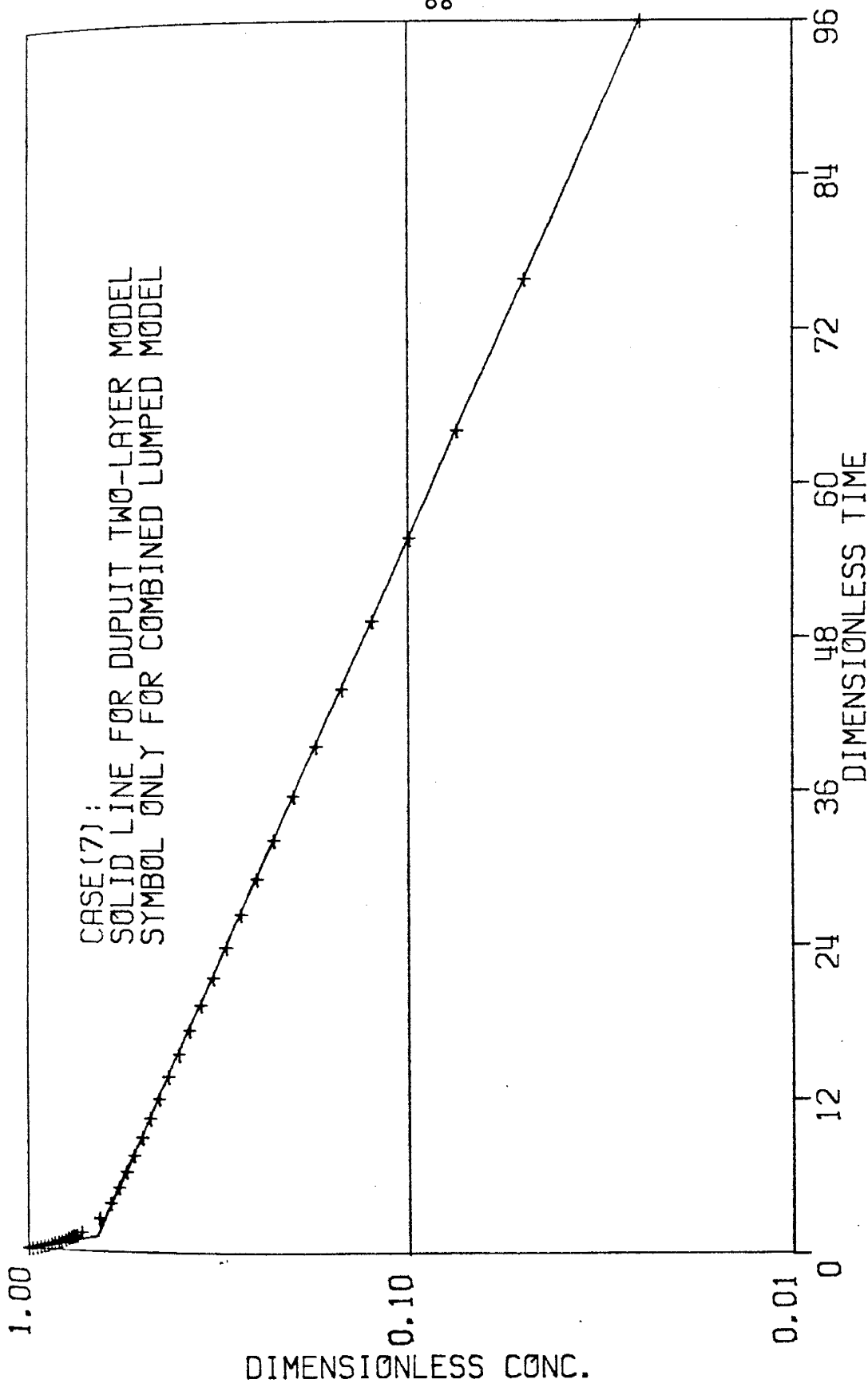


Figure 2.12: Log (c/c_{aq}) versus t/t_{c1} for case (7).

$= nh_1/\epsilon$ is the solute response time, or average solute residence time in the upper aquifer.

As was done in (2.2.25), we may compute the total mass leaving the system of Figure 2.5 utilizing (2.3.61), (2.3.64), and the relationship

$$\int_0^{\infty} (c/c_{aq}) d(t/t'_{c_1}) = (c_{aq} V_{aq}/Q)(t'_{c_1}/c_{aq})$$

While evaluation of this integral presents no problem, we would like to compare the results to those previously obtained in (2.2.26). Hence we must relate (t/t'_{c_1}) to (t/t_{c_1}) since they are different. Thus from Figures 2.4 and 2.5 we observe that

$$\begin{aligned} \frac{t'_{c_1}}{t_{c_1}} &= \frac{n(h_0 + \ell/2)}{\epsilon} \frac{\epsilon}{nh_1} = \frac{h_1 + h_2 + \ell/2}{h_1} \\ &= 1 + \frac{h_2}{h_1} + \frac{b_L(K_L - K_U)}{h_1 K_U} \\ &= 1 + \frac{h_2}{h_1} + \frac{h_2 K_L}{h_1 K_U} - \frac{h_2}{h_1} = 1 + r \end{aligned}$$

where $r = h_2 K_L / h_1 K_U$. Thus

$$\int_0^{\infty} \left(\frac{c}{c_{aq}}\right) d\left(\frac{t}{t'_{c_1}}\right) \left(\frac{t'_{c_1}}{t_{c_1}}\right) = (1+r) \int_0^{\infty} \left(\frac{c}{c_{aq}}\right) d\left(\frac{t}{t_{c_1}}\right) \quad (2.3.67)$$

Substituting (2.3.61) and (2.3.64) into (2.3.67) yields

$$\int_0^{\infty} \frac{c}{c_{aq}} d\left(\frac{t}{t'_{c_1}}\right) = (1+r) \int_0^{t/t'_{c_1}} c_1 \exp(-t/t'_{c_1}) d\left(\frac{t}{t'_{c_1}}\right) \\ + (1+r) \int_{t/t'_{c_1}}^{\infty} \left[\frac{K_L B}{K_u(1+r)}\right] \exp\left[-\frac{K_L(t-t_1)}{t'_{c_1}}\right] d\left(\frac{t}{t'_{c_1}}\right)$$

Performing the indicated integration and simplifying, noting that $Z = B$ at $t = t_1$ in (2.3.61) yields

$$\int_0^{\infty} \left(\frac{c}{c_{aq}}\right) d\left(\frac{t}{t'_{c_1}}\right) = (1+r) \left(1 - \frac{B + \frac{\hat{\xi}}{\xi}}{1 + \frac{\hat{\xi}}{\xi}} + \frac{B}{1 + \frac{\hat{\xi}}{\xi}}\right) \\ = (1+r)/(1 + \frac{\hat{\xi}}{\xi}) = 1 + m \quad (2.3.68)$$

since $\frac{\hat{\xi}}{\xi} = b_L(K_L - K_u)/h_o K_u = h_2(K_L - K_u)/(h_1 + h_2) K_u$ and $r = h_2 K_L / h_1 K_u$.

Equation (2.3.68) is exactly that obtained from (2.2.26); hence, we conclude that the combined lumped and linearized multi-layered Dupuit analyses yield exactly the same total mass convected from the system in the same time.

Figures 2.6 through 2.12 demonstrate that the output concentration history from the combined lumped parameter model is very nearly identical to that from a linearized multi-layered Dupuit aquifer analysis when the parameter μ in equation (2.2.23) is calculated according to equation (2.3.66). While we still have not demonstrated that either model accurately represents the physical situation depicted in Figure 2.2, we have shown that the combined lumped model is approximately equivalent to a two-dimensional representation of convective water and mass transport through a multi-aquifer system; furthermore, the total mass outflow from both systems is exactly the same. Thus if we conclude that a Dupuit

approximation is a valid assumption in a particular system analysis, we must also conclude that a combined lumped system analysis will approximately represent the Dupuit model, but in a hydraulically averaged sense.

In the comparison between the combined lumped model and multi-layer Dupuit aquifer model we used a simple flushing situation; that is, at time $t \geq 0$ all input contamination in recharge waters was taken to be zero. The reverse situation would be for initially uncontaminated aquifers at $t < 0$, and recharge waters carrying some constant contaminant input concentration c_0 at $t \geq 0$; all other conditions would remain unchanged. The resulting output concentration history in the stream would simply be identical to Figures 2.6 to 2.12, with the ordinate inverted, i.e. $c = c_0(1-y)$. A third hypothetical case may also be easily verified. If the input concentration at $t \geq 0$ is given by $c_i(t) = c_0$, a constant where $c_0 < c_{aq}$, then the output concentration history in the stream would also appear identical to Figures 2.6 to 2.12. However the ordinate would be translated upward by the constant c_0/c_{aq} , i.e. $c = (c_{aq} - c_0)y + c_0$.

2.4 Laplace Aquifer Model

2.4.1. Flow Field Description

In the preceding sections we showed that the lumped parameter model in the form of a linear reservoir could be related to a linearized Dupuit aquifer system. In the case of a single reservoir we saw that both models could be made to produce identical convective water and mass

transport results under simple flow conditions. When the system consisted of multiple aquifers, the lumped model with linear reservoirs approximated the linearized Dupuit analysis quite well. It should be clear, however, that neither model can be expected to accurately represent a given physical situation in which the original Dupuit assumptions are violated, or where net system stresses are highly localized. The first situation would lead us to at least consider the vertical flow effects on contaminant outflow concentration that were neglected in the horizontal Dupuit and lumped parameter models; the second situation would imply that hydrodynamic dispersion effects might be approaching the same order of magnitude as convective transport effects. Hydrodynamic dispersion has been investigated by many workers (for example, see Bear, 1972; Fried and Combarous, 1971; Fried, 1975; and references therein), and will not be addressed here. Subsurface vertical flow effects have also been the topic of considerable research effort. From this work one is likely to ask exactly when do the prediction errors in stream concentration resulting from neglecting these vertical effects actually become critical? Based on the preceding analysis, we might suspect that such errors will grow in some direct relation to which the Dupuit assumptions are violated. For example, Murray and Monkmeyer (1973) reported that for steady flow conditions, a free surface water table slope of less than 0.10 will produce less than 1 percent difference in the predicted water table shapes of the Dupuit and Laplace aquifer models. Verma and Brutsaert (1971) showed that the predicted nonlinear Dupuit and Laplace aquifer free surfaces were in reasonable agreement as long as the aquifer length was four times its initial saturated thickness. Neglecting any hydrodynamic dispersion and

hydrochemical interaction effects between the fluid and porous media, we might also expect the predicted outflow concentrations from such aquifer models to also be similar under such conditions.

Since these aquifer models are not exactly equivalent (the Dupuit aquifer model assumes a fully penetrating stream while the Laplace aquifer as used here has only a partially penetrating stream of some finite width), we would not expect to analytically relate differences in predicted free surface shapes to predicted outflow concentrations. We can gain some insight on the physical behavior of these models, however, by comparing their concentration outflow histories under simple flow conditions.

Figure 2.13 shows a schematic diagram of some idealized phreatic aquifer hydraulically connected to a partially penetrating stream or tile drain of some finite width. We will refer to this diagram as the idealized Laplace aquifer since some form of the so called Laplace equation, subject to appropriate initial and boundary conditions, will describe the spatial distribution of the piezometric head, h , or the streamline distribution. The Laplace equation describes these distributions since it is obtained by combining the continuity equation with Darcy's law (Bear, 1972). Using the vector summation convention, Darcy's law will be given by

$$q_i = - K_{ij} \frac{\partial h}{\partial x_j} ; (i, j = 1, 2, 3)$$

where the principal axes of the hydraulic conductivity tensor, K_{ij} , may be assumed to lie in the same direction as our orthogonal coordinate

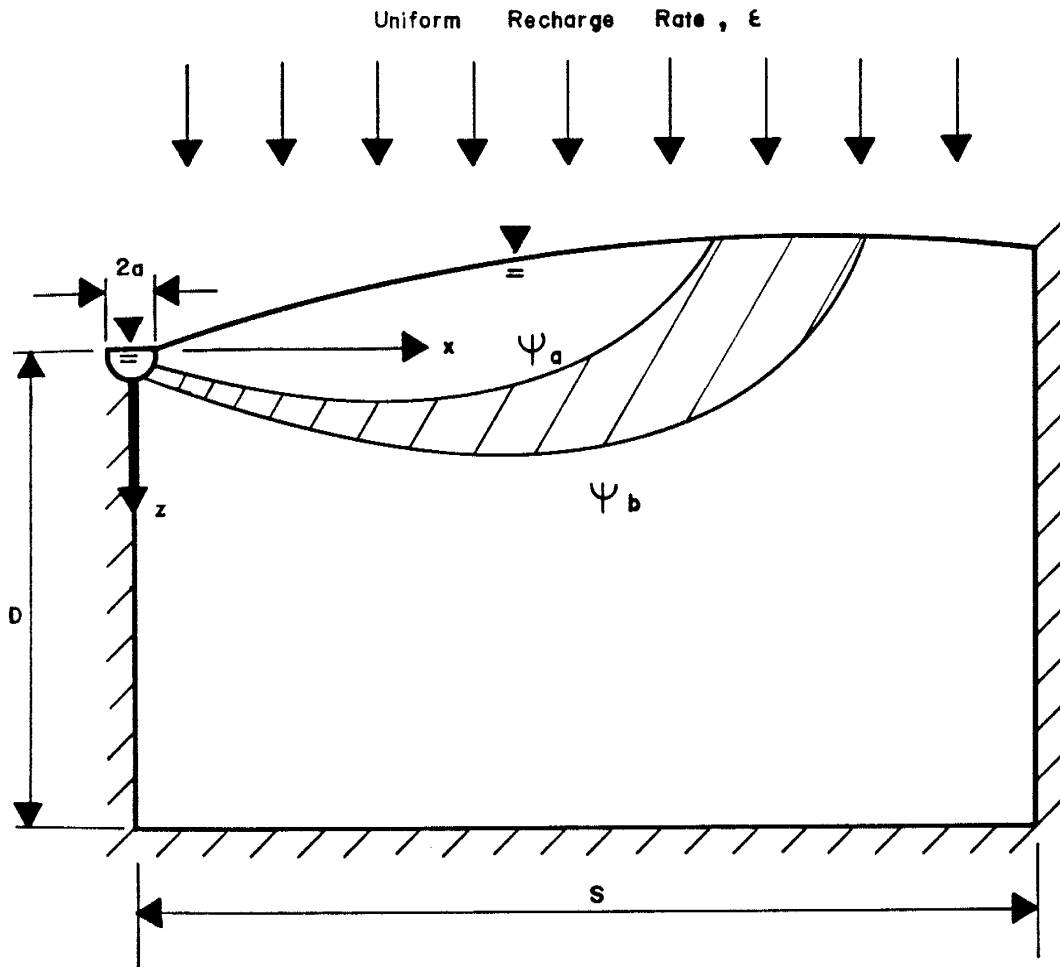


Figure 2.13 Schematic diagram for an idealized Laplace aquifer showing one possible streamtube.

system. The continuity equation for a non-deforming, saturated porous media with an incompressible fluid is given by (Bear, 1972)

$$\frac{\partial q_i}{\partial x_i} = 0; \quad (i = 1, 2, 3)$$

where the system contains no sources or sinks. Combining these equations yields

$$\frac{\partial}{\partial x_i} \left(K_{ij} \frac{\partial h}{\partial x_j} \right) = 0; \quad (i = 1, 2, 3) \quad (2.4.1)$$

This last equation is sometimes referred to as Richard's equation in the soil science literature; however, appropriate coordinate transformations will produce the familiar Laplace form. Thus if we let

$$X_i = x_i (K_i/K_j)^{1/2}, \quad \text{where } i = 1, 2, 3 \\ \text{and } j = \text{either } 1, 2, \text{ or } 3$$

then (2.4.1) becomes

$$\frac{\partial^2 h}{\partial X_i^2} = 0; \quad (i = 1, 2, 3) \quad (2.4.2)$$

for a homogeneous medium and where the x_i are the principal axes of the conductivity tensor. Equation (2.4.1) is valid for steady or unsteady flow in heterogeneous, anisotropic porous media; (2.4.2) would apply to steady or unsteady flow in homogeneous, anisotropic media. Both equations are obviously subject to the implied limitation contained in

the continuity equation. Note that any time dependency will be introduced through the boundary conditions.

If we wish to describe the instantaneous spatial distribution of the loci of points which are everywhere tangent to the specific discharge vector, q_i , then we must find all of the infinitesimally small position vectors that are everywhere tangent to q_i . These small position vectors will, in the limit, reduce to the required loci of points and describe the streamline. Mathematically a streamline will be defined by the vector cross product $\hat{q} \times d\hat{r} = 0$, where $d\hat{r}$ is some small position vector tangent to \hat{q} . Thus $\hat{i}(q_z dy - q_y dz) = \hat{j}(q_x dz - q_z dx) = \hat{k}(q_y dx - q_x dy) = 0$, where \hat{i} , \hat{j} , and \hat{k} are the unit vectors in the x, y, z coordinate system, respectively. This last relation may be written as

$$\frac{dx}{q_x} = \frac{dy}{q_y} = \frac{dz}{q_z}$$

For two dimensional flow in the x - z plane we obtain

$$q_z dx - q_x dz = 0.$$

The solution to this equation is given by the stream function $\Psi = \Psi(x, z)$ = a constant, which describes the instantaneous streamline distribution; Ψ is an exact differential only if $\partial q_i / \partial x_j = 0$, ($i=1, 2$). Since $\Psi(x, z)$ is to be exact then

$$d\Psi = 0 = \frac{\partial \Psi}{\partial x} dx + \frac{\partial \Psi}{\partial z} dz = q_z dx - q_x dz$$

Thus $q_x = -\frac{\partial \Psi}{\partial z}$ and $q_z = \frac{\partial \Psi}{\partial x}$. For macroscopic irrotational flow in

porous media where the principal axes of the hydraulic conductivity tensor are parallel to the cartesian coordinate system, $\nabla_x(\nabla h) = 0$ (Bear, 1972). If (x, z) are principal axes of the hydraulic conductivity this relationship may be combined with Darcy's law to yield $\nabla_x(-q_x/K_x, q_z/K_z) = 0$. For two-dimensional flow in non-homogeneous, anisotropic media this last expression becomes

$$\frac{\partial}{\partial x} \left(\frac{1}{K_z} \frac{\partial \psi}{\partial x} \right) + \frac{\partial}{\partial z} \left(-\frac{1}{K_x} \frac{\partial \psi}{\partial z} \right) = 0 \quad (2.4.3)$$

For homogeneous, isotropic media, (2.4.3) is simply Laplace's equation, or

$$\frac{\partial^2 \psi}{\partial x^2} + \frac{\partial^2 \psi}{\partial z^2} = 0 \quad (2.4.3a)$$

subject to the implied boundary conditions depicted in Figure 2.13. Thus for steady flow we have

$$\begin{aligned} (1) \quad \psi &= \psi_0 \text{ on } x = 0, 0 \leq z \leq D \\ (2) \quad \psi &= \psi_0 \text{ on } z = D, 0 \leq x \leq L \\ (3) \quad \psi &= \psi_0 \text{ on } x = L, 0 \leq z \leq D \\ (4) \quad \psi &= \left(\frac{x-a}{S-a} \right) \psi_0 \text{ on } z = 0, a \leq x \leq L \\ (5) \quad \psi &= \left(\frac{a-x}{a} \right) \psi_0 \text{ on } z = 0, 0 \leq x \leq a \end{aligned} \quad (2.4.3b)$$

where $\psi_0 = \epsilon L$ and the medium is isotropic and homogeneous. Using separation of variables and superposition of solutions, Kirkham (in

Kirkham and Powers, 1972, eq. 3-71; or see Carslaw and Jaeger, 1959, p. 167, eq. 9) obtained the solution

$$\psi(x,z) = \psi_0 - \frac{2\psi_0}{\pi} \sum_{m=1}^{\infty} \frac{\sin\left(\frac{m\pi x}{L}\right) \sinh\left[\frac{m\pi}{L}(D-z)\right]}{m \sinh\left(\frac{m\pi D}{L}\right)}$$

which is valid when $a \ll L$. If we let

$$\begin{aligned} \bar{\psi} &= \psi / \psi_0 \\ X &= x/L \\ Z &= z/D \\ n &= L/D \end{aligned} \tag{2.4.4}$$

Then we may express the solution to (2.4.3a) in dimensionless form, or

$$\bar{\psi}(X,Z) = 1 - \frac{2}{\pi} \sum_{m=1}^{\infty} \frac{\sin(m\pi X) \sinh\left[\frac{m\pi}{n}(1-Z)\right]}{m \sinh(m\pi/n)} \tag{2.4.5}$$

The boundary conditions employed in arriving at the solution to (2.4.3a) imply that the water table height above $x = 0$ in Figure 2.13 is zero; that is, (2.4.5) neglects the height of the water table above the stream or tile drain elevation. If this height is small, then (2.4.5) will approximate the streamline distribution adequately. Thus our solution to the Laplace model will differ from the Dupuit or lumped parameter aquifer analyses previously given because vertical flow effects resulting from partial stream penetration have been included.

2.4.2. Convective Outflow Concentration from the Laplace Aquifer Model

We have previously established the convective water and mass transport equivalence between the lumped parameter model and the linearized Dupuit model under steady flow conditions. We will now expand our analysis to include a quantitative comparison of convective mass transport through the steady flow field of the Laplace aquifer model. Again we may conceptually visualize a small parcel of contaminated water following some given space-time trajectory through the steady flow field of the Laplace aquifer; these trajectories may be approximately described by (2.4.5). We can determine the total change in quality of all such parcels at the partially penetrating stream or tile drain where all parcels begin to mix together. As before we must of course assume that no chemical interaction effects between adjacent water parcels of different quality or between water parcels and the porous medium occur. Hydrodynamic dispersion effects are likewise neglected.

The streamlines described by (2.4.5) represent the actual flow path followed by the moving parcels of contaminated water through the aquifer model. If conservation of mass is to be maintained, then the total fluid discharge between any two adjacent streamlines must remain constant. Thus the discharge between any two streamlines is

$$Q_{AB} = \int_A^B (q_x dz - q_z dx) = \int_A^B (-d\psi) = \psi_A - \psi_B$$

where $d\psi = q_z dx - q_x dz$ was obtained in the previous section. The dimensions of ψ are volume per unit time per unit of width. According to the sign convention employed here, $\psi_B > \psi_A$, so Q_{AB} is negative.

In Figure 2.13 a hypothetical streamtube bounded by the streamlines ψ_A and ψ_B is illustrated as the shaded area. Recharging water entering this streamtube at the water table will eventually discharge in the partially penetrating stream or tile drain. The volume of water entering this tube per unit time, assuming a unit length in the third dimension, is given by $Q = \epsilon \Delta x \hat{j}$. The fluid velocity at any point within this streamtube will be $ds/dt = Q/nA$, where ds is the differential arclength along the flow path, n is the effective porosity of the porous medium, and A is the unit cross-sectional area of the tube. The time required for a parcel of contaminated water to pass between two points s_1 and s_2 in the tube is therefore

$$\int_{t_1}^{t_2} dt = \int_{s_1}^{s_2} (nA/Q) ds = \frac{n}{\epsilon \Delta x} \int_{s_1}^{s_2} A ds$$

where the integral $n \int A ds$ is the volume of water (the x - z plane area times a unit thickness in the third dimension) in the streamtube between s_1 and s_2 . Therefore

$$t_2 - t_1 = t = n \Delta A / \Delta \psi$$

Written in differential form this last expression is

$$t = \frac{n dA(\psi)}{d\psi}$$

where $dA(\psi)$ is a single valued function of ψ . Now if we let

$$\begin{aligned}
 A_0 &= LD, \text{ the aquifer cross-sectional area,} \\
 \Psi_0 &= \epsilon L \text{ as before,} \\
 \bar{A}(\bar{\Psi}) &= A(\Psi)/A_0, \text{ a dimensionless area, and} \\
 \bar{\Psi} &= \Psi/\Psi_0, \text{ a dimensionless streamline,}
 \end{aligned}$$

then we see that

$$t = \frac{nA_0 d\bar{A}(\bar{\Psi})}{\Psi_0 d\bar{\Psi}} = \frac{nLDd\bar{A}(\bar{\Psi})}{\epsilon Ld\bar{\Psi}}$$

Thus,

$$t/t_c = d\bar{A}(\bar{\Psi})/d\bar{\Psi} \quad (2.4.6)$$

where $t_c = nD/\epsilon$ was previously defined as the solute response time for the lumped parameter model. Equation (2.4.6) says that the dimensionless travel time, t/t_c , of some parcel of water following some specified streamline will be given by the slope of the dimensionless $d\bar{A}$ versus $d\bar{\Psi}$ curve. If we know how the streamline cross-sectional area changes for different streamtubes, then we can find the travel time of some parcel following the intermediate streamline contained in that streamtube. Note that for the aquifer depicted in Figure 2.13, $\bar{A}(\bar{\Psi})$ is only a function of $\bar{\Psi}$. Using (2.4.5) we may numerically integrate between streamlines to obtain the area. Jury (1975) used a Gaussian quadrature technique in a similar procedure (Jury, 1976, personal communication) to obtain solute travel time estimates for tile-drained fields where (2.4.5) approximately defines the streamline distribution. His approach, however, is limited to those idealized situations where an analytical solution to the Laplace

equation is possible. These solutions are not generally practical, or even possible, when irregular boundaries are present or when the porous medium is anisotropic and non-homogeneous.

The dimensionless concentration appearing in the stream or tile drain of Figure 2.13 at the dimensionless time given by (2.4.6) can be found as before. Thus if some particle enters the aquifer system at $\bar{\psi} = x/L$ at time $t/t_c = 0$, it will follow the space trajectory described by (2.4.5) to the stream or tile drain. If the aquifer is initially contaminated at some uniform concentration level, c_{aq} , at time $t/t_c \leq 0$, and for $t/t_c \geq 0$ no new contamination is introduced with the recharge waters, then the dimensionless concentration in the stream or drain will be given by

$$c/c_{aq} = 1 - \bar{\psi} \quad (2.4.7)$$

This concentration level will appear in the stream after the time given by (2.4.6).

Unless we have some analytical function that relates $\bar{A}(\bar{\psi})$ to $\bar{\psi}$ in (2.4.6), we must evaluate this derivative numerically. One possibility to obtaining the desired solution is to develop a finite difference or finite element program that will solve for the spatial distribution of $\bar{\psi}(X,Z)$. Once this distribution is known, then we may proceed to contour in the streamlines and evaluate the area under each of several specified values of $\bar{\psi}$. Finally we may plot a curve of $\bar{A}(\bar{\psi})$ versus $\bar{\psi}$ and graphically or numerically evaluate the slope along various points; this information will enable us to develop concentration break-through curves (i.e., c/c_0 vs. t/t_c) for various aquifer geometry configurations. This procedure is illustrated in Figures 2.14 and 2.15 below. Note in

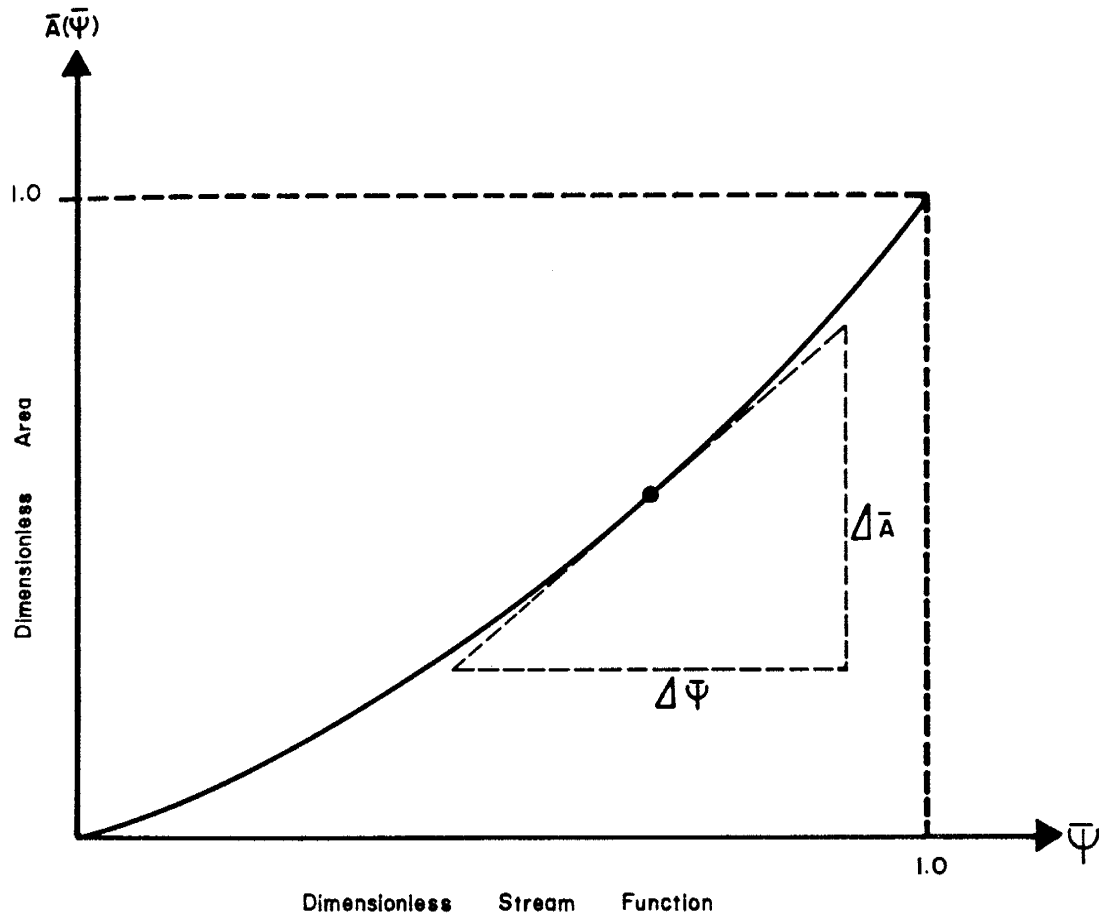


Figure 2.14 Idealized relationship indicating the total dimensionless area swept out by a given dimensionless streamline.

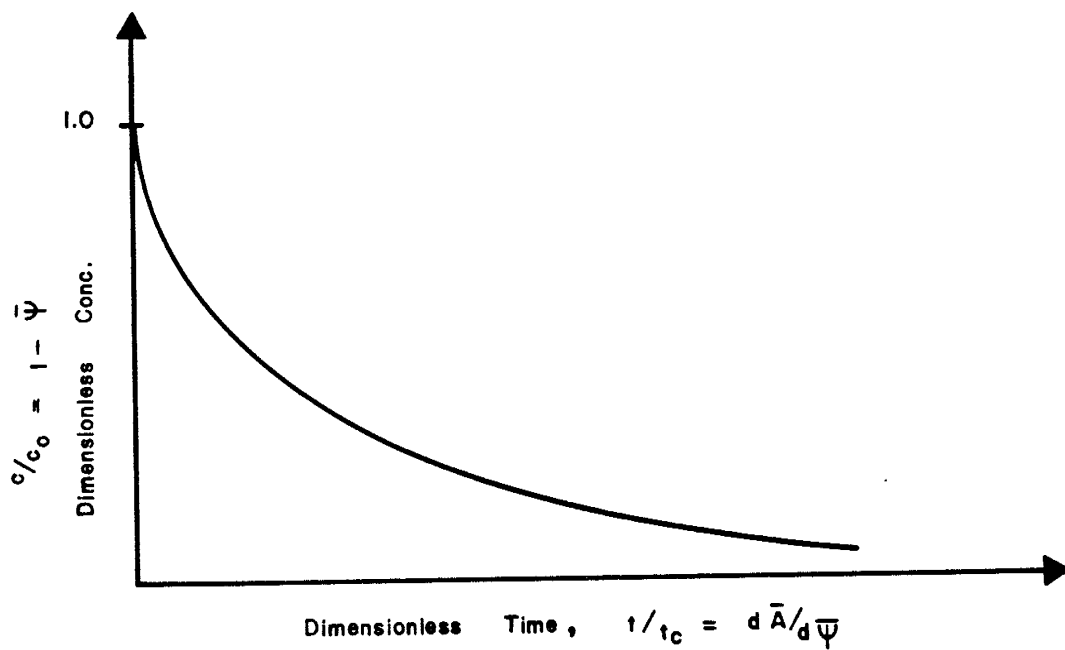


Figure 2.15 Idealized breakthrough curve computed from the area versus streamline relationship.

Figure 2.14 that the dimensionless area swept out by the dimensionless streamline $\bar{\psi} = 0$ is zero, and that at $\bar{\psi} = 1$, $\bar{A} = 1$. These values correspond to a dimensionless concentration of 1 in the stream at $t/t_c = 0$ (the onset of non-contaminated recharge water infiltration) in Figure 2.15; similarly for $\bar{\psi} = 1$ we have $c/c_0 = 0$, which requires an infinite travel time. In other words a particle flowing along $\bar{\psi} = 1$ (the aquifer boundary) will require an infinitely large travel time to reach the stream since $d\bar{A}/d\bar{\psi} = \infty$ at $\bar{\psi} = 1$.

In arriving at a solution to our problem, the critical step is obtaining an accurate $\bar{A}(\bar{\psi})$ versus $\bar{\psi}$ relationship. Most existing computer models for subsurface hydrology, however, are designed to solve (2.4.1) for the spacial distribution of piezometric head, h . One existing finite element program (U.S. Corps of Engineers, 1972) may be easily modified to solve (2.4.3); however, care must be used to also modify the implied boundary conditions. This particular program also has the advantage that anisotropic, non-homogeneous problems containing sources and sinks are easily handled for any irregular flow zone geometry.

The governing partial differential equation for an area containing no sources or sinks is given by (2.4.3), subject to the boundary conditions implied in Figure 2.13, and given by (2.4.3b). Recall that we have assumed that $a \ll L$. We may use the above procedure and the modified U.S. Corps of Engineers program to solve several problems, varying L , D , K_x , and K_z in combination. Since the same coordinate grid network may be used in all of the analyses, an internal grid generator was incorporated into the model program. Finally, the program was further modified so as to track specified streamlines from $x=0$ to the stream-aquifer boundary. A linear interpolation scheme was employed to

track streamline coordinates between grid lines so that the area under a given $\bar{\psi}$ could be simultaneously calculated using a simple trapezoidal summation. The complete modified U.S. Corps of Engineers program is listed in Appendix B along with a partial sample output. Output results were modified so that $\bar{A}(\bar{\psi})$ could be plotted against $\bar{\psi}$; the corresponding c/c_0 versus t/t_c curves could then be easily computed for various $(L/D)(K_z/K_x)^{1/2}$ ratios. In the program a 41 by 21 mesh was used in conjunction with quadrilateral (rectangular) elements; Figure 2.16 shows a diagram of this grid network.

2.4.3 Effects of Partial Stream Penetration

Using the procedure outlined in the previous section, we may quantify the concentration outflow history from the Laplace aquifer of Figure 2.13 for various ratios of L/D and a step change in input concentration at time $t/t_c = 0$ (the flushing system previously employed). The ratio, $n = L/D$, may be referred to as an aspect ratio that characterizes the aquifer geometry. The stream or tile drain is assumed to be a point sink located at the (x,z) coordinates $(0,0)$. The Laplace model will yield a streamline distribution similar to that shown in Figure 3-8 of Kirkham and Powers (1972, p. 95), from which the $\bar{A}(\bar{\psi})$ versus $\bar{\psi}$ curves for various values of n may be constructed. Equations (2.4.6) and (2.4.7) of the previous section will yield the corresponding c/c_0 versus t/t_c curves for a flushing aquifer system. These curves are shown in Figure 2.17 for values of $n = 1$ through 5, and $n = 10$. For the sake of clarity only, the curves for values of $n = 1, 5,$ and 10 are shown in Figure 2.18. Also shown in Figure 2.18 as a reference is the

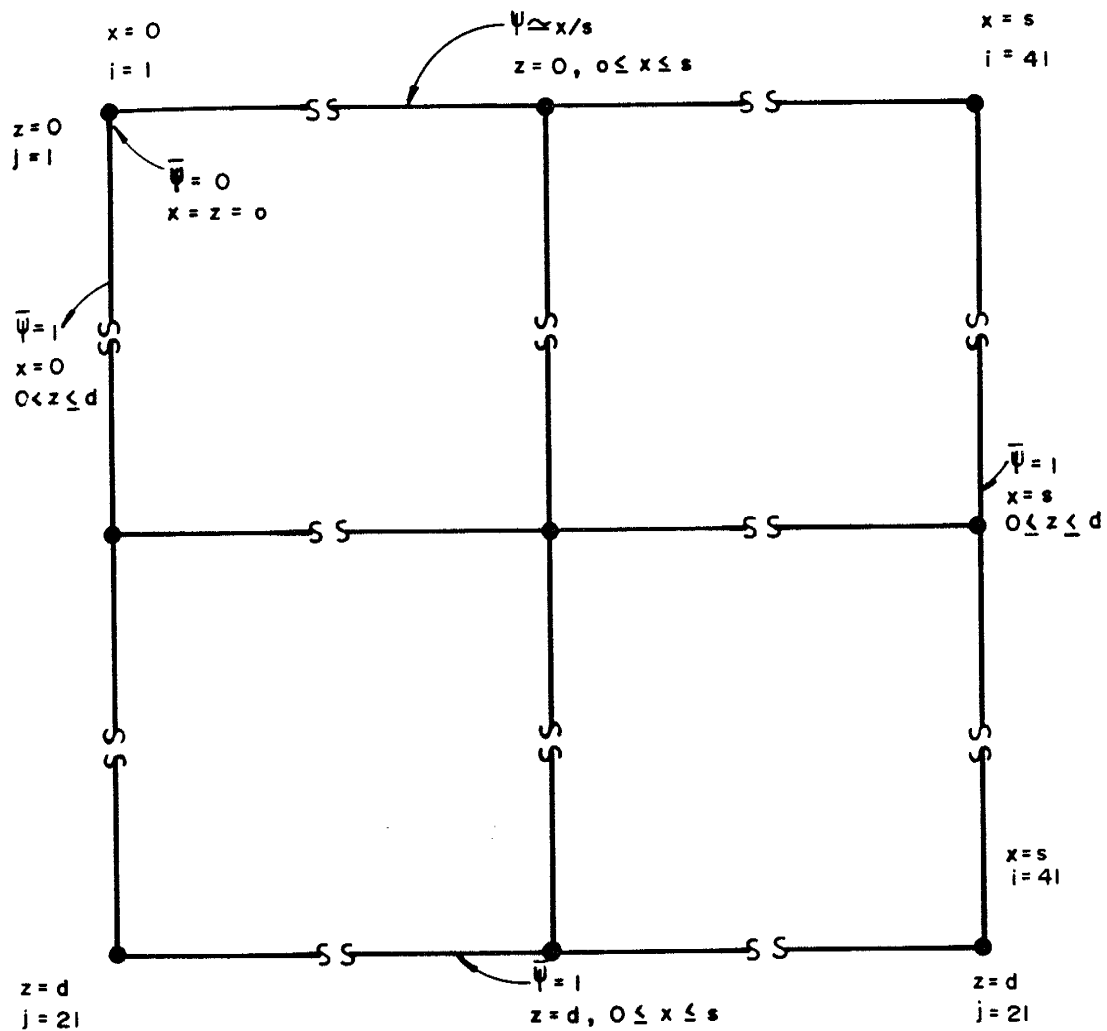


Figure 2.16 Program grid network showing boundary conditions.

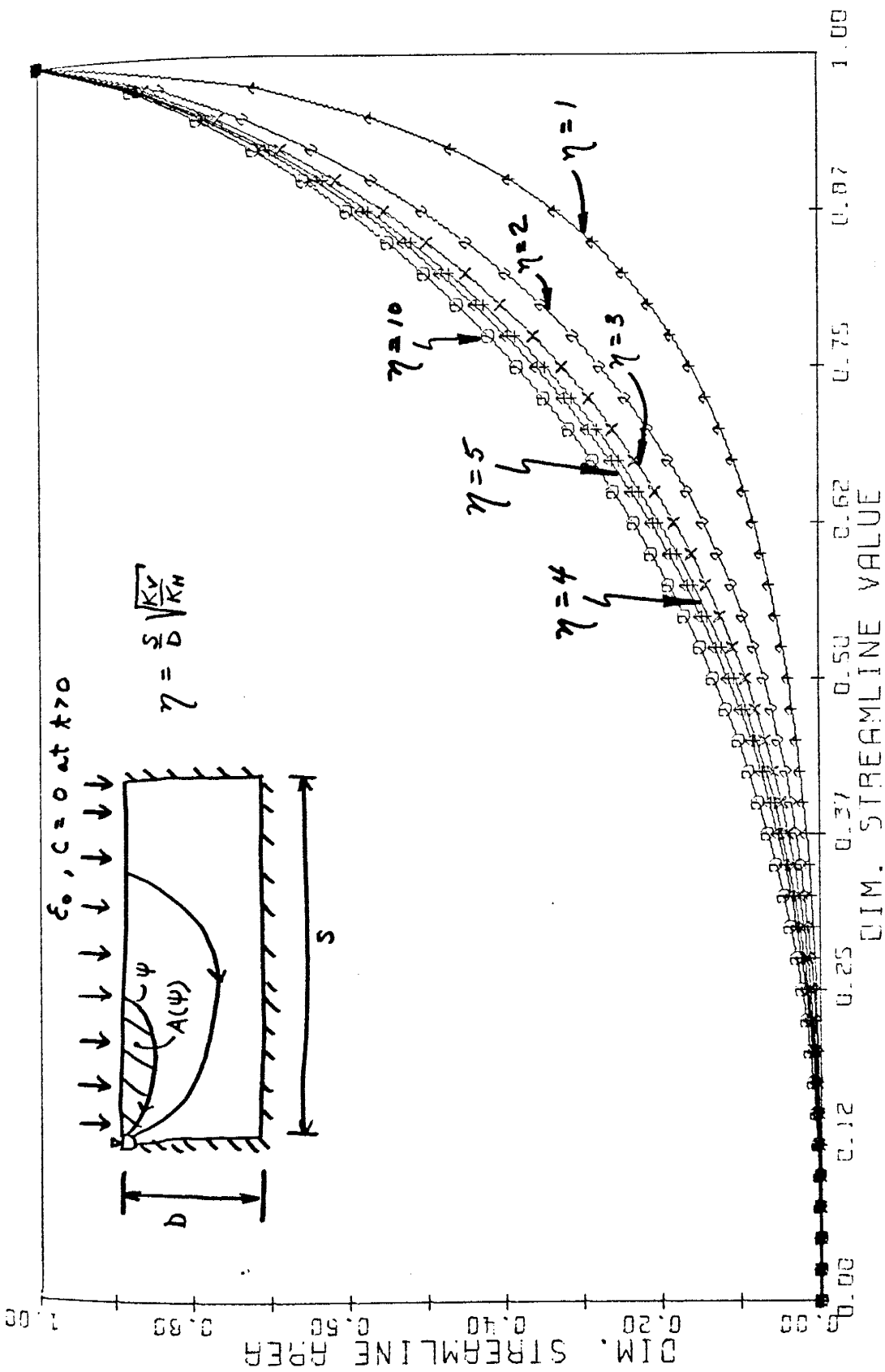


Figure 2.17: $\bar{A}(\bar{\psi})$ as a function of $\bar{\psi}$ for different values of η .

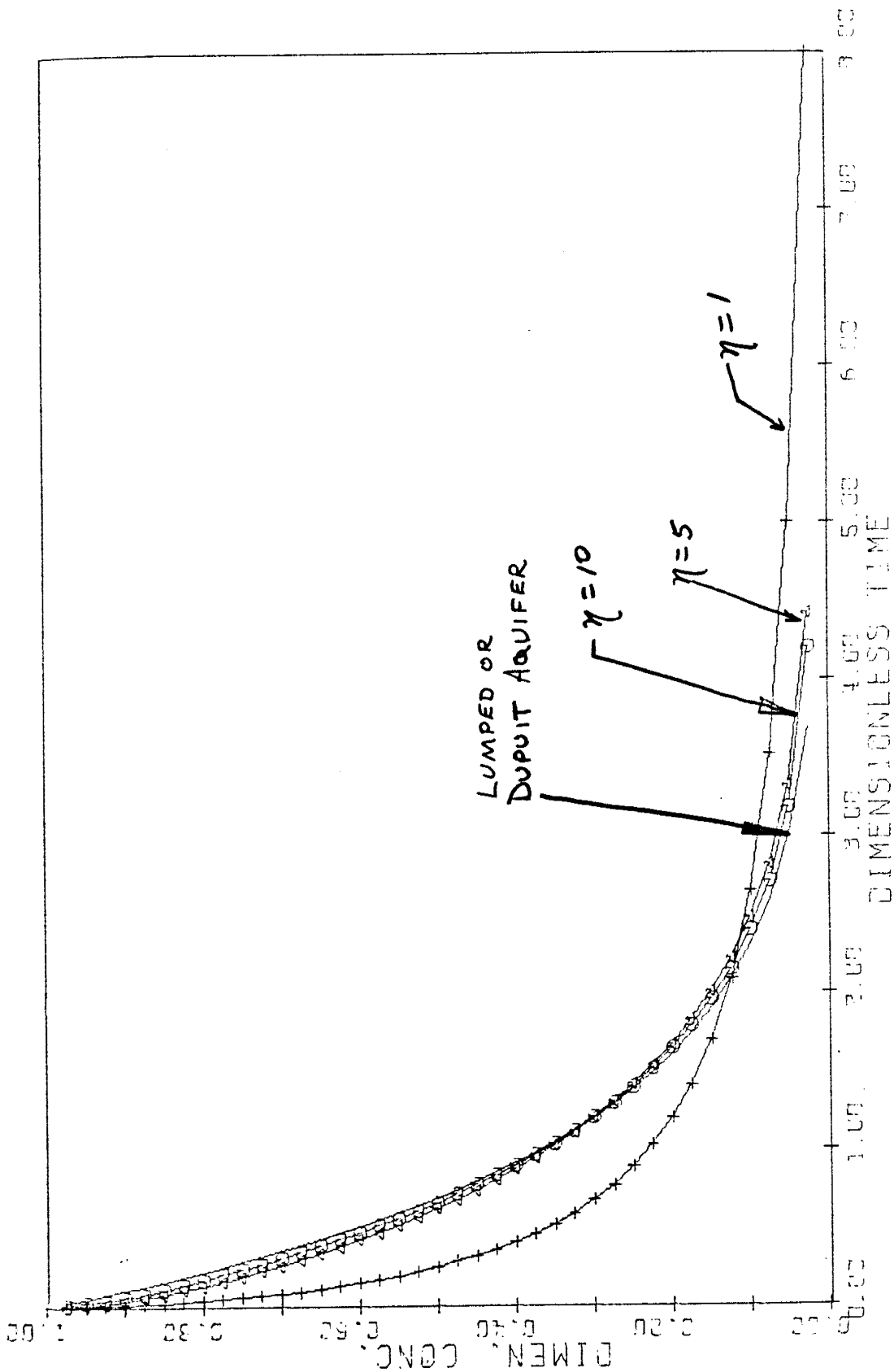


Figure 2.18: c/c_0 as a function of t/t_c for different values of η .

lumped parameter outflow concentration previously calculated; this lumped concentration was shown to be identical to that obtained in the Dupuit aquifer analysis. From Figure 2.18 we may conclude that the larger the aspect ratio, the smaller the difference in outflow concentrations from the Laplace and lumped parameter (or Dupuit) models. Another way of stating this conclusion is that the vertical flow effects on the outflow concentration resulting from partial stream penetration become small for a larger drain spacing to aquifer depth ratio. Actually we would expect these concentration differences to be small since we already know that the vertical flow effects (Kirkham and Powers, 1972; and others) become less pronounced with increasing values of n . As seen in Figure 2.18 for a value of $n \geq 5$, there is practically no difference in the predicted outflow concentrations for the models and flow conditions used in this analysis. A similar value of $n \geq 4$ has also been obtained for differences in the water table shape predicted from several distributed models applied to a fully penetrating well in a phreatic aquifer (Elbakhbekhi, 1976).

The above results are similar to those obtained by Jury (1975a) using a slightly different solution technique; however, he chose to plot the dimensionless distance from the tile drain, x/L , versus the dimensionless travel times. He did compare his model results to observed chloride effluent concentration as a function of time for a ponded leaching experiment and obtained remarkably good results. He did not relate his model behavior to that of the lumped parameter model. If one were to plot his results depicted in Figure 1 (Jury, 1975b) on a semi-logarithmic scale, the exponential behavior characteristic of the lumped parameter model with a linear reservoir could be approximated.

This type of exponential behavior is shown in Figure 2.19, plotting the results shown in Figure 2.18 on a semi-log scale. Note that the lumped parameter concentration break-through curve plots as a straight line, while Laplace model curves are distorted away from this line for different values of n .

2.4.4 Effects of an Anisotropic Porous Medium

Again using the same procedure previously employed, along with similar flow boundaries and input concentration initial conditions, we may quantify the concentration outflow history from an anisotropic porous medium with a partially penetrating stream. The only change required in the modified Corps of Engineers program (Appendix B) is in the specified input data for hydraulic conductivity. Mathematically we must now solve (2.4.3), which for a homogeneous porous medium becomes

$$K_x \frac{\partial^2 \Psi}{\partial (x')^2} + K_z \frac{\partial^2 \Psi}{\partial (z')^2} = 0 \quad (2.4.8)$$

where the primed notation refers to the real Cartesian coordinate system. Equation (2.4.8) is subject to the same boundary conditions given by (2.4.3b). Note, however, that the primed notation for x and z should be used. We may transform (2.4.8) to an equivalent isotropic system (EIS) using the well-known transformations given by

$$x = x' (K_z/K_x)^{1/2} \quad \text{and} \quad z = z' \quad (2.4.9)$$

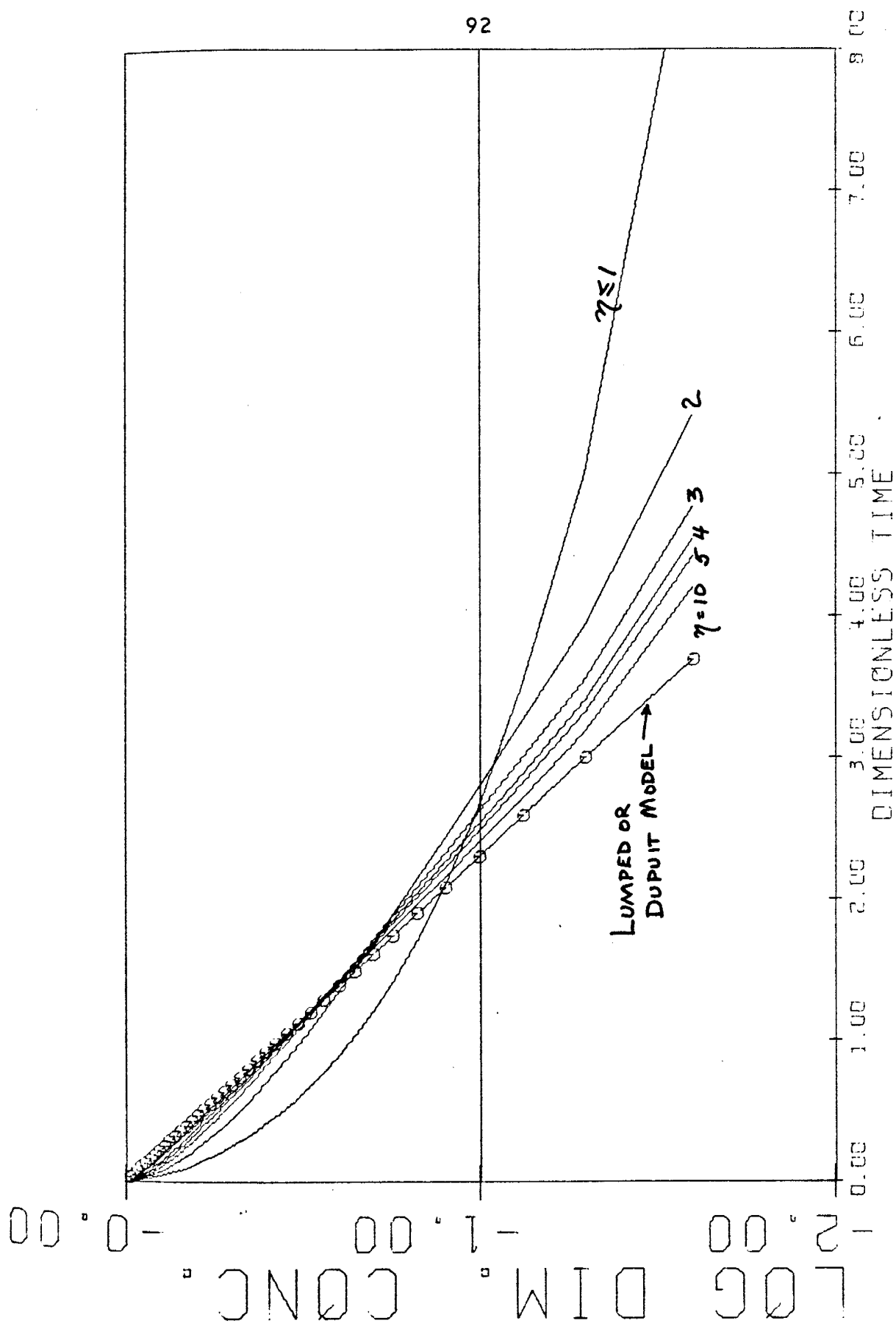


Figure 2.19: $\text{Log } (c/c_0)$ as a function of t/t_c for different values of η .

where x and z represent the EIS Cartesian coordinate scale. These coordinate transformations are equivalent to a shrinkage or expansion of the real x' coordinate scale so that the resulting perzometric head contours may be graphically drawn perpendicular to streamlines in the transformed x - z coordinate system. Using (2.4.9), (2.4.8) is transformed to

$$\frac{\partial^2 \psi}{\partial x^2} + \frac{\partial^2 \psi}{\partial z^2} = 0 \quad (2.4.10)$$

subject to the transformed boundary conditions given in (2.4.3b) without the primed notation. The solution to (2.4.10) in the EIS scale, subject to the appropriate boundary conditions, is given by (2.4.5) using the dimensionless parameters defined in (2.4.4). We may again use (2.4.9) to transform our EIS solution back to the real x' - z' coordinate system. We, therefore, obtain the solution in dimensionless form as

$$\bar{\psi}(X', Z') = 1 - \frac{2}{\pi} \sum_{m=1}^{\infty} \frac{\sin(m \pi X') \sinh[m \pi (1-Z')/\eta']}{m \sinh(m \pi / \eta')} \quad (2.4.11)$$

where the primed notation refers to the real Cartesian coordinate system and $\eta' = L'/D' (K_z/K_x)^{1/2}$. Dropping the primes we see that (2.4.11), the solution for an anisotropic and homogeneous media, is identical to (2.4.5) for the isotropic homogeneous case with the exception that $\eta = L/D$ in (2.4.5). Actually the solution given by (2.4.5) is a special case (2.4.11) with $K_x = K_z$.

We may still use (2.4.6) and (2.4.7) to calculate concentration break-through curves for anisotropic, homogeneous media. Thus for the flushing system used in the previous section, we obtain identical

solutions to those depicted in Figures 2.17 through 2.19, with the exception that $n = L/D (K_z/K_x)^{1/2}$. Again we may conclude that as long as $n \geq 5$, there is practically no difference in the predicted outflow concentrations for the models and flow conditions used in the analysis.

A crude conservation of mass check may be used for each of the break-through curves developed from the Laplace model analysis. Thus, recalling that we are employing the flushing system boundary and initial conditions here, the area under each c/c_0 versus t/t_c curves must approach 1.0 given sufficient time. This is simply another way of saying that all of the mass originally present in aquifer waters must be flushed out after some long time t . The dimensionless areas calculated for each of the aspect ratios are summarized in Table 2.2; however, in computing these areas, the integration procedure was terminated when $c/c_{aq} = 0.025$.

Table 2.2: Comparison between the Laplace model and the generalized lumped parameter model.

<u>n</u> (1)	<u>LAPLACE</u> (2) <u>AREA</u>	<u>LUMPED</u> <u>AREA</u>
1	0.964	0.977
2	0.985	0.977
3	0.986	0.977
4	0.986	0.977
5	0.986	0.977
10	0.976	0.977

(1) The aspect ratio is defined as $L/D(K_z/K_x)^{1/2}$

(2) Areas under the dimensionless concentration (c/c_0) versus dimensionless time (t/t_c) curves for (c/c_0) greater than 0.025 as obtained from the Laplace model.

2.4.5 Impermeable Basement at Infinite Depth

Up to now we have examined the effects of vertical flow on the aquifer outflow concentration resulting from varying the drain half-spacing aquifer depth ratio and the hydraulic conductivity. These vertical flow effects were lumped into a single parameter, defined as $n=(L/D)(K_z/K_x)^{1/2}$. In the analysis, we always assumed that an impermeable basement was located at some finite distance, D , below the stream or tile drain. An obvious question that arises is, What happens to the concentration outflow history when D approaches infinity? Figure 2.20 schematically depicts the flow field in question. The governing partial differential equation is identical to (2.4.8) for an anisotropic, homogeneous medium subject to the implied boundary conditions in Figure 2.20. Thus,

$$\begin{aligned}
 (1) \quad & \psi = \psi_0 = \epsilon L \text{ on } x = 0; 0 \leq z \leq \infty \\
 (2) \quad & \lim_{z \rightarrow \infty} \psi = \psi_0 \text{ on } z \rightarrow \infty; 0 \leq x \leq L \\
 (3) \quad & \psi = \psi_0 \text{ on } x = L; 0 \leq z \leq \infty \\
 (4) \quad & \psi = \left(\frac{x-a}{L-a} \right) \psi_0 \text{ on } z = 0; a \leq x \leq L \\
 (5) \quad & \psi = \left(\frac{a-x}{a} \right) \psi_0 \text{ on } z = 0; 0 \leq x \leq a
 \end{aligned}$$

Using the transformations in (2.4.9), the governing differential equation in the EIS coordinate scale looks like that given in (2.4.3a). The solution has been found to be (Carslaw and Jaeger, 1959, p. 163, eq. 6).

$$\bar{\psi} = 1 - \frac{2}{\pi} \sum_{m=1}^{\infty} \left(\frac{1}{m} \right) \exp(-m \pi Z) \sin(m \pi X) \quad (2.4.12)$$

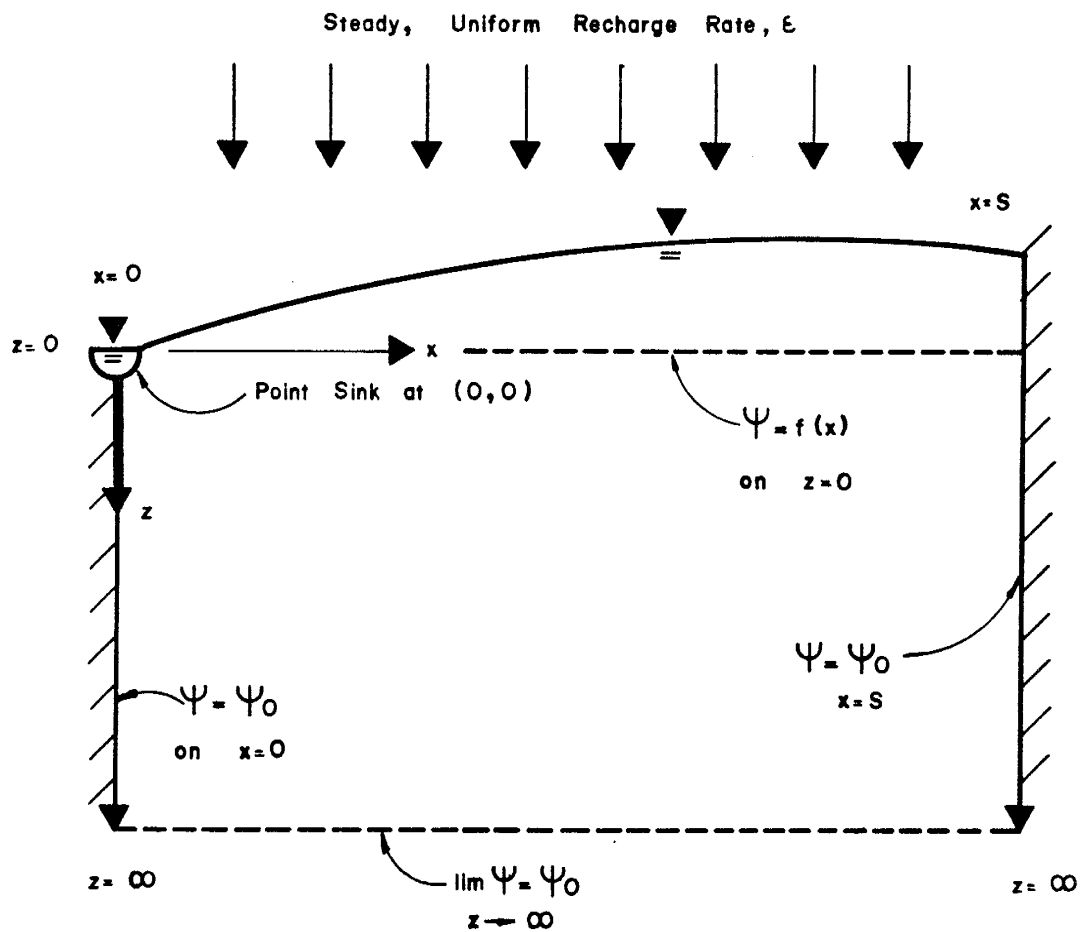


Figure 2.20 Schematic Laplace aquifer with impermeable basement at an infinite depth below the drain.

where the dimensionless parameters previously defined in (2.4.4) are used. In (2.4.12) we have assumed that $a \ll L$, as was done earlier.

The Fourier series solution found by Kirkham, and given here as (2.4.5), may be written for $n = 1$ as

$$\bar{\psi} = 1 - \frac{2}{\pi} \sum_{m=1}^{\infty} \frac{\sin(m\pi X) \sinh[m\pi(1-Z)]}{m \sinh(m\pi)} \quad (2.4.13)$$

Using the trigometric identity $\sinh(A-B) = \sinh(A) \cosh(B) - \cosh(A) \sinh(B)$ we see that

$$\frac{\sinh[m\pi(1-Z)]}{\sinh(m\pi)} = \frac{\sinh(m\pi) \cosh(m\pi Z) - \cosh(m\pi) \sinh(m\pi Z)}{\sinh(m\pi)}$$

which may be simplified to

$$\frac{\sinh[m\pi(1-Z)]}{\sinh(m\pi)} = \cosh(m\pi Z) - \frac{\sinh(m\pi Z)}{\tanh(m\pi)}$$

where the term $1/\tanh(m\pi) = 1.0037$ for $m = 1$, and rapidly converges to 1.0 for all $m > 1$. Therefore, (2.4.13) may be approximately written as

$$\bar{\psi} = 1 - \frac{2}{\pi} \sum_{m=1}^{\infty} \frac{\sin(m\pi X)}{m} [\cosh(m\pi Z) - \sinh(m\pi Z)]$$

where $\cosh(m\pi Z) - \sinh(m\pi Z) = \exp(-m\pi Z)$. Thus (2.4.13) approximates the expression

$$\bar{\psi} \approx 1 - \frac{2}{\pi} \sum_{m=1}^{\infty} \left(\frac{1}{m}\right) \exp(-m\pi Z) \sin(-m\pi X) \quad (2.4.14)$$

This is exactly the solution found in (2.4.12) for the infinite depth aquifer. We may then conclude that the streamline distribution for $n = (L/D) K_z/K_x)^{1/2} = 1$ is nearly identical to that obtained for an infinitely deep impermeable basement. Muskat (1946) has also found nearly identical streamline distributions for the cases of $z \rightarrow \infty$ and $z = L$. The corresponding concentration outflow history for $n \rightarrow 0$ is approximately the same as for $n = 1$. Thus in Figures 2.17 through 2.19 these curves are indicated as $n \leq 1$.

2.4.6 Multiple Aquifer Systems

In sections 2.4.2 through 2.4.5, we have obtained the concentration outflow histories for a partially penetrating stream or tile drain for a variety of $(L/D)(K_z/K_x)^{1/2}$ ratios; we have compared these solution to that obtained from the lumped parameter model with a linear reservoir term (or Dupuit aquifer model with a fully penetrating stream). Explicit assumptions made in the foregoing analyses include: (1) steady flow; (2) the height of the water table above the top of the stream was quite small, and its slope nearly horizontal; (3) that the drain width was much smaller than the drain half-spacing ($a \ll L$); and (4) the porous media is anisotropic and homogeneous. The first three restrictions are often met in actual field situations. However stream-connected phreatic aquifer systems are seldom isotropic and homogeneous. The above analyses addressed the problem of anisotropy but did assume a homogeneous porous media (ie, where K_x and K_z are both independent of the spatial

coordinates). The modified Corps of Engineers program (Appendix B) may be used to solve for the concentration break-through curves resulting from any spatial distribution of K_x and K_z . For brevity, however, only the relative importance of several simple combinations of aquifer geometry (i.e., the L/D ratio) and heterogeneity (i.e., the spatial variation of the K_z/K_x ratio) were examined in a geologically layered system. Figure 2.21 depicts the general Laplace aquifer system examined, using the same modified Corps of Engineers program described earlier and subject to the implied limitations previously mentioned. Kirkham, et al., (1974) have analytically solved similar problems for the potential distribution in multi-layered systems. Thus the same 41 by 21 rectangular grid was retained. This portion of the research was designed to investigate the differences in concentration outflow history resulting from variations in: (1) the upper layer thickness; (2) the isotropic, homogeneous hydraulic conductivity contrast between the upper and lower aquifers; (3) anisotropy in each aquifer; and (4) an isotropic, homogeneous hydraulic conductivity reduction near the drain. For all specific cases examined values for the aspect ratio, defined here as $n = L/D$, were the same as previously used (i.e., $n = 1$ through 5, and $n = 10$). Table 2.3 summarizes the cases examined; the "REMARKS" column refers to the figures where concentration break-through curves are presented. The notation used in the table corresponds to that used in Figure 2.21. For cases (1) through (5), the aquifers were considered to be isotropic and homogeneous; for case (6), anisotropic, homogeneous effects were included. Case (5) in the table examines the effect of stream clogging on concentration outflow histories. Thus K_{aq} represents the isotropic, homogeneous hydraulic conductivity of the entire model aquifer, while

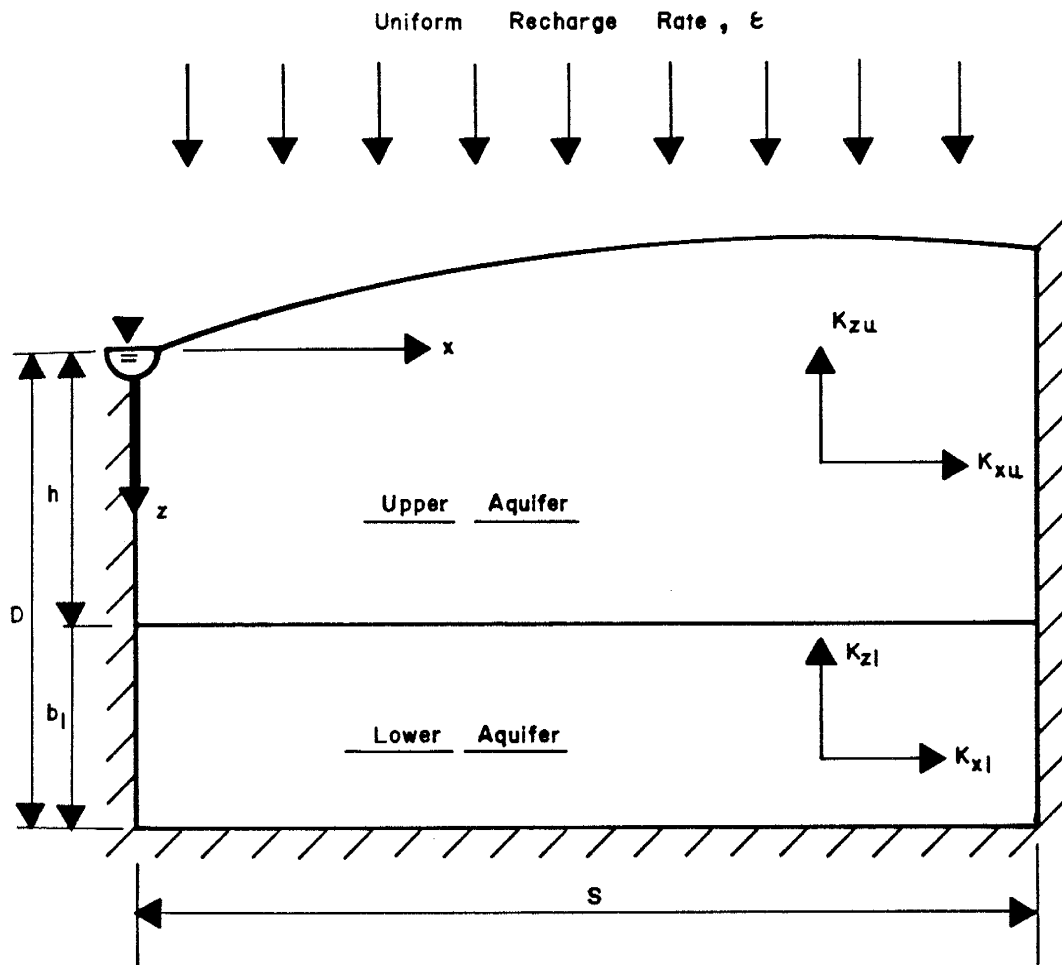


Figure 2.21 Schematic diagram showing the Laplace aquifer configuration for geologic layering.

K_C represents that value in the immediate vicinity of the stream or tile drain where clogging occurs (represented in the model program as the single finite element adjacent to the stream).

Table 2.3: Cases examined for geological layering effects; see Figure 2.21 for notation definitions.

<u>CASE EXAMINED</u>	<u>$m=b_L/(h_0-b_L)$</u>	<u>$N=D/h$</u>	<u>RANGE OF α</u>	<u>REMARKS</u>
1. $K_U = \alpha K_L$	3	4	20, 10, 5, 2	See Figs. 2.22-2.33
2. $K_U = \alpha K_L$	9	10	10	See Figs. 2.34-2.36
3. $K_U = \alpha K_L$	19	20	10	See Figs. 2.37-2.39
4. $K_L = \alpha K_U$	3	4	10	See Figs. 2.40-2.42
5. $K_{aq} = \alpha K_C$	0	1	10, 100	See Figs. 2.43-2.46
6. $K_{xu} = \alpha K_{uL}$	3	4	10	See Figs. 2.47-2.49
$K_{zu} = \alpha K_{zL}$	3	4	1	

In the calculation of the dimensionless concentration (c/c_0) versus dimensionless time (t/t_c) curves, (2.4.6) and (2.4.7) may still be used. These concentration break-through curves will reflect the vertical flow effects on the outflow concentration history resulting from the various hydraulic and aquifer geometry factors summarized in Table 2.3. We would like to compare these c/c_0 versus t/t_c curves with that obtained previously using a lumped parameter modeling approach that characterizes only the upper aquifer of Figure 2.21. The reason for this comparison is to qualify those situations where vertical flow through multiple reservoirs may have relatively little effect on outflow concentration. The application of the lumped modeling approach to only the upper aquifer constitutes the simplest tool available for analyzing such problems, and hence retains its primary advantage. In order to make such a comparison, however, we must modify our definition of the solute response time, $t_c = nD/\epsilon$, such that the distance D is the same for all concentration break-through curves. Since we have chosen the lumped

parameter model to characterize only the upper aquifer, its solute response time will be given by $t_c = nh/\epsilon$. For the specific cases examined, the relationship between h and D is given in Table 2.3. We therefore see that $D = Nh$ where N is some specified integer for a particular case of interest. The lumped parameter solute response time for the upper aquifer of Figure 2.21 may therefore be related to the dimensionless travel time found from (2.4.6) for any case shown in Table 2.3. Thus

$$\frac{t}{t_c} = \frac{t\epsilon}{nD} = \left(\frac{t\epsilon}{nh}\right) \left(\frac{h}{D}\right) = \frac{d\bar{A}(\bar{\psi})}{d\bar{\psi}}$$

or

$$t/t_{c_1} = \frac{d\bar{A}(\bar{\psi})}{d\bar{\psi}} \quad (2.4.15)$$

where $t_{c_1} = \frac{nh}{\epsilon} \frac{D}{h} t_c$. This new t_{c_1} characterizes the average solute travel time through the multiple reservoir system of Figure 2.21 times the ratio D/h ; it is identical to the lumped parameter solute response time for a single reservoir of thickness h . Equations (2.4.7) and (2.4.15) therefore define the dimensionless concentration break-through curves characterizing a flushing aquifer system depicted in Figure 2.21; furthermore, these curves may be directly compared to the resulting dimensionless concentration break-through curves obtained from the lumped parameter model previously developed for a single reservoir. Results of these analyses are shown in Figures 2.22 through 2.49, and correspond to the specific cases outlined in Table 2.3.

Again, the same conservation of mass check procedure previously employed may be used for the above analyses. Thus the area under each

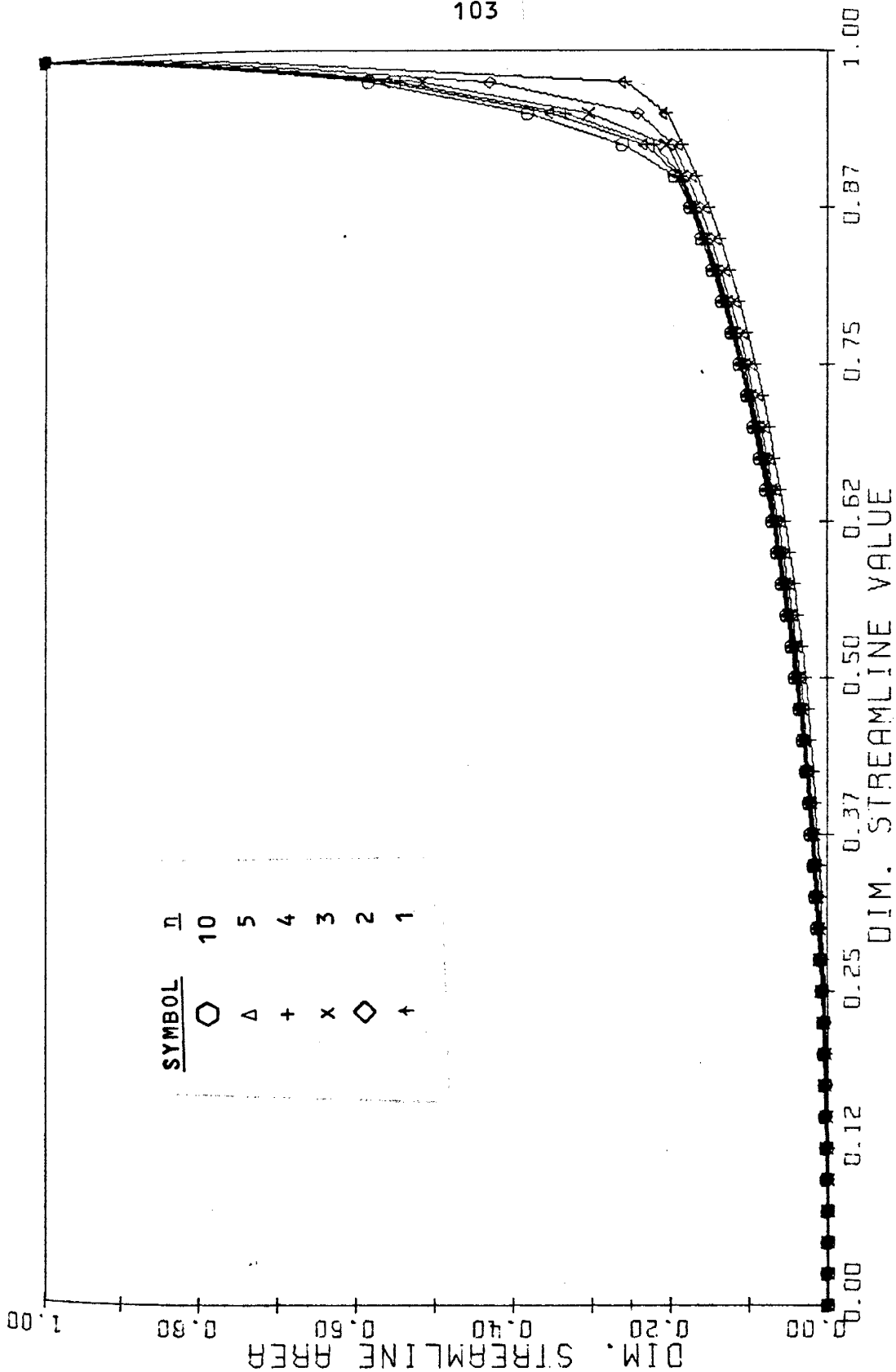


Figure 2.22: $\bar{A}(\bar{\psi})$ vs. $\bar{\psi}$ for different n ; case (1) in Table 2.3 with $\alpha = 20$.

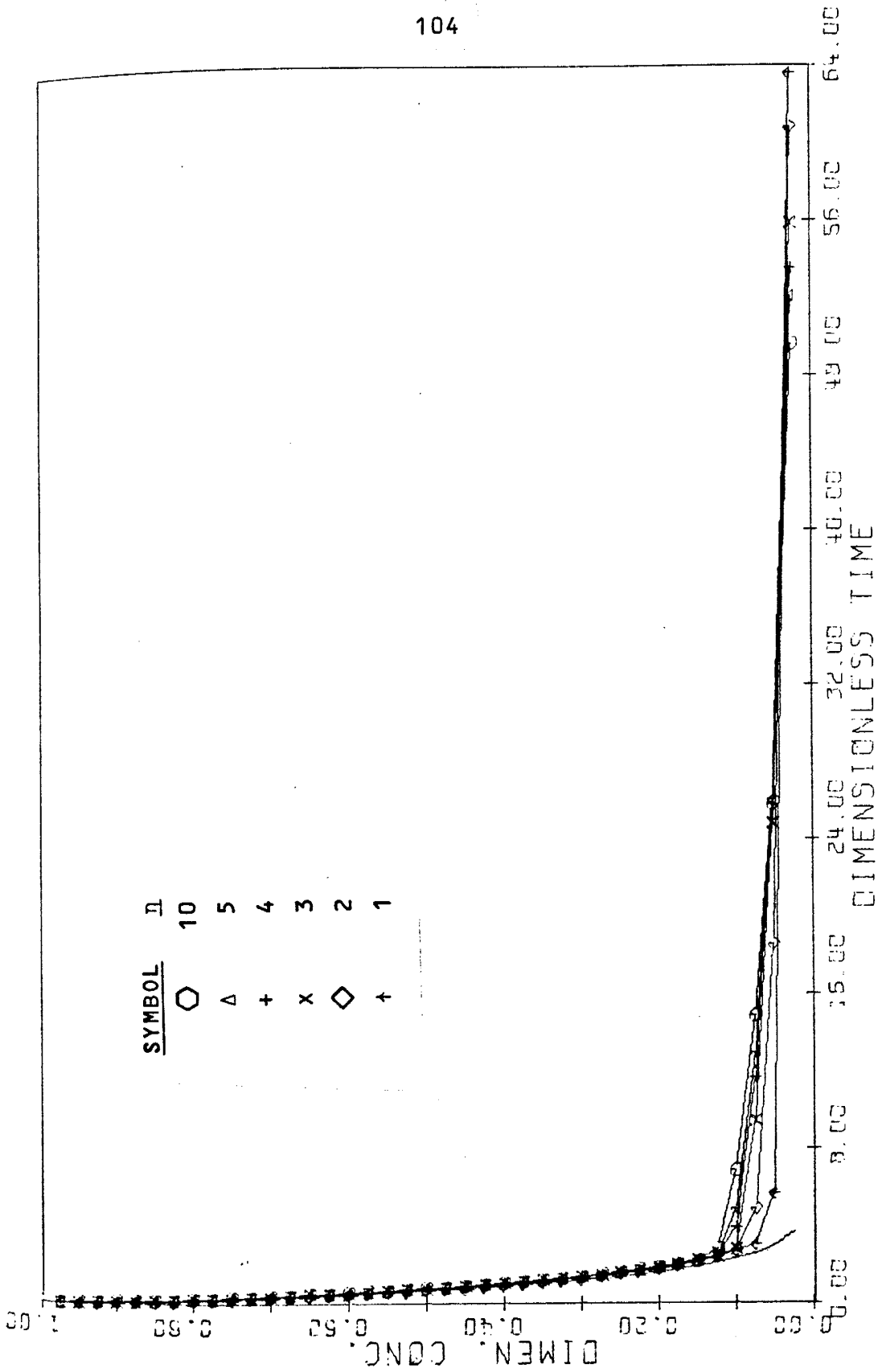


Figure 2.23: c/c_0 vs. t/t_{c1} for different n ; case (1) in Table 2.3 with $\alpha = 20$.

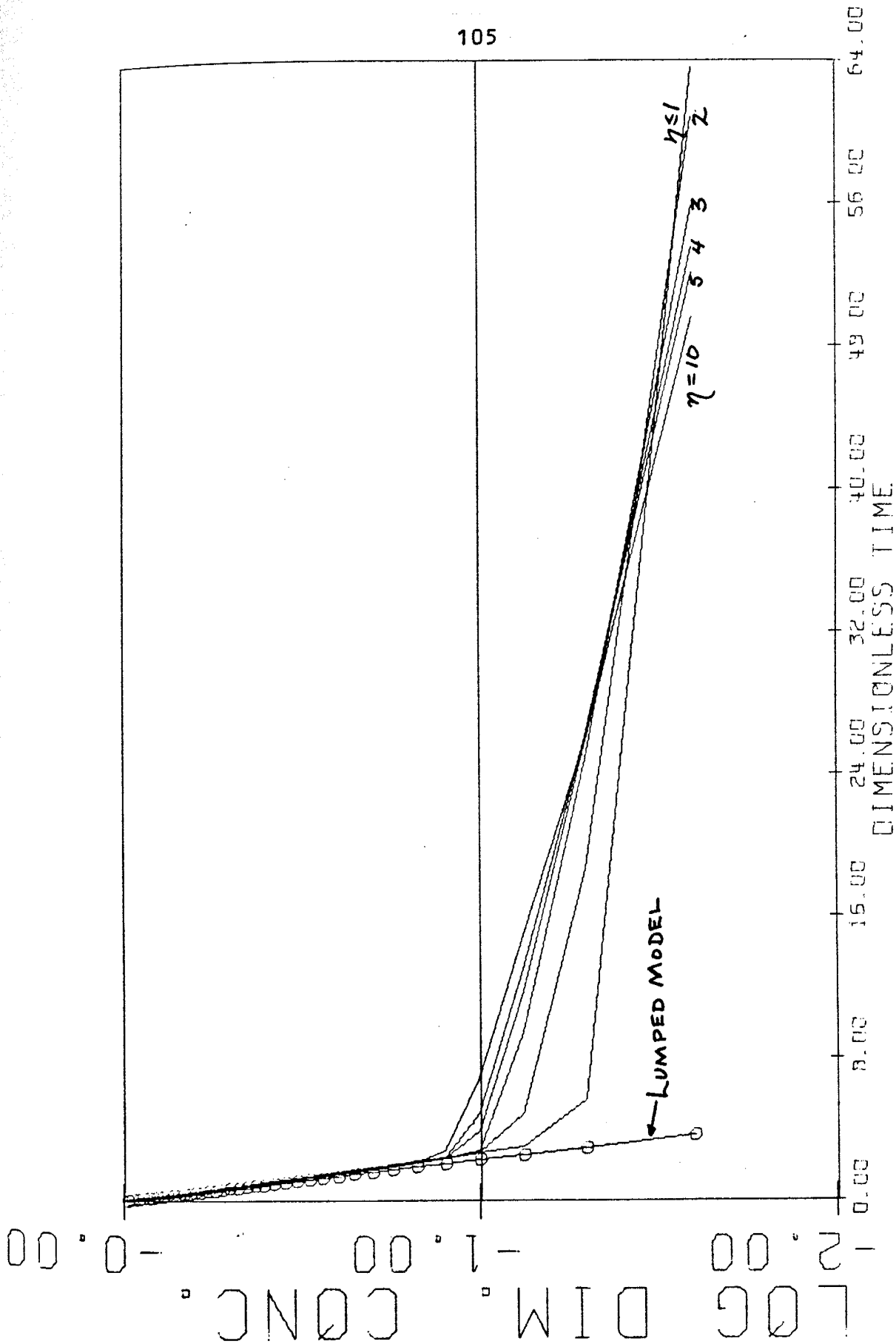


Figure 2.24: Log c/c_0 vs. t/t_{c1} for different n ; case (1) in Table 2.3 with $\alpha = 20$.

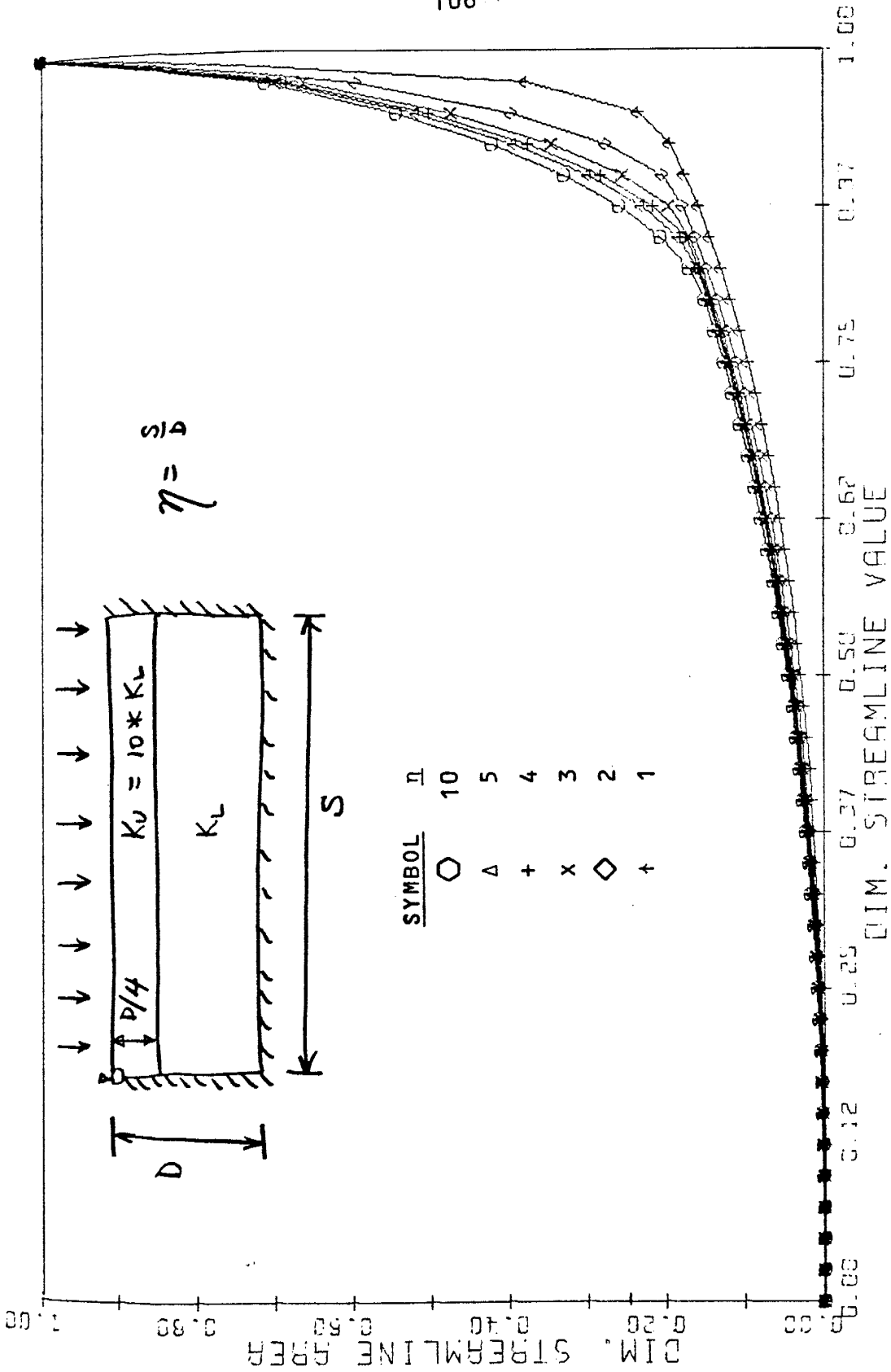


Figure 2.25: $\bar{A}(\bar{\Psi})$ vs. $\bar{\Psi}$ for different n ; case (1) in Table 2.3 with $\alpha = 10$.

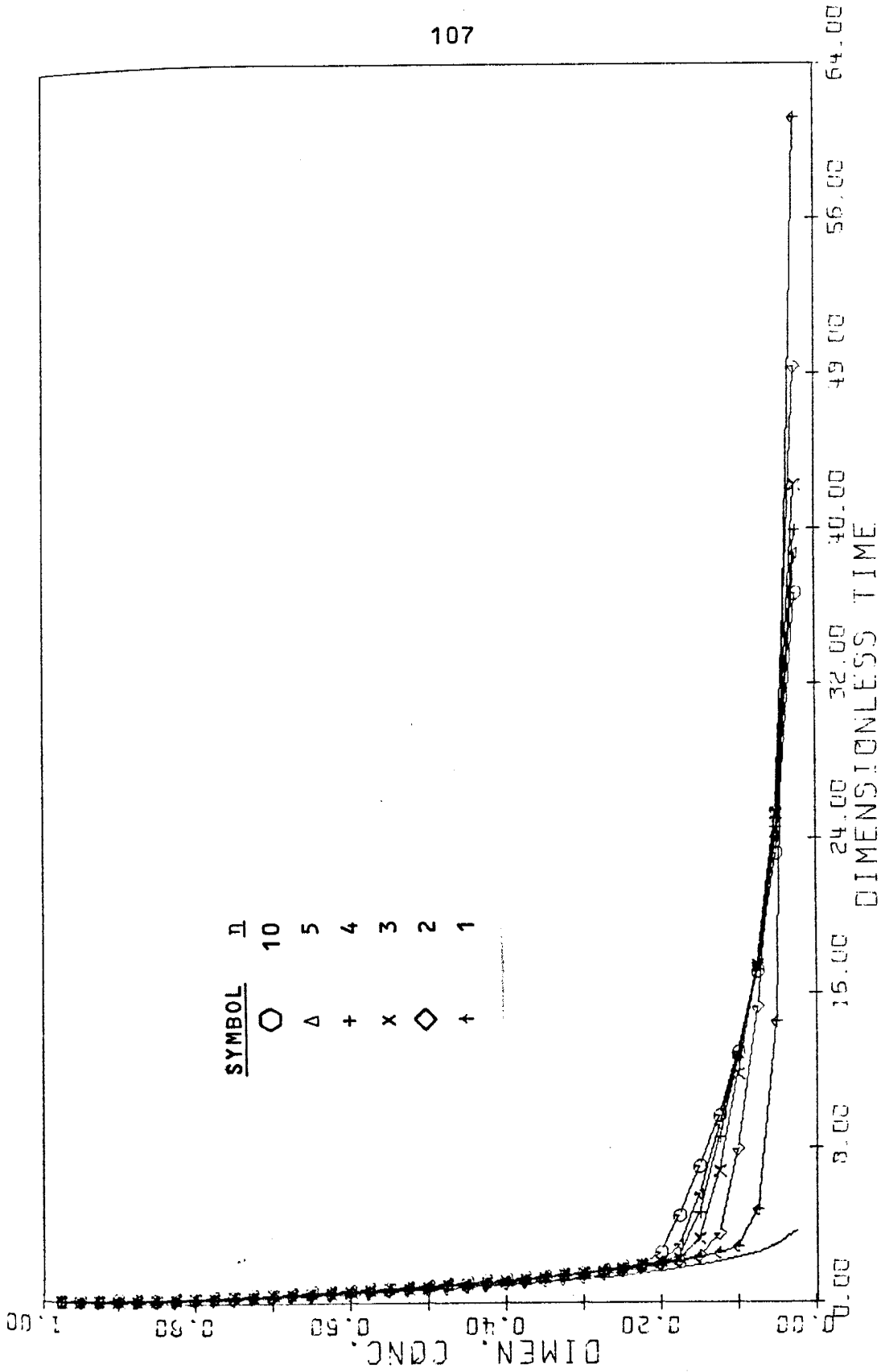


Figure 2.26: c/c_0 vs. t/t_{c1} for different n ; case (1) in Table 2.3 with $\alpha = 10$.

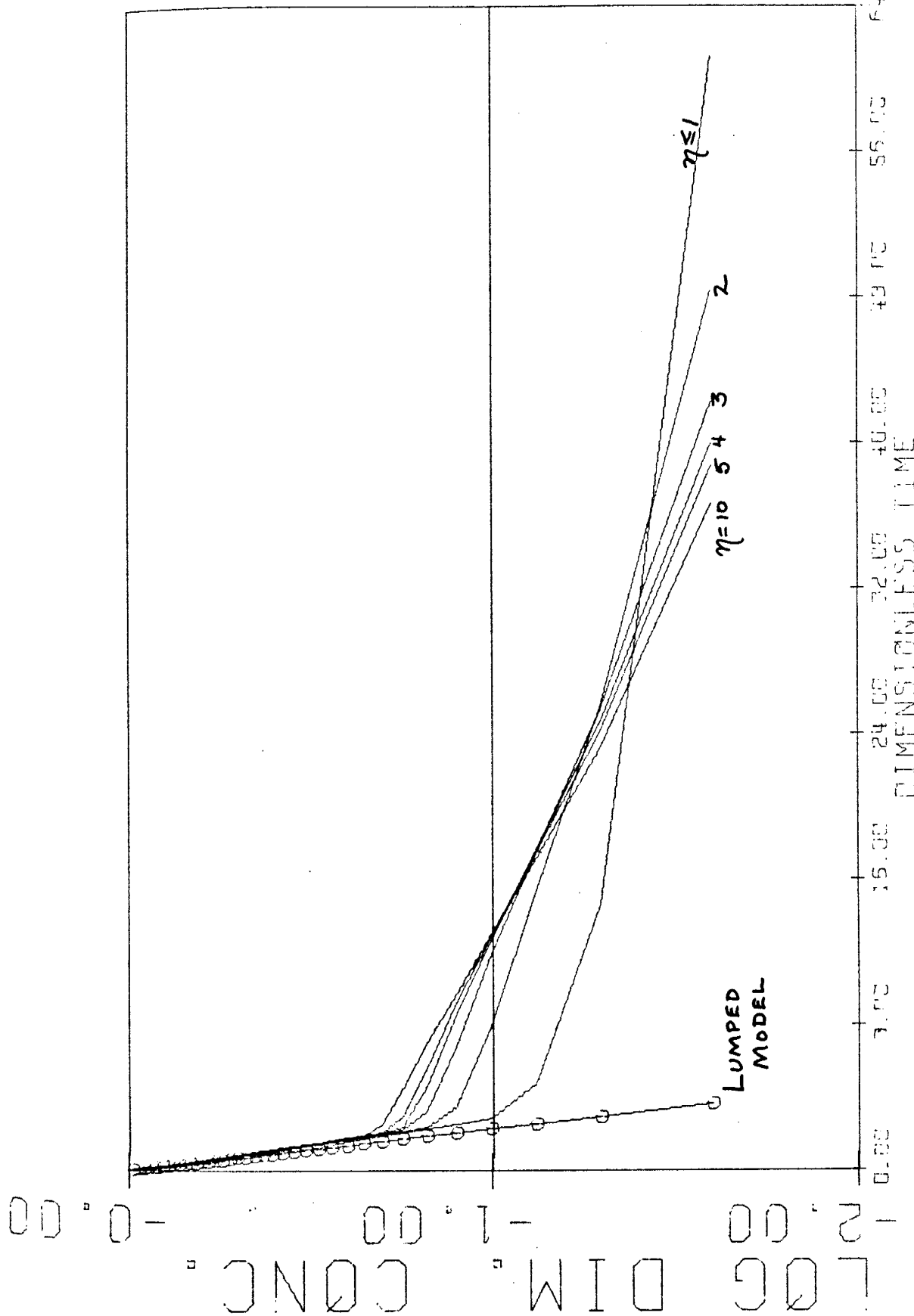


Figure 2.27: Log c/c_0 vs. t/t_{c1} for different n ; case (1) in Table 2.3 with $\alpha = 10$.

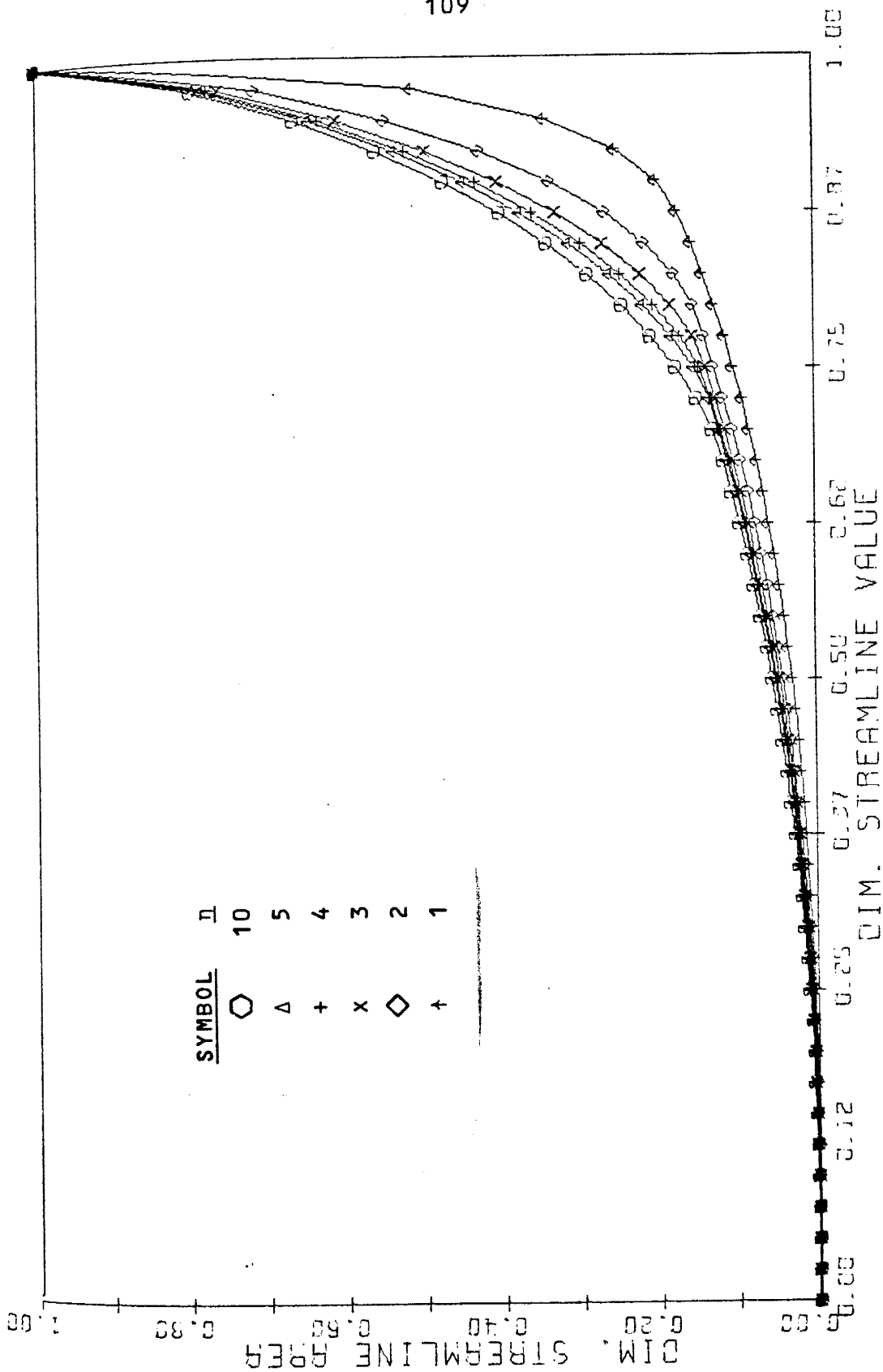


Figure 2.28: $\bar{A}(\bar{\Psi})$ vs. $\bar{\Psi}$ for different n ; case (1) in Table 2.3 with $\alpha = 5$.

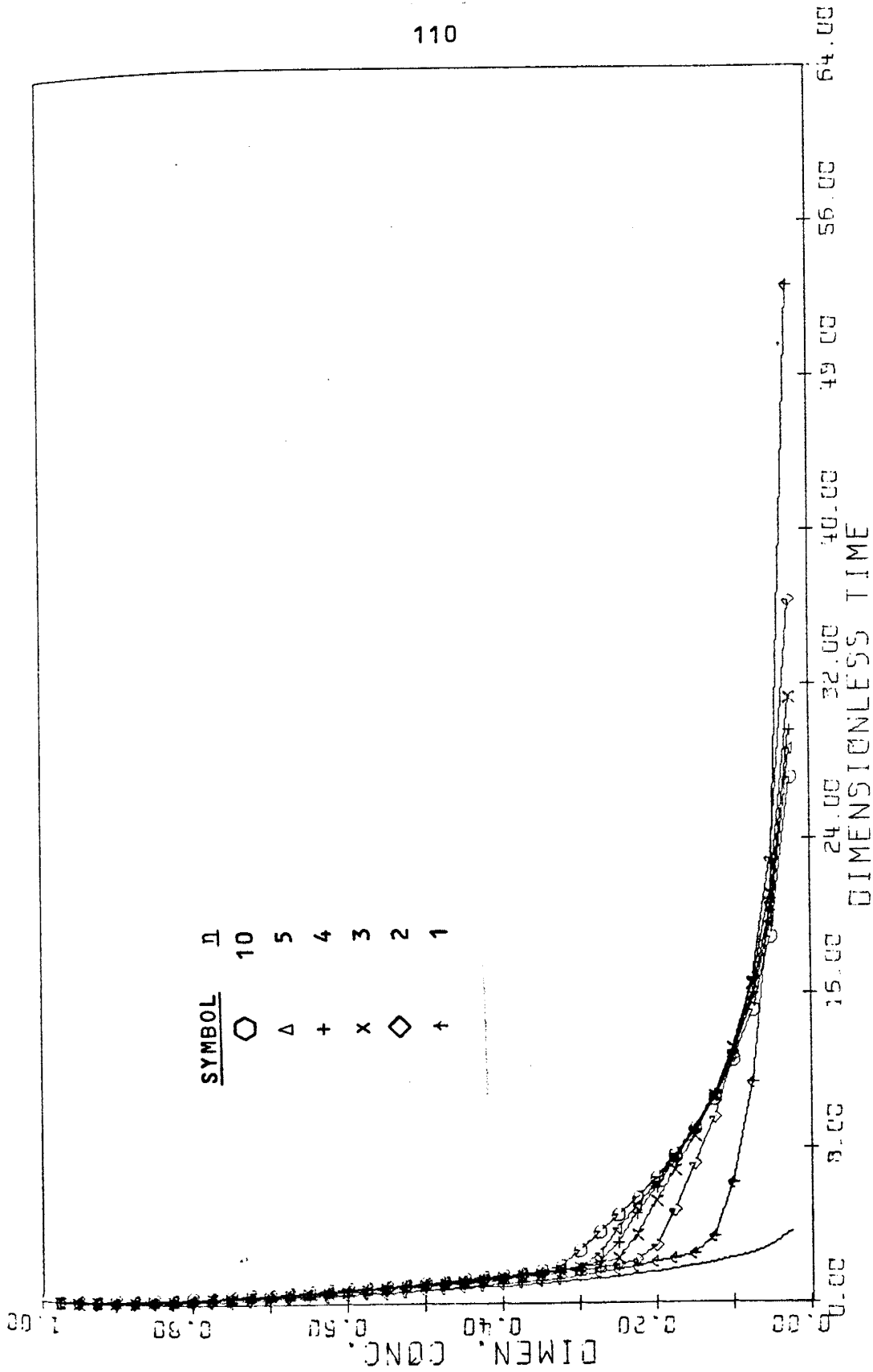


Figure 2.29: c/c_0 vs. t/t_{c1} for different n ; case (1) in Table 2.3 with $\alpha = 5$.

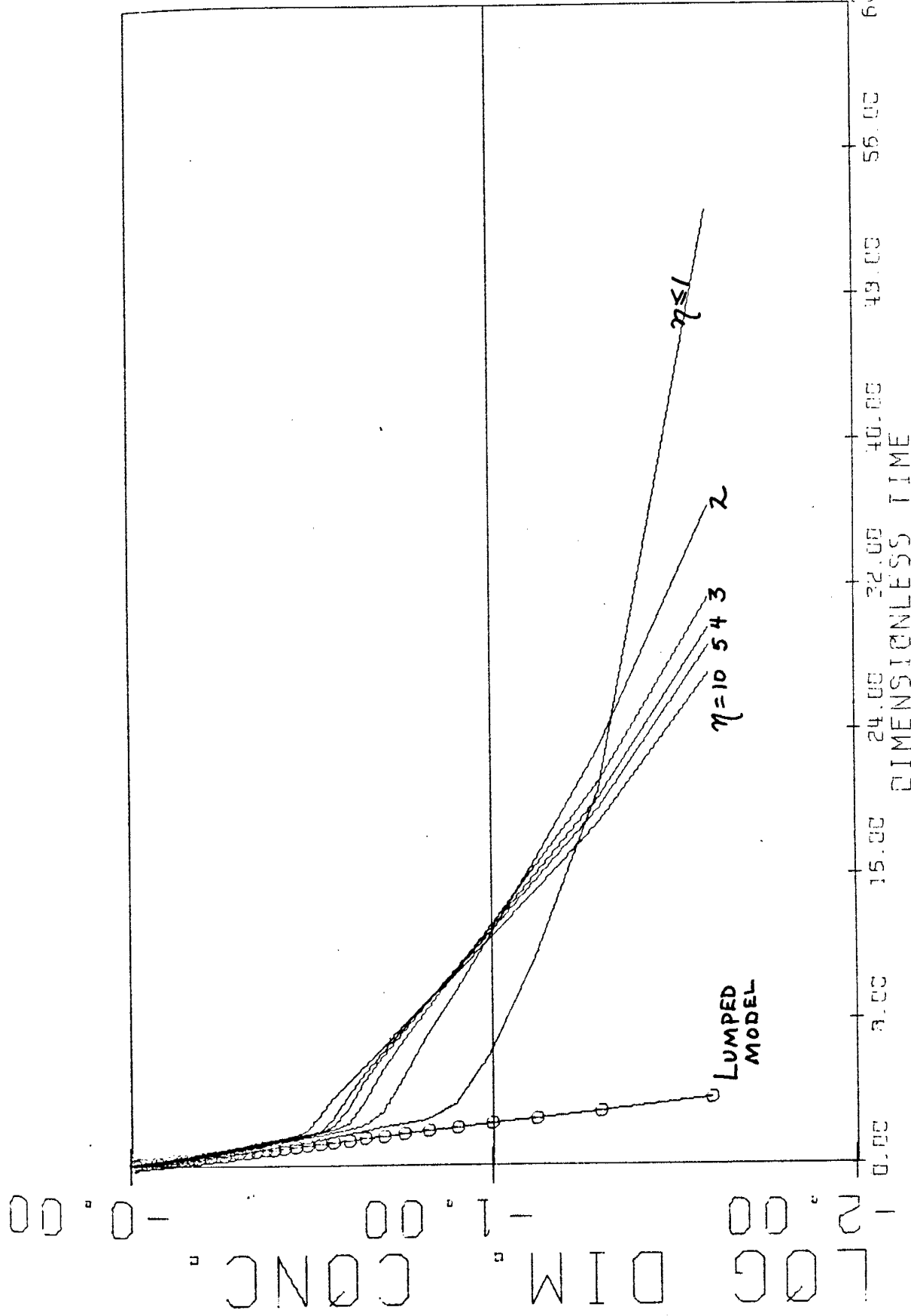


Figure 2.30: Log c/c_0 vs. t/t_{c1} for different n ; case (1) in Table 2.3 with $\alpha = 5$.

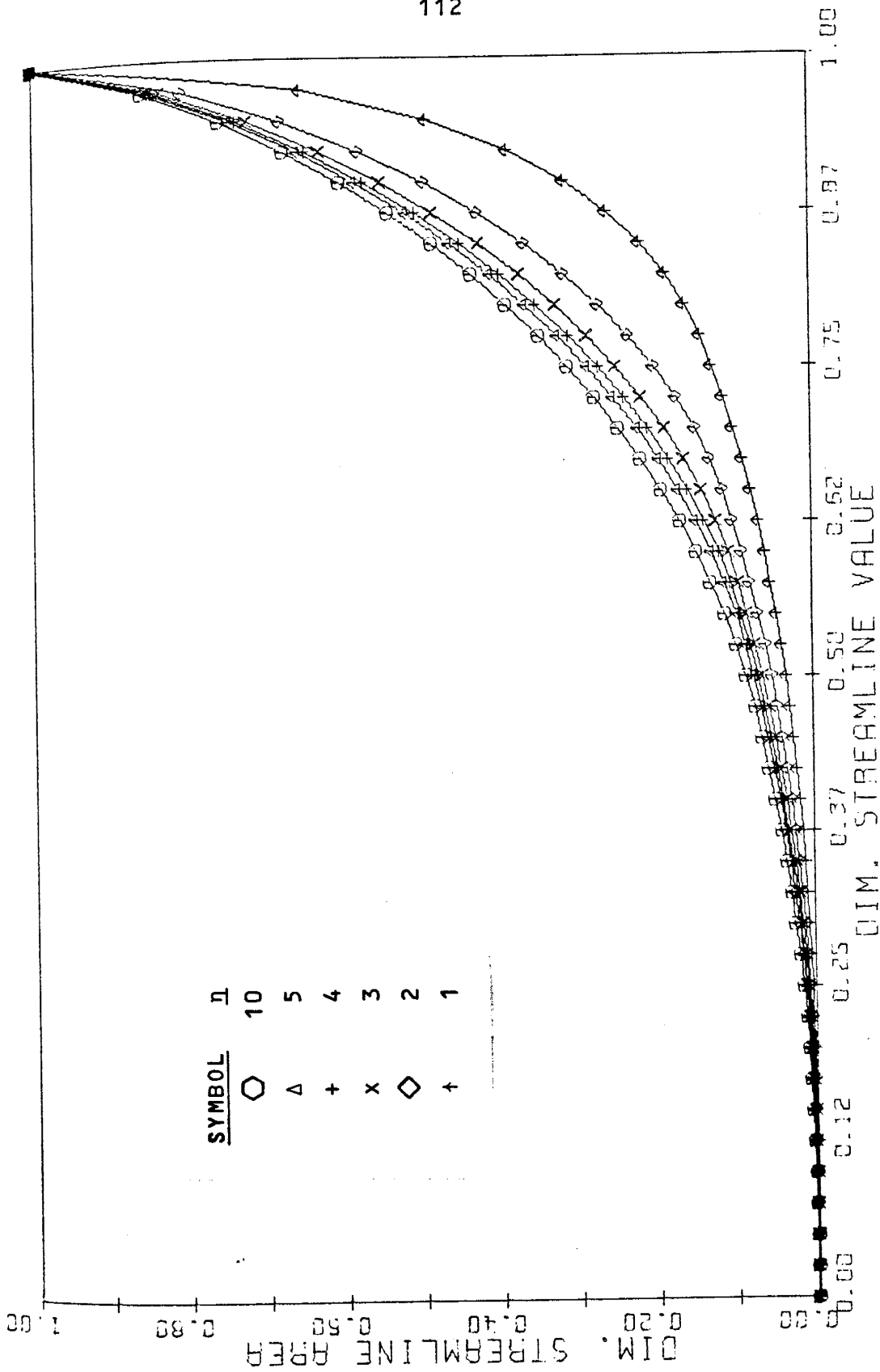


Figure 2.31: $\bar{A}(\bar{\Psi})$ vs. $\bar{\Psi}$ for different n ; case (1) in Table 2.3 with $\alpha = 2$.

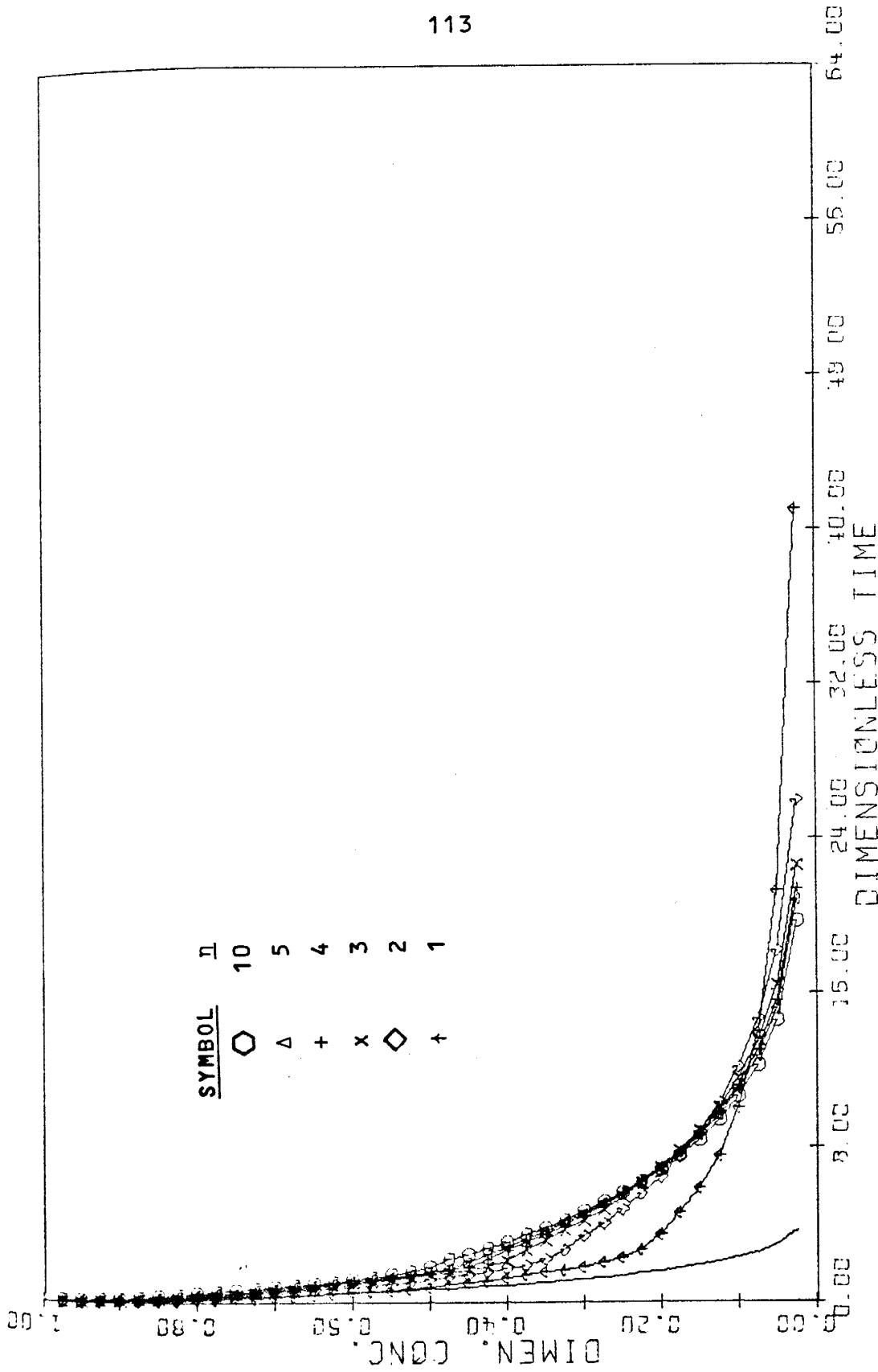


Figure 2.32: c/c_0 vs. t/t_{c1} for different n ; case (1) in Table 2.3 with $\alpha = 2$.

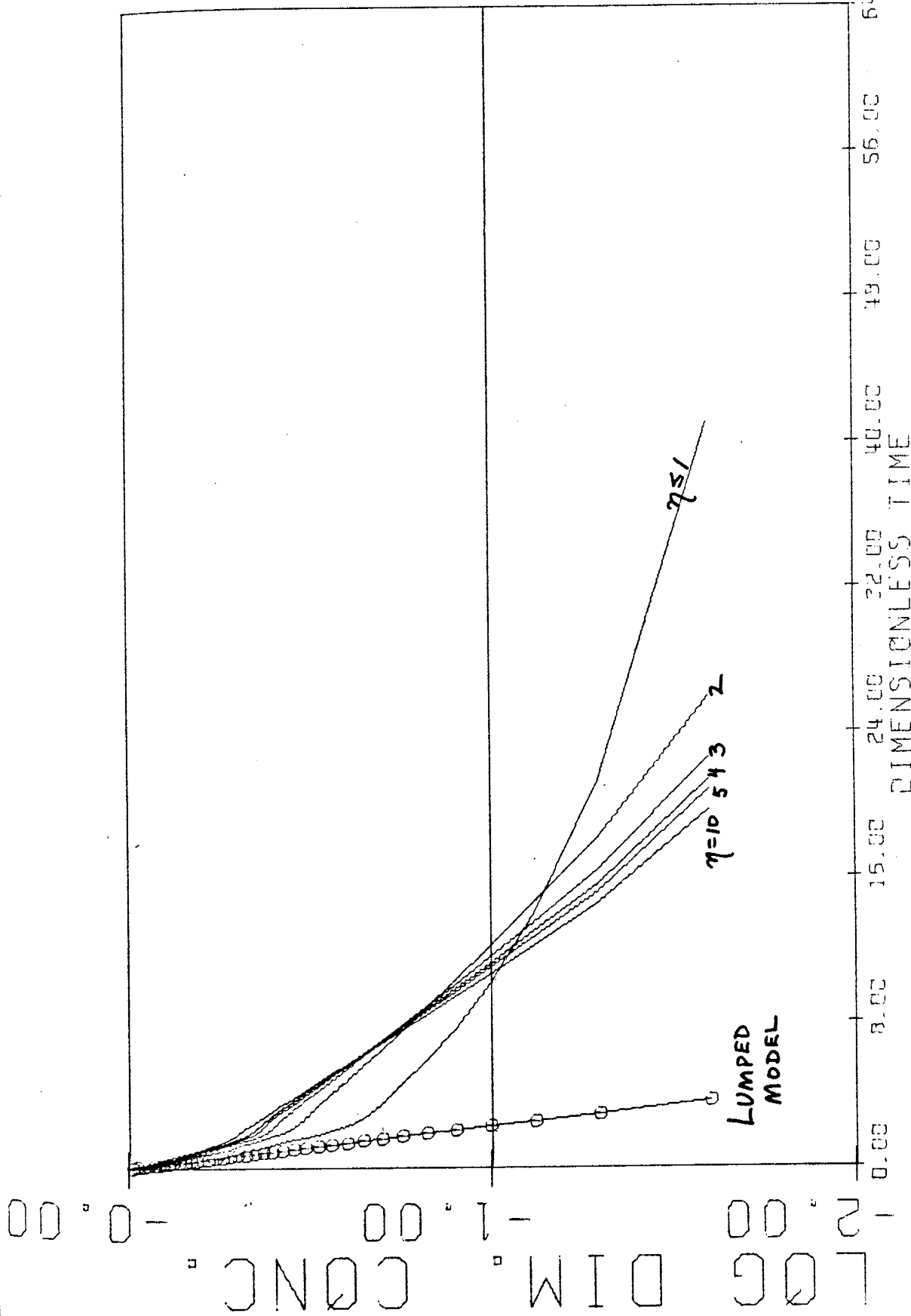


Figure 2.33: Log c/c_0 vs. t/t_{c1} for different η ; case (1) in Table 2.3 with $\alpha = 2$.

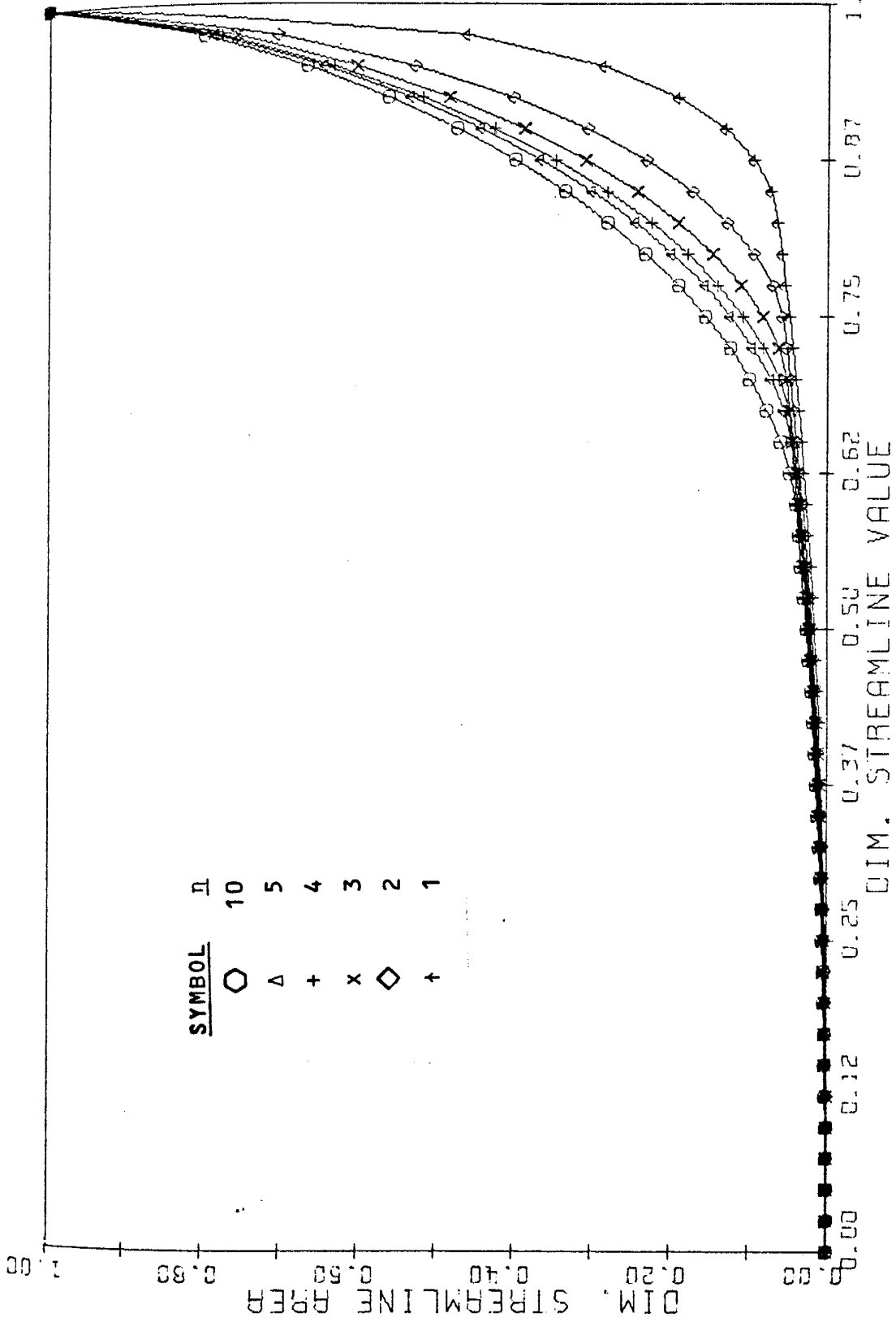


Figure 2.34: $\bar{A}(\bar{\Psi})$ vs. $\bar{\Psi}$ for different n; case (2) in Table 2.3 with $\alpha = 10$.

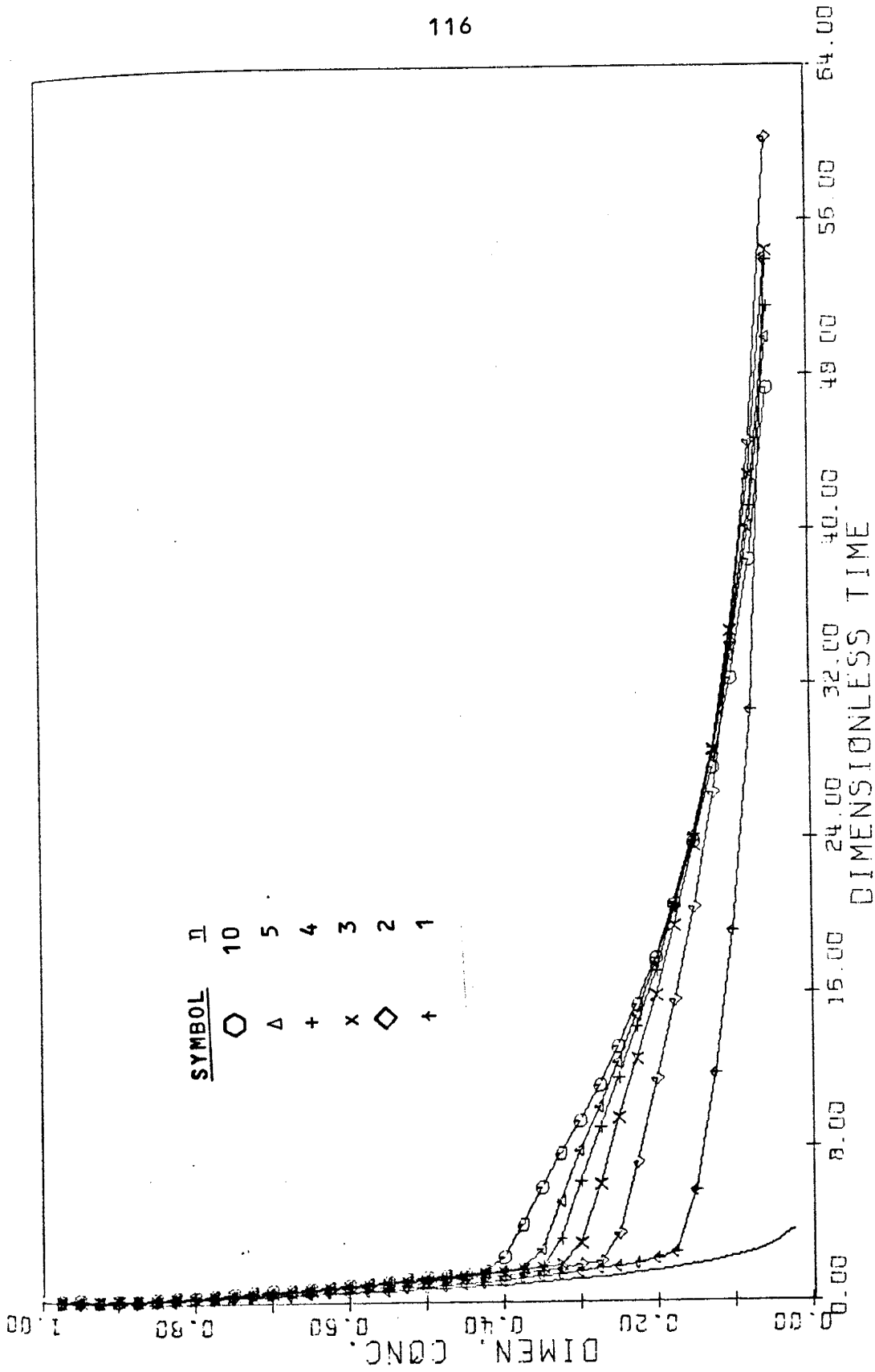


Figure 2.35: c/c_0 vs. t/t_{c1} for different n ; case (2) in Table 2.3 with $\alpha = 10$.

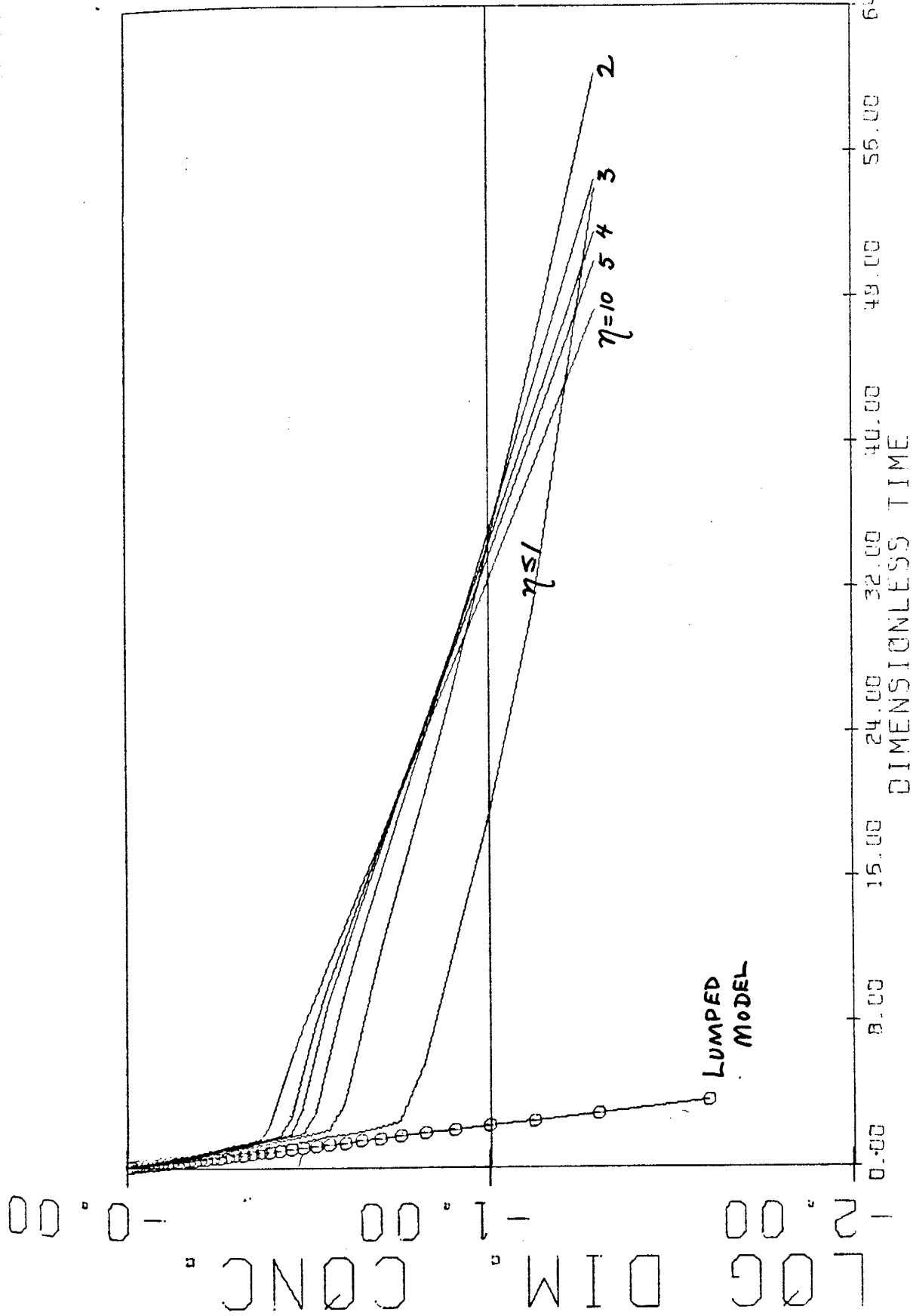


Figure 2.36: $\log c/c_0$ vs. t/t_{c1} for different η ; case (2) in Table 2.3 with $\alpha = 10$.

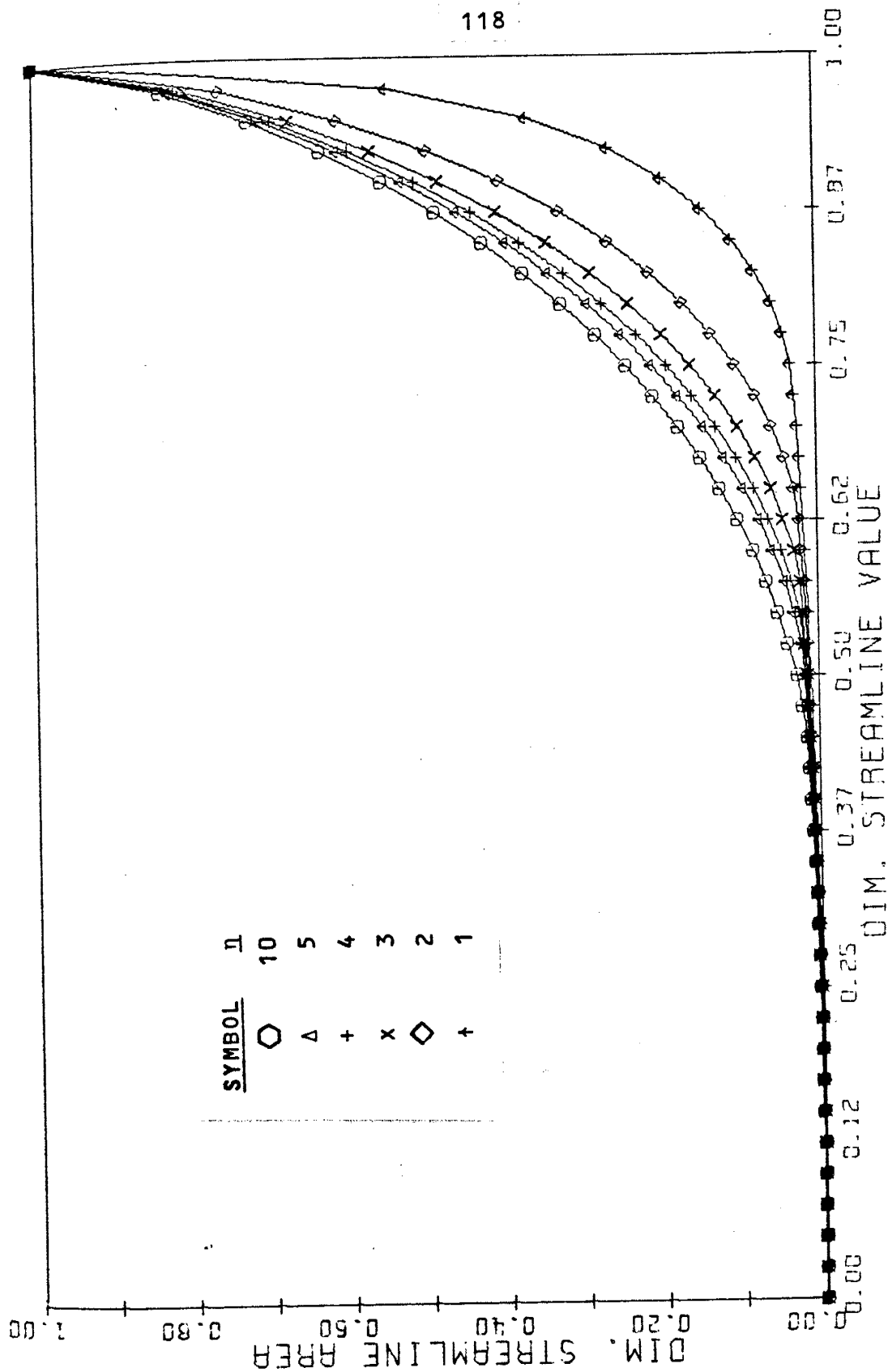


Figure 2.37: $\bar{A}(\bar{\psi})$ vs. $\bar{\psi}$ for different n ; case (3) in Table 2.3 with $\alpha = 10$.

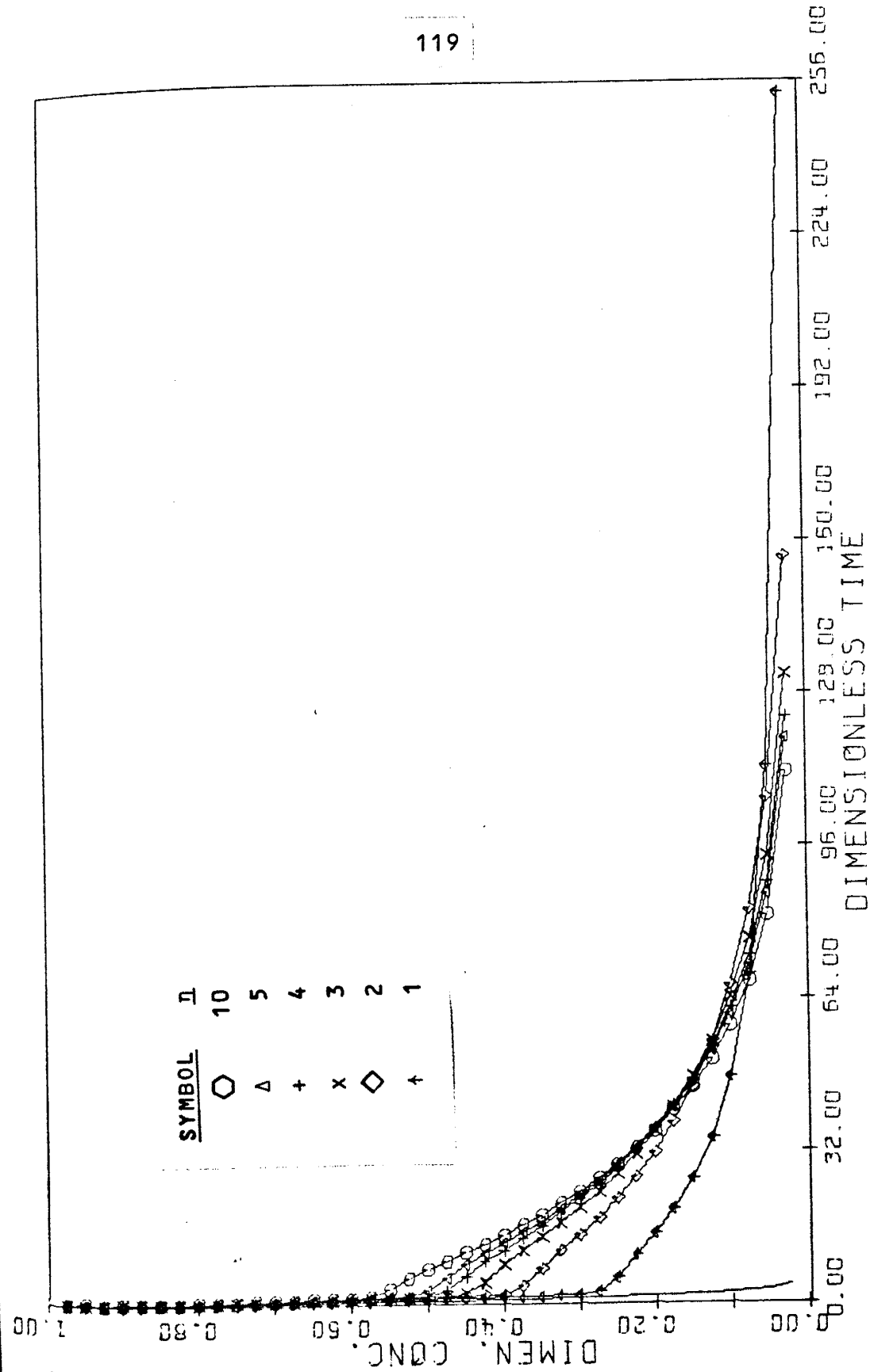


Figure 2.38: c/c_0 vs. t/t_{c1} for different n ; case (3) in Table 2.3 with $\alpha = 10$.

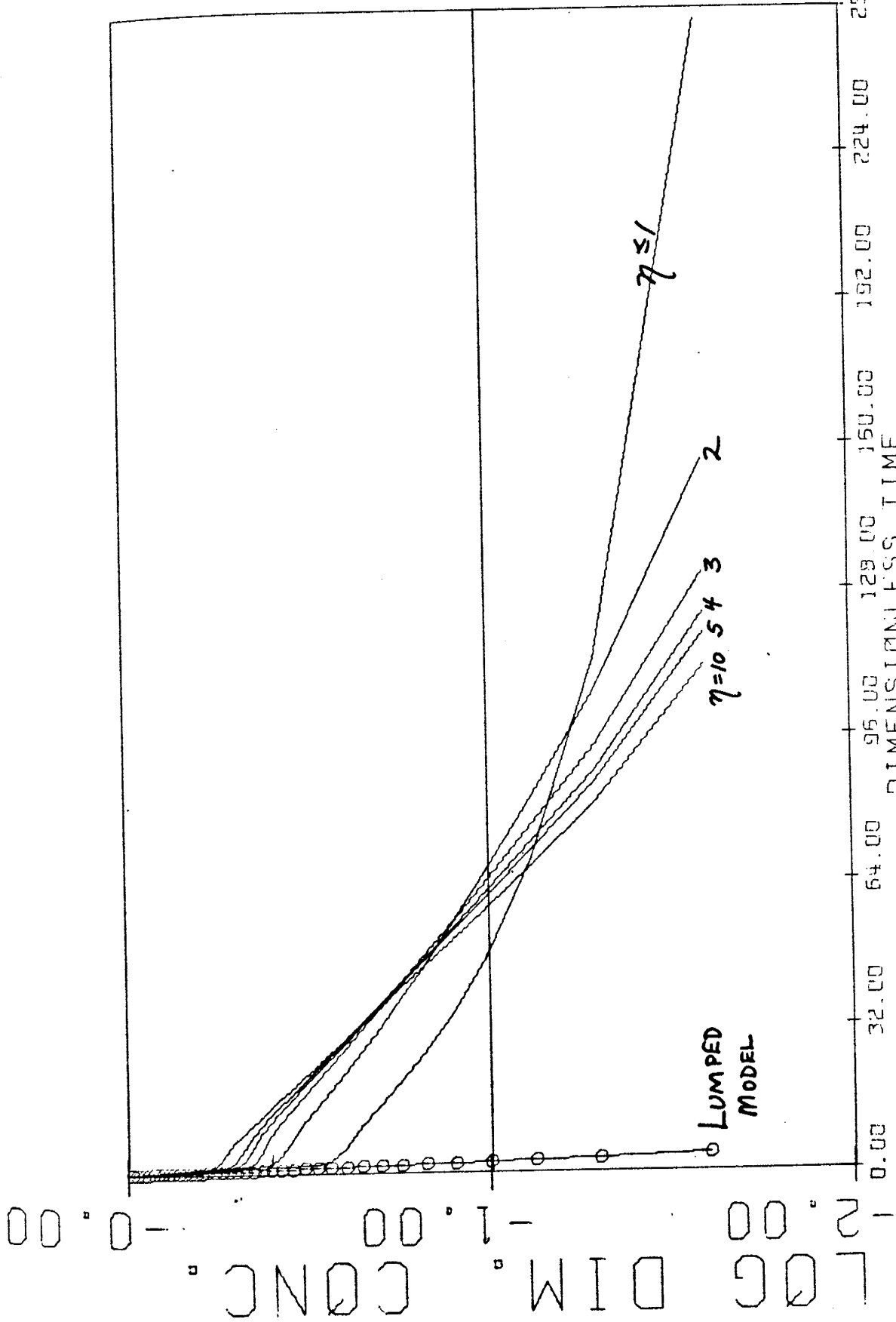


Figure 2.39: Log c/c_0 vs. t/t_{c1} for different n ; case (3) in Table 2.3 with $\alpha = 10$.

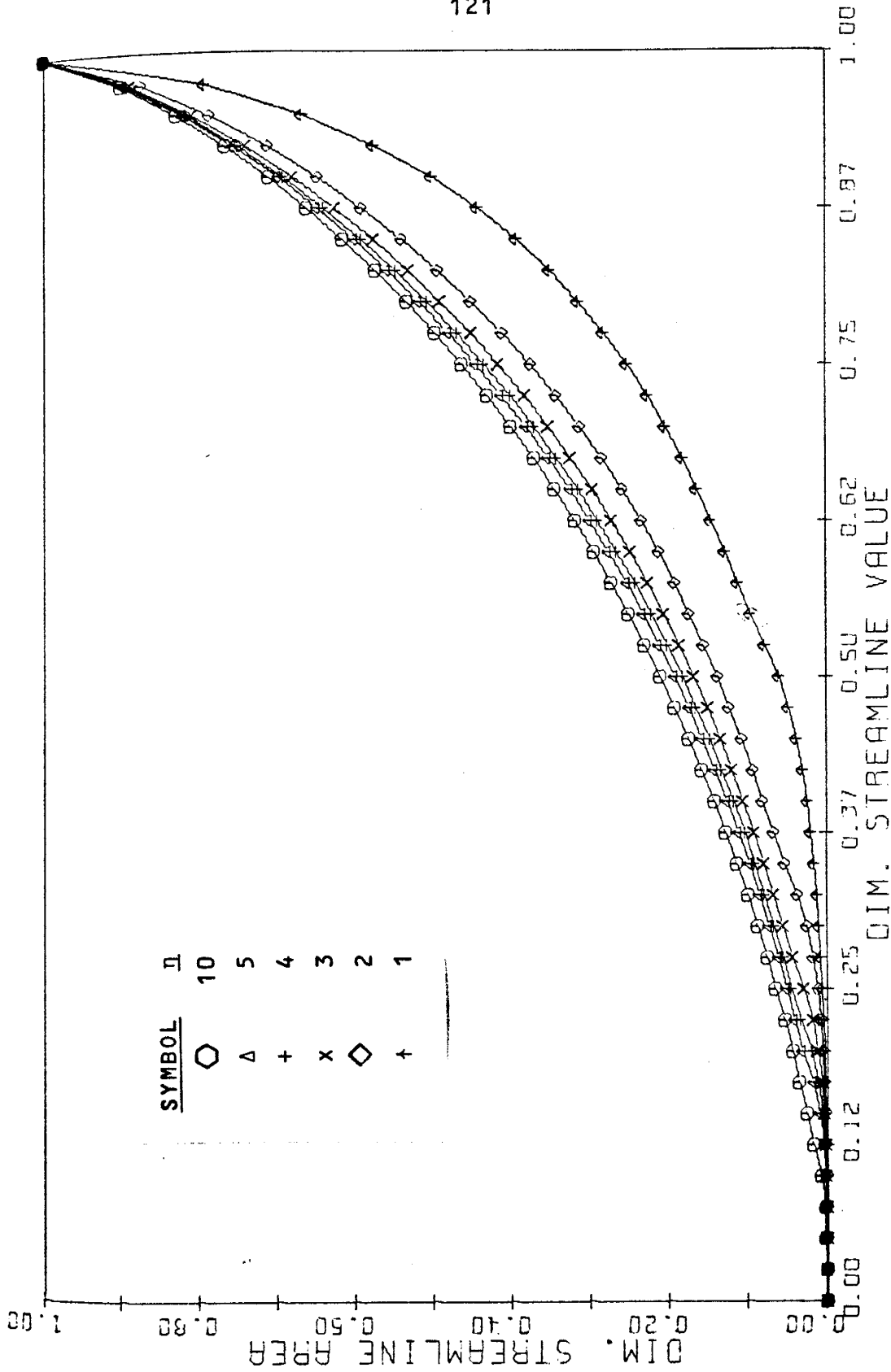


Figure 2.40: $\bar{A}(\bar{\Psi})$ vs. $\bar{\Psi}$ for different n ; case (4) in Table 2.3 with $\alpha = 10$.

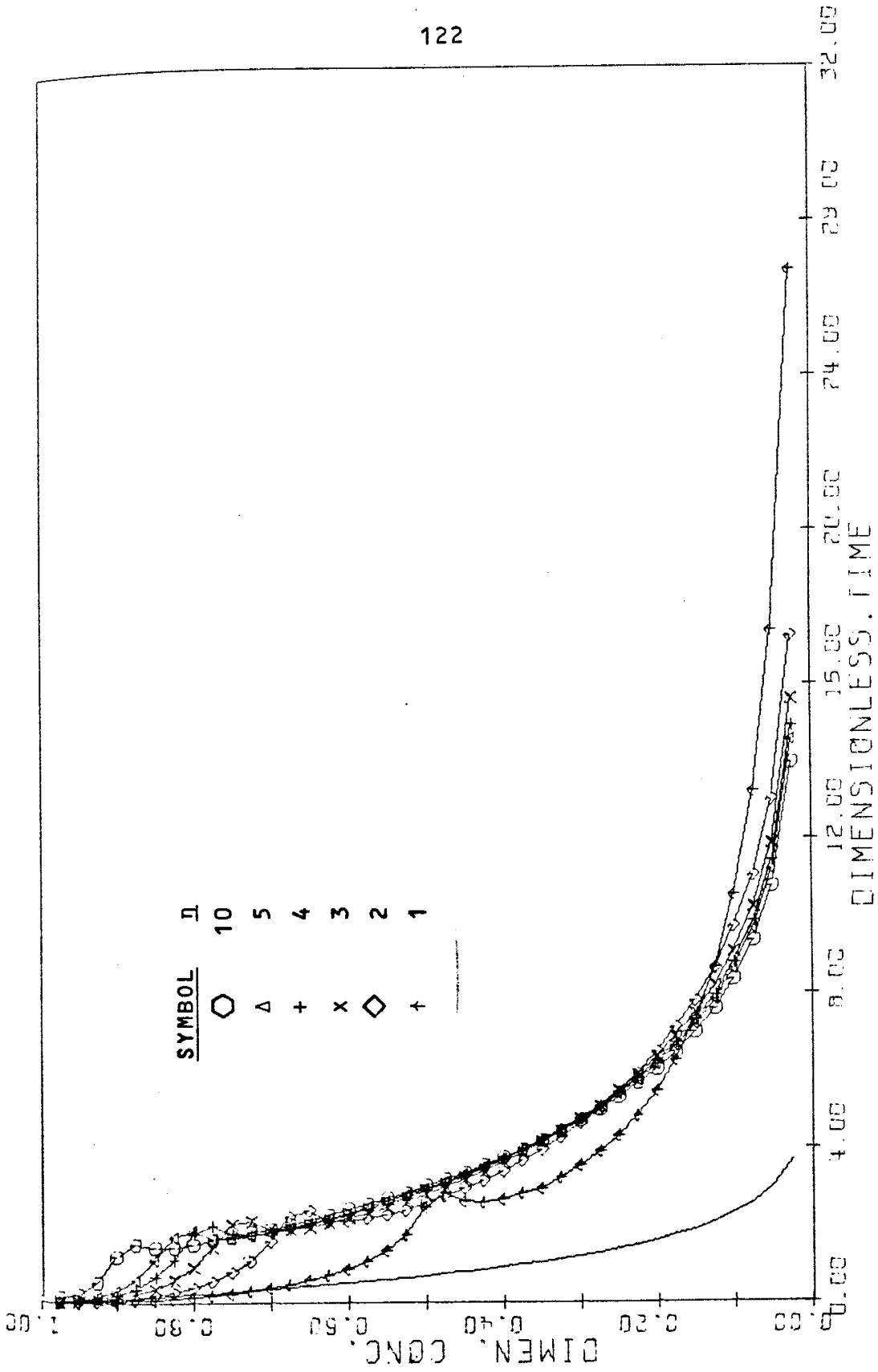


Figure 2.41: c/c_0 vs. t/t_{c1} for different n ; case (4) in Table 2.3 with $\alpha = 10$.

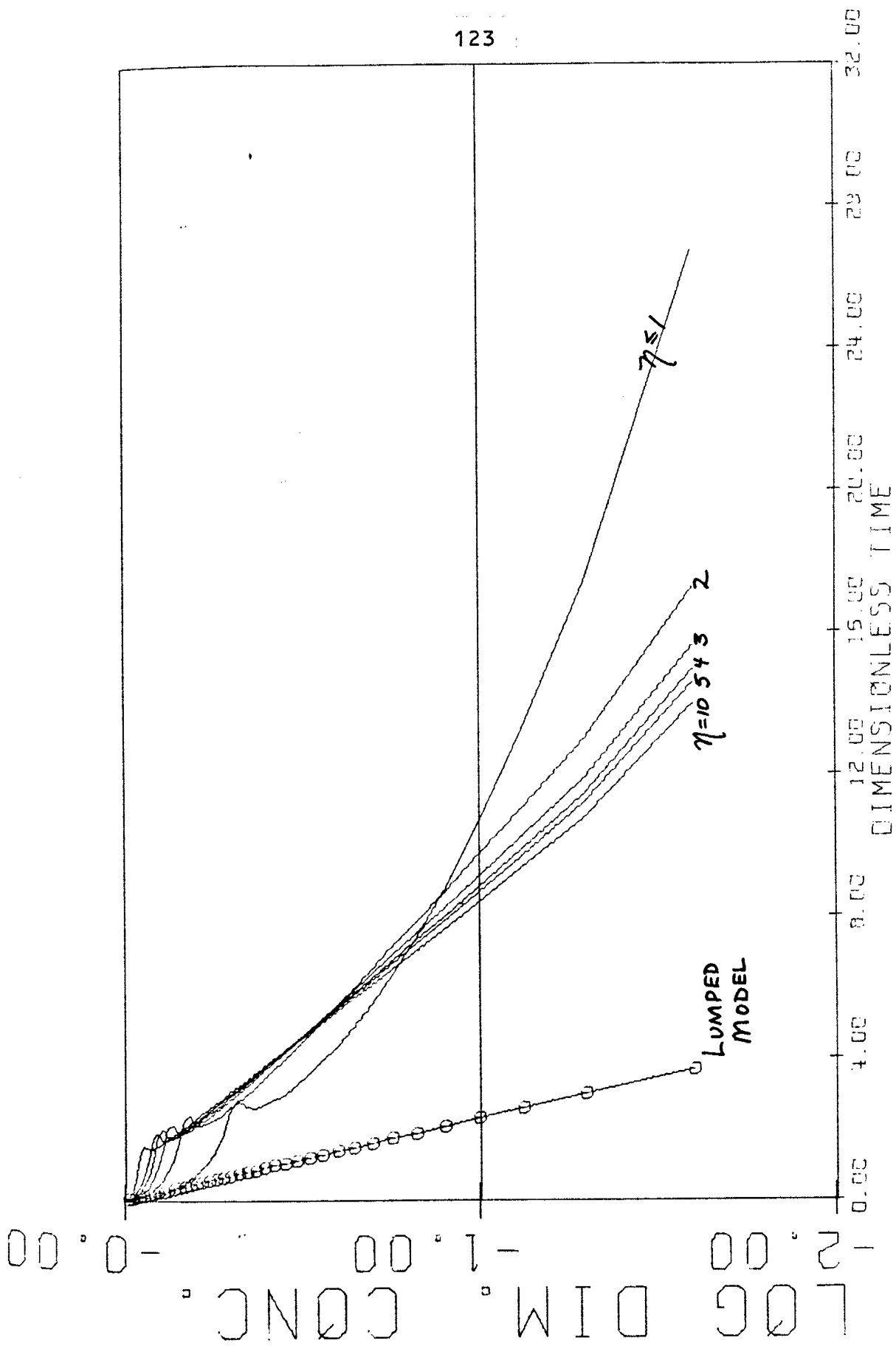


Figure 2.42: $\log c/c_0$ vs. t/t_{c1} for different n ; case (4) in Table 2.3 with $\alpha = 10$.

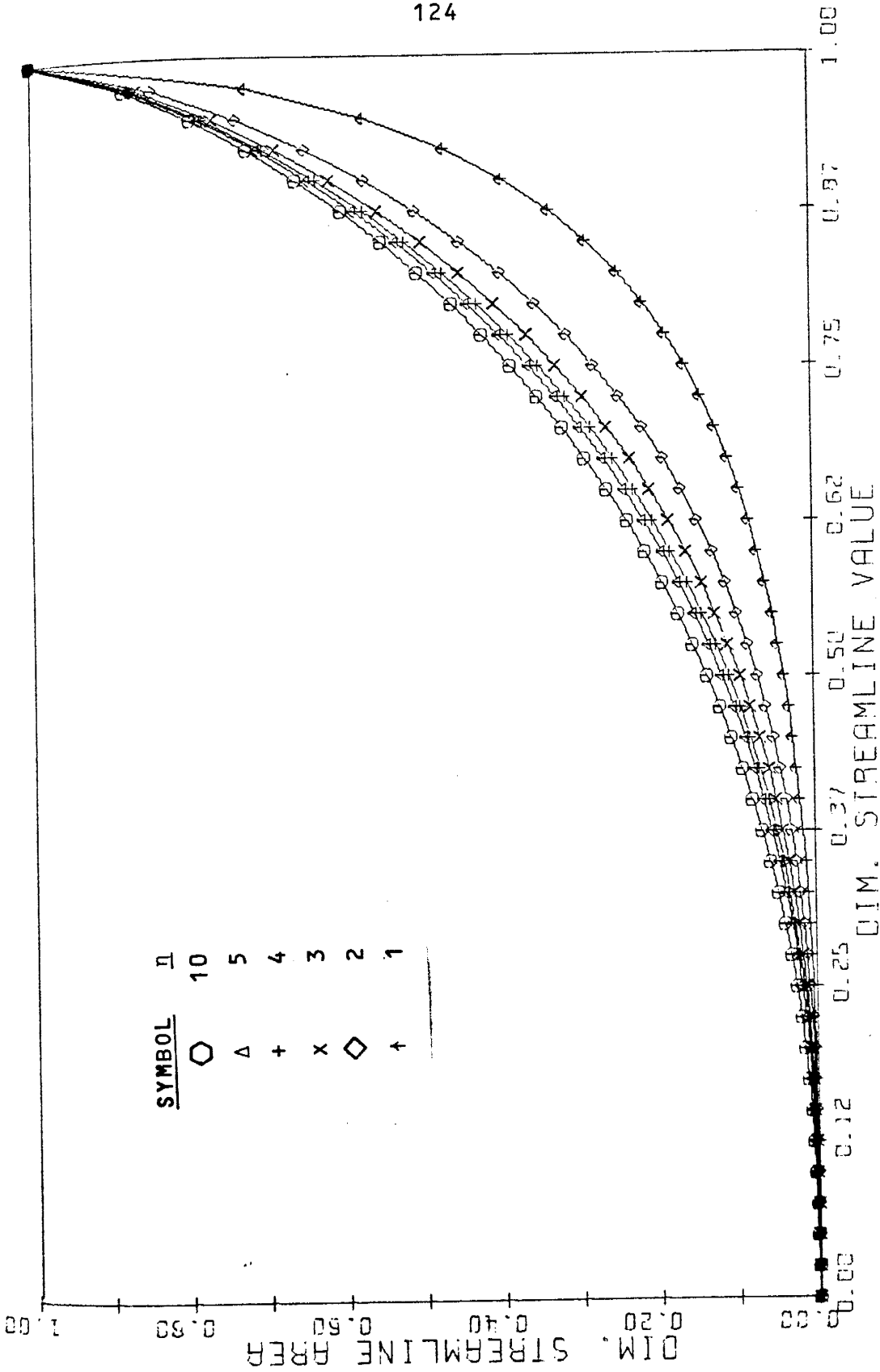


Figure 2.43: $\bar{A}(\bar{\Psi})$ vs. $\bar{\Psi}$ for different n ; case (5) in Table 2.3 with $\alpha = 10$.

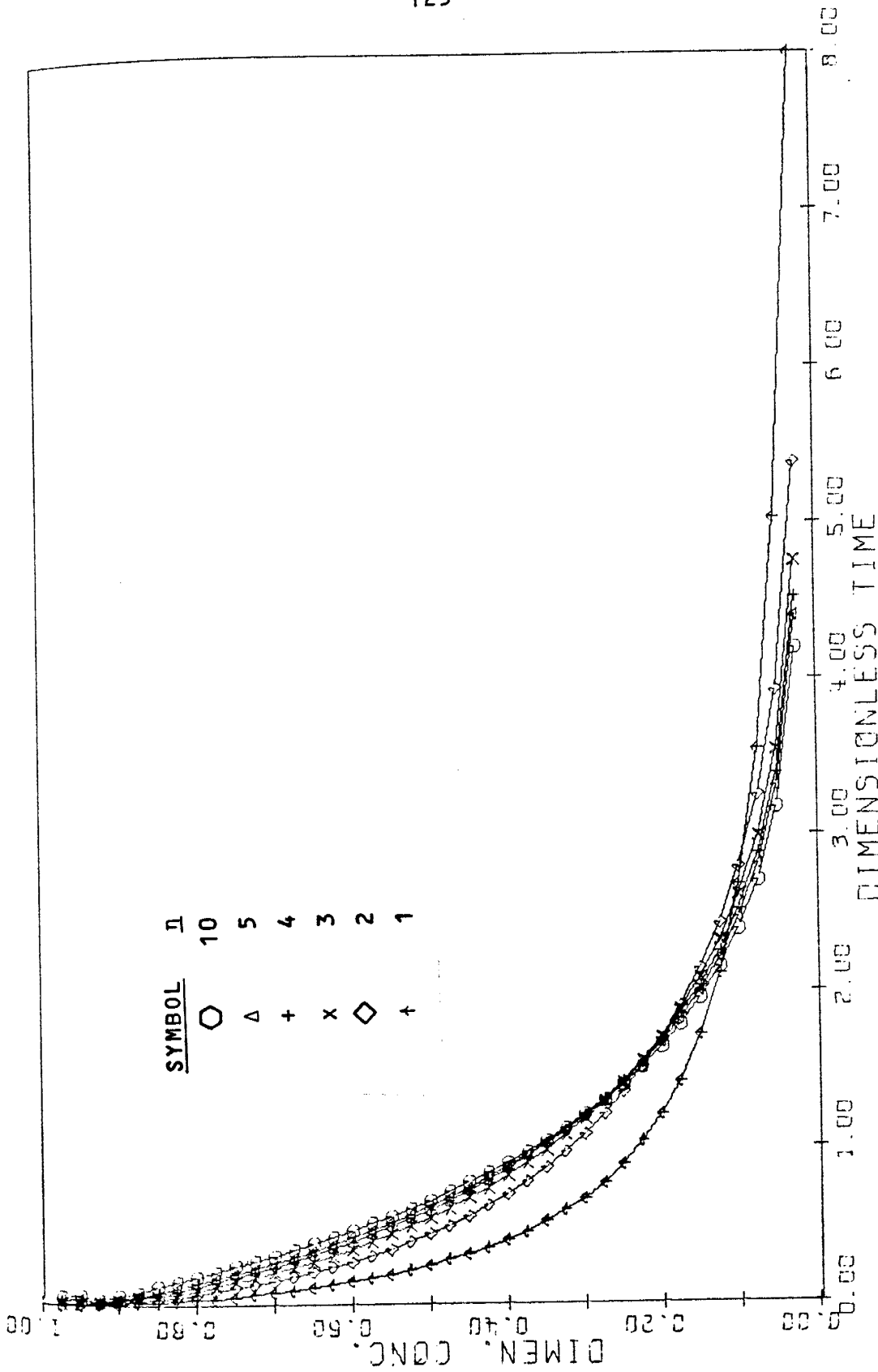


Figure 2.44: c/c_0 vs. t/t_{c1} for different n ; case (5) in Table 2.3 with $\alpha = 10$.

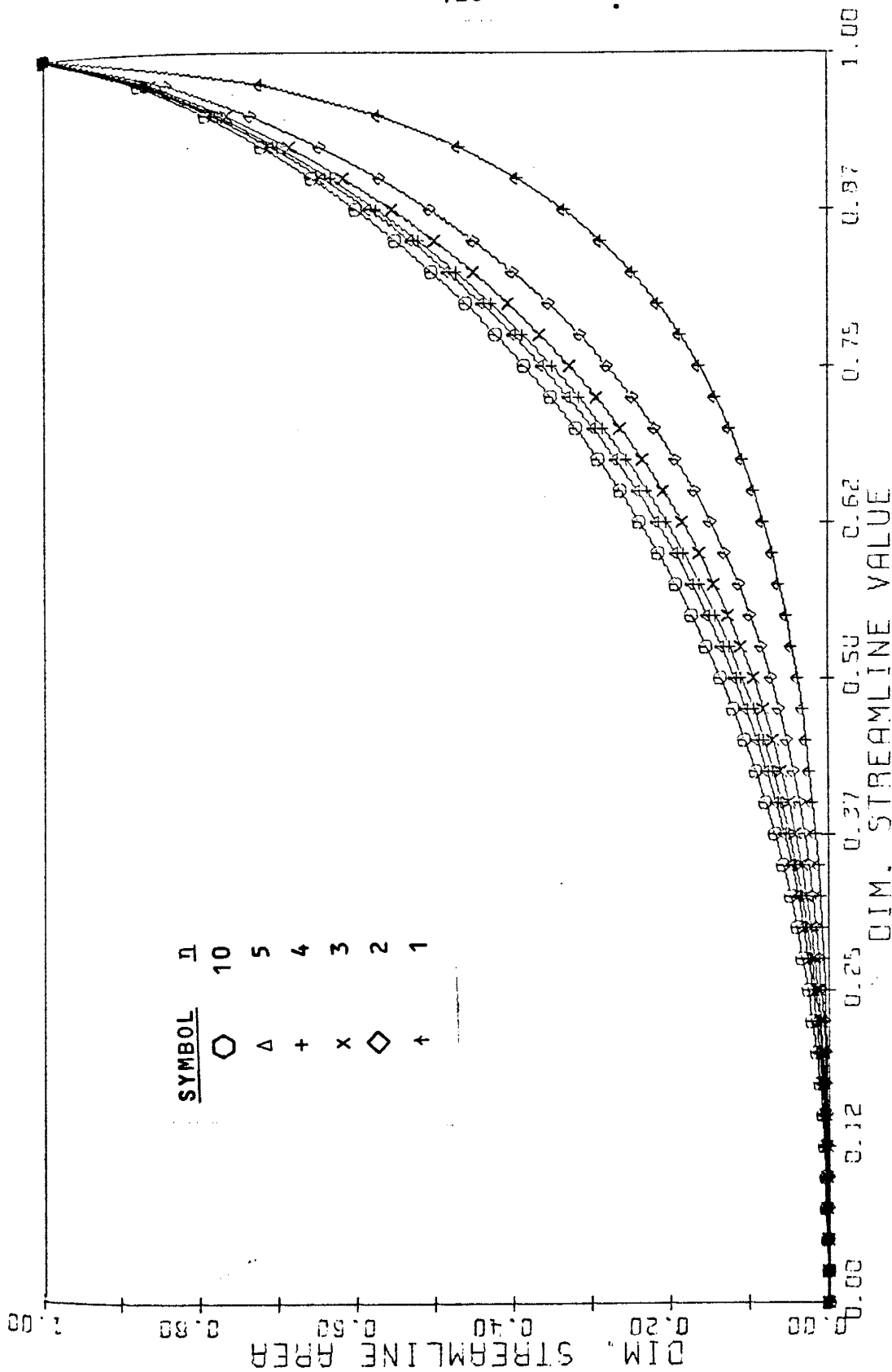


Figure 2.45: $\bar{A}(\bar{\Psi})$ vs. $\bar{\Psi}$ for different n ; case (5) in Table 2.3 with $\alpha = 100$.

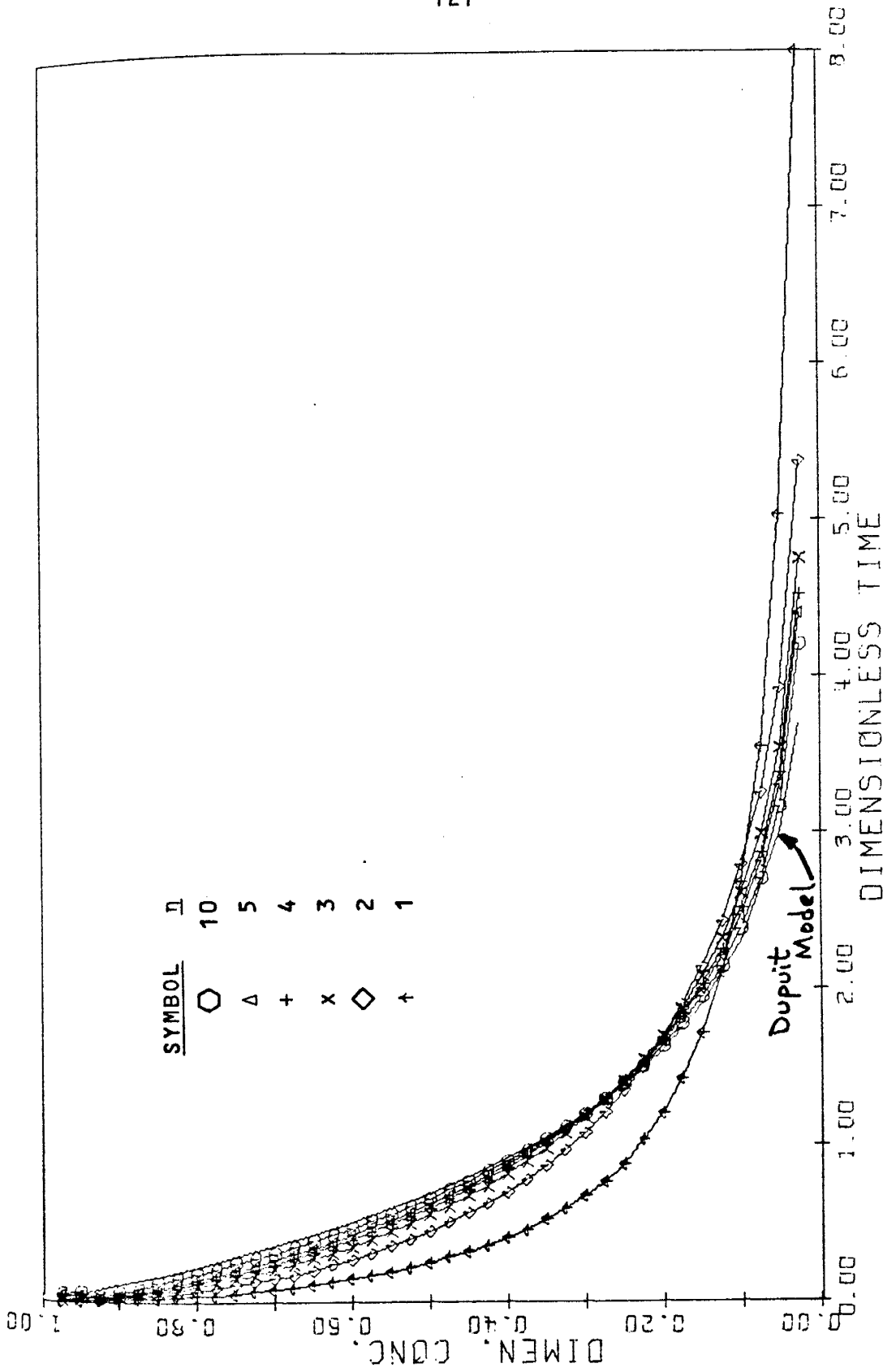


Figure 2.46: c/c_0 vs. t/t_{c1} for different n ; case (5) in Table 2.3 with $\alpha = 100$.

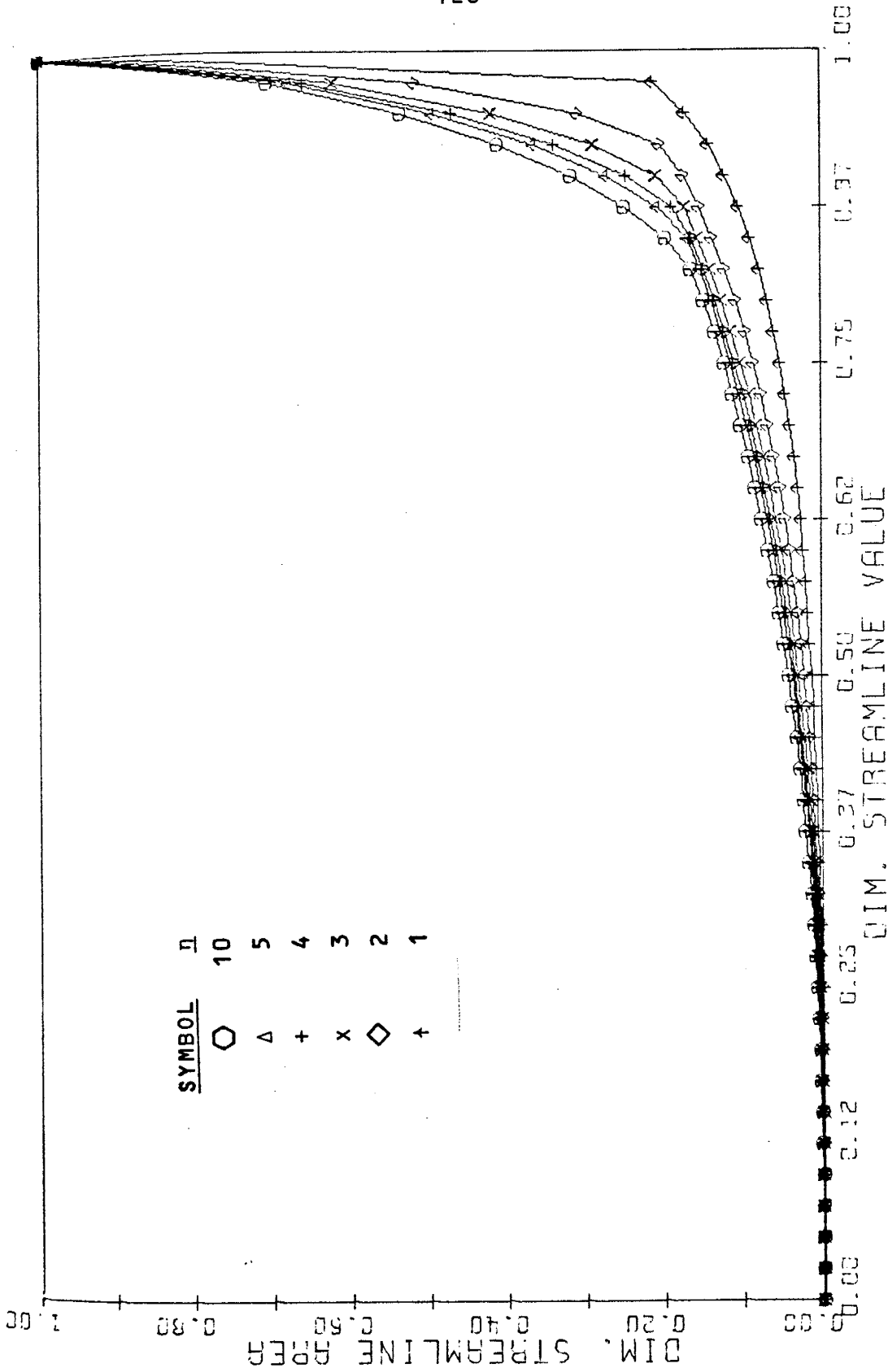


Figure 2.47: $\bar{A}(\bar{\Psi})$ vs. $\bar{\Psi}$ for different n ; case (6) in Table 2.3 with $\alpha = 10$.

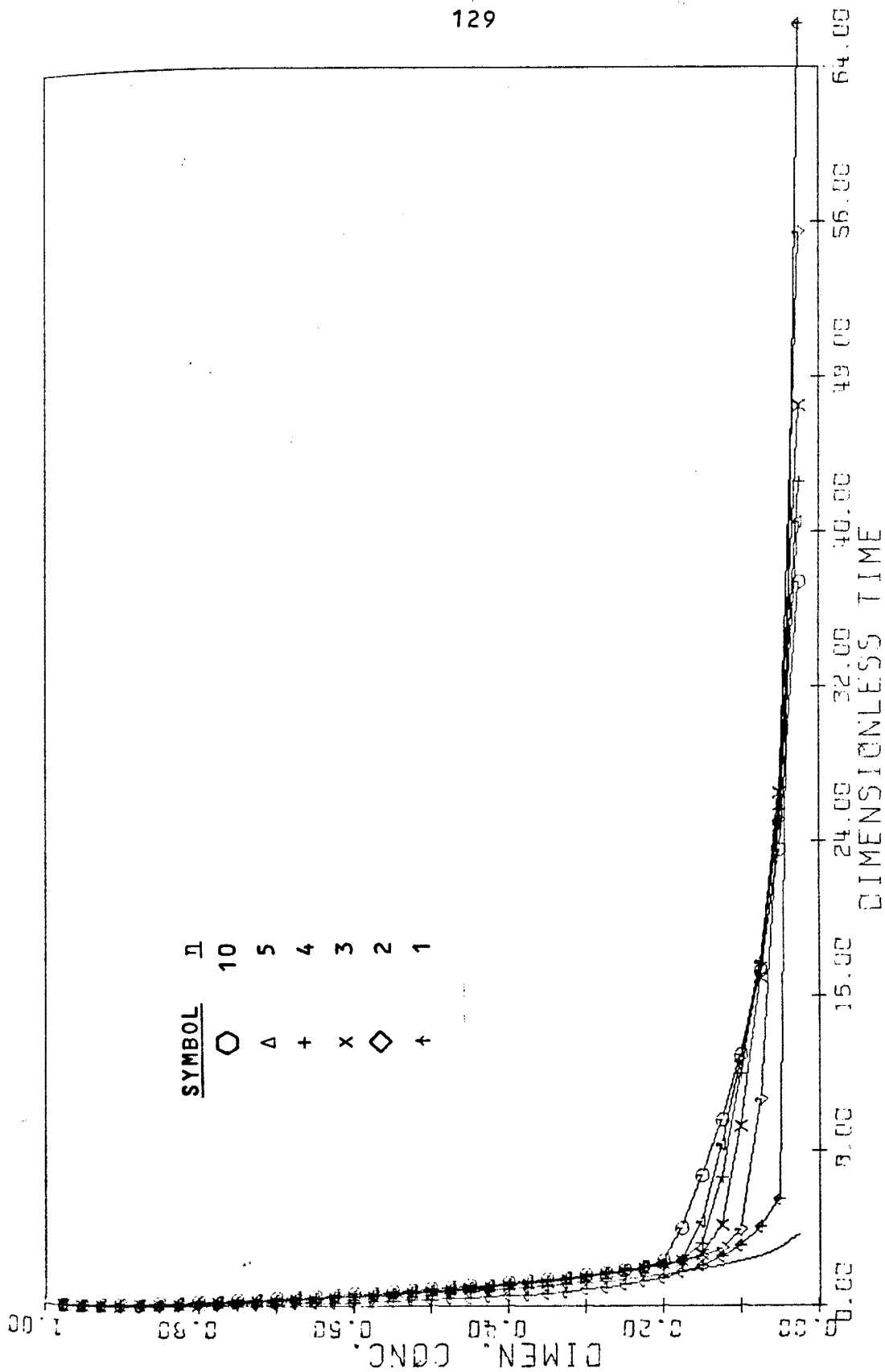


Figure 2.48: c/c_0 vs. t/t_{c1} for different n ; case (6) in Table 2.3 with $\alpha = 10$.

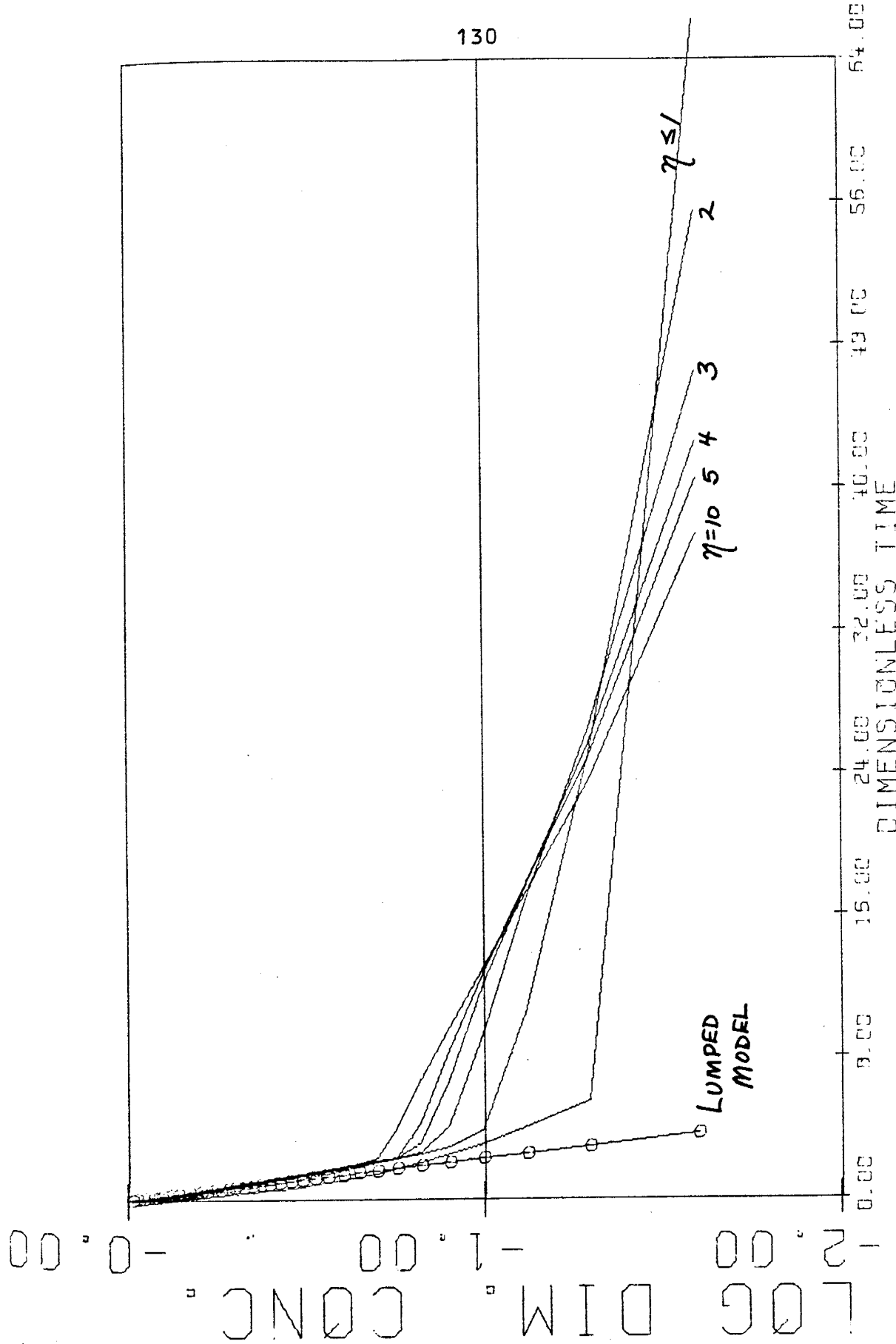


Figure 2.49: $\log c/c_0$ vs. t/t_{c1} for different n ; case (6) in Table 2.3 with $\alpha = 10$.

dimensionless c/c_0 versus t/t_c curve should approach 1.0 after some sufficiently long time. These areas were calculated for each of the cases examined in Table 2.3, and are summarized in Tables 2.4a and 2.4b.

2.4.7 Relationship Between the Laplace and Combined Lumped Models

We are now in a position to compare concentration outflow histories between the multi-layered Laplace and combined lumped parameter models. Again, we are still using the same steady recharge condition with a flushing initial condition for input concentration. We have already seen the Laplace outflow curves for several multi-layered aquifer configurations in the previous section; similarly, the combined lumped results were also previously given. Plotting these results on the same scale then, will enable a direct visual comparison between models. These graphs are presented in Figures 2.50 through 2.55 for several conditions already given in Tables 2.2. and 2.3. Recall that the combined lumped results were found using (2.2.23) with uniform porosity in both aquifers. Thus

$$c/c_{aq} = (1-\beta) \exp(-t/t_{c1}) + \beta \exp(-\gamma t/t_{c1}) \quad (2.4.16)$$

where $\beta = u/(1-\gamma)$

$$\gamma = u/m$$

$$m = b_L/(h_0 - b_L)$$

$$u = (mk_L/K_u)/(1+mk_L/K_u)$$

Table 2.4a. Areas Under c/c_0 vs. t/t_c For Case (1) shown in Table 2.3.

$\eta=L/D$	$\alpha=20$	$\alpha=10$	$\alpha=5$	$\alpha=2$
1	0.825	0.878	0.918	0.952
2	.902	.946	.968	.980
3	.929	.960	.975	.983
4	.935	.962	.976	.984
5	.938	.964	.976	.984
10	.940	.963	.974	.978
See Fig. No.	2.23	2.26	2.29	2.32

Table 2.4b. Areas Under c/c_0 vs. t/t_c For Case (1) shown in Table 2.3.

$\eta=L/D$	Case (2)	Case (3)	Case (4)	Case (5)		Case (6)
	$\alpha=10$	$\alpha=10$	$\alpha=10$	$\alpha=10$	$\alpha=100$	$\alpha=10$
1	0.908	0.930	0.977	0.965	.965	.812
2	.968	.976	.988	.984	.984	.930
3	.975	.981	.989	.984	.980	.954
4	.977	.982	.989	.964	.958	.961
5	.977	.982	.988	.957	.952	.963
10	.972	.980	.962	.944	.937	.966
See Fig. No.	2.35	2.38	2.41	2.44	2.46	2.48

NOTE: To compute the actual area under the c/c_0 vs. t/t_{c1} curves depicted in Figures 2.22 to 2.49, multiply the appropriate area by the corresponding N value given in Table 2.3. The areas listed here are for the c/c_0 vs. t/t_c curves and are intended for comparison to the lumped parameter curve area previously found to be 0.977. Furthermore, the integration procedure used in these area determinations are terminated when $c/c_{aq} = 0.025$.

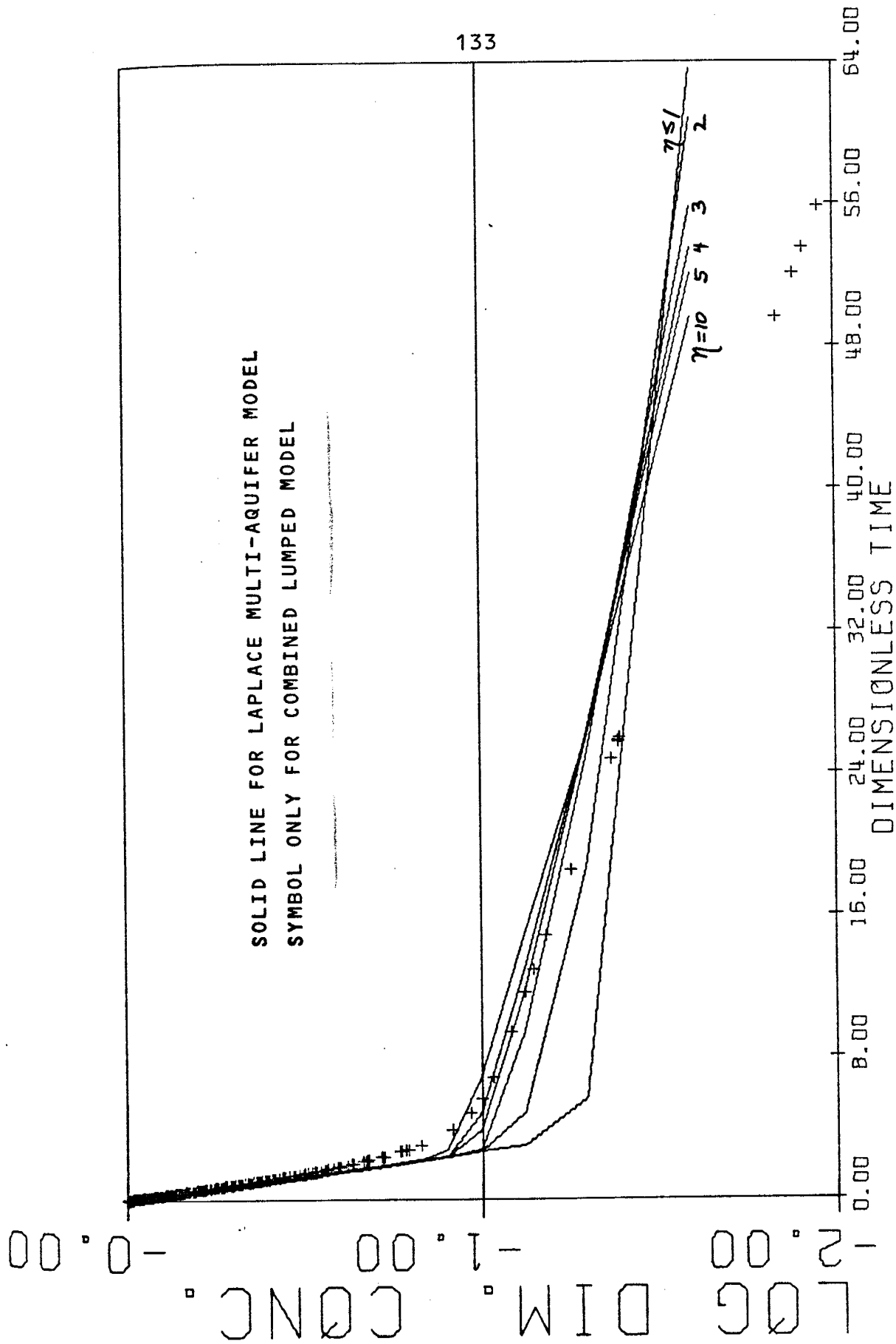


Figure 2.50: Laplace and combined lumped model c/c_{aq} vs. t/t_{c1} curves (see Figures 2.6 and 2.24).

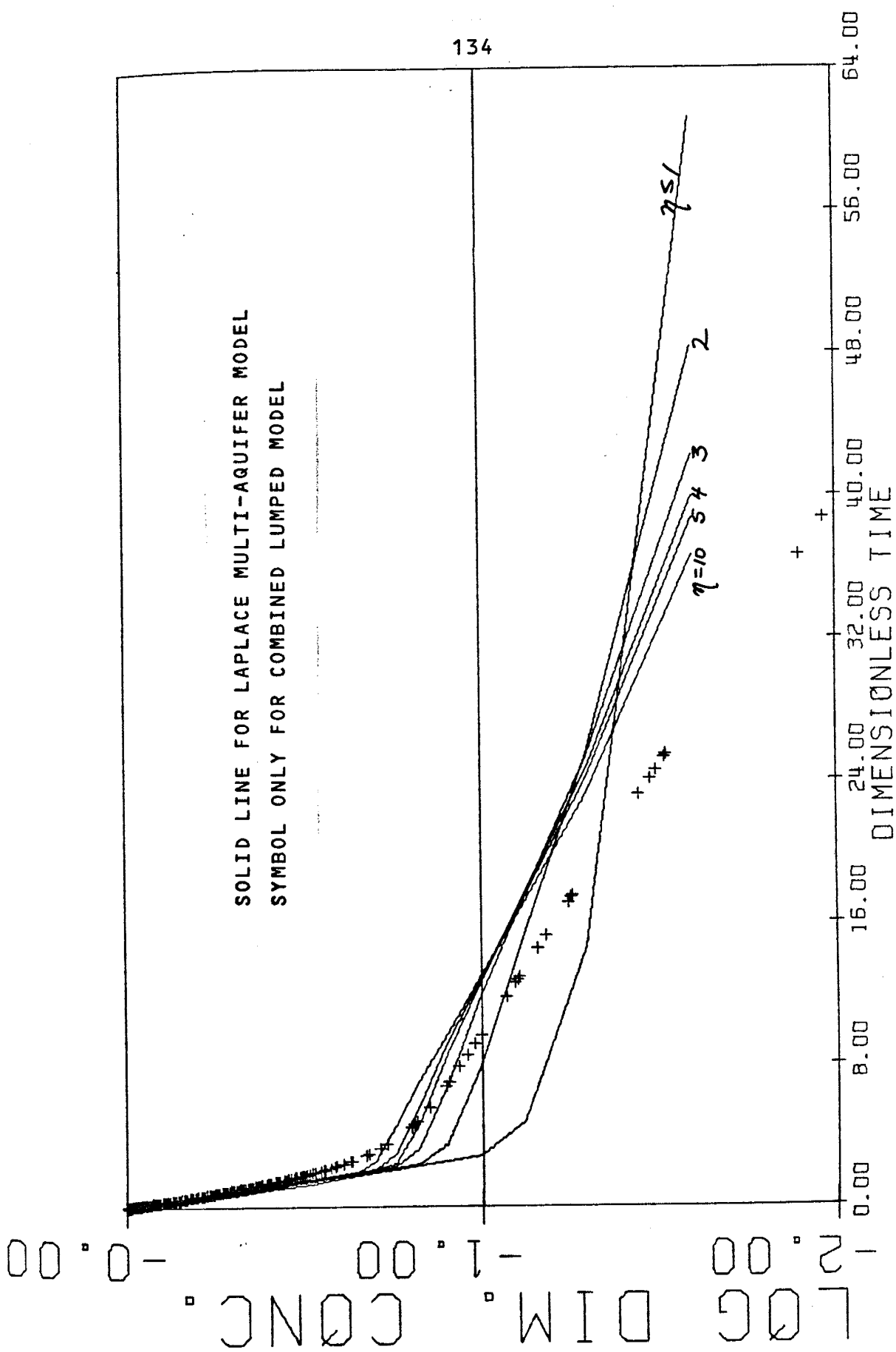


Figure 2.51: Laplace and combined lumped model c/c_{aq} vs. t/t_{c1} curves (see Figures 2.7 and 2.27).

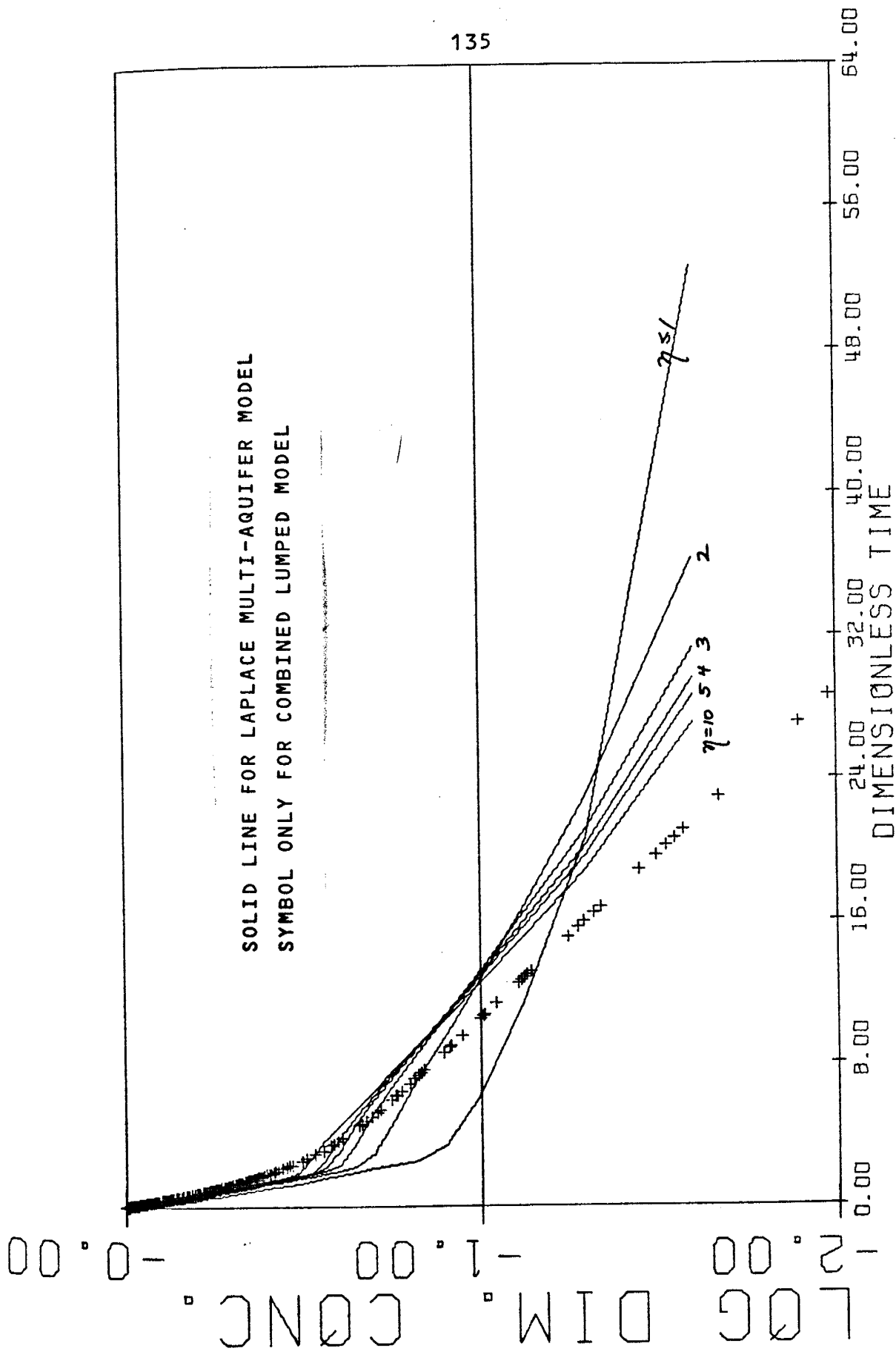


Figure 2.52: Laplace and combined lumped model c/c_{aq} vs. t/t_{c1} curves (see Figures 2.8 and 2.30).

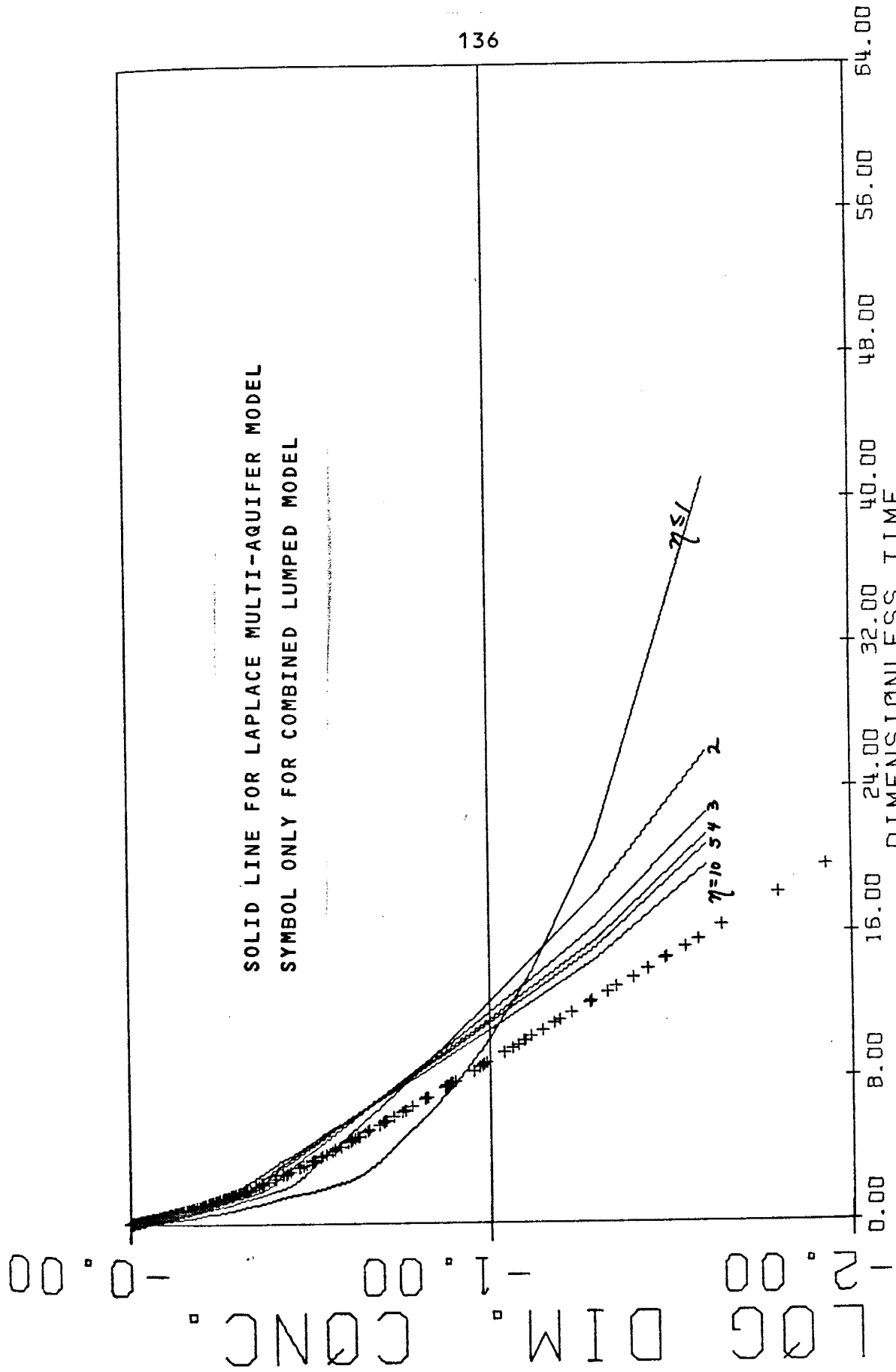


Figure 2.53: Laplace and combined lumped model c/c_{aq} vs. t/t_{c1} curves (see Figures 2.9 and 2.33).

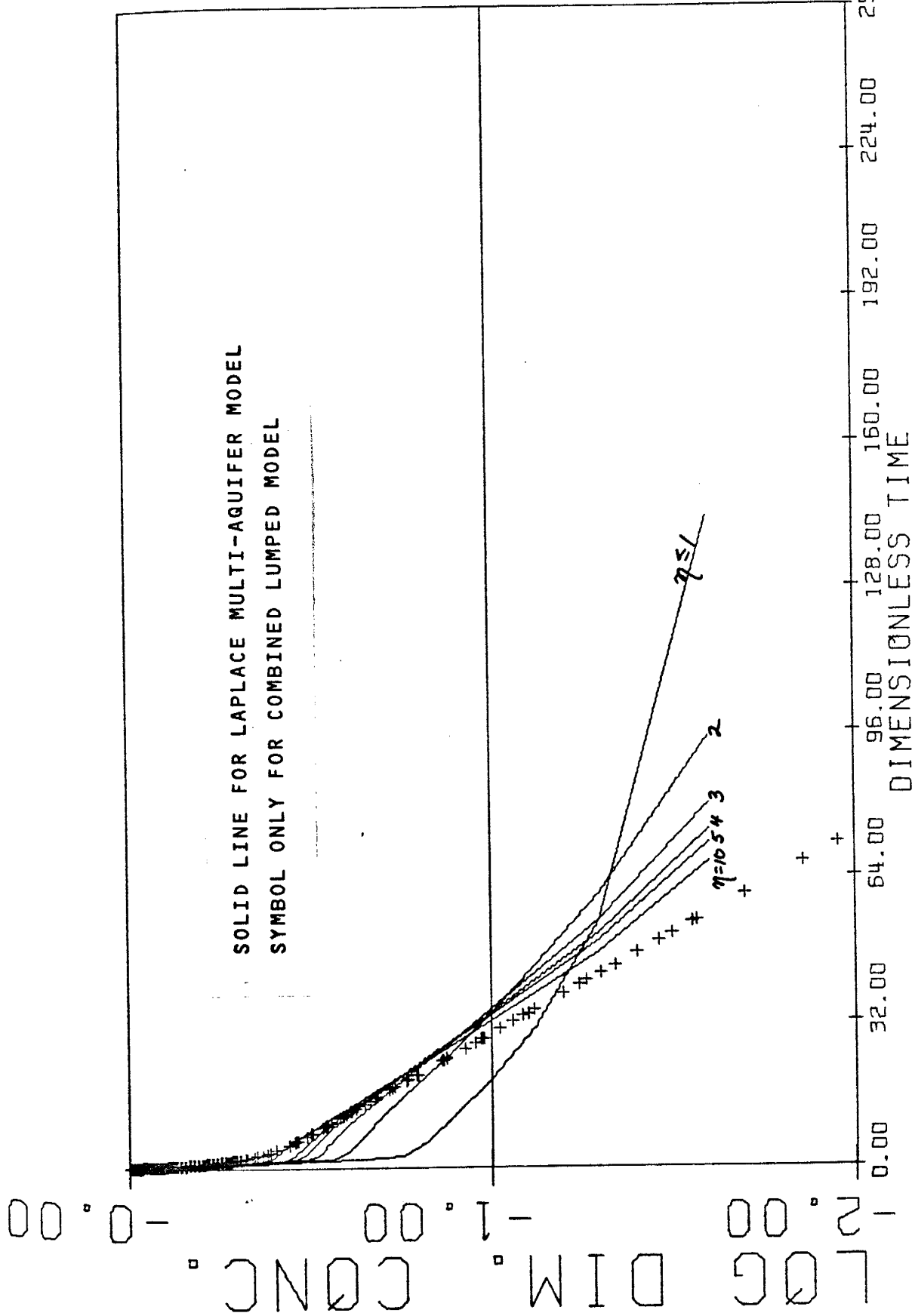


Figure 2.54: Laplace and combined lumped model c/c_{aq} vs. t/t_{c1} curves (see Figures 2.11 and 2.36).

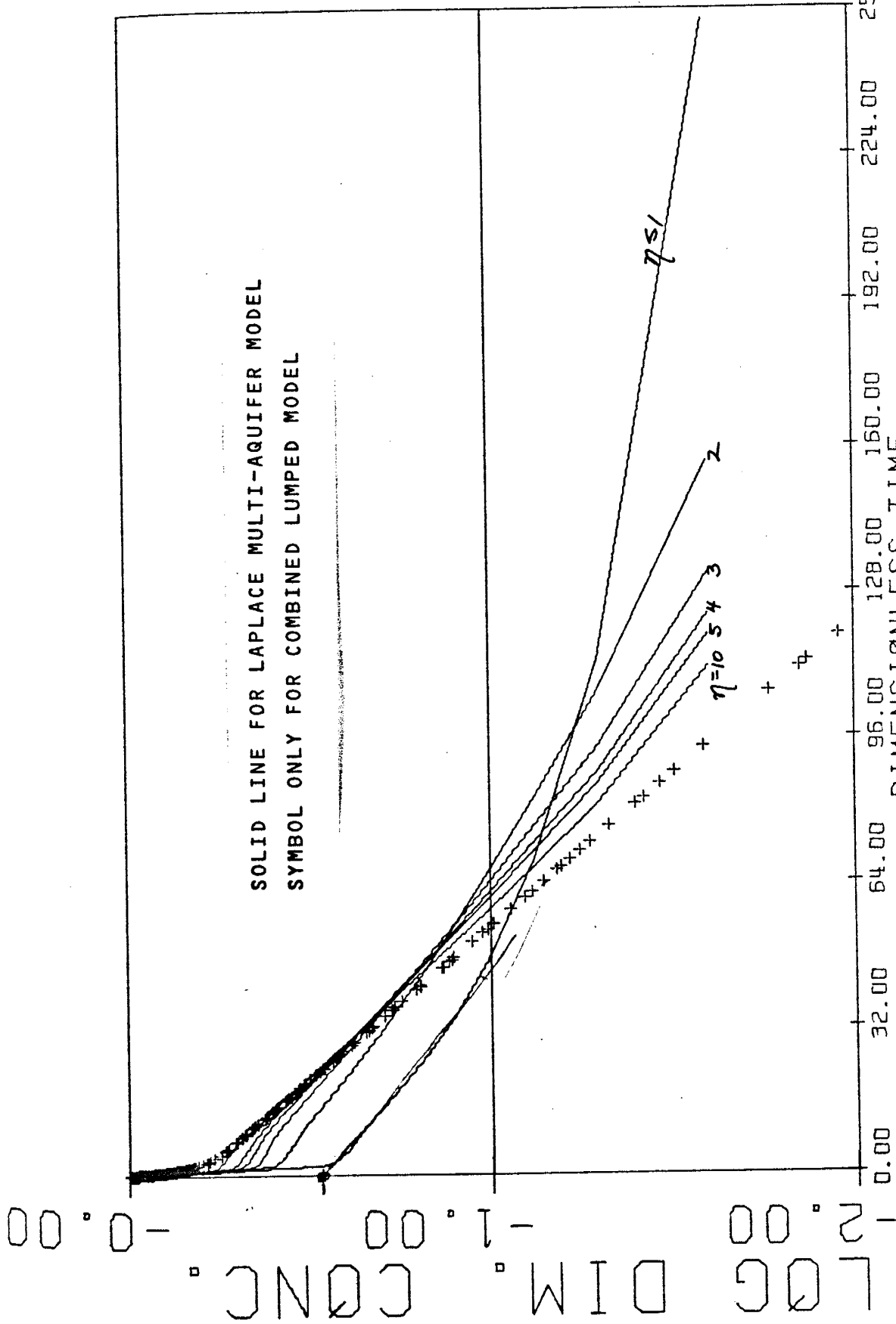


Figure 2.55: Laplace and combined lumped model c/c_{aq} vs. t/t_{c1} curves (see Figures 2.12 and 2.39).

The relation for μ was found using the Dupuit approximation. For convenience, we have also assumed that the upper and lower reservoirs have identical initial average concentrations. Figures 2.50 through 2.55 show that the multi-layered Laplace results are approximated fairly well by (2.4.16), although differences in concentration resulting from vertical flow are not considered. Even when these vertical flow components become quite pronounced (Figure 2.55, for example) the combined lumped model still predicts concentration outflow behavior surprisingly well.

Equation (2.4.16) may be thought of as simply two exponential curves summed together to give the concentration in the stream. The first exponential term represents convective travel time through the upper reservoir, while the second term represents travel time in the lower aquifer. At early time, the first exponential term predominates since γ is less than one in all cases examined here. Similarly, the second term dominates at late time. By "adjusting" the parameters β and γ we may alter the shape of the combined lumped parameter breakthrough curves. It may easily be shown that the second term in (2.4.16) has a slope of $-\gamma$ and a c/c_{aq} intercept of β when the dimensionless concentration versus dimensionless time values are plotted on semi-logarithmic scales respectively. For example, the first term in (2.4.16) has a slope of -1 and an intercept of $+1$. Since the Laplace concentration break-through curves also appear to exhibit this exponential behavior, we should be able to read these intercept and slope values of the lower portion of each of the Laplace curves. Substitution of these resulting new β and γ values into (2.4.16) will then "correct" the combined lumped parameter outflow concentration so that vertical flow effects are approximately

taken into account. Figures 2.56 and 2.57 show the relationship between these respective slope and intercept values. To use the figures, we must first calculate μ , m , and γ for the lumped model using the previously derived Dupuit relationships given along with (2.4.16). The lumped model intercept value will ideally be μ , while its slope will be given by γ . Once we know these values, we can refer to Figures 2.56 and 2.57. Reading upward from the appropriate abscissa to the corresponding $n=L/D$ ratio in the Laplace model and over to the ordinate, we obtain the "corrected" Laplace β and γ values to be used in (2.4.16). If the Laplace and lumped models agreed perfectly, then the n curves would fall on the "theoretical" line indicated in each figure. We can see that for large n , and hence decreased importance of vertical flow, these curves do approach this theoretical limit. Thus vertical flow effects may be incorporated into the seeming crude lumped parameter model with reasonable success. For example, with a K_L/K_U ratio of 0.20 and a value of $m=3$, Table 2.5 lists the corresponding values of γ from Figure 2.56 and values of β from Figure 2.57; these γ and β values may be used in (2.4.16) to correct the combined lumped model for vertical flow effects. Figures 2.58 through 2.63 show the resulting concentration breakthrough curves so obtained; the corresponding Laplace curves previously found are also shown for reference.

Table 2.5. Values of γ and β from Figures 2.56 and 2.57 with $K_L/K_U = 0.20$, $m=3$, and $\mu = 0.375$.

<u>n</u>	<u>γ</u>	<u>β</u>	<u>REMARKS</u>
10	0.096	0.322	See Fig. 2.58
5	.084	.280	See Fig. 2.59
4	.081	.266	See Fig. 2.60
3	.074	.242	See Fig. 2.61
2	.062	.202	See Fig. 2.62
1	.038	.121	See Fig. 2.63

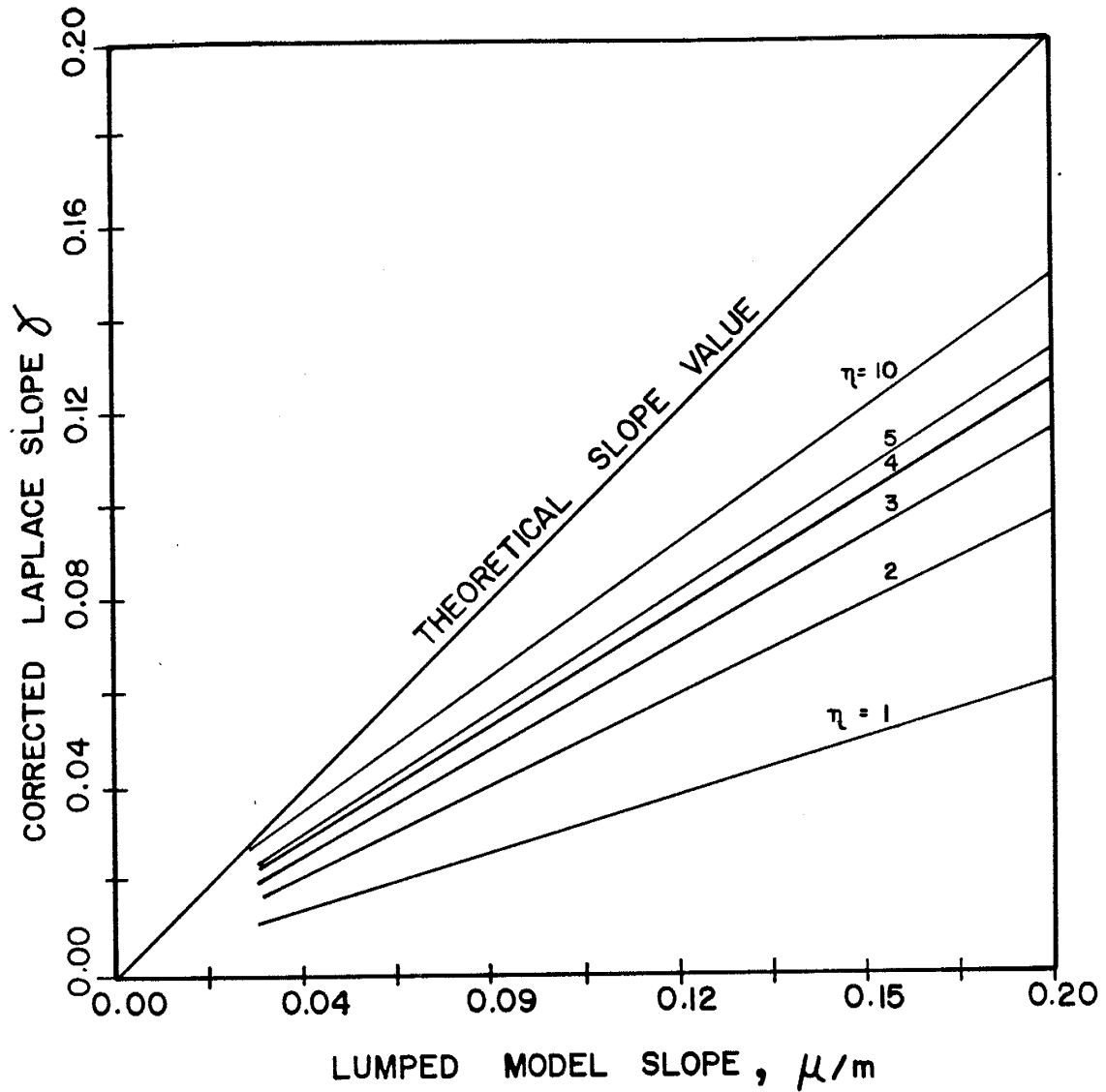


Figure 2.56 Lumped (μ/m) vs. Laplace (γ) model slopes for different values of η .

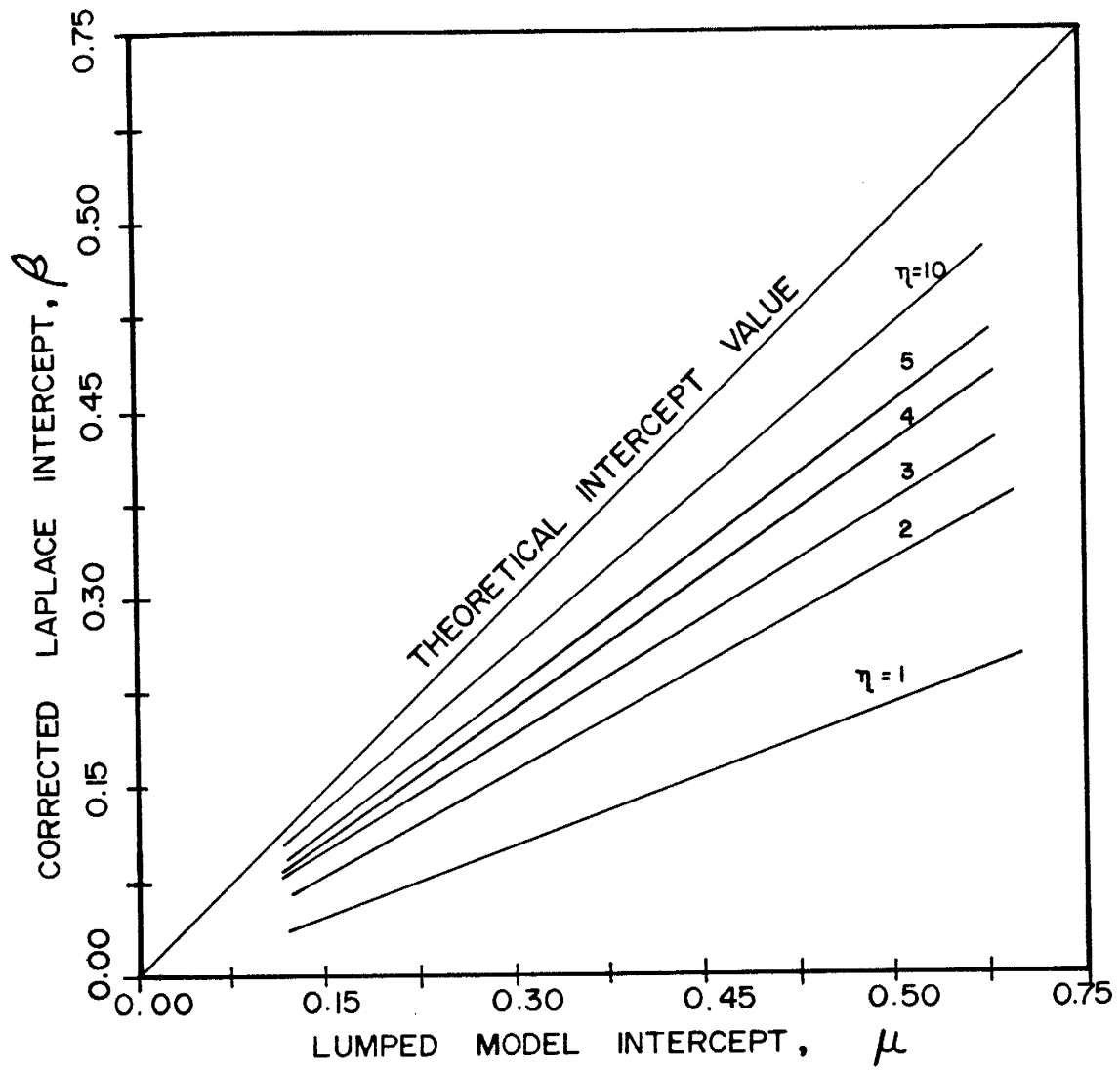


Figure 2.57 Lumped (μ) vs. Laplace (β) model intercepts for different values of n .

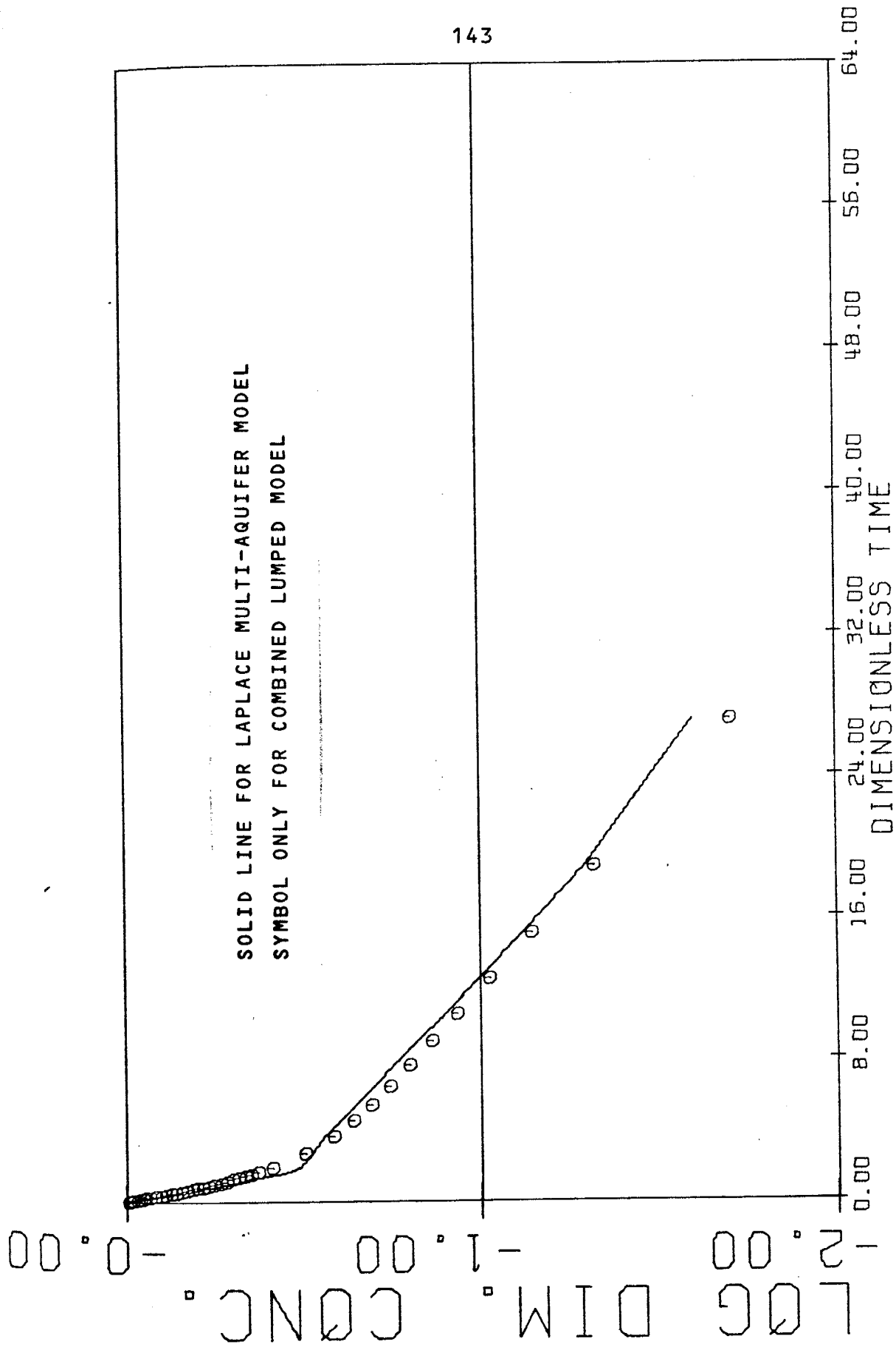


Figure 2.58: Corrected lumped vs. Laplace c/c_{aq} vs. t/t_{c1} curves for $n = 10$.

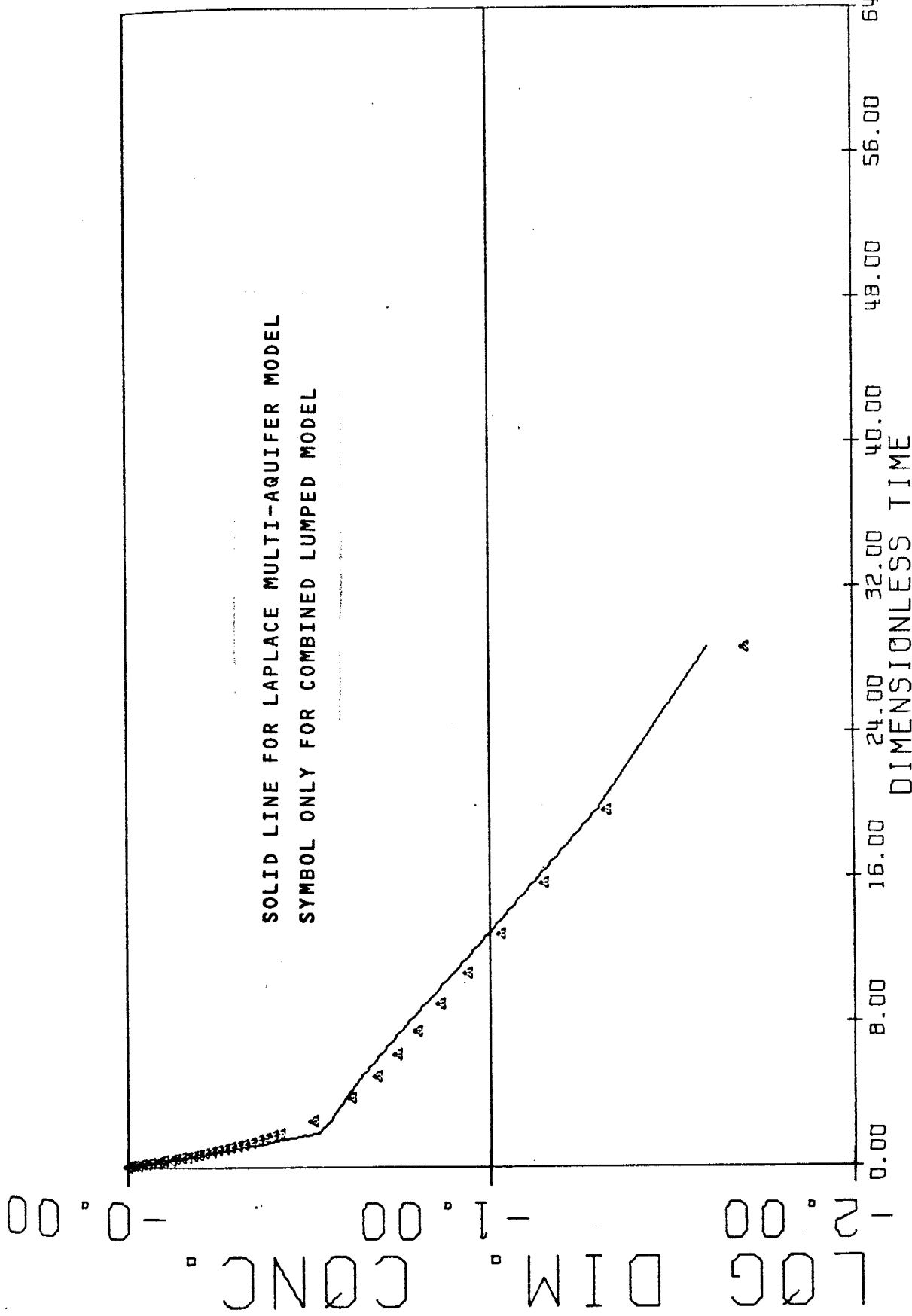


Figure 2.59: Corrected lumped vs. Laplace c/c_{aq} vs. t/t_{c1} curves for $n = 5$.

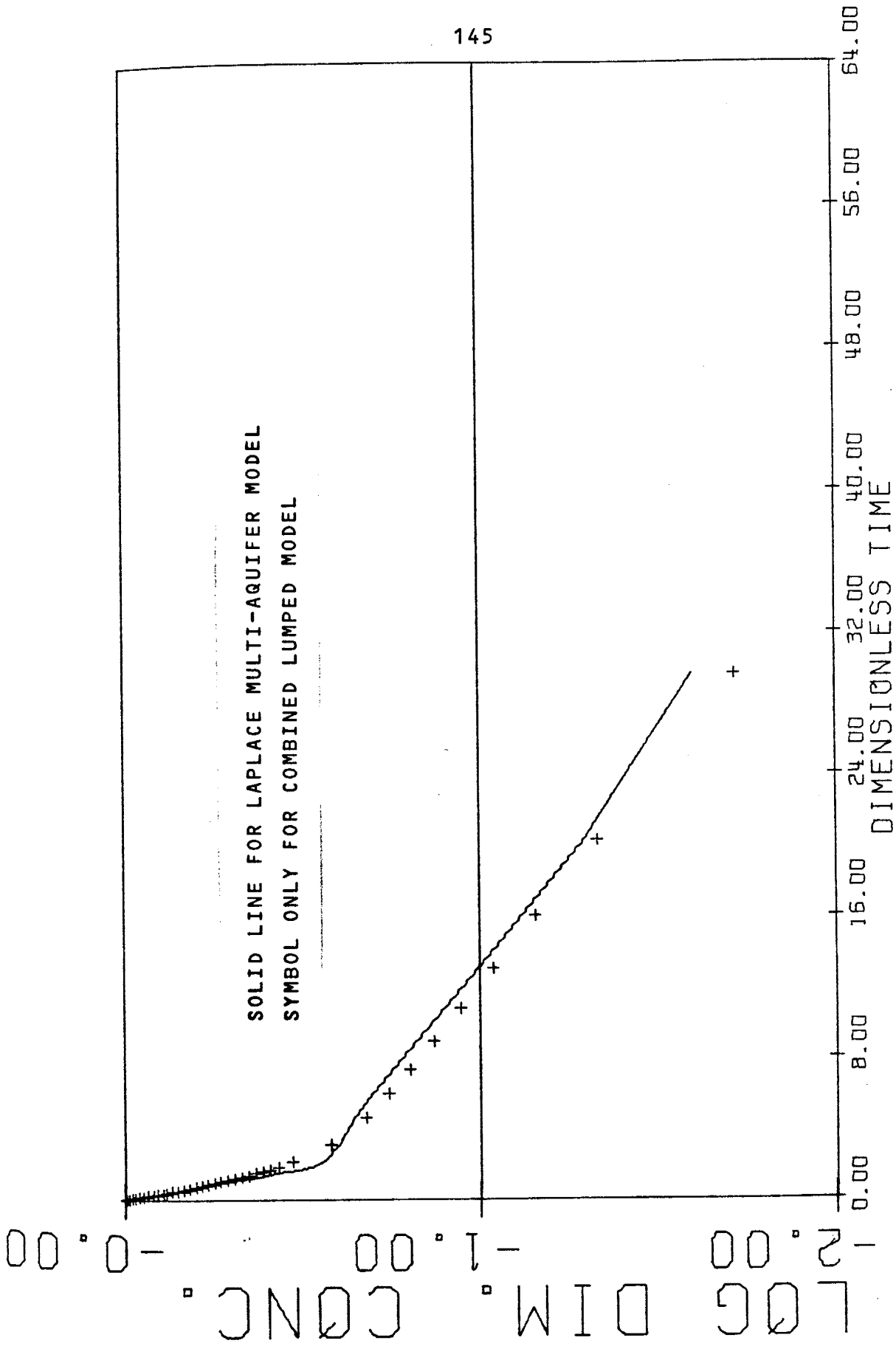


Figure 2.60: Corrected lumped vs. Laplace c/c_{aq} vs. t/t_{c1} curves for $n = 4$.

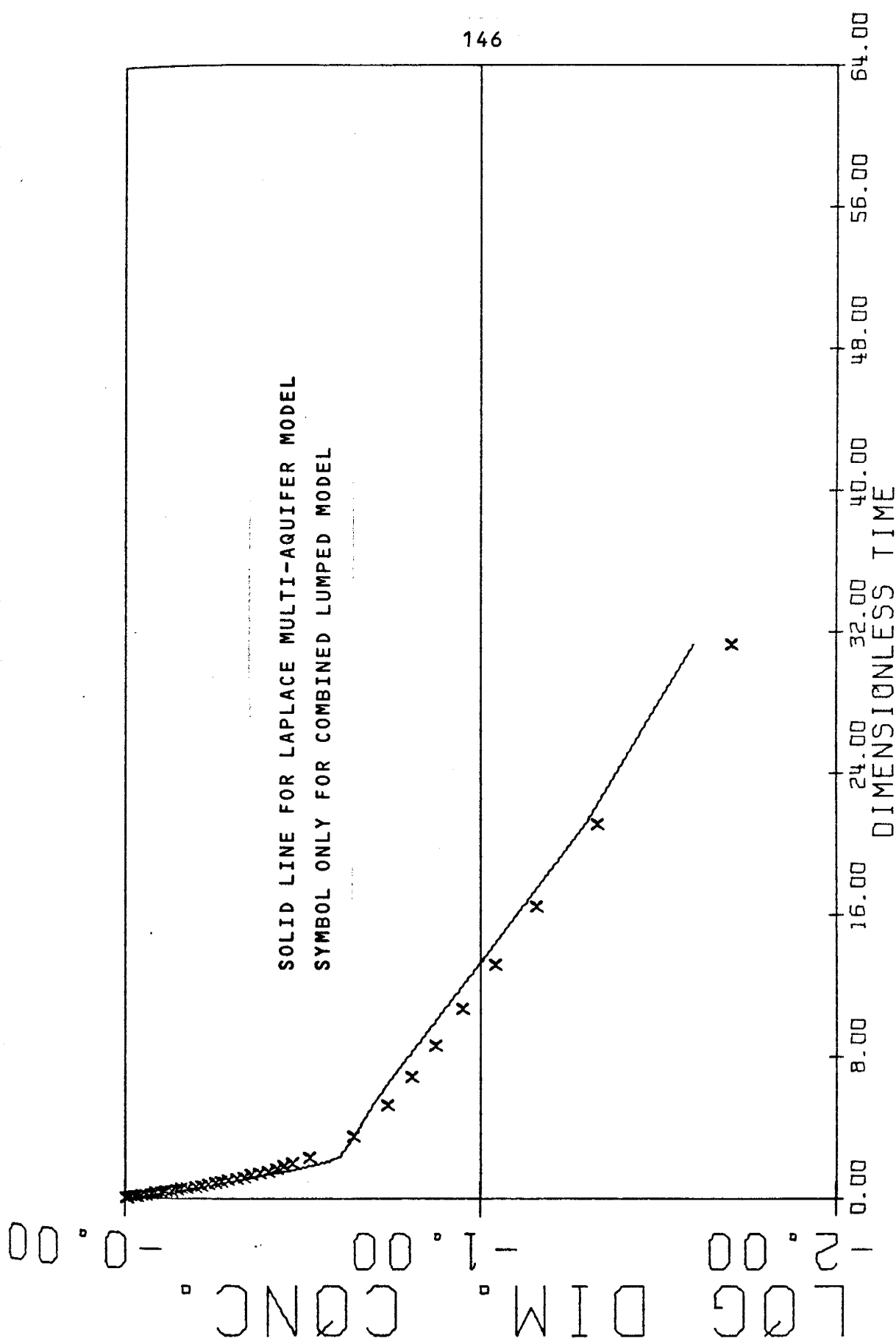


Figure 2.61: Corrected lumped vs. Laplace c/c_{aq} vs. t/t_{c1} curves for $n = 3$.

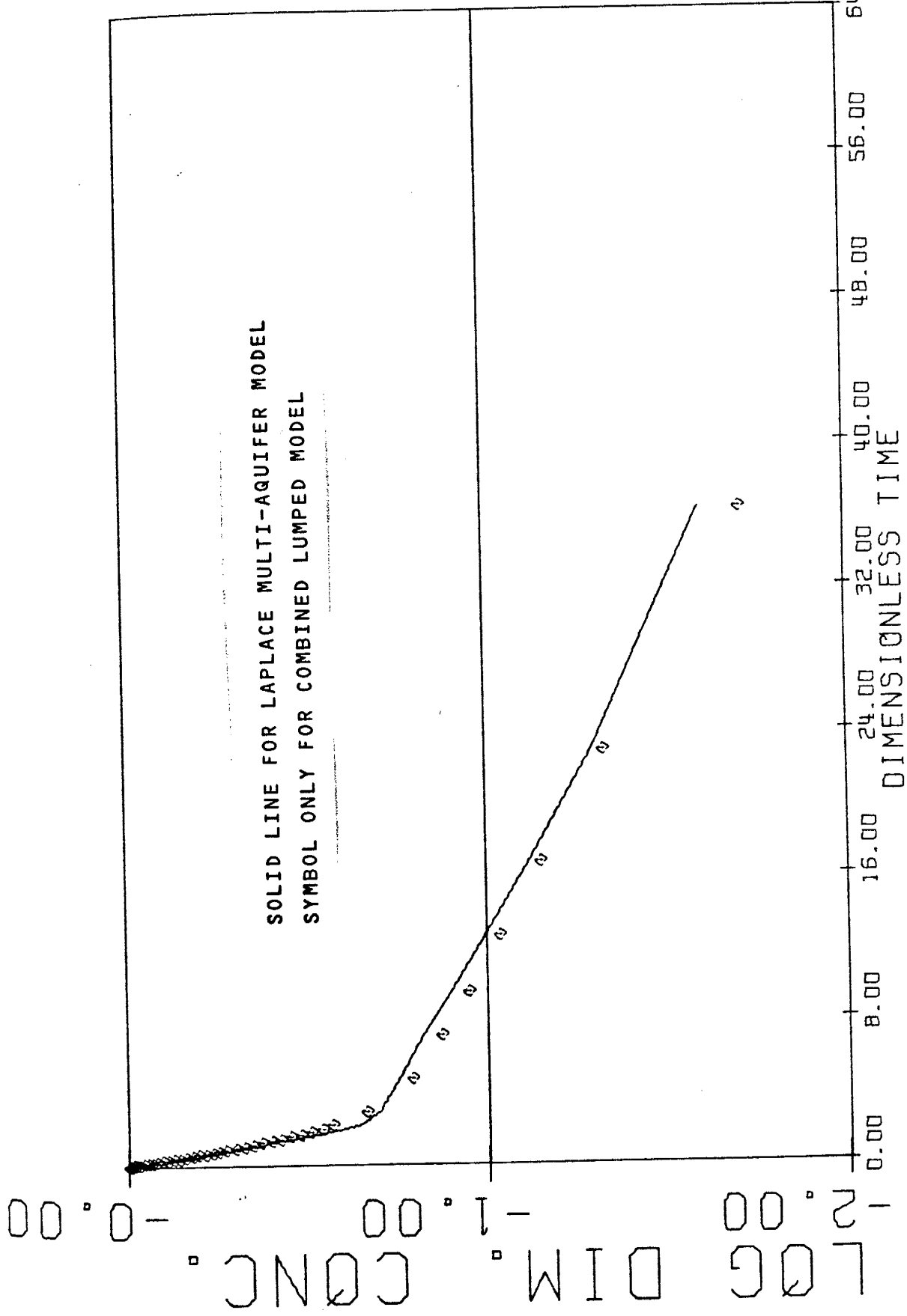


Figure 2.62: Corrected lumped vs. Laplace c/c_{aq} vs. t/t_{c1} curves for $n = 2$.

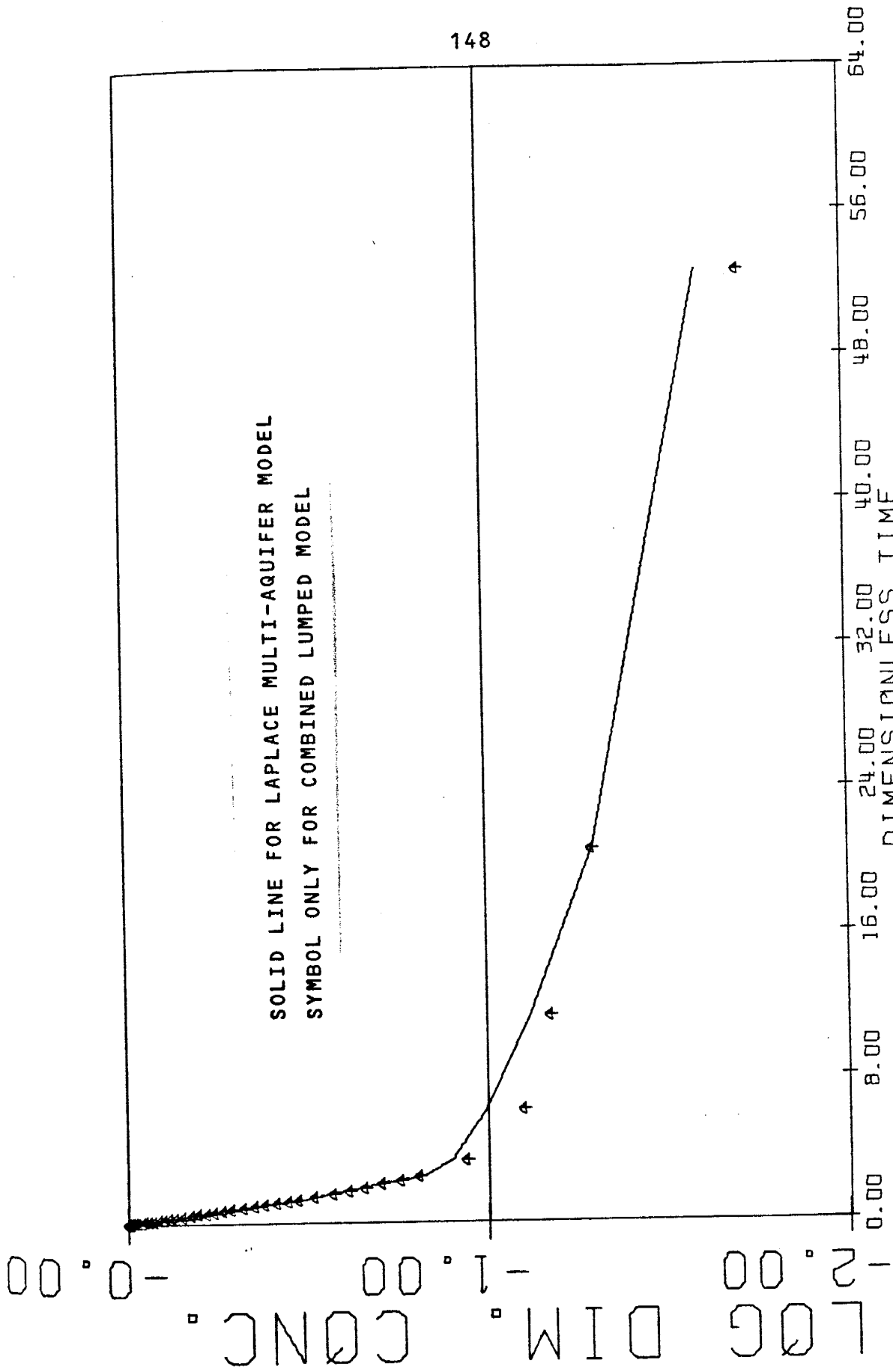


Figure 2.63: Corrected lumped vs. Laplace c/c_{aq} vs. t/t_{c1} curves for $n = 1$.

2.5 General Conclusions

The following general conclusions are made based on the foregoing analyses conducted in this chapter.

1. For a constant uniform recharge and a step change in input concentration, the preceding analyses have shown that the lumped parameter model in the form of a linear reservoir yields an identical convective outflow concentration in the stream compared to the one-dimensional Dupuit aquifer analysis ((2.2.15) and (2.3.23) are identical). Furthermore, the instantaneous average concentration of contaminants being transported through the Dupuit aquifer was shown to be identical to this stream concentration. Hence the well-mixed reservoir assumption inherent to the lumped parameter approach is justified, based on these two-dimensional modeling considerations. A similar conclusion was also reached by Gelhar and Wilson (1974).

2. For multi-layered aquifer systems with constant uniform recharge and a step change in input concentration, the multi-layered Dupuit analysis produced a nearly identical convective outflow concentration to two lumped parameter models linked in a parallel configuration. Thus (2.2.23) and (2.3.64) show nearly identical concentration break-through curves as seen in Figures 2.6 through 2.12. Furthermore, the total mass leaving each system was shown to be identical.

3. Vertical flow effects on convective outflow concentration were examined from a two-dimensional Laplacian point of view using the steady uniform recharge condition and a step change in contaminant input concentration. These vertical flow components of specific discharge are a consequence of the assumed partial stream penetration and an

anisotropic, homogeneous porous medium. Concentration break-through curves for different values of the aspect ratio, $n=(L/D)(K_z/K_x)^{1/2}$, are shown in Figure 2.19; the parameter n combines anisotropy and partial penetration into a single variable. For values of n greater than about five, the lumped parameter model (or equivalent Dupuit model) approximates these Laplace break-through curves quite well. However, for values of n less than about five, there are considerable differences between curves. At early time, the Laplace model concentration curves show faster arrival times at the stream or tile drain than does the lumped model curve. At late time, however, these Laplace curves are "retarded" in comparison to the lumped curve. These arrival times simply reflect the longer contaminant flow paths resulting from pronounced vertical flow effects. Furthermore, when the impermeable basement is taken at some infinite depth below the stream or tile drain, the resulting concentration break-through curve is nearly identical to that for $n = 1$. Thus the vertical flow effects seem to reach a maximum for near a value of n near one. In other words, the shallow aquifer circulation pattern does not significantly change for n values less than about one; the resulting break-through curves therefore do not significantly change either. In these analyses, the degree of partial stream penetration was assumed to be very small in relation to total saturated thickness. Thus, the Laplace results presented here, along with the fully penetrating Dupuit or equivalent lumped parameter analysis, should be viewed as extreme situations. When the degree of stream penetration is not small in comparison to the total saturated thickness, the Laplacian analysis should yield concentration break-through curves somewhere between the above mentioned limits.

4. The effects of stream clogging, resulting from reductions in hydraulic conductivity up to two orders of magnitude in the small zone surrounding the stream or tile drain, have negligible effects on concentration break-through curves (see Figure 2.18 compared to Figures 2.44 and 2.46).

5. The vertical flow effects associated with multi-layered aquifer systems were shown to have an influence on the convective outflow concentration in the stream or tile drain (see Figures 2.24, 2.27, 2.30, 2.33, 2.36, 2.39, 2.42, and 2.49). These results indicate large systematic departures from the well-mixed linear lumped parameter curve at late time, even though the early time concentrations are in fair agreement. These departures are most pronounced for relatively thin upper layers (i.e., the upper layer is less than 1/4 the total saturated thickness); and for the upper layer having nearly the same hydraulic conductivity as the lower layer. Vertical flow effects obviously are important under such circumstances. However, for an upper to lower hydraulic conductivity (K_U/K_L) equal to 20, and a total saturated thickness to upper layer thickness ratio (D/h) equal to four, the lumped model will accurately predict the stream concentration level down to about $c/c_0 = 0.15$ for all values of n (Figure 2.24). In all cases examined, decreasing the drain half spacing to total saturated thickness ratio (L/D) results in large departures from the lumped break-through curves. The pronounced vertical flow effects associated with smaller L/D ratios are at least partially offset by increases in the K_x/K_z ratio.

6. Outflow concentration from the lumped parameter model used in a parallel type configuration as given by (2.4.16) closely approximates the multi-layered concentrations predicted from the two-dimensional analysis

quite well (see Figures 2.50 through 2.55), although vertical flow effects do result in some differences between the Laplace break-through curves for different physical situations (i.e., different ratios for K_x/K_z , L/D , and D/h). Certain parameters in the combined lumped curves predicted from (2.4.16) may be corrected, however, using Figures 2.56 and 2.57. The resulting concentration break-through curves will closely approximate those from the multi-layered Laplace analysis (see Figures 2.58 through 2.63). Thus vertical flow effects resulting from aquifer layering may be approximately incorporated into the seeming crude lumped parameter model with surprisingly good results.

7. The foregoing analyses have attempted to justify certain simplifying assumptions inherent to the lumped modeling approach, and to demonstrate the overall flexibility for incorporating consideration of complex physical phenomena into this format. Though appearing to be superficially crude, this approach implicitly incorporates many multi-dimensional effects found in more elaborate models presently available.

CHAPTER 3

FIELD APPLICATION OF THE LINEAR RESERVOIR MODEL

3.1 Introduction

The predictive capabilities of the lumped parameter linear reservoir model were examined under actual field conditions in the Arkansas River Valley near La Junta, Colorado, where a distributed convective-dispersive model was previously tested (Konikow and Bredehoeft, 1974a; 1974b). Application of the lumped model under these conditions served a dual purpose. First the linear model could be field tested so as to demonstrate the applicability of this approach in representing average temporal hydrological and water quality variations occurring in an irrigated stream-aquifer system under a semiarid climatic setting. Secondly, the lumped model's performance could be directly compared to a conceptually more complex digital model. Of course one would not expect this lumped temporal approach to produce the complete system-wide description inherent to a time dependent spatial model; however, if one's primary objective is to simulate downstream changes in quantity and quality of surface flows, the lumped approach may be sufficient.

The area selected for this detailed field investigation includes an 11 mile reach of the Arkansas River in southeastern Colorado between La Junta and the Bent-Otero County line. This study area has been previously described by Konikow and Bredehoeft (1974a; 1974b; and references therein); Appendix D shows a site location map of the study area. Briefly, in this 11 mile study reach of the Arkansas River, downstream changes in surface water total dissolved solids and shallow

groundwater quality have been observed from some time. These changes have been attributed primarily to return flows from irrigated agriculture, but surface runoff from summer thundershowers also contributes to the surface water quality deterioration. The alluvial valley in the study reach is about 1.5 miles wide and covers a gross area of about 17 square miles. Southeastern Colorado has a predominantly semiarid climate characterized by low humidity and scant rainfall. The mean annual precipitation is about 13 inches, two-thirds of which occurs during the growing season (May through September). Irrigation waters are diverted from the Arkansas River into the Fort Lyon Canal just above the study area and west of La Junta. Several minor tributaries and the La Junta municipal sewage effluent enter the river within the study reach. When surface diversions are not sufficient to meet irrigation requirements, additional shallow groundwaters are pumped from the many irrigation wells located throughout the 11 mile reach. The La Junta municipal water supply also depends entirely upon groundwater withdrawn from the valley alluvium inside the study area. The saturated thickness of the recent alluvium in the study reach varies from only a few feet to over forty feet and averages about 18 feet. This shallow alluvium is underlain by a thick sequence of relatively impermeable bedrock.

The computer-based finite difference model used by Konikow and Bredehoeft (1974a) in their study is based on a discretized form of the partial differential equation governing groundwater flow, subject to appropriate initial and boundary conditions. For nonsteady, two-dimensional flow in a nonhomogeneous, anisotropic aquifer, this equation is given by

$$\frac{\partial}{\partial x_i} [T_{ij} \frac{\partial h}{\partial x_j}] = S \frac{\partial h}{\partial t} + W(x,y,t); \quad i,j = 1,2 \quad (3.1.1)$$

where T_{ij} is the transmissivity tensor (having dimensions of L^2/T); h is the hydraulic head (having dimensions of L); S is the storage coefficient (dimensionless); t is time (having dimensions of T); and W is the volume water flux per unit area (having dimensions of $L^3/T/L^2$). In (3.1.1) W is given by

$$W(x, y, t) = Q(x, y) + \frac{K_z}{m} (H_f - h) \quad (3.1.2)$$

where Q represents groundwater pumpage or recharge (dimensions of $L^3/T/L^2$); K_z is the vertical hydraulic conductivity of the streambed (dimensions of L/T); m is the streambed thickness (dimensions of L); and H_f is the hydraulic head in the river (dimensions of L). Using an iterative alternating direction implicit procedure, the authors were able to solve a discretized form of (3.1.1) for the time dependent spatial distribution of hydraulic head throughout the study reach. These values of head were then used to calculate the seepage velocity throughout the finite difference grid network using Darcy's law in the form of

$$v_i = - \frac{K_{ij}}{n} \frac{\partial h}{\partial x_j}; \quad i, j = 1,2 \quad (3.1.3)$$

where n is the effective porosity of the aquifer (dimensionless); K_{ij} is the hydraulic conductivity tensor (dimensions of L/T); and v_i is the seepage velocity (dimensions of L/T). Once this seepage velocity was found, the mass transport equation describing the time dependent spatial

distribution of dissolved chemical constituents in the saturated porous media could be solved. This mass transport equation is given by

$$\frac{\partial c}{\partial t} = \frac{\partial}{\partial x_i} \left(D_{ij} \frac{\partial c}{\partial x_j} \right) - \frac{\partial (v_i c)}{\partial x_i} - W_i; \quad i, j = 1, 2 \quad (3.1.4)$$

where c is the mass concentration of dissolved solids (dimensions of M/L^3); D_{ij} is the hydrodynamic dispersion tensor (dimensions of L^2/T); and W_i is the mass flux of any sources or sinks of dissolved solids (dimensions of M/L^3T). In the study by Konikow and Bredehoeft, contributions of molecular and ionic diffusion to the hydrodynamic dispersion tensor were neglected since seepage velocities were sufficiently large so as to overwhelm these effects. The hydrodynamic dispersion tensor was further decomposed into longitudinal and transverse components taken parallel and perpendicular, respectively, to a given streampath. These components were described by the relationship

$$D_L = \alpha_1 v \quad \text{and} \quad D_T = \alpha_2 v \quad (3.1.5)$$

where D_L and D_T are these respective components; α_1 and α_2 are the longitudinal and transverse dispersivities; and v is the seepage velocity. The authors further assumed that $\alpha_2 = 0.3 \alpha_1$, but pointed out that sensitivity of the output results to variations in this relationship were not evaluated. Finally the authors used the method of characteristics to solve the discretized version of (3.1.4) since this method does not introduce any artificial dispersion associated with the numerical procedure. The interested reader is referred to the original

references already given for a more complete description of these procedures used in the Arkansas River Valley study.

3.2 Arkansas River Linear Reservoir Model

Derivation of the lumped parameter linear reservoir model for application to the irrigated stream-aquifer configuration along the Arkansas River Valley is nearly identical to that previously derived in Chapter 2. However, slight modifications to the general form of the model are often required for specific case studies; the derivation here will thus serve to demonstrate the flexible nature of the lumped modeling approach.

Based on the data summary provided by Leonard F. Konikow (personal communication, 1977), the irrigated system along the Arkansas River near La Junta may be depicted as a series of observed and estimated inputs and outputs as shown in Figure 3.1 below. One major difference between the model representation here and that given previously in Chapter 2 is the water source of q_{et} ; the reason for this depicted difference will become obvious later. In the figure, q_s represents the volumetric surface water application rate per unit surface area and has a concentration level c_0 ; q_l represents the volumetric surface water seepage per unit surface area lost to the aquifer; q_b represents the net volumetric boundary flux per unit surface area having a concentration level c which is assumed identical to the average aquifer concentration level; q_p and q_m represent the volumetric irrigation and municipal pumpage rates per unit surface area, respectively; q_{et} represents the volumetric phreatophyte evapotranspiration rate per unit surface area

which is assumed to contain no dissolved contaminants; ϵ represents the net aquifer recharge rate (that is, the total applied water minus the crop evapotranspiration) per unit surface area and has a concentration level of c_i ; and q_r represents the natural aquifer outflow rate to the river or aquifer drains per unit surface area having a concentration level of c' . If q_r is positive as depicted in the figure, then $c' = c$, the average aquifer concentration. Using the linear reservoir assumption previously introduced, q_r may be approximated by the relationship:

$$q_r = a_r (h - h_r) \quad (3.2.1)$$

which is analogous to (2.2.2). If the average aquifer water level, h , falls below the average stream reference level, h_r , then q_r will be negative and water will be transferred from the river to the aquifer with $c' = c_r$, the concentration level in the river. Note that for this situation, one would generally expect $c_r = c_0$, the concentration level of surface irrigation waters; however, in the Arkansas River Valley near La Junta, surface waters are diverted from the river above the actual irrigation site. The river water quality actually deteriorates somewhat before arriving at the study reach while concentration levels in the canal waters remain essentially unchanged. It should also be pointed out that during the study calibration period, there was sufficient water available in the river to maintain the occasional "reverse flow" situation predicted in (3.2.1); that is, for those months when $h < h_r$.

Again it is implied in Figure 3.1 that the aquifer is a well-mixed reservoir, and that any outflows (other than ET) leaving the aquifer will carry this average concentration level. It should also be noted that the valley drainage term q_0 in (2.2.1) has been lumped together with the

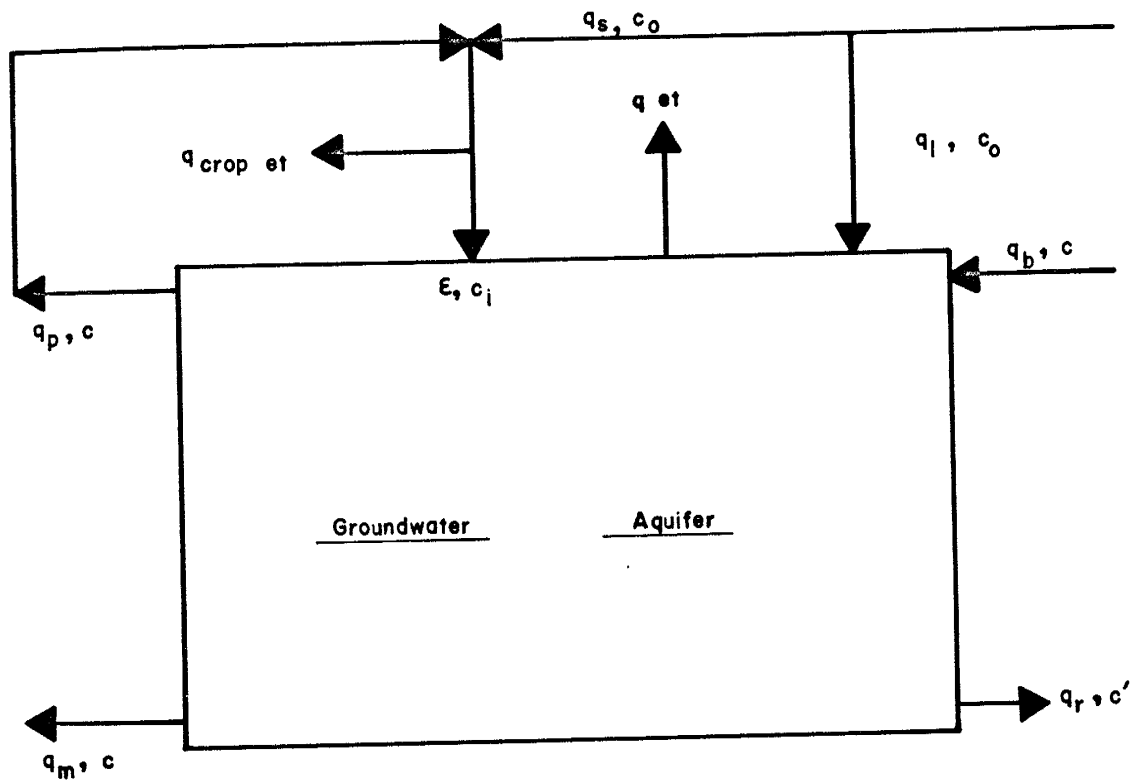


Figure 3.1. Lumped parameter representation of the irrigated stream-aquifer system near La Junta, Colorado.

river-aquifer leakage term, q_r , since observed drain flow data for the study period were not available.

If S is the average storage coefficient (specific yield) of the aquifer, the water balance equation for Figure 3.1 may be written as

$$S \frac{dh}{dt} = \epsilon + q_1 + q_b - (q_p + q_m + q_{et}) - q_r$$

which may be simplified to

$$\frac{dh}{dt} + (a_r/S)h = \frac{\epsilon_{net} + a_r h_r}{S} \quad (3.2.2)$$

where q_r is given by (3.2.1) and where $\epsilon_{net} = \epsilon + q_1 + q_b - (q_p + q_m + q_{et})$.

In the data provided by Konikow (personal communication, 1977), the terms in parentheses were not reported individually but rather were all combined. Hence in Figure 3.1, the q_{et} term is depicted as having an aquifer source. This representation will not affect the mass balance equation because the total dissolved solids content of the q_{et} waters was assumed to be zero. The general solution to (3.2.2) for a pulsed net recharge, ϵ_{net} , may be easily found by utilizing an integration factor of $\exp(-a_r t/S)$. Thus

$$h = (h_r + \frac{\epsilon_{net}}{a_r}) + (h_0 - h_r - \frac{\epsilon_{net}}{a_r}) \exp(-a_r t/S)$$

where h_0 is the average aquifer water level at time $t = 0$. This equation may be discretized over any time interval Δt , as shown in Figure 3.2, by considering the pulsed net recharge ($\epsilon_{net, i+1/2}$) to be

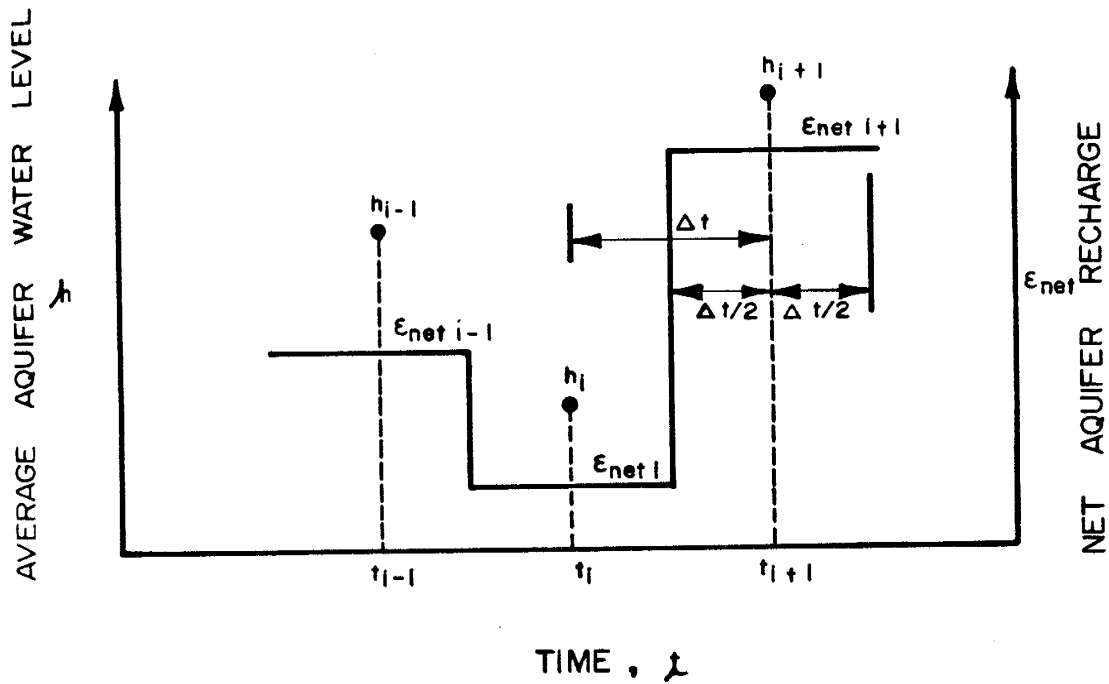


Figure 3.2 Discretized subscript notation for water balance equation.

measured in the middle of the i th and $(i+1)$ time steps. Thus we may write the relationships

$$h_{i+1/2} = \left(h_r + \frac{\epsilon_{\text{net}_i}}{a_r} \right) + \left(h_i - h_r - \frac{\epsilon_{\text{net}_i}}{a_r} \right) \exp(-\Delta t/2t_h)$$

$$h_{i+1} = \left(h_r + \frac{\epsilon_{\text{net}_{i+1}}}{a_r} \right) + \left(h_{i+1/2} - h_r - \frac{\epsilon_{\text{net}_{i+1}}}{a_r} \right) \exp(-\Delta t/2t_h)$$

where $t_h = S/a_r$, the hydraulic response time of (3.2.2). Combining these last two equations yields

$$\begin{aligned} h_{i+1} = & \left(h_r + \frac{\epsilon_{\text{net}_{i+1}}}{a_r} \right) + \left(h_i - h_r - \frac{\epsilon_{\text{net}_i}}{a_r} \right) \exp(-\Delta t/2t_h) \\ & + \frac{1}{a_r} (\epsilon_{\text{net}_i} - \epsilon_{\text{net}_{i+1}}) \exp(-\Delta t/2t_h) \end{aligned} \quad (3.2.3)$$

A mass balance relationship for Figure 3.1 may also be written. Thus

$$n \frac{d(hc)}{dt} = \epsilon c_i - q_{\text{et}} c_{\text{et}} + q_1 c_0 + q_b c - q_p c - q_m c - q_r c' \quad (3.2.4)$$

where $c_{\text{et}} \approx 0$. Combining (3.2.2) and (3.2.4), and assuming that $S = n$, yields

$$Sh \frac{dc}{dt} + c(\epsilon + q - q_p - q_{\text{et}} - q_r) = c_0 (q_s + q_1) - q_r c' \quad (3.2.5)$$

where $\epsilon c_i = q_s c_0 + q_p c$. If $q_r > 0$ (i.e., $h > h_r$), then $c' = c$,

and (3.2.5) becomes

$$Sh \frac{dc}{dt} + (\epsilon_{TOT})c = (q_{TOT})c_o \quad (3.2.6)$$

where $\epsilon_{TOT} = \epsilon + q_1 - q_p - q_{et}$ and $q_{TOT} = q_s + q_1$. If $q_r < 0$ (i.e., $h < h_r$), then $c' = c_r$, and (3.2.5) becomes

$$Sh \frac{dc}{dt} + (\epsilon'_{TOT})c = (q'_{TOT})c'_o \quad (3.2.7)$$

where

$$\epsilon'_{TOT} = \epsilon + q_1 - q_p - q_{et} - q_r = \epsilon_{TOT} - q_r$$

$$q'_{TOT} = q_s + q_1 - q_r = q_{TOT} - q_r$$

$$c'_o = [c_o(q_s + q_1) - q_r c_r] / q'_{TOT}$$

A general analytical solution to (3.2.5) for a pulsed recharge of constant concentration is not readily available when h is not constant. However, a numerical approximation may be obtained via an explicit discretization. Hence, for (3.2.6) and from Figure 3.3, we obtain the backward difference relationship given by

$$\begin{aligned} \left[\frac{c_i - c_{i-1}}{\Delta t} \right] + c_i \left[\frac{\epsilon_{TOT_i} + \epsilon_{TOT_{i-1}}}{2} \right] / \left[\frac{S(h_i + h_{i-1})}{2} \right] \\ = c_o \left[\frac{q_{TOT_i} + q_{TOT_{i-1}}}{2} \right] / \left[\frac{S(h_i + h_{i-1})}{2} \right] \end{aligned}$$

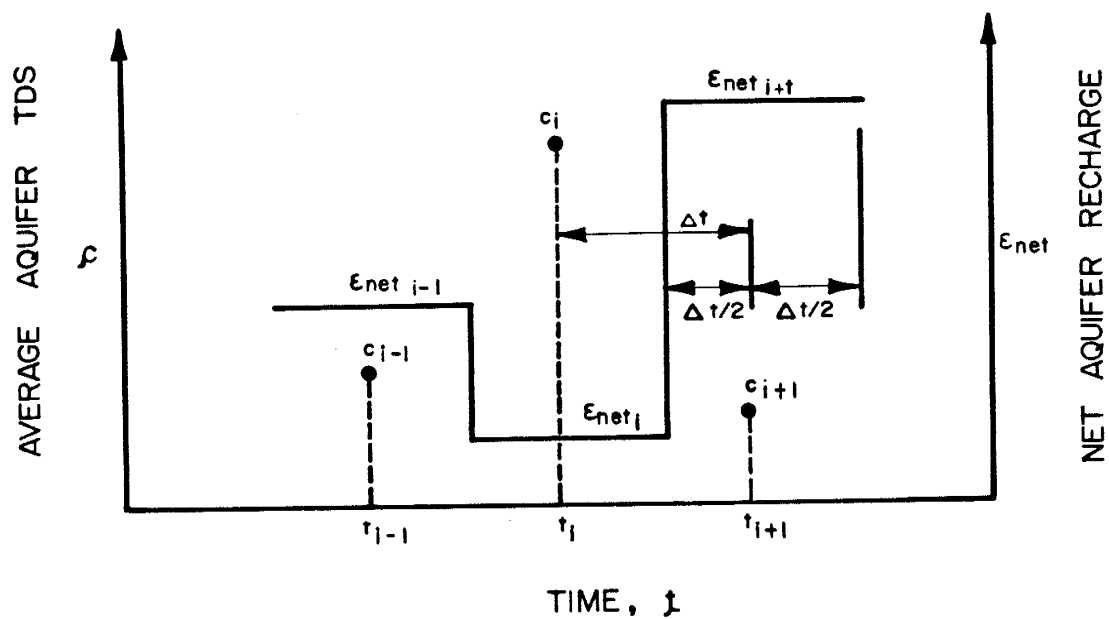


Figure 3.3 Discretized subscript notation for mass balance equation.

This expression may be rearranged to yield

$$c_i = [c_{i-1} + \frac{\Delta t (q_{TOT_i} + q_{TOT_{i-1}}) c_{o_i}}{S(h_i + h_{i-1})}] / [1 + \frac{\Delta t (\epsilon_{TOT_i} + \epsilon_{TOT_{i-1}})}{S(h_i + h_{i-1})}]$$

A similar expression for (3.2.7) may also be written.

Equations (3.2.1), (3.2.2), (3.2.6), and (3.2.7) form a coupled linear description of convective water and mass transport through the irrigated stream connected aquifer system along the Arkansas River Valley near La Junta, Colorado. Although the lumped modeling technique previously given in Chapter 2 would probably form an adequate system-wide temporal description if observed drain flow data were available, the model alterations given here serve to demonstrate the adaptability of this general approach. In addition, these alterations allow the direct comparison of the lumped approach to a physically more complex solution presented by Konikow and Bredehoeft, since their spatially averaged input data may be used directly. These data are summarized in Appendix D. The computer code for the lumped parameter simulations of the Arkansas River Valley is listed in Appendix C.

3.3 Parameter Estimation Procedure

In the proceeding analyses (3.2.2) will yield a general solution for the predicted average water table elevation above some reference datum for the lumped parameter linear reservoir model; (3.2.5) will similarly yield a corresponding aquifer outflow concentration. The lumped model is therefore dynamic when these equations are coupled together; hence it is

a truly predictive management tool. Thus if future input stresses can be estimated, then the corresponding system outputs can be found. However, certain physical parameters which characterize the system response must first be estimated. Updegraff and Gelhar (1978) proceeded to estimate a_r and h_r in (3.2.1) for a linear reservoir model using a least squares regression technique for observed average water level and drain flow data in the Mesilla Valley of southcentral New Mexico. Once these parameters were found, they discretized an equation similar to (3.2.2), except that a net aquifer recharge term (ϵ_{net}) and a constant river leakage term (q_1) were separately identified. They obtained a least squares estimate of n and q_1 for periods of no irrigation (i.e., no net recharge). Their equation could then be resolved for net recharge using the above parameters and observed values of average water level elevations for periods of active irrigation assuming the river leakage remained constant for a given recharge-recession period. Their procedure worked extremely well in the Mesilla Valley for the five year period 1946-1951. However, in the Arkansas River Valley study, no drain flow data is currently available; thus their technique does not apply here. Instead, an alternate but similar procedure was employed since the data of Konikow and Bredehoeft includes water level elevations from 40 observation wells (see Appendix D) and their calculated net aquifer recharge. Furthermore, they have estimated the effective porosity and storage coefficient as $n = S = 0.20$.

Rewriting (3.2.2) in integral form and using a Simpson integration approximation yields

$$n (h_{i+2} - h_i) + \frac{a_r \Delta t}{3} (h_i + 4h_{i+1} + h_{i+2}) - 2a_r h_r \Delta t \quad (3.3.1)$$

$$= \frac{\Delta t}{3} (\epsilon_{\text{net}_i} + 4 \epsilon_{\text{net}_{i+1}} + \epsilon_{\text{net}_{i+2}})$$

where i represents a given time step and Δt the time interval. In this equation, h and ϵ_{net} are known over twelve time intervals for $\Delta t = 1$ month. Thus we may represent the best estimate of (3.3.1) as

$$Y = \beta_0 + \beta_1 X_1 + \beta_2 X_2 + E \quad (3.3.2)$$

where

$$Y = \Delta t/3 (\epsilon_{\text{net}_i} + 4\epsilon_{\text{net}_{i+1}} + \epsilon_{\text{net}_{i+2}})$$

$$\beta_0 = -2a_r h_r \Delta t,$$

$$\beta_1 = n$$

$$\beta_2 = a_r$$

$$X_1 = (h_{i+2} - h_i)$$

$$X_2 = \Delta t/3 (h_i + 4h_{i+1} + h_{i+2})$$

and where E represents the residual between the observed value of the dependent variable Y and its "true" value, and where $\epsilon_{\text{net}} = \epsilon + q_1 + q_b - (q_p + q_m + q_{\text{et}})$, from (3.2.2). Equation (3.3.2) may be viewed as a parametric equation describing the "true" value of some dependent variable Y , while (3.3.1) describes the corresponding observed value. Since ϵ_{net} and h are known in (3.3.1), the parameters β_0 , β_1 , and β_2 may be estimated via a least squares regression so as to minimize

the difference between the two equations; that is, to minimize E. This procedure is well known (Draper and Smith, 1966, for example) and will not be fully described here. However, the procedure basically involves solving the simultaneous equations given below for b_0 , b_1 , and b_2 .

$$\sum_{i=1}^m Y_i = b_0 m + b_1 \sum_{i=1}^m X_{1_i} + b_2 \sum_{i=1}^m X_{2_i} \quad (3.3.3)$$

$$\sum_{i=1}^m X_{1_i} Y_i = b_0 \sum_{i=1}^m X_{1_i} + b_1 \sum_{i=1}^m X_{1_i}^2 + b_2 \sum_{i=1}^m X_{1_i} X_{2_i}$$

$$\sum_{i=1}^m X_{2_i} Y_i = b_0 \sum_{i=1}^m X_{2_i} + b_2 \sum_{i=1}^m X_{1_i} X_{2_i} + b_2 \sum_{i=1}^m X_{2_i}^2$$

In these equations b_0 , b_1 , and b_2 represent the best estimates for β_0 , β_1 , β_2 , respectively. Since the Konikow-Bredehoeft data has 12 monthly water level and recharge observations and the Simpson approximation implies a moving three point averaging process, $m = 10$ in the above relationships. For a complete data description see Appendix D. These "observations" used in (3.3.3) are actually the mean water level values as found using a Thiessen polygon weighting technique for the 40 observation wells and the average valley recharge (ϵ) estimated by Konikow and Bredehoeft. This information together with (3.3.2) and (3.3.3) yields

$$\begin{aligned} S &= 0.098 \\ a_r &= 0.0791/\text{month} \\ h_r &= 4011.13 \text{ feet} \\ t_h &= 1.24 \text{ months} \end{aligned}$$

Since the S found above is less than half that used by Konikow and Bredehoeft, it was decided to let S equal their value (i.e., $S = 0.20$), and resolve (3.3.3) to obtain

$$\begin{aligned} S &= 0.20 \\ a_r &= 0.0720/\text{month} \\ h_r &= 4011.16 \text{ feet} \\ t_h &= 2.78 \text{ months} \end{aligned}$$

The differences in the two techniques probably reflect some errors in calculating the net aquifer recharge; furthermore, some substantial differences were found between the Thiessen weighted average water levels and the spatially averaged values obtained from the Konikow-Bredehoeft model. While some variation is expected, the differences encountered (see Table 3.1 or Figure 3.7) were rather large and may have resulted from the location of the observation wells. That is, if too many of these wells are located near drains, pumping wells, or conveyance canals, they may not show a representative water level. It is interesting to note that if the Konikow-Bredehoeft average water levels were used in (3.3.3) above instead of the Thiessen values, one would find that

$$\begin{aligned} S &= 0.233 \\ a_r &= 0.0204/\text{month} \\ h_r &= 4013.67 \text{ feet} \\ t_h &= 11.42 \text{ months} \end{aligned}$$

Since $S = 0.233$ as compared to 0.20, one might expect this third parameter set to be the most representative of the data; however, the large aquifer response time leads one to suspect that the parameter a_r is too small. Whatever the shortcomings of the data or parameter estimation procedure may be, the second set of parameters was chosen for all subsequent simulations. The data used in the above parameter estimation procedures are briefly summarized in Table 3.1

Table 3.1: Data summary for the parameter estimation.

	<u>Konikow-Bredehoeft ave aq. water level</u> (1)	<u>Thiessen aq. water level</u> (2)	<u>Net Aq. Recharge</u> (3)
March 1971	4013.74 feet	4011.80 feet	0.011 ft/mo
April	4013.60	4011.24	-0.013
May	4013.64	4011.13	-0.016
June	4013.74	4010.57	-0.071
July	4013.76	4010.82	-0.058
August	4013.45	4010.50	-0.061
September	4013.35	4010.57	-0.036
October	4013.46	4010.64	0.016
November	4013.64	4011.21	0.067
December	4013.66	4011.61	0.046
January 1972	4013.68	4011.70	0.043
February	4013.68	4011.65	0.036

(1) The arithmetically averaged water levels as predicted from the uniformly spaced, block-centered, rectangular finite difference grid used by Konikow and Bredehoeft (1974a), and consisting of 20 rows and 44 columns.

(2) Thiessen averaged water levels based on forty observation wells located within the study reach, as depicted in Appendix D.

(3) The lumped net recharge, ϵ_{net} , as defined in (3.2.2); a negative sign indicates a net valley withdrawal (from Konikow and Bredehoeft, 1974a, and listed in Appendix D).

3.4. Comparative Study Results

Figures 3.4 through 3.8 show a comparison between the spatially averaged data for the 40 observation wells obtained from data provided by Leonard F. Konikow (personal communication, 1977), the Konikow-Bredehoeft spatially averaged model output, and that obtained here for the study period March 1971 through February 1972 using the second parameter set. This calibration period had sufficient input data available to enable a model simulation of the river valley. It can be seen in these figures that both the Konikow-Bredehoeft spatial model and the lumped temporal model reproduce the average observed data in a similar fashion, with the exception of the changes in average aquifer water levels and the aquifer total dissolved solids concentration, in which case the Konikow and Bredehoeft results seem to predict somewhat smaller seasonal changes than that observed or simulated by the lumped parameter model. These later differences may be a result of the difference in spatial averaging (Thiessen polygons for the lumped parameter model versus grid averaging in the Konikow and Bredehoeft model; see Table 3.1).

As was done by Konikow and Bredehoeft with their simulations, the lumped model was used to predict any long-term changes that might occur within the stream-aquifer system. This was accomplished by simulating a five year period, assuming that all system averaged stresses that occurred during the calibration period (March 1971 to February 1972) would also occur in each of the succeeding four years. These simulation results are depicted in Figures 3.9 and 3.10. Extending the simulation period on the basis of the one year observed irrigation practices served three purposes: (1) to further demonstrate the lumped model's

OBSERVED DATA (+)
 KONIKOW & BREDEHOEFT (X)
 THIS STUDY (Δ)

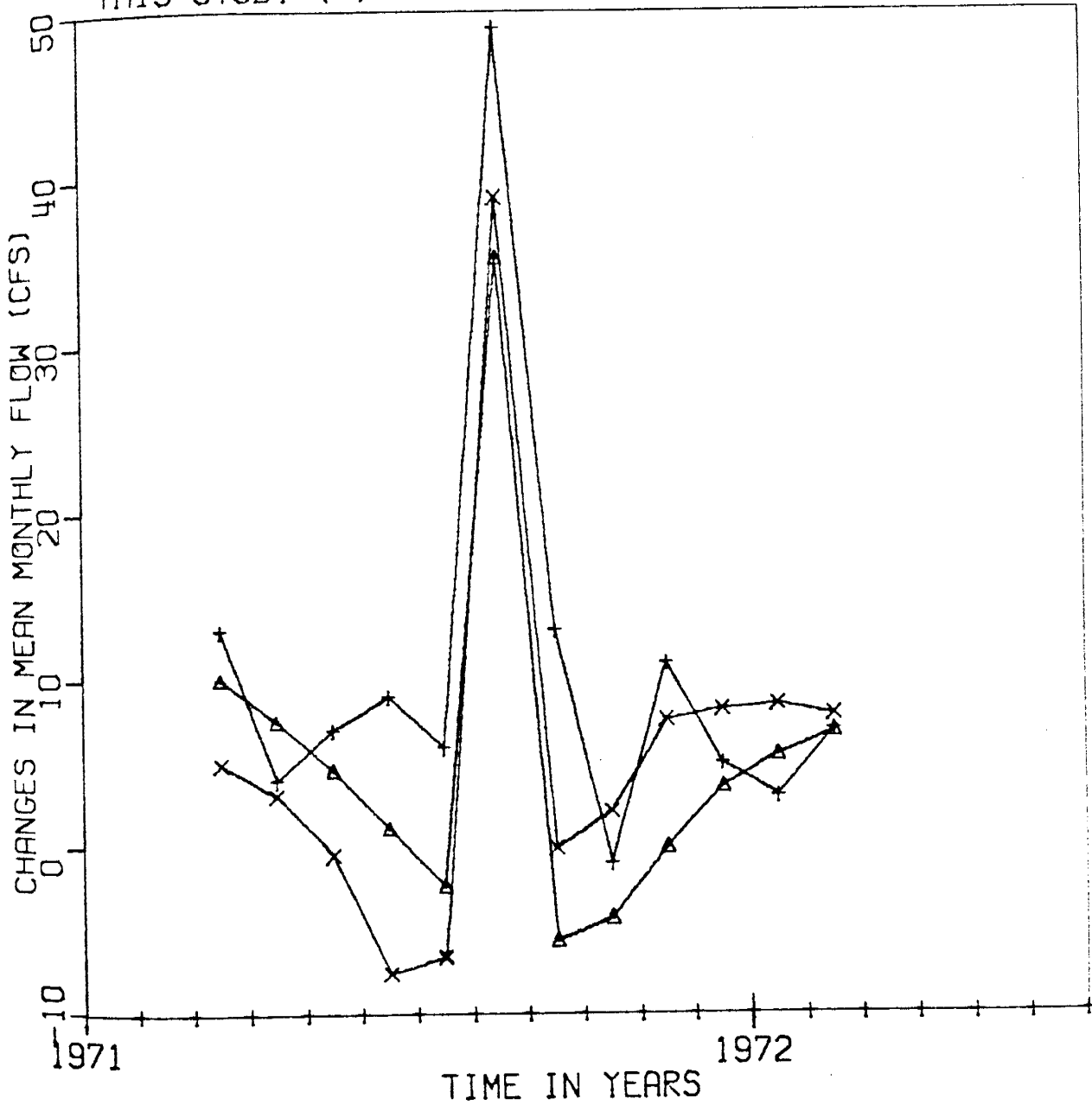


Figure 3.4 Changes in mean monthly river flow over the study reach.

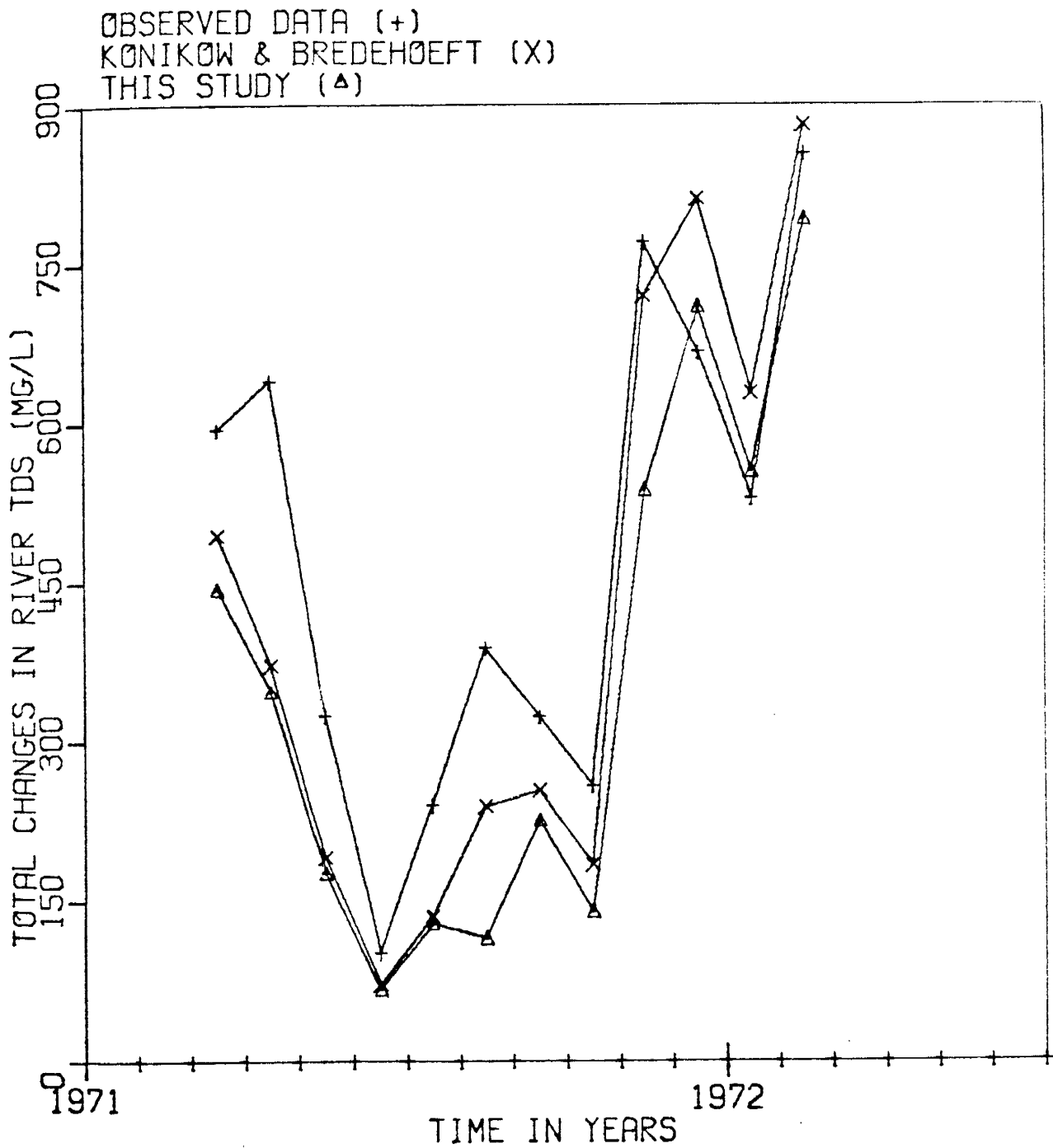


Figure 3.5 Total changes in river TDS between the Ft. Lyon canal diversion and the Bent-Otero County line.

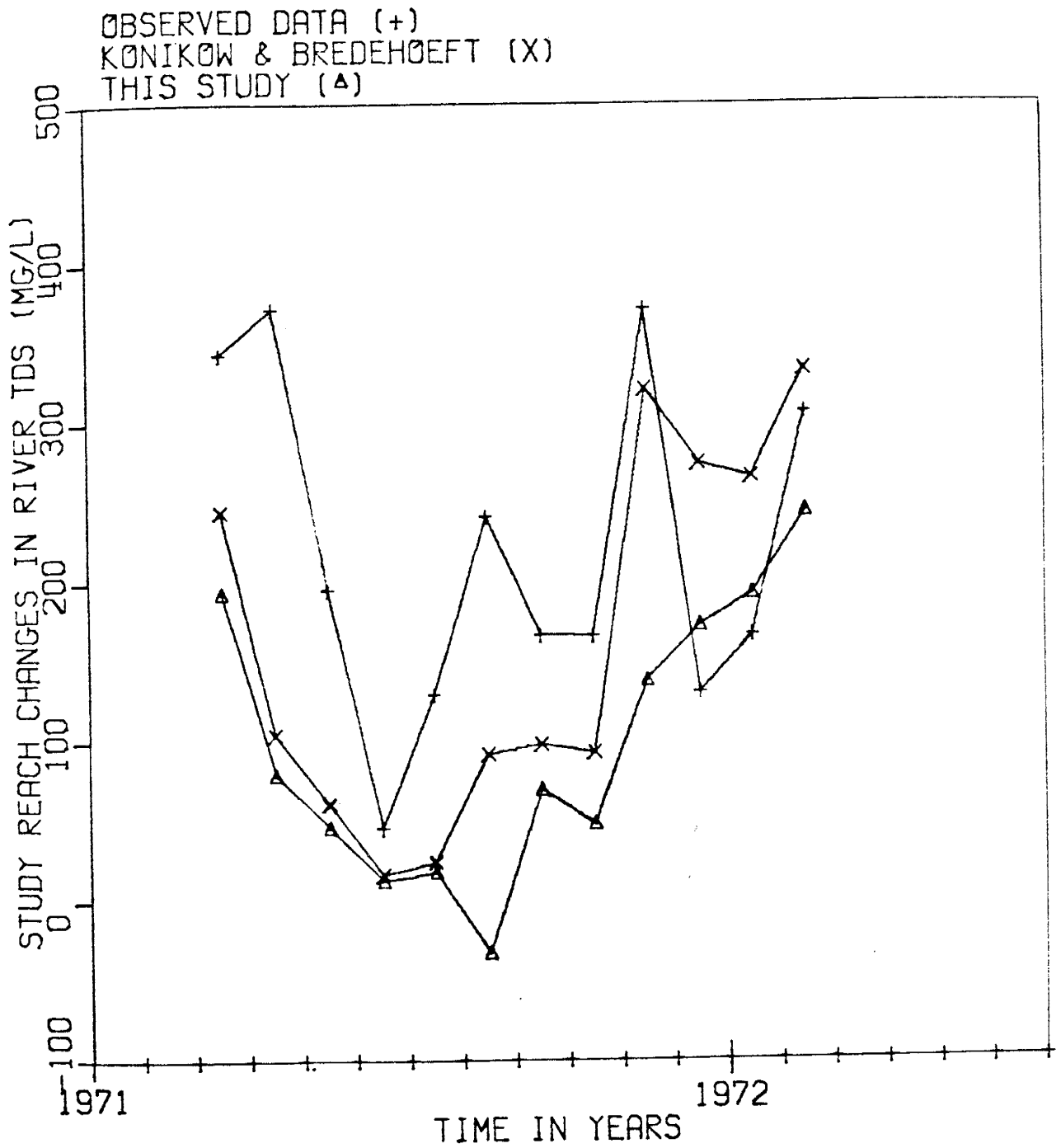


Figure 3.6 Study reach changes in river TDS.

Observed data (+)

Konikow & Bredehoeft (x)

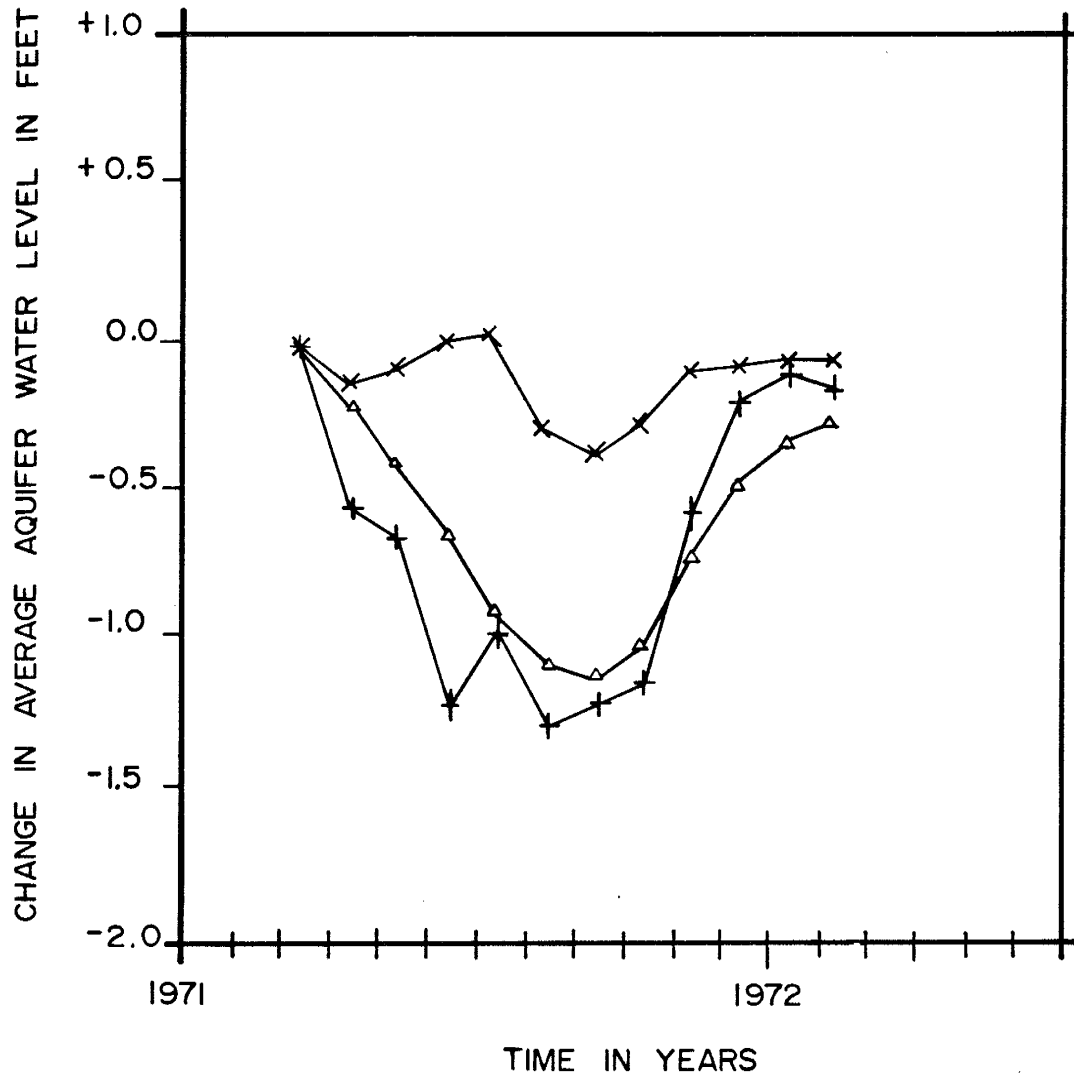
This study (Δ)

Figure 3.7 Changes in average aquifer water levels in the Arkansas River Valley study area.

Observed data (+)

Konikow & Bredehoeft (X)

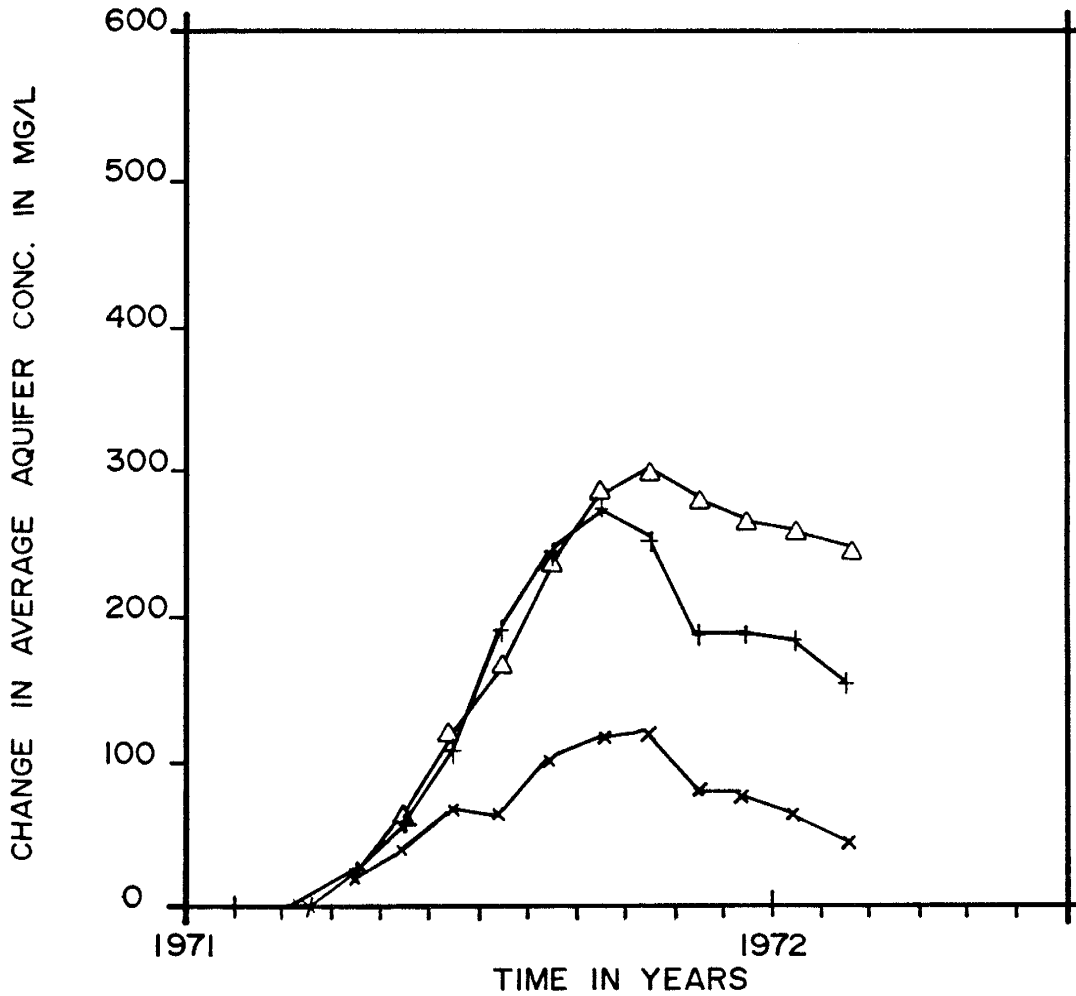
This study (Δ)

Figure 3.8 Changes in average aquifer TDS concentration in the study reach.

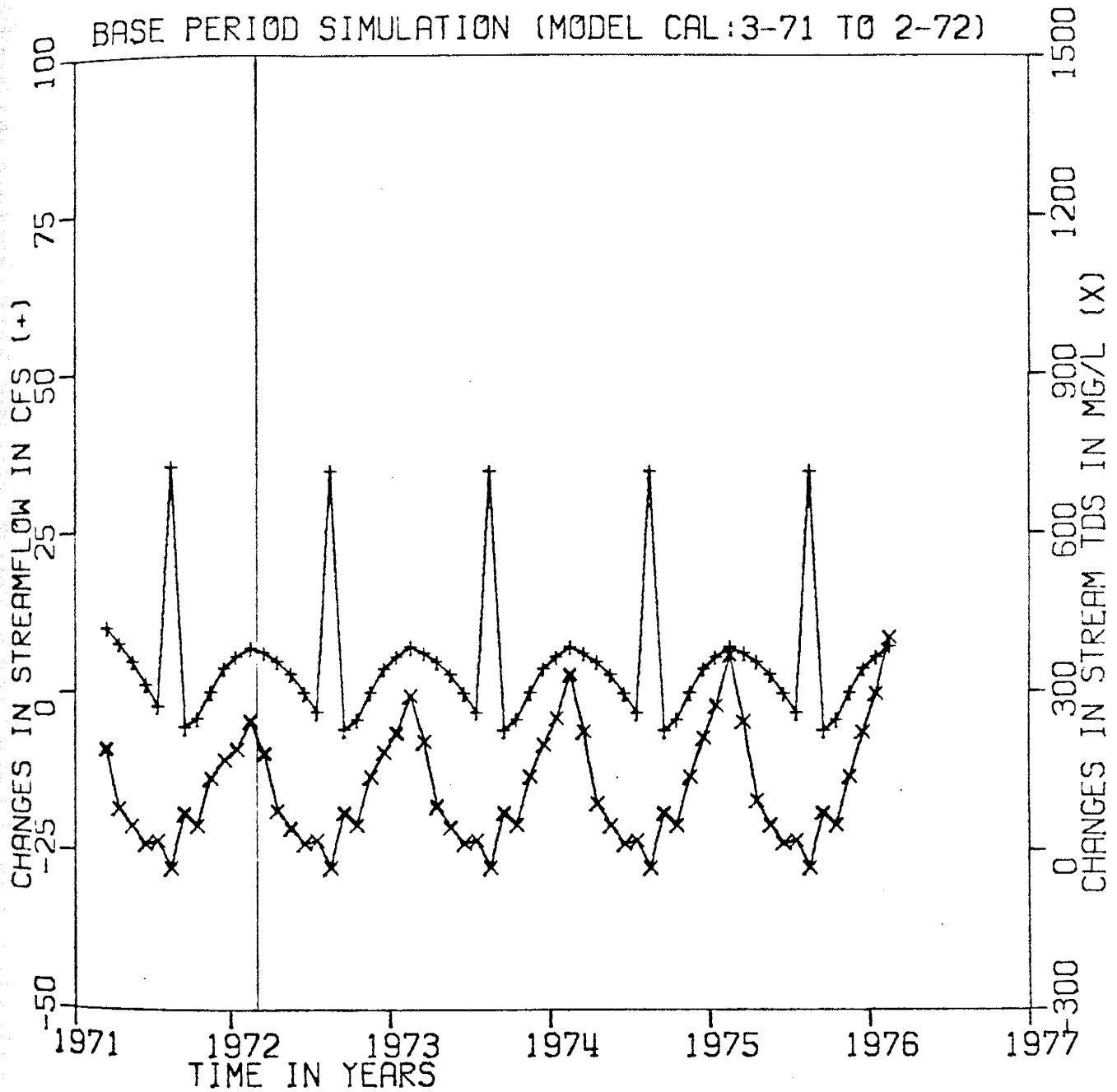


Figure 3.9 Study reach changes in streamflow and surface water quality as a result of extending the lumped model simulation period an additional four years

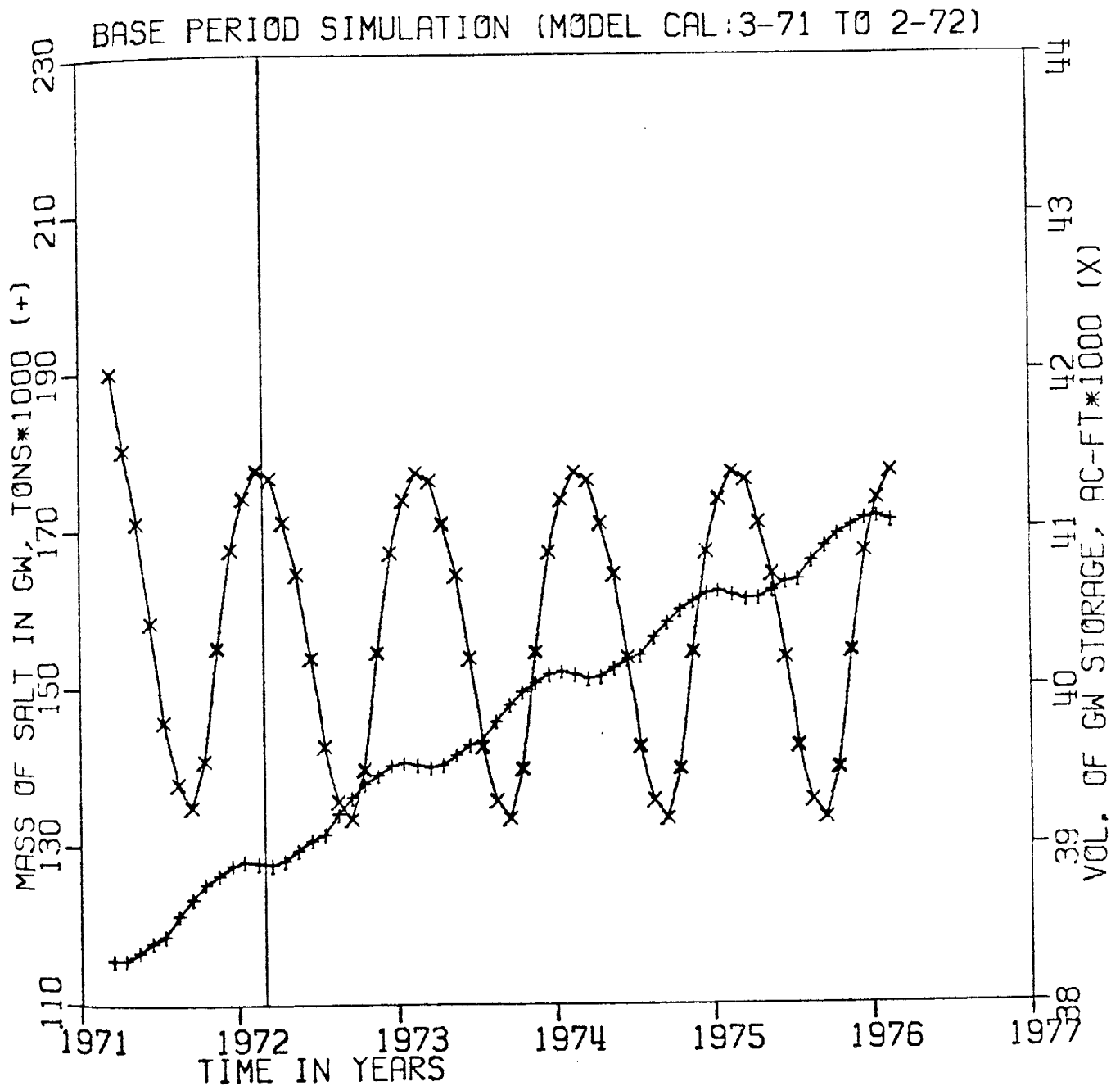


Figure 3.10 Study reach changes in groundwater storage and water quality as a result of extending the lumped model simulation period an additional four years.

capabilities, (2) to provide a basis for comparing the results of other water management options, and (3) to provide an additional basis of comparison with the Konikow-Bredehoeft model simulations.

The results of this five-year simulation are depicted in Figures 3.9 and 3.10; these data are summarized in Table 3.2 along with that from the Konikow-Bredehoeft model. Two hydrological parameters, the volume of groundwater in storage (Figure 3.10) and net stream gains or losses (Figure 3.9), show marked seasonal variations comparable to the Konikow-Bredehoeft model (see Figure 3.9, Konikow and Bredehoeft, 1974b, p. 50). Even though the aquifer parameter estimation procedure used here may only be a crude approximation of actual conditions, the model output still favorably compares to that of Konikow and Bredehoeft.

3.5 Water Management Options with the Linear Model

Following Konikow and Bredehoeft (1974b) the lumped model was evaluated for its ability to predict the long term response of the stream-aquifer system within the study reach associated with changes in irrigation practices or water management options. Because of the uncertainty associated with long term historical precipitation, groundwater withdrawals, surface water irrigation, and the ratio of groundwater to surface water applications, it was decided to present the simulation results on a relative rather than an absolute response; a similar presentation was used by Konikow and Bredehoeft (1974b). However, Appendix D lists both absolute and relative responses to the base period simulation. The four specific water management options are

Table 3.2: Summary comparison between the five year lumped output and the Konikow-Bredehoeft model output

	<u>Lumped Model (1)</u>	<u>Konikow-Bredehoeft Model</u>	<u>Observed(2)</u>
1. Change in groundwater storage in 1000's acre-feet per month:			
minimum	39.2	40.5	
maximum	42.0	42.0	
2. Study reach changes in streamflow in cubic feet per second:			
minimum	-6.3	-7.3	-1.0
maximum	+35.2	+38.9	+49.0
3. Changes in mass of salt stored in groundwater in 1000's tons:			
minimum	116.2	116.2	
maximum	171.2	128.8	
4. Total changes in stream concentration level in milligrams per liter:			
minimum	68.6	72.7	102.2
maximum	793.0	882.0	854.7
5. Study reach changes in stream concentration level in mg/l:			
minimum	-32.4	+17.0	+46.4
maximum	+244.9	+333.9	+371.6

(1) Using $S = n = 0.20$, $a_r = 0.0720/\text{month}$, $h_r = 4011.16$ feet. For this case study, $t_h = 2.8$ months, and $t_c = 10.4$ years.

(2) Based on observed data over the period March 1971 to February 1972.

identical to those of Konikow and Bredehoeft (1974b); these are: (1) increased groundwater use, (2) increased surface water use, (3) improved irrigation efficiency, and (4) lining the Fort Lyon Canal to prevent seepage losses.

The first case addressed the following questions: What would the hydrological and chemical effects on the stream-aquifer system be if surface water diversions were not available for irrigation use within the study reach? As was pointed out by Konikow and Bredehoeft, this represents an extreme situation that might occur during a very low flow period in the Arkansas River. As was done by them, it was assumed that both irrigated acreage and total applied water would not change, and that a sufficient number of wells existed to supply the irrigation demand entirely from groundwater pumpage. The results of this test as shown in Figures 3.11 and 3.12, and indicated the following:

1. The water table elevation throughout the study reach near the end of the growing season (September) would decrease by an average 0.8 foot. Konikow and Bredehoeft also reported an average decline of 0.8 foot.

2. A serious deterioration in groundwater quality would occur during the first year; the TDS in the aquifer throughout the study period would increase by an average of 343 mg/l (420 mg/l according to Konikow and Bredehoeft).

3. Streamflow gains resulting from increased groundwater flow directly to the river within the study reach would decrease by an average of 7.2 CFS (cubic feet per second) during the first year. Konikow and Bredehoeft also reported a decrease in flow of 7.2 CFS.

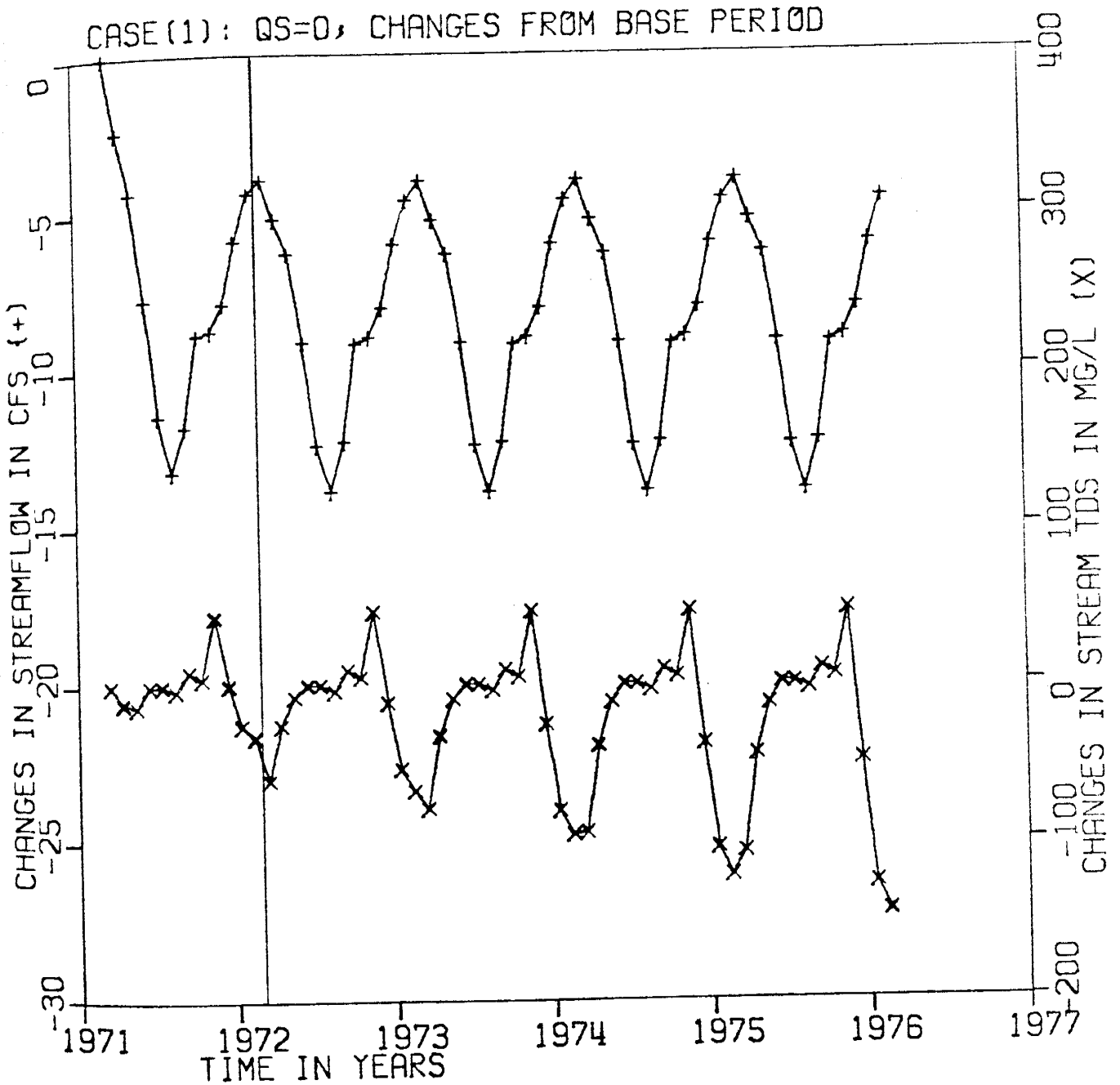


Figure 3.11 Changes in streamflow and surface water quality relative to the base simulation for no surface water diversion for irrigation.

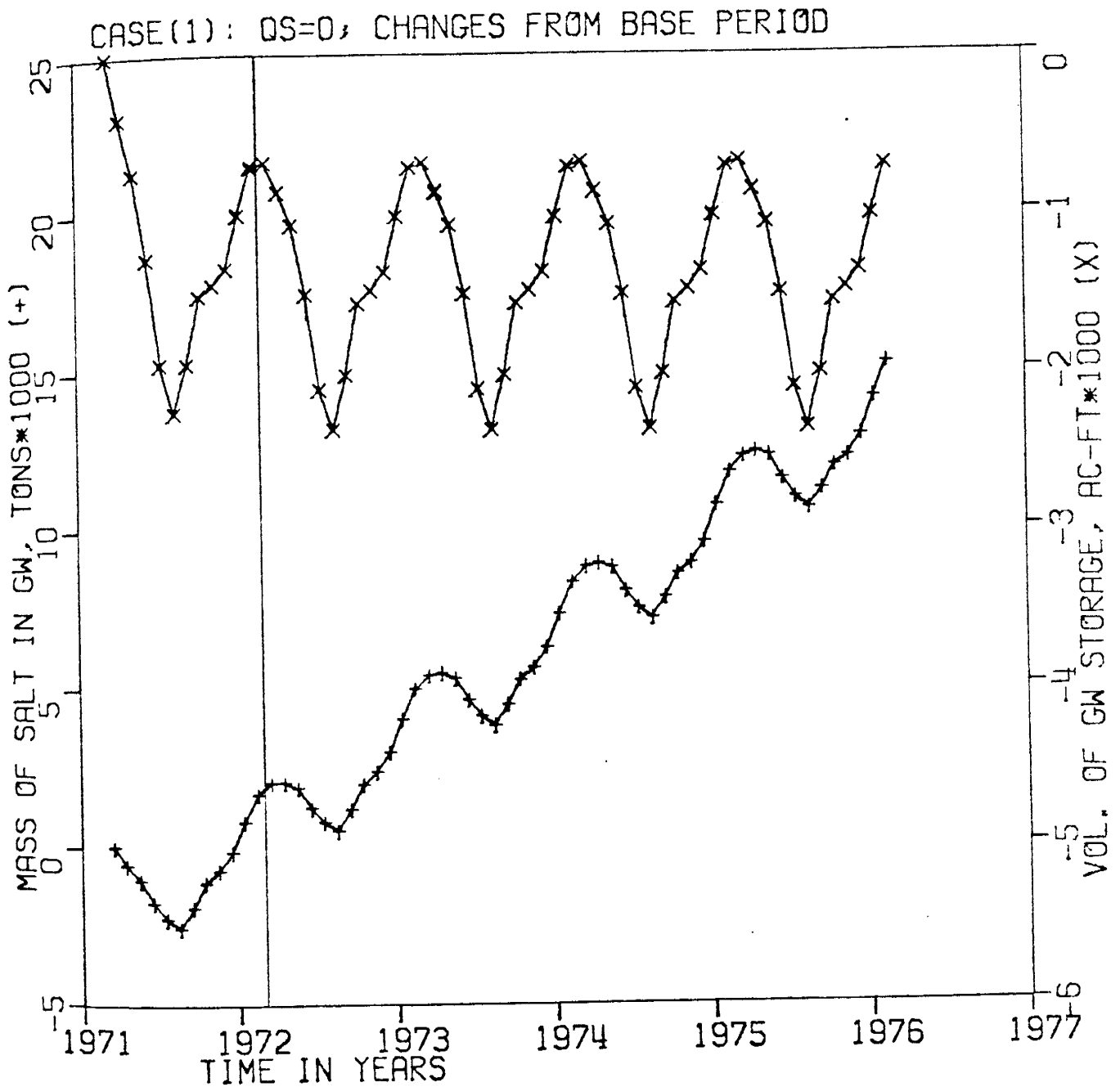


Figure 3.12 Changes in groundwater storage and water quality relative to the base simulation for no surface water diversion for irrigation.

4. The dissolved solids concentration in the Arkansas River at the Bent-Otero County line would decrease by an average of 2.1 mg/l during the first year, and show an average decrease of 35.4 mg/l after five years. Konikow and Bredehoeft reported an average decrease of 20 mg/l in their study.

As was reported in Konikow and Bredehoeft, when all of the irrigation demand is supplied by groundwater, the aquifer TDS concentration level rises dramatically. These increases are due primarily to a higher concentration level in recharge waters, but are also partially related to the decrease in groundwater storage throughout the year. Because average aquifer water levels would fall below the average stream level, the river loses water to the aquifer and a smaller salt load enters the river. This extreme situation demonstrates that an increased ratio of groundwater pumpage to surface water diversions in order to supply irrigation demand will probably result in deteriorating groundwater supplies, although the surface water quality may actually be improved somewhat if sustained pumpage results in a significant lowering of the water table near the river.

The second case also addressed a somewhat extreme situation by asking the question: What would be the effect on the stream aquifer system if all irrigation demands were supplied from surface diversions and no irrigation wells were pumped? It was again assumed that the total irrigated acreage and total applied water remained the same as the base period simulation; Figures 3.13 and 3.14 summarize these results. The results of this test indicated the following:

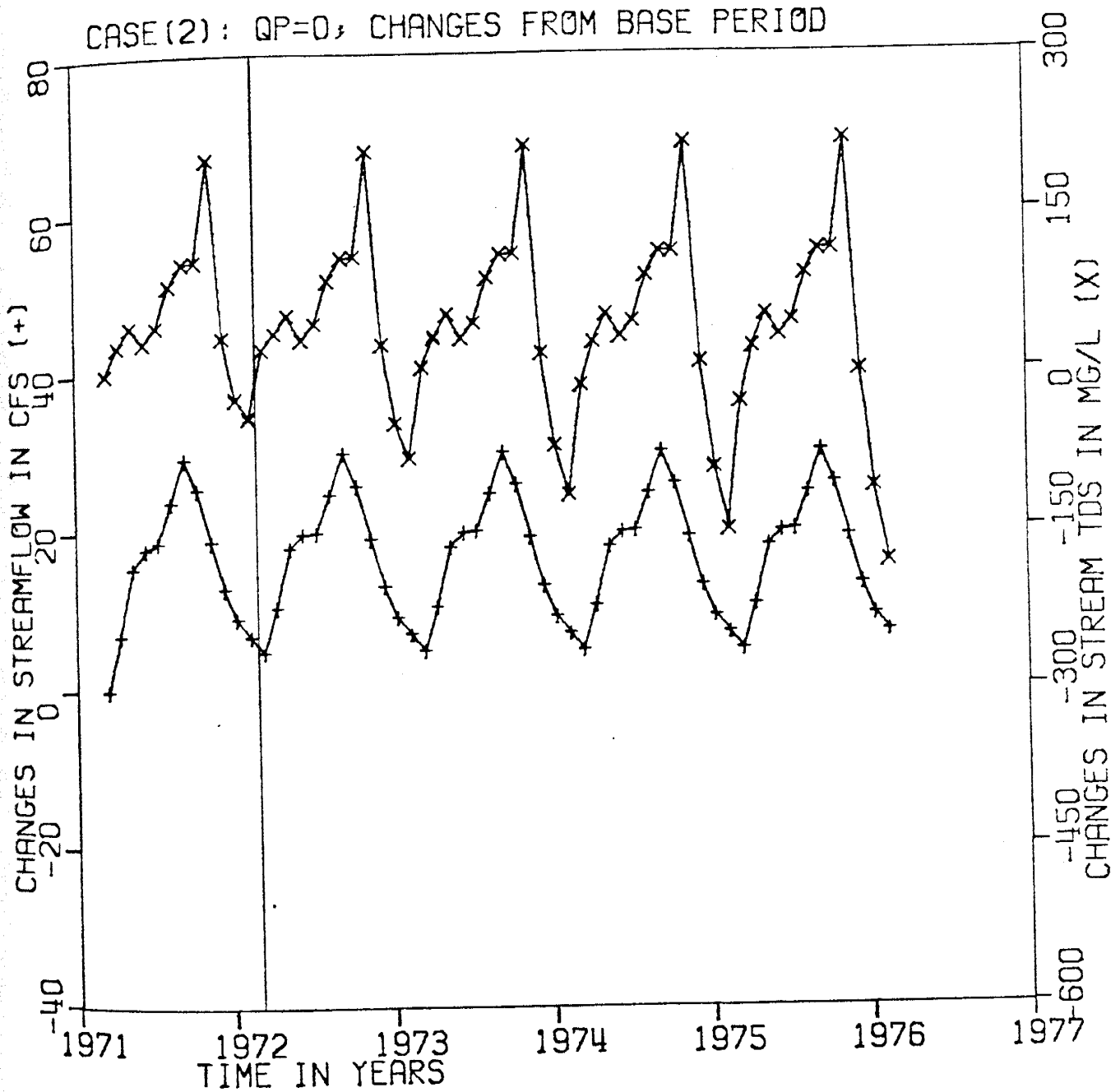


Figure 3.13 Changes in streamflow and surface water quality relative to the base simulation for no groundwater pumpage for irrigation.

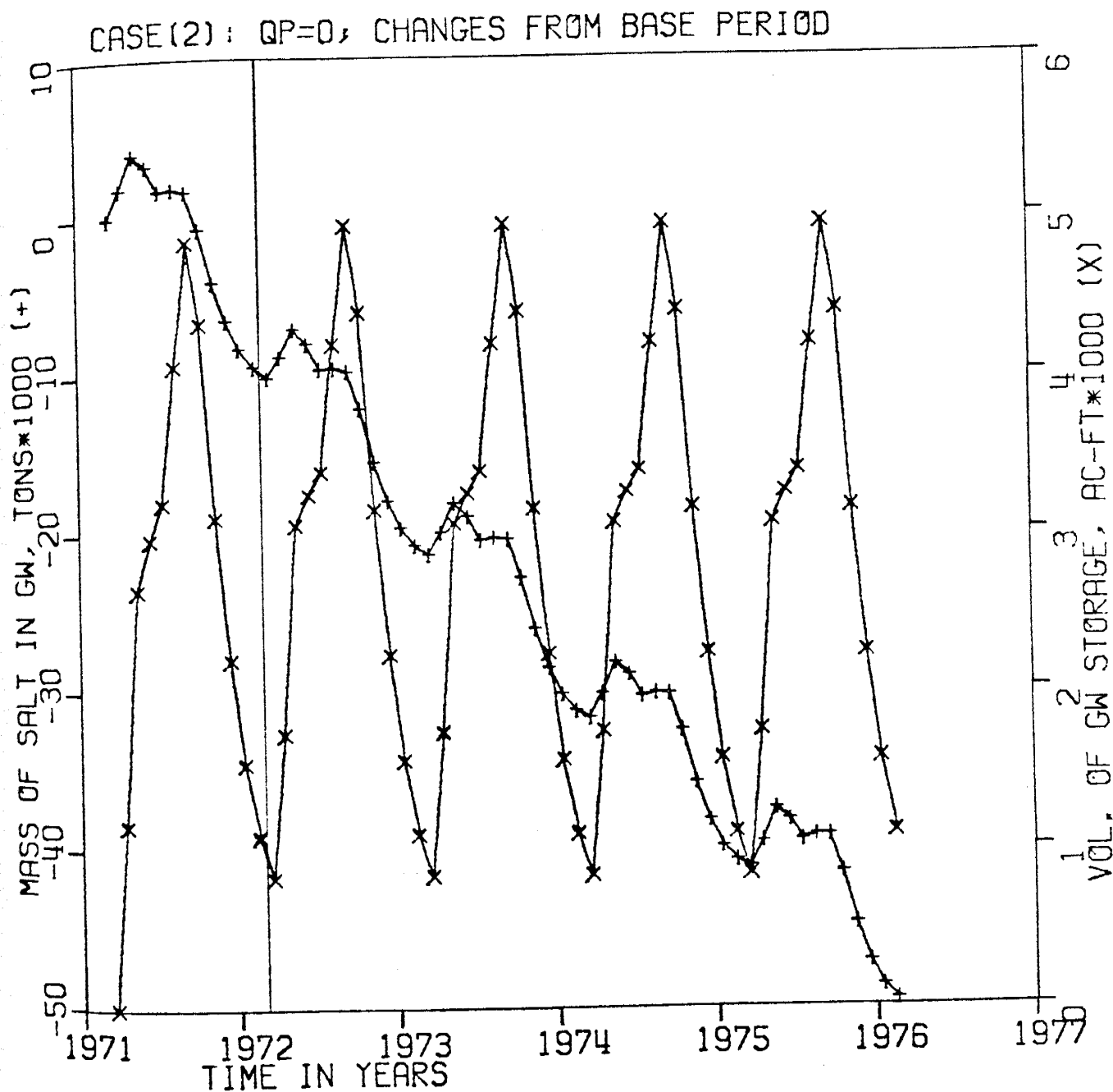


Figure 3.14 Changes in groundwater storage and water quality relative to the base simulation for no groundwater pumpage for irrigation.

1. The water table elevation throughout the study area would be increased by an average of 2.0 feet at the end of the growing season (September) as compared to an average increase of 1.6 feet reported by Konikow and Bredehoeft. However, all but about 0.4 feet of this increase had completely dissipated by the end of each simulation year (March); Konikow and Bredehoeft reported a complete dissipation in their simulation. The increase in volume of groundwater in storage would vary between zero and about 5000 acre-feet as compared to a reported increase of zero to about 4000 acre-feet in Konikow and Bredehoeft.

2. The most significant change was seen in the average decrease in the total dissolved solids concentration of groundwater, where an average decline of 156 mg/l was predicted during the first year; this decrease averaged 918 mg/l during the fifth year, and 553 mg/l over all five years. Konikow and Bredehoeft reported an average decrease of about 300 mg/l after one year throughout the study area.

3. The flow in the Arkansas River at the Bent-Otero County line would be increased by over 15.4 CFS as compared to an increase of 15.0 CFS according to the model of Konikow and Bredehoeft.

4. The dissolved solids concentration in the river at the Bent-Otero County line would increase by about 51 mg/l during the first year, but would only increase by about 30 mg/l during the fifth year. Konikow and Bredehoeft reported an average increase of over 40 mg/l during the first year, but an average decrease of more than 60 mg/l after five years.

When all of the irrigation demand is supplied from better quality surface diversions, a lower salinity level of recharge waters results; furthermore, additional dilution is provided by the greater volume of

groundwater in storage. However, as Konikow and Bredehoeft pointed out, this improvement in groundwater would be of little benefit since no water was pumped for irrigation use. The benefit of this management option would probably be only minor and apply primarily to downstream water users only. After a new dynamic equilibrium is established, the temporary flushing effect of aquifer waters would decrease as seen in Figure 3.13, and a corresponding decrease in river water quality would result. However, this effect would take considerable time according to both the lumped model, and that of Konikow and Bredehoeft.

The third test addressed the following question: What would be the hydrological effects on the stream-aquifer system as a result of improving irrigation efficiency by 20 percent (that is, 20 percent less irrigation water is required to produce the same crop yields from the same irrigated acreage)? It was assumed that both groundwater pumpage and surface diversions for irrigation were reduced by 20 percent from the base period; furthermore, it was assumed that recharge was reduced by 39.3 percent (Konikow and Bredehoeft, 1974b, p. 55). The actual or predicted effects on crop yield from this improvement in water use were not explored. The results of this test are graphically depicted in Figures 3.15 and 3.16, and indicated the following:

1. Konikow and Bredehoeft reported that the total annual water requirement for irrigation was reduced by about 3600 acre-feet, while the annual recharge to the aquifer was reduced by some 2900 acre-feet. The average fraction of net applied water that was recharged was reduced from 0.30 to 0.27. The dissolved solids concentration of the recharged waters increased by an average of about 15 percent over that of the base period. Similar changes were used in the lumped parameter model for this

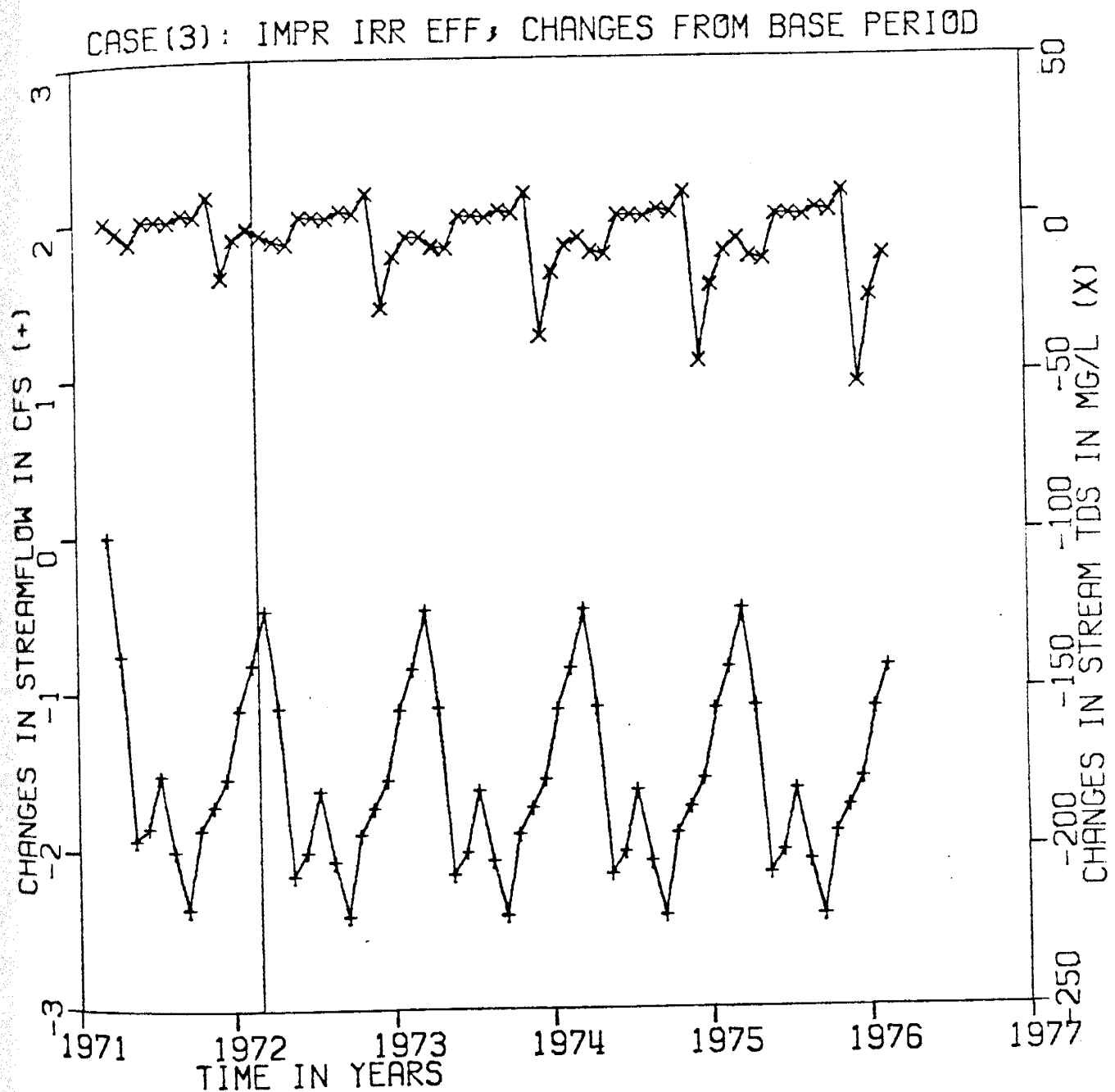


Figure 3.15 Changes in streamflow and surface water quality relative to the base simulation for a 20 percent improvement in irrigation efficiency.

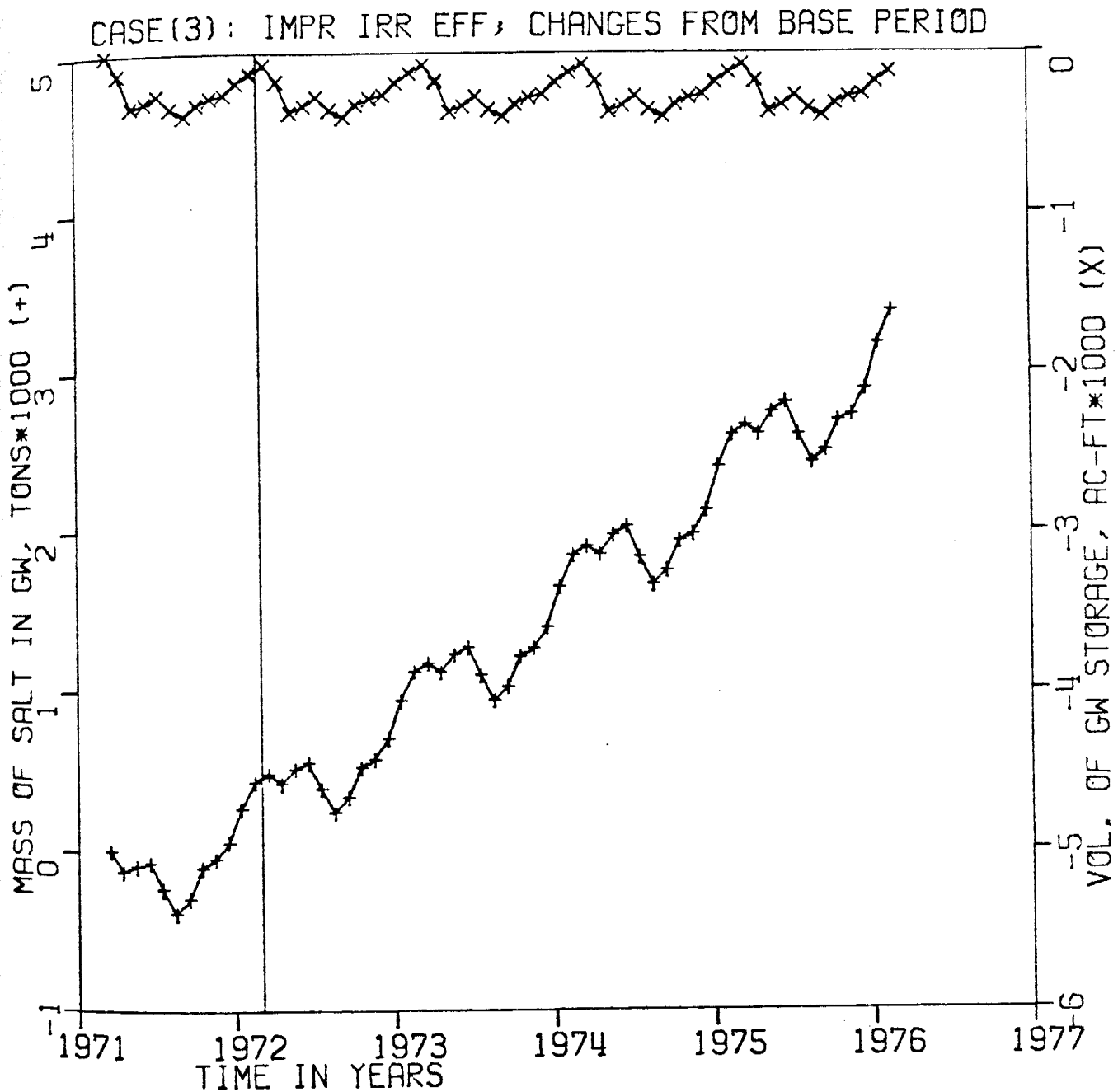


Figure 3.16 Changes in groundwater storage and water quality relative to the base simulation for a 20 percent improvement in irrigation efficiency.

case. Konikow and Bredehoeft assumed that tailwater would decrease from 15 to 7.5 percent of the applied water; however, tailwater quantities are not accounted for in the lumped model as formulated in this particular application.

2. The water table elevation throughout the study area near the end of the irrigation season averaged 0.2 foot lower than the base period, while Konikow and Bredehoeft reported an average increase of 0.1 foot.

3. The dissolved solids concentration of groundwater increased by 13 mg/l after one year, and 71 mg/l after five years. Konikow and Bredehoeft reported an average increase of about 50 mg/l after five years. The total mass of salt stored in the aquifer showed an initial decrease relative to the base period of about 0.04 kilotons after the first year, but increased during the second through fifth years, resulting in an average increase of 2.8 kilotons. Konikow and Bredehoeft's model showed a similar trend, but they reported the mass increase after the third year. Their model showed an average mass increase of about 2.5 kilotons after five years. The initial mass decrease predicted by the lumped model during the first year was apparently due to the decrease in the volume of water in storage; however, this decreased aquifer storage was eventually offset by the higher increase of salinity in recharged waters.

4. The flow of the Arkansas River at the Bent-Otero County line was decreased by an average of 1.5 CFS, while Konikow and Bredehoeft reported an average increase of less than 0.2 CFS.

5. The increase in the dissolved solids concentration in the river was less than 3 mg/l after one year, and less than 9 mg/l after five years. Konikow and Bredehoeft also reported small magnitude changes in

river concentration, but their changes more closely paralleled groundwater salinity fluctuations.

As was stated by Konikow and Bredehoeft, and verified here with the lumped model, improvements in irrigation efficiency would in general have a negative effect on groundwater quality. These effects, however, are partially offset by the benefits that might be derived from the salvaged water. Furthermore, reductions in pumping costs resulting from lower water applications would be a direct and immediate benefit to local irrigators, even though somewhat higher lifts are predicted by the lumped model.

The fourth case addressed the following question: What would be the hydrological effects on the stream-aquifer system as a result of lining the Fort Lyon Canal to prevent leakage losses? While lining canals has frequently been regarded as a management technique to improve irrigation efficiency, it is so often cited that it was decided to explore its individual effects separately. The results of this simulation are shown in Figures 3.17 and 3.18, and indicated the following:

1. The discharge of the Fort Lyon Canal at the Bent-Otero County line would be increased by an average of 12.4 CFS, which was also reported by Konikow and Bredehoeft.

2. The water table elevation throughout the study reach would decline by an average of about 0.7 foot during the first year, and by about 0.9 foot after five years. Konikow and Bredehoeft reported an average decline of about two feet after one year, and no significant additional declines after five years.

3. The total dissolved solids concentration in the aquifer increased by 67 mg/l during the first year, and 443 mg/l after five

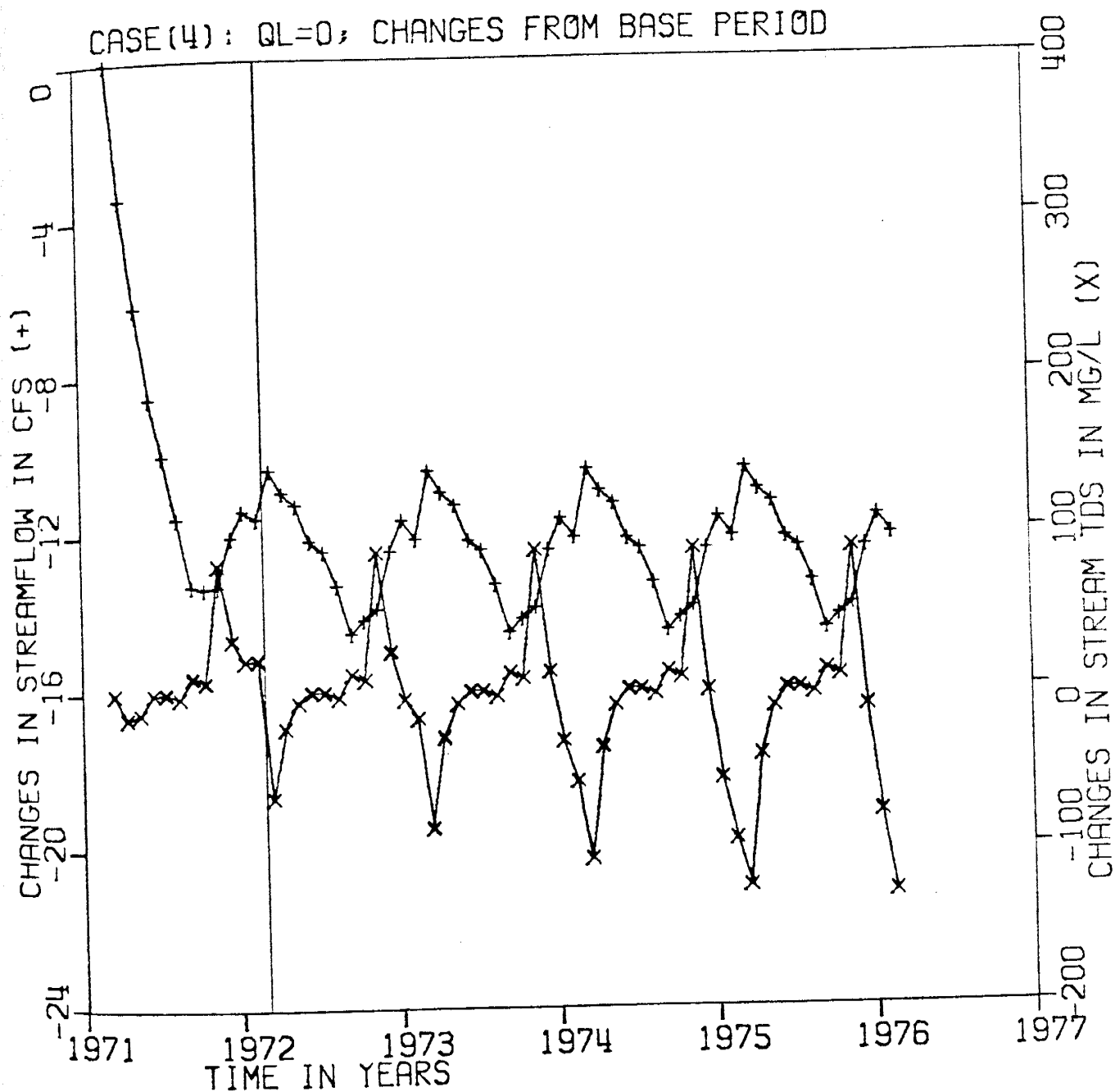


Figure 3.17 Changes in streamflow and surface water quality relative to the base simulation as a result of canal lining to prevent leakage.

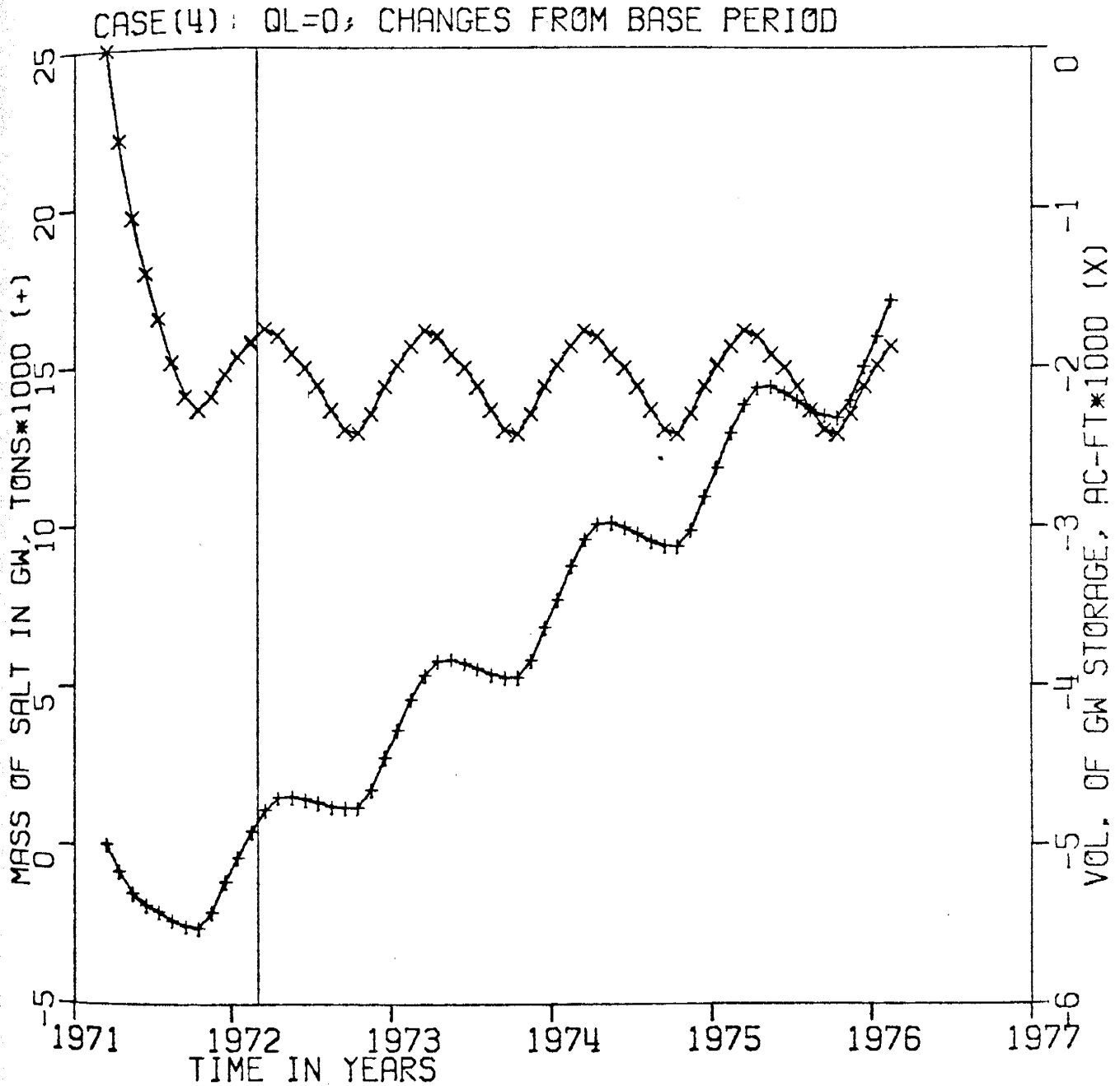


Figure 3.18 Changes in groundwater storage and water quality relative to the base simulation as a result of canal lining to prevent leakage.

years. Konikow and Bredehoeft reported a five average increase of more than 600 mg/l.

4. The flow in the Arkansas River at the Bent-Otero County line would decrease by 9.6 CFS after one year, and 12.4 CFS after five years. Konikow and Bredehoeft reported an average decrease of almost 5 CFS during the first year, and about 11 CFS during the fifth year.

5. The total dissolved solids concentration in the Arkansas River at the Bent-Otero County line would increase by an average of 12 mg/l after one year, but would decrease by an average of 24 mg/l after five years. Konikow and Bredehoeft did not report these changes for this case.

As was pointed out by Konikow and Bredehoeft, these results show that while seepage losses from the Fort Lyon Canal do represent water losses to downstream users, no losses are produced from a systems standpoint. Instead these seepage losses actually serve to maintain water levels in the study reach; furthermore, these higher quality waters tend to dilute the lower quality groundwaters. Canal lining would produce lower water table elevations that would result in higher pumping costs and lower well yields.

3.6 Conclusions

The complex water quantity and quality variations that exist in the irrigated stream-aquifer system along the Arkansas River Valley near La Junta, Colorado, have been successfully simulated by Konikow and Bredehoeft (1974a, 1974b). Their spatially distributed finite difference model solves the appropriate two-dimensional flow equation in order that the convective-dispersive mass transport equation may be solved. As a

result, they obtained monthly distributions of aquifer water levels and total dissolved solids concentration levels throughout the 11 mile reach. They were also able to demonstrate the model's usefulness in evaluating alternative water management options. But as they pointed out (Konikow and Bredehoeft, 1974a), changes in the dissolved solids concentration of groundwater were predominately controlled by convective transport and by mixing with recharged waters of different quality. In their two-dimensional depth averaged model representation, the influence of hydrodynamic dispersion was relatively minor. This may indeed be close to physical reality, or perhaps only a characteristic of input data limitations describing the hydrodynamic dispersion tensor. As future research may show, it may also be that hydrodynamic dispersion effects cannot be adequately represented in a two-dimensional model where input stresses are uniformly distributed. In any case, their results still reasonably reproduce observations, even if some doubt remains about the controlling physics that governs the spatial and temporal distribution of dissolved solids.

Konikow and Bredehoeft also pointed out the relatively large amount of input data required for this and other spatially distributed models. In other words there is a price to pay for obtaining a highly detailed hydrologic description of a given area. If one's primary interest is in simulating the average temporal fluxuations in hydrological and chemical parameters in a system-wide stressed study reach where hydrodynamic dispersion effects are secondary to convective transport, the generalized lumped parameter model may prove useful. The field application of the linear reservoir model to the Arkansas River Valley near La Junta, Colorado, has demonstrated its ability to adequately represent these

average temporal system responses. The predictive capabilities of this lumped approach produce results which are comparable with that obtained from the more complex, distributed, convective-dispersive model of Konikow and Bredehoeft (1974a, 1974b). Furthermore input data requirements and computational effort are greatly reduced, but one sacrifices a more detailed system wide description of water levels and contaminant distributions. While this lumped modeling approach has received relatively little attention in groundwater hydrology, its performance here in the Arkansas River Valley does suggest some potential. For example, in those situations where input data are severely limited, a spatially distributed model may provide no better answers to problems than the lumped model. In other cases the detail provided by more complex techniques may not be warranted. This is not to say the lumped approach is without its own limitations. Environmental contamination and possible solute migration problems are indeed of utmost concern; a spatial and temporal contaminant distribution may be highly desirable. The generalized lumped modeling technique simply provides one with an additional choice in the problem solution.

CHAPTER 4

FIELD APPLICATION OF THE NONLINEAR RESERVOIR MODEL

4.1 Introduction

The predictive capabilities of the lumped parameter reservoir model were examined under a second field evaluation in the Mesilla Valley of southcentral New Mexico. However, this second application simulated a nonlinear stream-connected phreatic aquifer system. This second field test should serve to further demonstrate the general applicability of the lumped modeling approach. Because the required data base covers the thirty-one year period from 1946 to 1976, a better estimate of aquifer parameters can be obtained. Actual simulations were made for the ten year period from May 1967 to December 1976, since an initial estimate of average aquifer water quality was available for this period only. The performance of the lumped model may thus be compared to observed temporal hydrological and chemical concentration variations occurring in the Rio Grande at the valley outflow point as seen in Figure 4.1. Furthermore, these simulation results can also be directly compared to results from the U.S. Bureau of Reclamation Conjunctive Use and Water Quality Model (USBR, 1977) previously tested in the same area under identical conditions (Gelhar and McLin, 1979; McLin and Gelhar, 1979).

The Mesilla Valley is an alluvial valley adjacent to the Rio Grande in southcentral New Mexico; it extends approximately 60 miles from Seldon Canyon (near Radium Springs, New Mexico) to El Paso Canyon (near El Paso, Texas). The valley encompasses a gross area of about 108,600 acres; between 70 and 80 percent of this area is in agricultural production

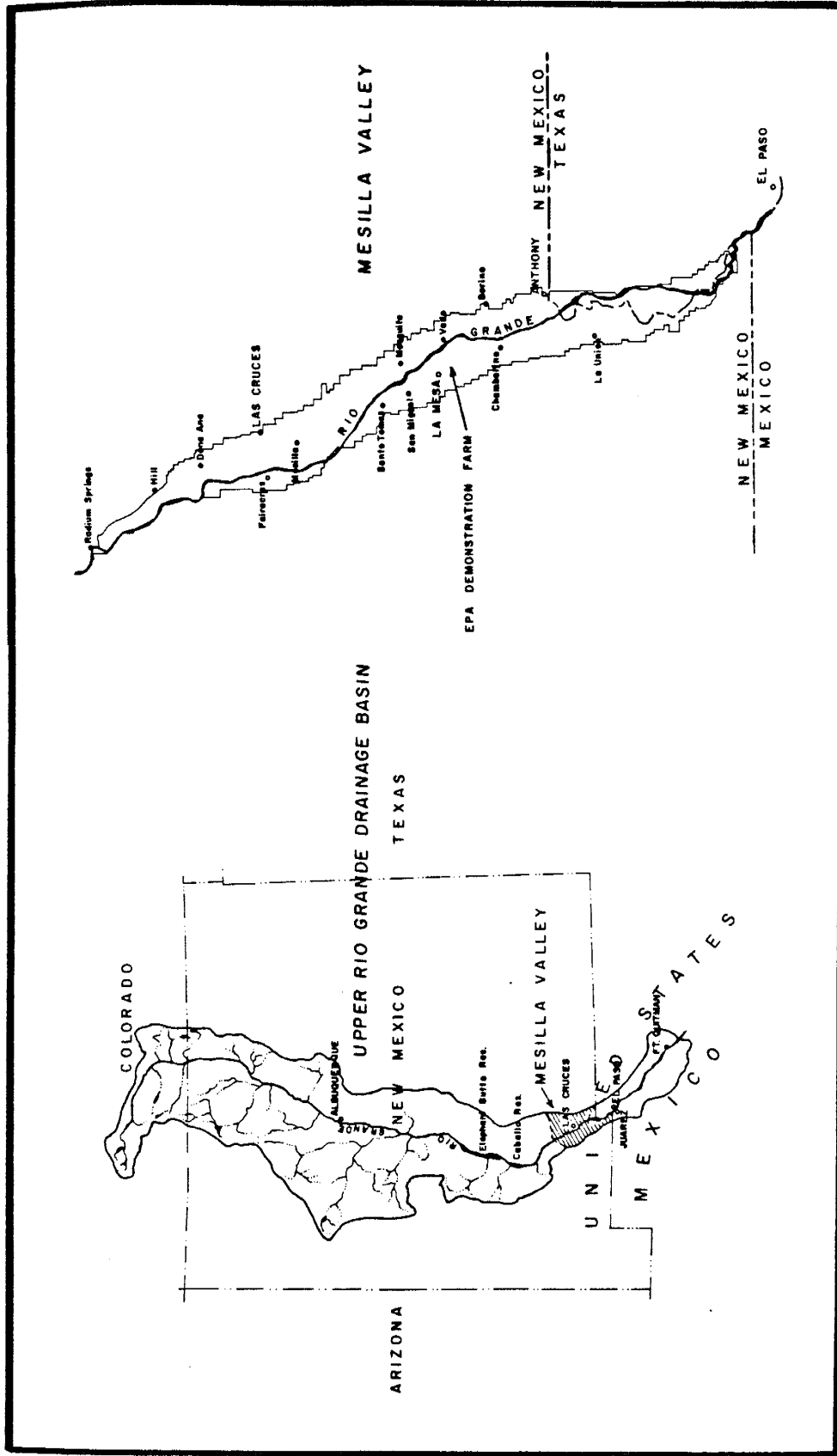


Figure 4.1: Location of the Mesilla Valley and Upper Rio Grande drainage basin, New Mexico.

during the average growing season. The area has had a relatively reliable supply of surface water for irrigation since the completion of Elephant Butte Reservoir in 1916. Groundwater was extensively developed to supplement surface water during the severe drought in the 1950's and continues to be an important source. Major population centers include Las Cruces, New Mexico, and El Paso, Texas. The valley has fairly level topography with a relatively flat alluvial floodplain ranging in width from about 100 yards to about 5 miles. It is bordered by steep bluffs 50 to 100 feet high and composed of loosely cemented sand, silt, clay, and gravel. Recent alluvium on the order of 100 feet thick forms the primary aquifer; this aquifer is underlain by basin fill deposits of the Santa Fe Group which may be several thousands of feet thick. This lower aquifer also yields substantial amounts of higher quality water. The climate of the valley is predominantly semiarid; it is characterized by clear and sunny days, large diurnal temperature ranges, low humidity, and scant rainfall. The mean annual precipitation averages less than 10 inches, with a recorded maximum of about 20 inches and a minimum of about 3 inches. The summer months are generally wettest when tropical air masses move in from the Gulf of Mexico or the Baja Peninsula and cause thundershowers.

Two separate hydrosalinity models of irrigation related water quality were applied to the Mesilla Valley. The USBR-EPA (U.S. Bureau of Reclamation and U.S. Environmental Protection Agency) hydrosalinity model would be classified as a multicell lumped parameter model. Its structure is based on a simple water balance for a given time interval which is computed for each cell. Nondynamic transfers of water between the river and aquifer are used to maintain this balance. A corresponding mass

balance is simultaneously computed according to the volumes of water mixed; simulated chemical exchange reactions in the unsaturated zone may also be included. The generalized lumped parameter model in the form of a nonlinear subsurface reservoir formed the second model; this model dynamically couples the water and mass balance relations in its convective transport description. This second model does not consider any soil-water exchange effects; however, it has advantages of reduced complexity and input data requirements. The nonlinear lumped model may, in addition, be more easily utilized in evaluating the effects of various system management options.

4.2 USBR-EPA Hydrosalinity Model

An extensive description of the USBR Conjunctive Use and Water Quality Model (herein referred to as the USBR-EPA model) is given in the user's manual (U.S. Bureau of Reclamation, 1977); here we emphasize only major irrigation related features. This model is a computer-based mathematical description of an irrigated area represented by a series of small subunits, or nodes, as seen in Figure 4.2. The model structure is based on a water balance for a given time interval which is computed and maintained for each node; nondynamic transfers of water between the river and aquifer are essential to maintain this balance. A model soil column is included to simulate the chemical exchange between the soil and water as water percolates down through the vadose zone. This chemical exchange in the soil is represented by a series of chemical equilibrium reactions for major ions which contribute to agricultural salinity; this segment of the model is based on the development by Dutt, et al., (1972). The model

SCHEMATIC FLOW CHART

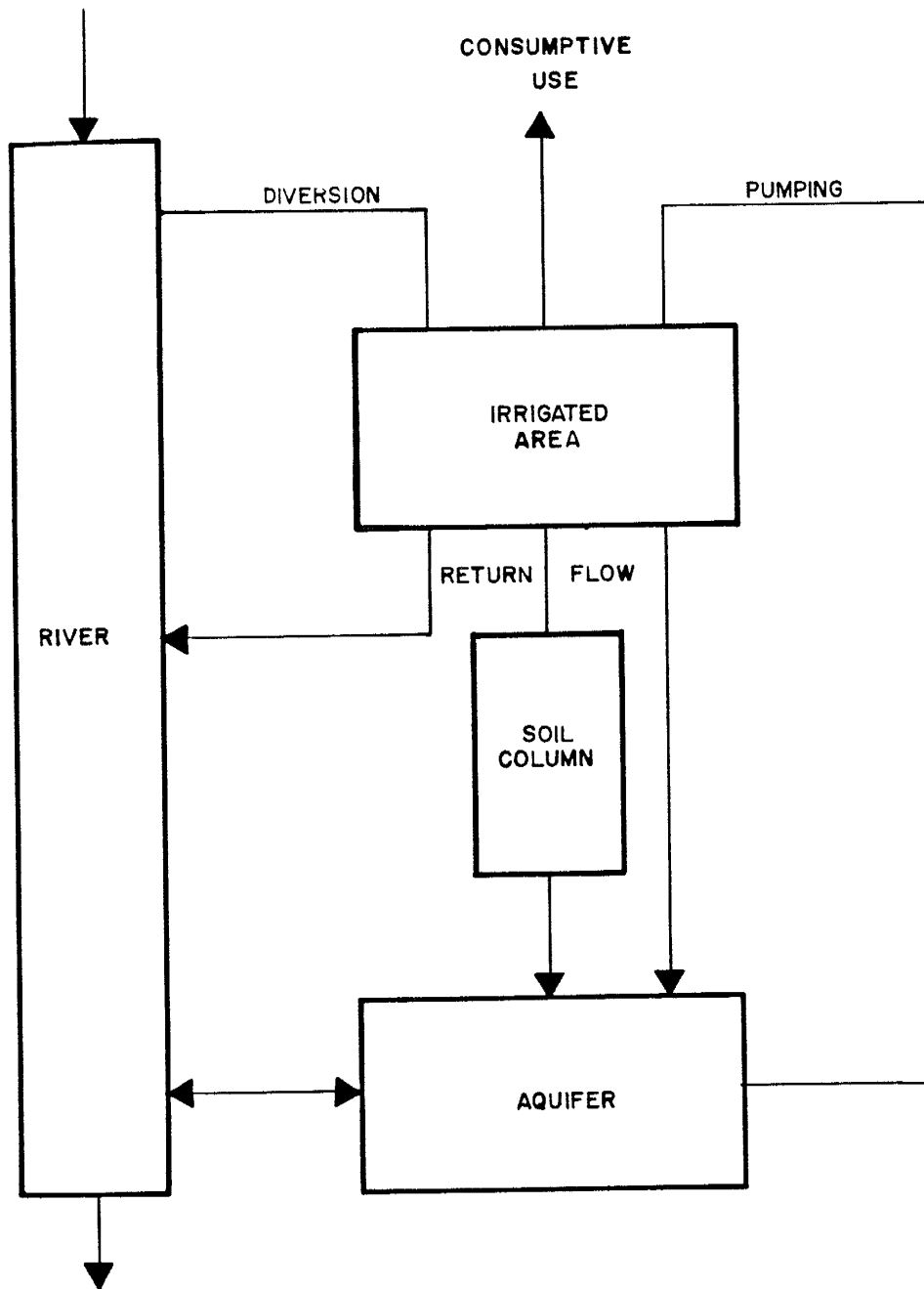


Figure 4.2: Schematic nodal representation showing some features simulated in the USBR-EPA model.

allows the mixing of one water with another, and computes the chemical quality of mixed waters in proportion to the volumes mixed. Hydrodynamic dispersion effects are not considered. All computations are performed for one node at a time and progress from the upstream to the downstream nodes.

A nodal operation includes the simulation of magnitude and quality of river flow, diversions to meet the irrigation demand, water transfers between the river and aquifer, and irrigation return flows. The river flow at the start of a nodal operation is a known quantity, a portion of which is diverted to meet the irrigation demand. The areal irrigation crop demand is computed independently, using the crop consumptive use (ET) and the overall average irrigation efficiency. If the amount of surface water diversion minus canal leakage is not sufficient to meet this demand, additional water is pumped from the model aquifer. The irrigation return flow is then computed by subtracting the ET from the irrigation demand. Simulation of irrigation return flow directly to the river, through the soil to the aquifer, and directly to the aquifer are user specified as a percent of total return flow and may vary for any given time interval. Once a portion of the return flow mixes with the river flow, the chemical quality is updated according to the volumes mixed; this river flow is identified as the predicted river flow. By transferring water to or from the model aquifer, the predicted river flow at the nodal outflow boundary is adjusted to agree with that observed, thereby maintaining an overall water balance. Finally, at the end of a nodal operation, the model yields the river flow and its chemical quality. In the model operation, no consideration is given to soil and aquifer hydraulic properties within a node. Input data needed to operate

the model include irrigation demand, crop consumptive use, diversions, river water quality, initial aquifer water storage and chemical quality, soil moisture content and its chemical quality, and precipitation. Table 4.1 summarizes the major features of the USBR-EPA model as applied to the Mesilla Valley.

In the Mesilla Valley irrigated acreage is served by two diversion dams; these diversion points serve as natural nodes for the USBR-EPA model, since quantity and chemical quality of flows are observed here. Preliminary simulations were made with this internal structure for the period July 1967 through June 1968 in order to examine the model's sensitivity to observed and calculated input data (Gelhar and McLin, 1979). The analyses showed that the monthly predicted TDS outflow concentration history in the Rio Grande was most sensitive to the combined effects of average ET and irrigation efficiency, and to the average initial concentration of aquifer waters. The model output was much less sensitive to aquifer saturated thickness and soil chemistry. These results influenced the emphasis in data collection and final simulations.

Because of the general uncertainty in evaluating ET, we chose the simple Blaney-Criddle method (Blaney and Hanson, 1965) to obtain these initial estimates. In the USBR-EPA model simulations, these initial ET estimates were further adjusted so that the simulated and observed average aquifer water levels followed similar trends. This modification was made using an average ET multiplier that tended to minimize the root mean square (RMS) error between predicted and observed average valley water levels as found from a Thiessen polygon weighting technique.

Table 4.1: Summary of major USBR-EPA model features.

<u>MODEL FUNCTION</u>	<u>IMPLIED LIMITATION</u>	<u>DATA REQUIRED</u>	<u>COMMENTS</u>
Multi-cell lumped parameter model will accept up to 20 cells (or nodes)	All variables are functions of time only	See following	Number of nodes depends on available data and physical features of area
Irrigation water balance to compute return flow	Deals with monthly average over nodal area	Diversions, irrigation demand, GW pump-age, irr. eff., consumptive use, % of return to aquifer and to river	Much of required data is hypothetical
Irrigation water chemistry represented by eight chemical constituents and IDS (Ca^{++} , Mg^{++} , Na^+ , Cl^- , SO_4^- , HCO_3^- , CO_3^- , NO_3^-)	Deals with monthly average over nodal area	Surface diversion chemistry and GW chemistry	Chemistry based on volume proportion of waters mixed
Soil moisture movement thru as many as 10 soil column segments	Steady forced displacement each month	Soil moisture content	Optional feature of model
Soil chemistry includes reactions involving Ca^{++} , Mg^{++} , Na^+ , SO_4^- , HCO_3^- , CO_3^- , and undissolved CaSO_4 and MgSO_4	Chemical equilibria approach where extended Debye-Huckel theory applies	Soil chemical analysis, cation exchange capacity, gypsum concentration	Neglects reactions involving Cl^- and NO_3^-
Aquifer balance transfers water to or from surface and to downstream nodal aquifer(s)	Transfer of flow between river and aquifer is forced to maintain water balance	Initial aquifer volume	River-aquifer transfer should depend on aquifer water level
Aquifer chemistry included in simulation	Uniform aquifer chemistry throughout an entire node	Average initial aquifer chemistry	Equivalent to a "well-mixed" reservoir assumption

Results of several simulation runs are summarized in Figure 4.3; these results indicated that this Blaney-Criddle ET multiplier has a value of about 1.7 for the Mesilla Valley.

The estimation of initial average aquifer water quality was made using the previously obtained Thiessen weight factors (Gelhar and McLin, 1979); however, this approach is complicated by the fact that there are major changes of concentration not only horizontally, but also vertically for any point in the valley. Very little detailed information is currently available, but preliminary studies (C. A. Wilson, personal communication, 1977) indicate a water quality transition zone occurs 50 to 150 feet below land surface. We have simply assumed that concentration is constant with depth down to the average base of the alluvial aquifer at 80 feet in calculating an average aquifer concentration. A complete listing of all data used in the USBR-EPA model is given in Gelhar and McLin (1979), along with the USBR-EPA computer program listing with identified modifications made by us.

The nearly ten-year long period from May 1967 to December 1976 was used in the final simulation to evaluate the USBR-EPA model in the Mesilla Valley. Observed and predicted concentration outflow histories in the Rio Grande are depicted in Figures 4.4 and 4.5 for two of several cases examined. The first case (Figure 4.4) shows the effects of soil chemistry exchange reactions on outflow concentration, using an actual irrigation efficiency estimated to be 0.50. It can be seen that the model simulates the observed seasonal variation in water quality quite well. This seasonal pattern seems to be related to the fact that the relatively saline groundwater, which drains from the aquifer after the irrigation season, is the predominant source of river flow during the

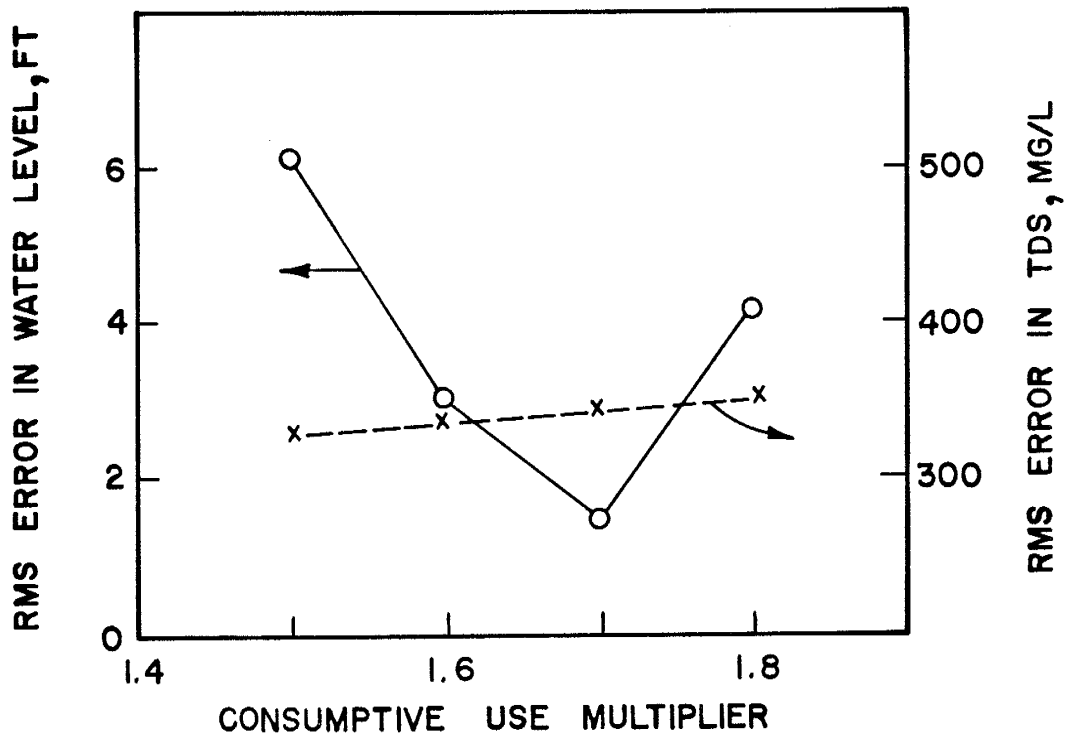


Figure 4.3: RMS error in simulated USBR-EPA aquifer water levels and simulated river outflow chemistry as a function of the consumptive use multiplier.

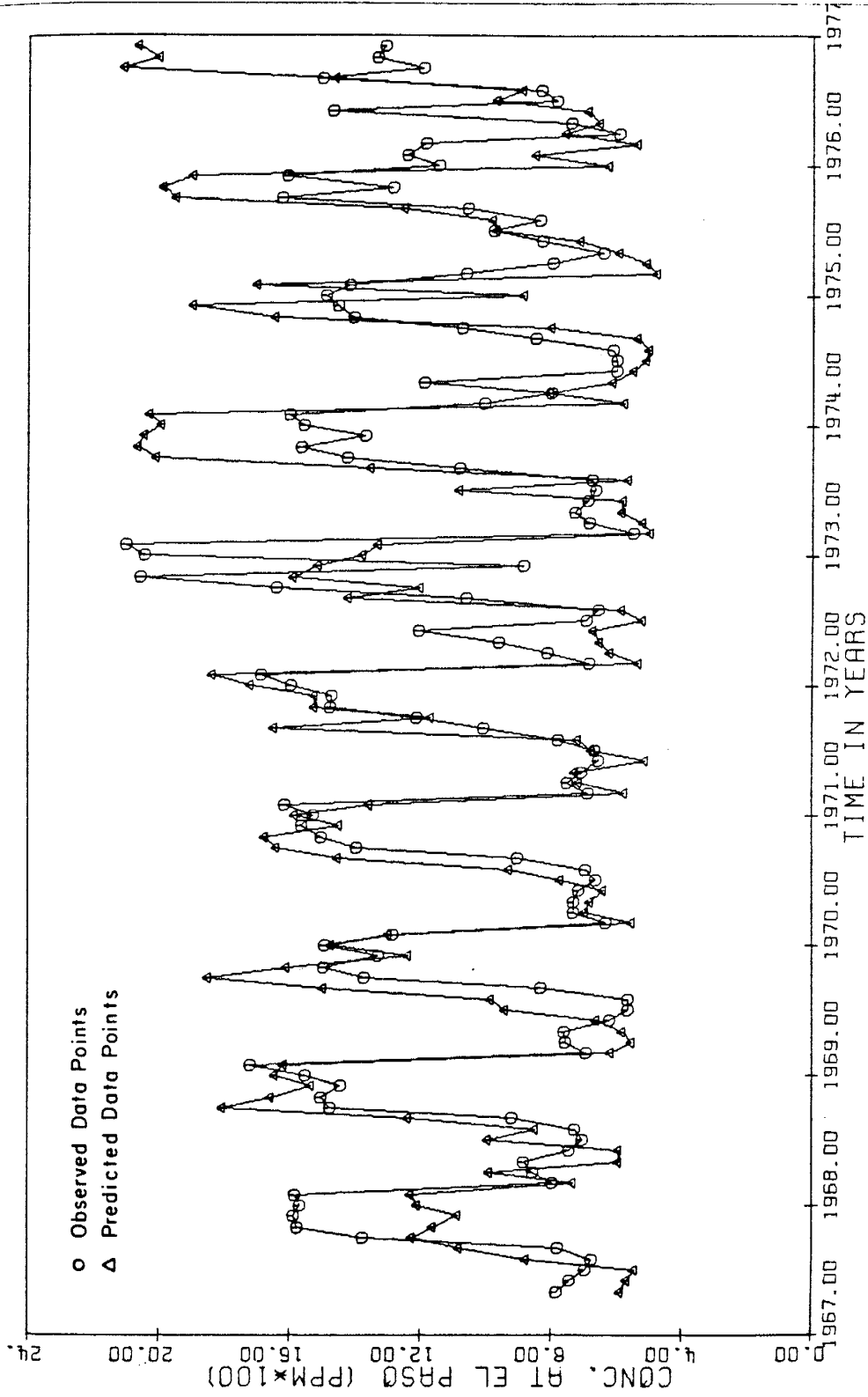


Figure 4.4: USBR-EPA model results using an irrigation efficiency of 50 percent; the RMS error between observed and predicted output is 344 mg/l.

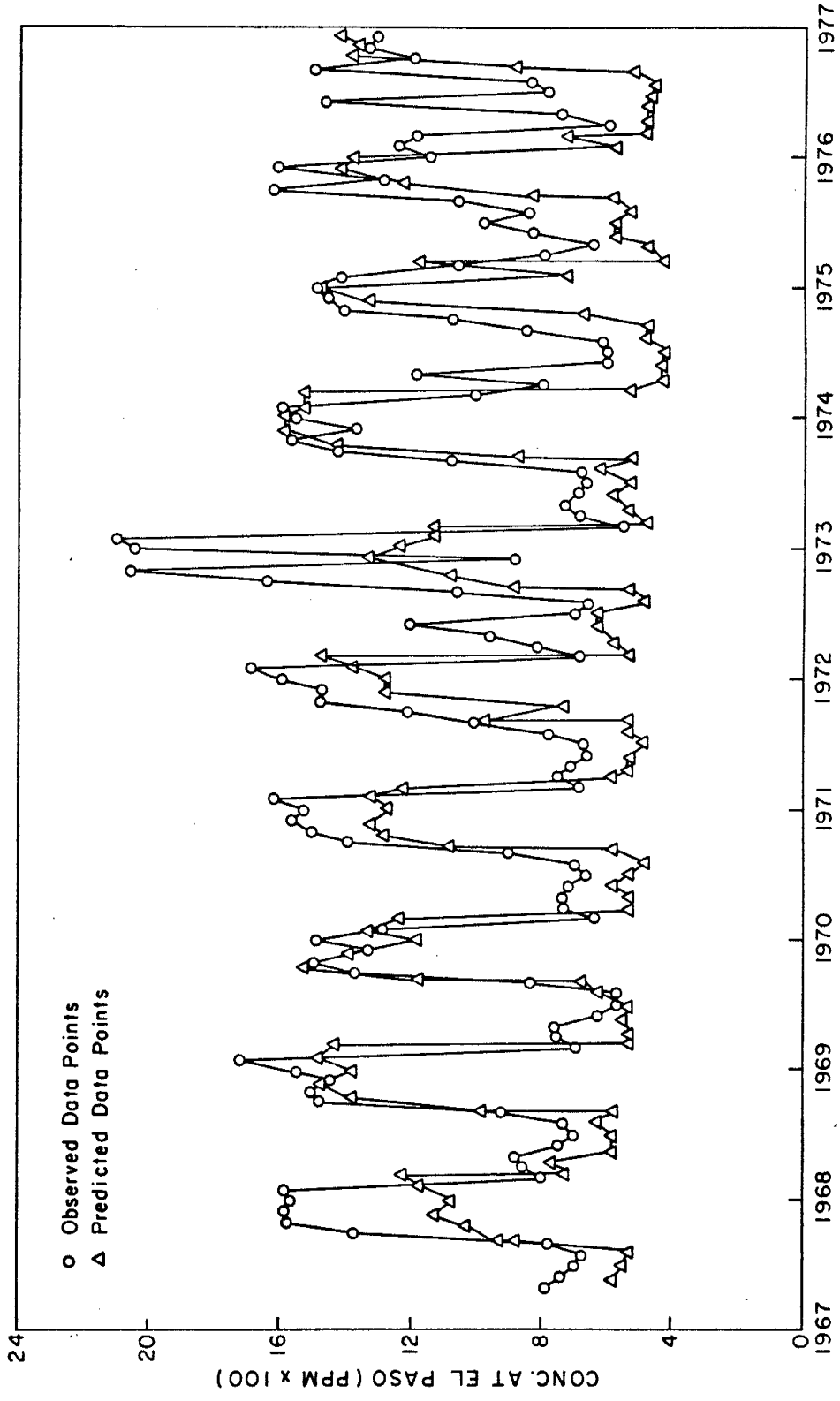


Figure 4.5: USBR-EPA model results using a hypothetical irrigation efficiency of 75 percent; the observed data are shown for reference only.

winter months. However, in the early part of a growing season, large quantities of relatively low salinity surface diversions are applied; these waters tend to dilute the drainage waters.

The second case (Figure 4.5) was run under identical conditions, but with a hypothetical irrigation efficiency of 0.75. These results indicate that a 25 percent improvement in irrigation efficiency over the ten years would have decreased the TDS concentration in the river, especially towards the end of the simulation period (the observed data is plotted in Figure 4.5 for reference only). Because of the relatively large solute response time of this system, several years are required before any noticeable improvements are seen in downstream water quality. In the Mesilla Valley this time period would apparently have been about six years and would have typically resulted in about a 550 mg/l lowering of TDS in the Rio Grande at El Paso, Texas.

4.3 Mesilla Valley Nonlinear Reservoir Model

Derivation of the lumped parameter nonlinear reservoir model for application to the irrigated stream-aquifer configuration along the Mesilla Valley requires only slight modification to the linear case already presented. Based on the data summary previously assembled (Gelhar and McLin, 1979; and references therein), this irrigated system may be represented as a series of observed and estimated inputs and outputs as depicted in Figure 4.6. Thus, the aquifer water balance equation would be

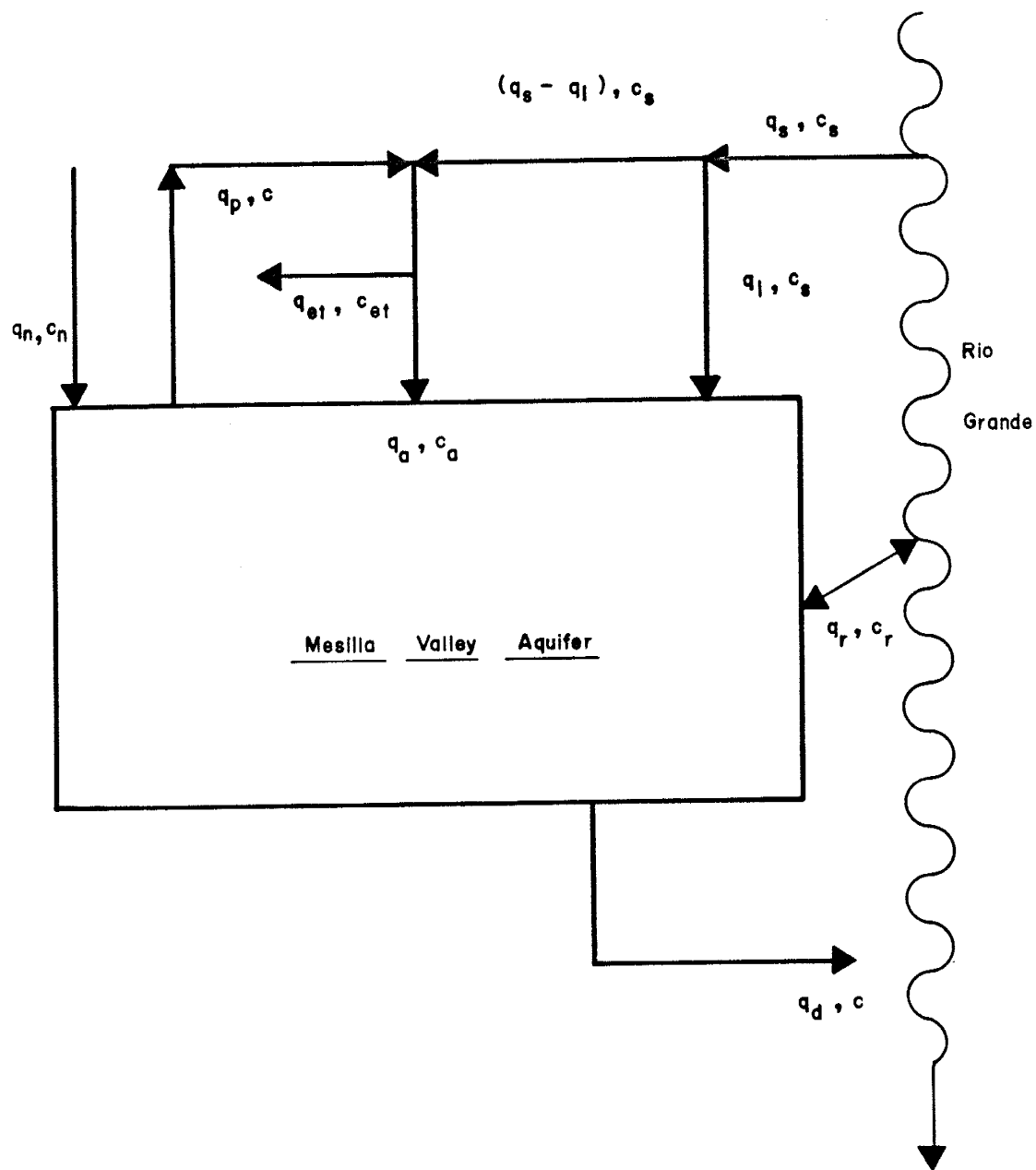


Figure 4.6: Lumped parameter representation of the irrigated stream-aquifer system in the Mesilla Valley.

$$S \frac{dh}{dt} = q_a + q_n + q_l \pm q_r - q_d - q_p \quad (4.3.1)$$

- where S = average specific yield of the aquifer,
 $h(t)$ = average thickness of the aquifer,
 $q_a(t)$ = average artificial recharge rate per unit surface area,
 $q_n(t)$ = average natural recharge rate per unit surface area,
 $q_l(t)$ = average canal leakage rate per unit surface area,
 $q_r(t)$ = average river leakage rate per unit surface area,
 $q_d(h,t)$ = natural aquifer outflow rate per unit surface area,
 $q_p(t)$ = aquifer pumping rate per unit surface area.

The river leakage term in (4.3.1) is positive for river to aquifer transfers and negative otherwise. Also shown in Figure 4.6 are the respective contaminant concentration levels corresponding to the various average temporal inputs and outputs. In the figure the evapotranspiration term, q_{et} , is depicted as a net deficit on the applied irrigation waters and never enters the model aquifer. The advantage of this representation will become clear later; briefly stated, however, we may actually avoid having to evaluate the crop consumptive use and groundwater pumpage terms in the final water and mass balance equations. Data for q_p is often unavailable since this quantity is usually unmetered; furthermore, q_{et} will carry large inherent errors no matter what technique is utilized in its determination. Thus, we actually may avoid many of the shortcomings of the typical water balance approach to this type of problem solving by using this formulation.

The manner in which the natural aquifer outflow term (or aquifer drain flow), q_d , is approximated will determine whether the lumped model represented by (4.3.1) is linear or nonlinear. Updegraff and

Gelhar (1978) previously examined a linear representation of the Mesilla Valley water balance system using an expression like that in (3.2.1). Their approach worked extremely well for a normal precipitation period; however, it did not adequately represent average conditions in the valley for a subsequent low flow period. Pinzon (1978) was able to overcome this shortcoming by replacing (3.2.1) with the nonlinear expression

$$q_d = \frac{1}{N} \ln \left\{ \exp [aN(h-h_d)] + 1 \right\} \quad (4.3.2)$$

where N is some system constant and h_d is the average drain reference water level; all remaining terms have been previously defined. This equation is linear for sufficiently large values of $(h-h_d)$, but is nonlinear when $(h-h_d)$ becomes small or slightly negative; actually (4.3.2) is nearly identical to (3.2.1) over most of the range of values for h . Recall that the parameter a in (4.3.2) is a lumped outflow constant having units of inverse time; it is inherently related to Darcy's law (Flores W., et al., 1978). In general, h_d can be a function of time, but is usually assumed constant since fluctuations in h_d are usually much smaller than those for h . Substituting (4.3.2) into (4.3.1) yields a differential equation of the form

$$S \frac{dh}{dt} + q_d = E \pm q_r \quad (4.3.3)$$

where E represents the lumped time dependent inputs minus outputs (that is, $E = q_a + q_n + q_l - q_p$); in (4.3.3) the term q_d is given by (4.3.2). When using a nonlinear aquifer outflow expression like (4.3.2), a numerical solution technique for (4.3.3) will usually be necessary.

This particular nonlinear expression worked well in the elongated Mesilla Valley, where average valley water levels sometimes fall below h_d ; drainage then ceases in the upper portion of the valley while the lower end may still experience some active drainage.

The corresponding mass balance equation for the stream connected aquifer system would be

$$n \frac{d(hc)}{dt} = q_a c_a + q_n c_n + q_l c_o + q_r c_r - q_d c - q_p c + nhr' \quad (4.3.4)$$

where r' is a volumetric source-sink term that accounts for contaminant additions or degradation within the flow zone, c is the average aquifer concentration, and c_n and c_a are the respective concentrations for natural and artificial recharge waters. If the river leakage (q_r) is positive, then waters are transferred from the river to the aquifer; c_r is identical to the river concentration. If q_r is negative, then waters are transferred from the aquifer to the river and c_r is identical to the average aquifer concentration, c . It is implied in (4.3.4) that the aquifer is a well-mixed linear or nonlinear reservoir (depending on q_d) and that any system outflows will carry this average aquifer concentration; this assumption has been related to the nature of two-dimensional convective transport for the linear model in a previous chapter and earlier by Gelhar and Wilson (1974). Using (4.3.1) with $S = n$ in (4.3.4) yields

$$\frac{dc}{dt} + c \left(-\frac{\epsilon}{nh} + k \right) = [q_n c_n + c_s (q_s + q)] / nh \quad (4.3.5)$$

where $\epsilon = E + q$

$q = q_r$, if $q_r > 0$

$q = 0$, if $q_r < 0$

$r' = -kc$, a first order decay process

and where q_s represents the volume rate of applied river water per unit area, and c_s is its concentration level. For irrigated systems the transported contaminant will usually have a decay constant (k) equal to zero. The combined parameter nh/ϵ may be referred to as the solute response time (t_c) since it characterizes the average response time of (4.3.5); physically it represents the average solute residence time in the aquifer. For the Mesilla Valley study period, yearly averages for t_c varied between 11.4 and 40.3 years.

The water and mass balance statements given by (4.3.3) and (4.3.4), together with (4.3.2), form a coupled systems description of subsurface water and mass transport in a nonlinear lumped parameter format. Net stresses are assumed to be uniformly distributed throughout the system as was done for the linear reservoir model; hydrodynamic dispersion effects are not considered. These equations are easily manipulated in discretized form; examples of numerical solutions may be found in Gelhar and Wilson (1974), Flores W., et al., (1978), and Updegraff and Gelhar (1978).

In the proceeding analyses (4.3.3) will yield a general solution for the predicted average water table elevation above some reference datum for the lumped parameter nonlinear reservoir model; (4.3.5) will similarly yield a corresponding aquifer outflow concentration. The lumped model is therefore dynamic when these equations are coupled

together; hence, it is a truly predictive management tool. Thus if future input stresses can be estimated, then the corresponding system outputs can be found. However, certain physical parameters must first be estimated which characterize the system response. Updegraff and Gelhar (1978), using a linear reservoir lumped parameter model, proceeded to estimate a and h_d in (3.2.1) using a least squares regression technique for observed average water level and drain flow data in the Mesilla Valley. Once these parameters were found, they discretized an equation similar to (4.3.3). They obtained a least squares estimate of n and q_r for periods of no irrigation (i.e., no net recharge). Their equation could then be resolved for net recharge using the above parameters and observed values of average water level elevations for periods of active irrigation assuming the river leakage remained constant for a given recharge-recession period. This linear approach worked extremely well in the Mesilla Valley for a five-year period (1946-1951) showing normal precipitation. The onset of several successive dry years (1952-1956) resulted in excessive groundwater pumpage and relatively large water level declines; their linear technique did not adequately represent average conditions in the valley for this low drain flow period. Pinzon (1978), using (4.3.2) instead of (3.2.1), was able to overcome this shortcoming of the linear model. The procedure used by him was nearly identical to that of Updegraff and Gelhar (1978). A least squares nonlinear solution of (4.3.2) using Mesilla Valley observed average water level and drain flow data for the period 1946-1976 therefore yielded estimates for N , a , and h_d . The discretized form of (4.3.3) was solved for n and q_r during periods of no irrigation using observed values of h (see Appendix F); this same equation was then resolved for net recharge

(E) during active irrigation periods (see Appendix E). Finally a predictor-corrector iteration procedure was used in conjunction with (4.3.3) to obtain the predicted average aquifer water levels for the above parameters. Constant river leakage was again assumed for a given recharge-recession period. Equation (4.3.2) and this simulated h then yielded a simulated drain flow. Figures 4.7 and 4.8 show these simulated results for the period 1967 to 1976, along with the corresponding observed data; Table 4.2 lists both the linear and nonlinear parameters for the Mesilla Valley. The nonlinear outflow term given by (4.3.2) obviously simulates the average hydrological conditions in the Mesilla Valley quite well, although changes in this expression might improve these results even further.

Table 4.2: Comparison of linear and nonlinear aquifer parameters for the Mesilla Valley.

<u>Model</u>	<u>Parameter Value</u>	<u>Source</u>
linear	a = 0.0812/month	Updegraff and Gelhar, 1978
	n = 0.210	
	h_d = 3824.85 feet	
	t_h = n/a = 2.59 months	
nonlinear	a = 0.0755/month	Pinzon, 1979, and here.
	n = 0.218	
	h_d = 3825.00 feet	
	N = 9.05	
	t_h = n/a = 2.89 months	

The generalized lumped parameter model in the form of a nonlinear subsurface reservoir was evaluated here for its ability to predict fluctuations in monthly river outflow total dissolved solids, using a portion of the same data base previously assembled for the USBR-EPA model, and the water balance parameters found by Pinzon (1978). This

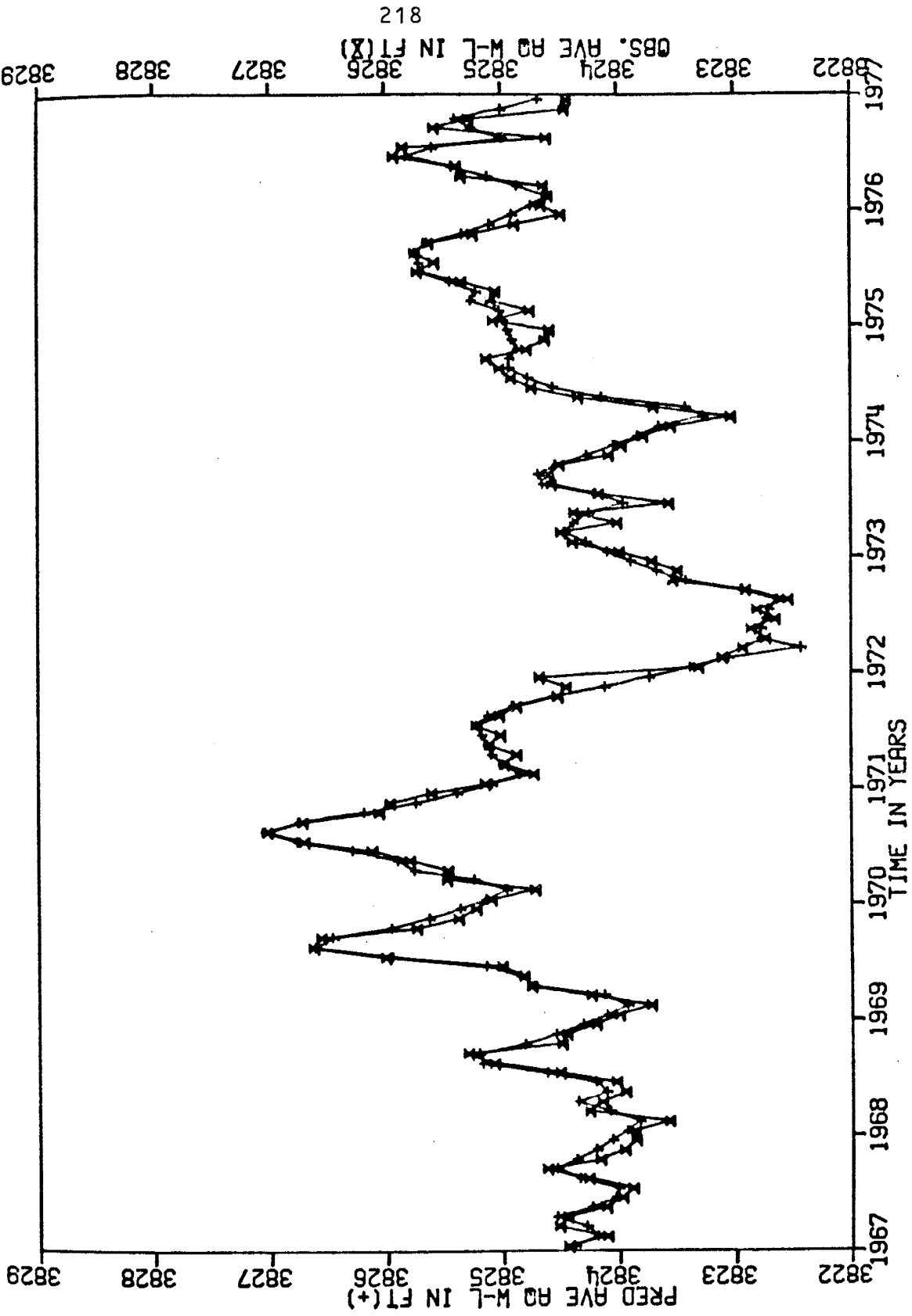


Figure 4.7: Monthly observed and simulated average aquifer water levels using the nonlinear lumped model; the RMS error is 0.18 feet (5.49 cm).

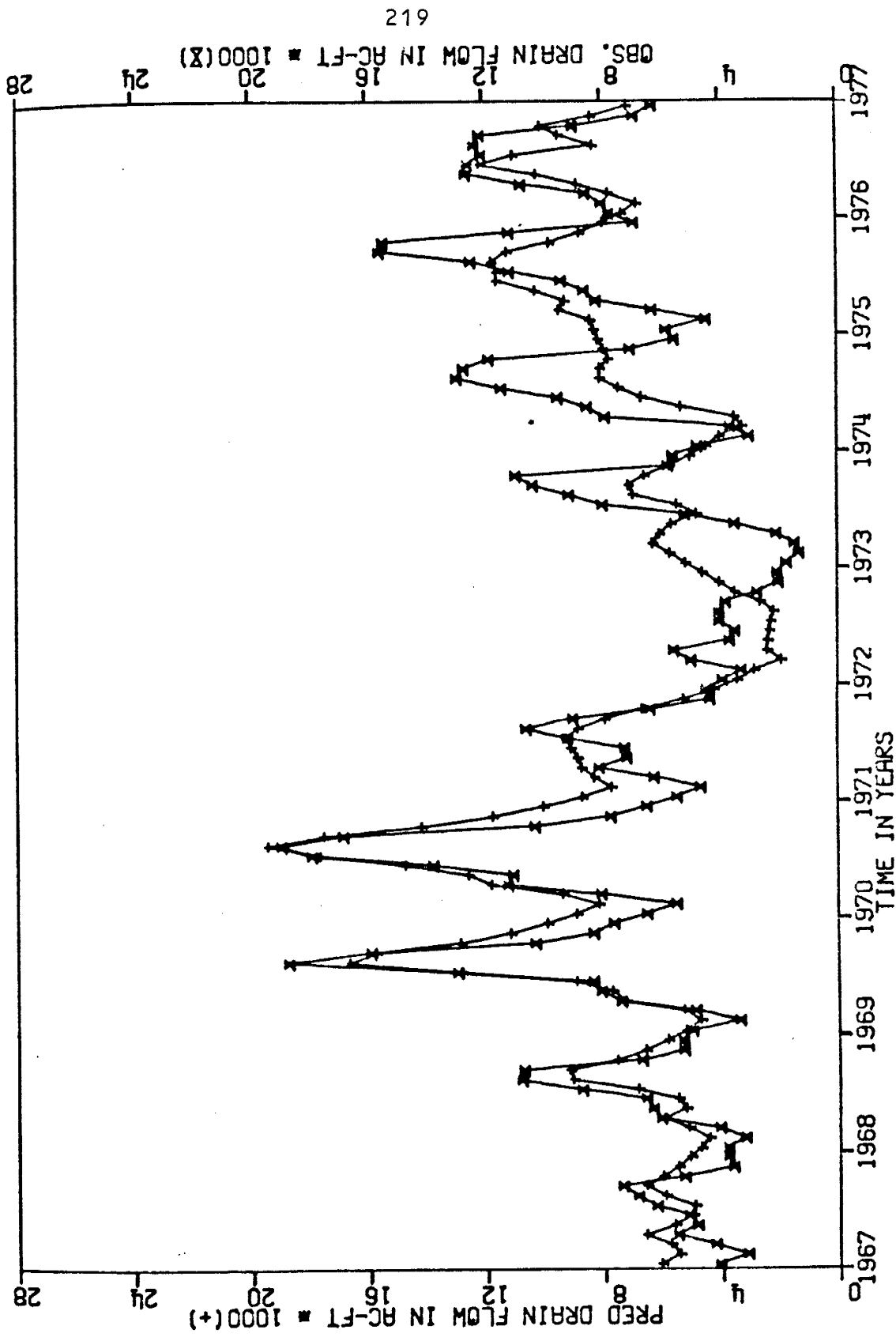


Figure 4.8: Lumped parameter monthly simulated and observed average valley drain flow; the RMS error is 0.02 feet per month (0.61 cm/month).

simulation was based on a discretized version of the mass balance relationship given by (4.3.5); thus, a backward difference approximation over the Δt time interval may be written as

$$\left(\frac{c_i - c_{i-1}}{\Delta t}\right) + \left(\frac{c_i + c_{i-1}}{2}\right) \left(\frac{E + q}{nh}\right)_{i-1} = c_{S_{i-1}} \left(\frac{q_s + q}{nh}\right)_{i-1}$$

where the decay constant k is zero and c_n is assumed to be negligible since $c_n \ll c_s$. In this expression we have considered the net recharge (E), the river leakage (q), and the surface diversion (q_s) to each be measured between the middle of the i th and $(i+1)$ th time steps, as was done in the previous field application (see Figures 3.2 and 3.3).

The above relationship may be solved to yield

$$c_i = \frac{c_{S_{i-1}} \left(\frac{q_s + q}{nh}\right)_{i-1} \Delta t - c_{i-1} \left(\frac{E + q}{nh}\right)_{i-1} \frac{\Delta t}{2}}{1 + \frac{\Delta t}{2} \left(\frac{E + q}{nh}\right)_{i-1}} \quad (4.3.6)$$

Equation (4.3.6) will yield the average aquifer concentration at each time step, assuming that the total net recharge, ϵ_i , is treated as a pulsed input of constant concentration during the i th month. In this discretized form (4.3.6) requires only q_s and c_s as input data, since all remaining terms may be found as a result of the parameter estimation procedure. Figure 4.9 shows predicted and observed changes in the Rio Grande TDS at El Paso, Texas (the Mesilla Valley outflow point), as found using the nonlinear lumped parameter model for the period 1967-1976. Other model output is shown in Figures 4.10 and 4.11; the complete computer program listings, input data and model output are listed in Appendices E, F, and G.

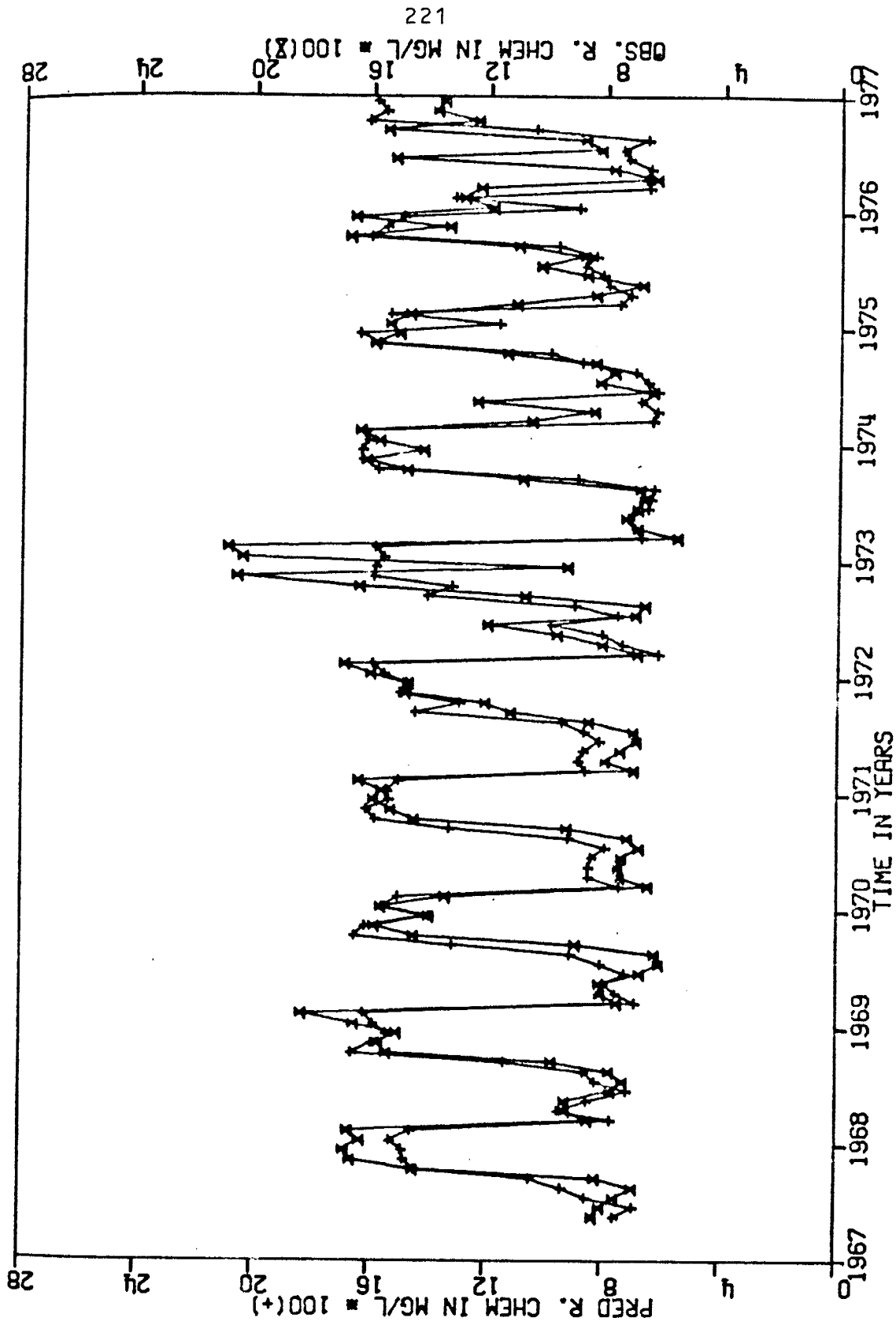


Figure 4.9: Lumped parameter monthly observed and simulated Rio Grande TDS at El Paso, Texas; the RMS error is 213 mg/l.

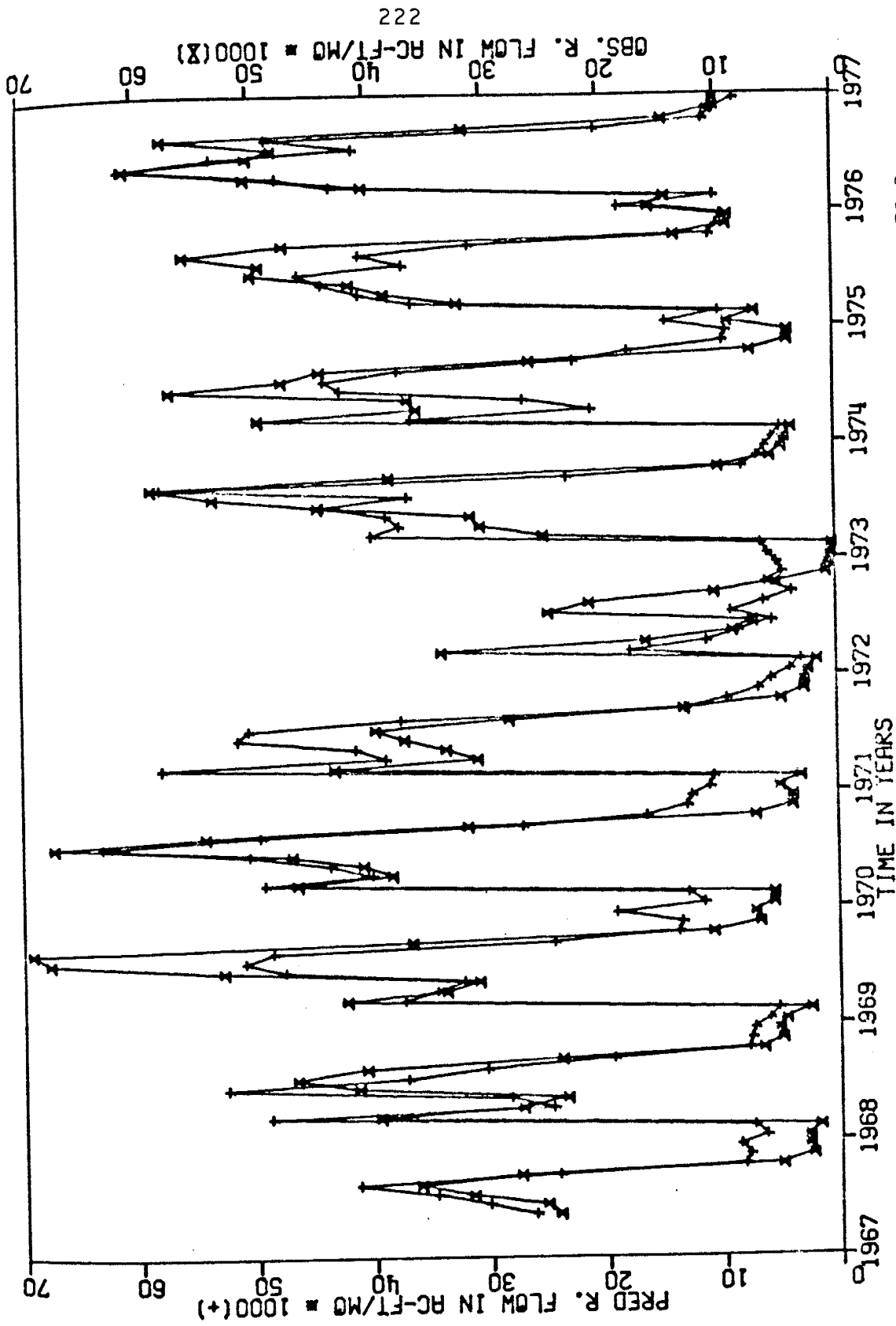


Figure 4.10: Lumped parameter monthly observed and simulated Rio Grande flow at El Paso, Texas; the RMS error is 7324 acre-feet per month (903 hectare-meters per month).

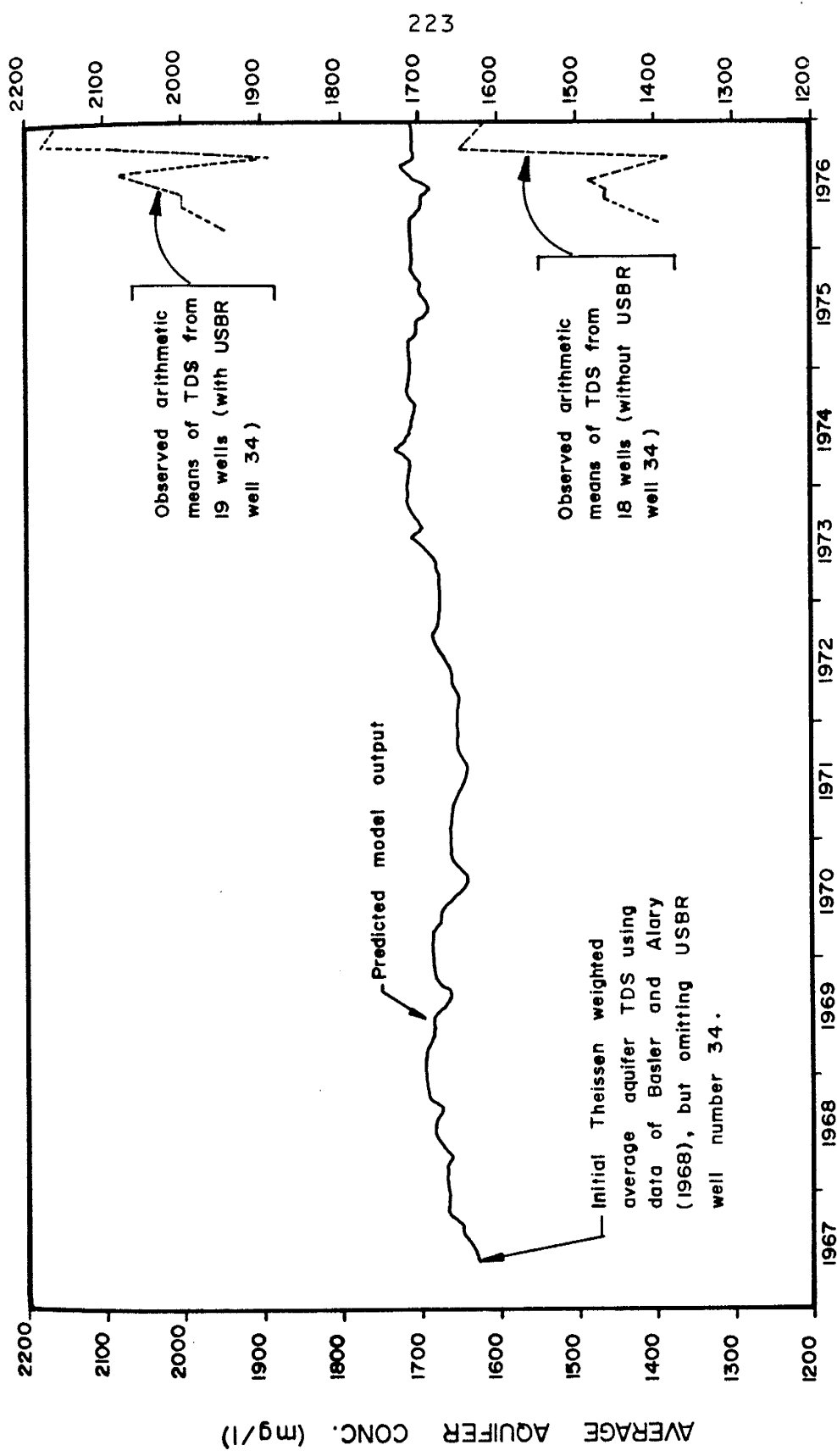


Figure 4.11: Lumped parameter monthly predicted average aquifer TDS; a few months of data were observed.

Figure 4.11 shows the lumped parameter model's prediction of average aquifer TDS over the ten year simulation period. These simulated results were obtained by using an initial average aquifer TDS value of 1626 mg/l. This initial value was computed from the observed water quality data reported by Basler and Alary (1968), and by utilizing the Theissen polygon weighing technique for 39 wells distributed throughout the 108,600 acre valley (Updegraff and Gelhar, 1978; and Gelhar and McLin, 1979). In this computation the data from USBR well number 34 was omitted (Gelhar and McLin, 1979, p. 153) because it was felt that the TDS value of about 11,000 mg/l was not representative. However, in Table 3 of Gelhar and McLin (1979, p. 30), this data was included in arriving at the Theissen weighed mean value of 1846 mg/l for TDS because of specific requirements there. Since water quality data from all of these same wells was not available after May 1967, no other Theissen averages appear on Figure 4.11; Table 4.3 summarizes these results. This figure does show, however, the results obtained from arithmetically averaging the observed TDS values from 19 wells having water quality data available over a portion of the simulation period (Gelhar and McLin, 1979, p. 128 to 153). No other water quality data is known to exist over this time interval. Furthermore, it is important to note that some of these 19 wells were installed after 1967, and therefore are not included in the original 39 wells used to establish initial conditions. Finally, if one well is omitted from this list of 19 wells (i.e., USBR well 34), the average aquifer TDS value is lowered by about 500 mg/l. Thus a band of these average TDS values is obtained, as shown in Figure 4.11.

As reported in Gelhar and McLin (1979, and references therein), the horizontal and vertical spatial variation in groundwater quality is very

Table 4.3: Arithmetic means of observed TDS values

<u>DATE</u>	UPPER (2) MEAN (mg/l)	LOWER (3) MEAN (mg/l)	PREDICTED (4) MODEL TDS (mg/l)
3-76	1961.7	1376.8	1715.6
4-76			1706.4
5-76	1998.4	1443.4	1698.3
6-76	2002.3	1465.8	1696.1
7-76	2039.6	1464.9	1683.2
8-76	2075.3	1485.9	1710.0
9-76			1722.2
10-76	1888.8	1388.7	1706.4
11-76	2183.7	1648.4	1707.6
12-76	2171.0	1630.0	1707.6
1-77	2134.9	1594.7	
2-77	1946.6	1450.0	
3-77			
4-77			
5-77	1996.0	1479.8	
6-77	2110.6	1587.6	
7-77			
8-77	1973.3	1500.5	
9-77			
10-77	2045.8	1577.7	
11-77	1993.3	1536.9	

- (1) See Figure 4.11 for time history plot of means; see Table 4.4 for list of wells used to compute arithmetic means.
- (2) Computed with data from USBR well no. 34.
- (3) Computed without data from USBR well no. 34.
- (4) No simulation results beyond December 1976.

Table 4.4: List of Wells used for Arithmetic means (1)

<u>SITE</u>	<u>WELL NUMBER</u>
3	USBR no. 19
5	USGS Well Nets, North Well
6	USBR no. 16
8	USBR no. 12
9	USGS Well Nest, North Well
10	USGS Well Nest, South Well
12	USBR no. 14
13	USBR no. 8
14	USBR no. 7
15	USBR no. 6
16	USBR no. 24
15	USBR no. 5
18	USGS Well Nest-2nd Well from East
19	USBR no. 22
20	USGS Well Nest-Middle Well
22	USBR no. 39
23	USBR no. 29
25	USBR no. 2
29	USBR no. 34

(1) The actual well locations and TDS data for each of the above wells is given in Gelhar and McLin (1979, p. 129-153).

complex in the Mesilla Valley. Any technique that is designed to yield a system wide average groundwater quality will yield results that are primarily dependent on the following: (a) well location and perforated interval, (b) total number of wells and their spatial distribution throughout the system, (c) time of year that samples are recovered, and (d) the precise technique employed in the averaging process.

Furthermore, once a given technique is selected for this averaging process, the same wells should be utilized in comparing temporal fluctuations in this spatial average. Even if such a procedure is followed, one may still wonder if a representative temporal trend in average water quality is obtained. While such a fundamental question is beyond the intended scope of this work, the potential effect must at least be qualitatively realized. The example in Gelhar and McLin (1979, p. 29 to 30) illustrates this point rather well.

4.4 Comparative Study Results

Figures 4.4 and 4.5 depict the USBR-EPA simulation results. The results of the ten-year simulation in the Mesilla Valley demonstrate that both the USBR-EPA hydrosalinity and generalized lumped parameter models can simulate pre-existing conditions quite well if adequate data are available to describe the overall water and mass balance conditions of the aquifer. Our experience with the Mesilla Valley system indicates that the data base is reliable and consistent. However, groundwater quality does show a high degree of horizontal and vertical spatial variability; furthermore, the vertical extent of this higher saline groundwater is not well defined. The actual predictive capabilities of

the USBR-EPA model are less clear than for the generalized lumped parameter model. In order to predict the hydrochemical effects of changes in water application, one must be able to synthesize the resulting changes of flow into and out of the system because this model completely ignores aquifer flow dynamics. In other words the aquifer hydraulic coefficients and changes in flow resulting from changes in water level are not used.

The generalized lumped parameter model, on the other hand, does incorporate the mixing cell structure plus the flow dynamics of the aquifer. This model can also be operated with a conceptually much simpler computer program; calculations can even be carried out on a programmable pocket calculator, though somewhat inconveniently for long simulation periods. The predictive capabilities of this same model for the Mesilla Valley are discussed in the next section.

4.5 Water Management Options with the Nonlinear Model

The same cases previously considered for the linear lumped parameter model (see Section 3.5) were used in this nonlinear application to demonstrate its ability to predict the long term response of the stream-aquifer system of the Mesilla Valley to changes in irrigation practices and water management options. It was again decided to present the simulation results on a relative rather than an absolute response; however, the original ten year simulation results of Section 4.4 are presented here in their absolute form for ease of comparison (see Figures 4.12 to 4.15). The same four specific water management options previously used in Section 3.5, and by Konikow and Bredehoeft, were

employed here; these cases were: (1) increased groundwater use, (2) increased surface water use, (3) improved irrigation efficiency, and (4) lining all conveyance canals.

In Section 4.3 the water balance equation was simplified to the form given by equation (4.3.3) in order to avoid calculation of the crop consumptive use and groundwater pumpage terms. The net lumped inputs minus outputs were identified as E ; these values were in turn determined as a result of the parameter estimation procedure. In order to utilize the nonlinear model to evaluate system responses to different water management options, we must be able to synthesize changes in E that result from changes in water application. These changes will in turn affect the water and mass balance equations, thereby yielding model predictions for any water management scheme envisioned.

We can see that in order to evaluate the effects of cases (1) through (4) above, appropriate source terms must appear explicitly in equations (4.3.3) and (4.3.5). Thus in case (1) for example, we need to know what fraction of the surface diversion (q_s) enters the aquifer as artificial recharge (q_a) in order to change this volume source in the mass balance relationship to reflect increased groundwater pumpage. Additional valley data or appropriate assumptions are therefore required in order to proceed. In the Mesilla Valley most of this type of information is costly and is therefore not generally available. Some estimations have been made, however, with regard to canal leakage and irrigation efficiency (Gelhar and McLin, 1979). Here we use the term irrigation efficiency to mean the percentage of the total amount of ground and surface waters diverted for use by agricultural crops. The irrigation demand is the total consumptive use (q_{et}) divided by the

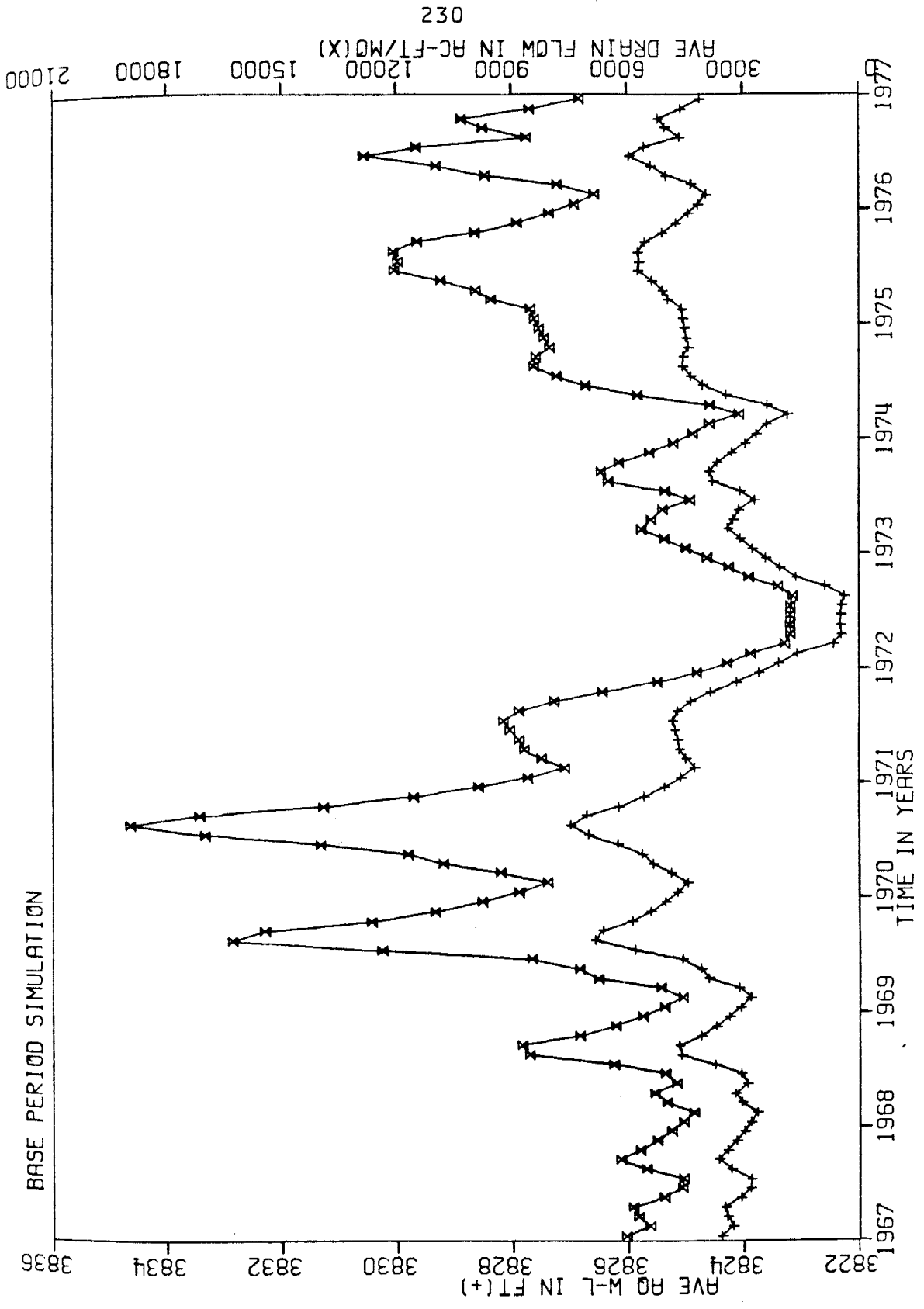


Figure 4.12: Base period simulation of average aquifer water levels and drain flow in the Mesilla Valley.

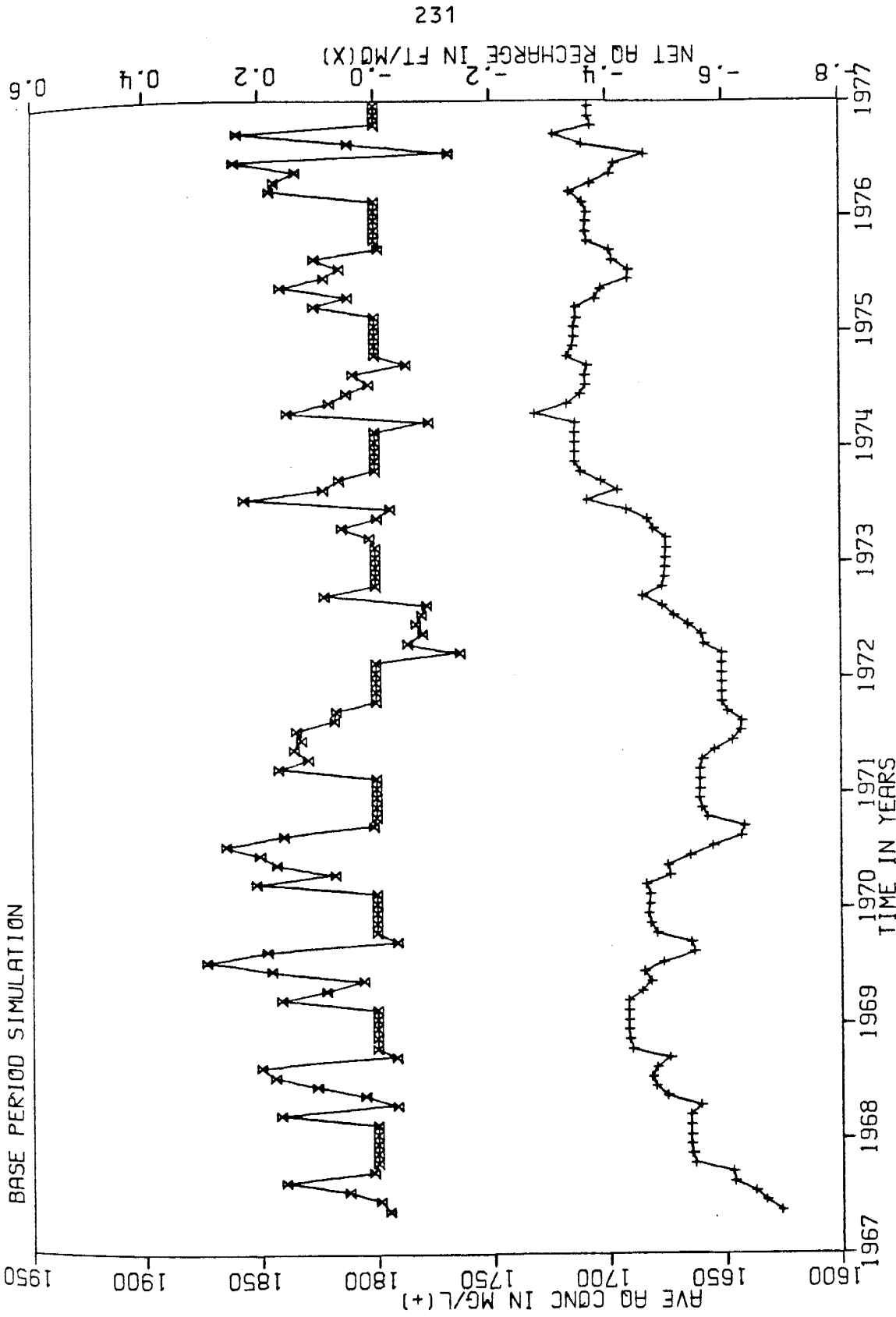


Figure 4.13: Base period simulation of average aquifer TDS and net aquifer recharge for the Mesilla Valley.

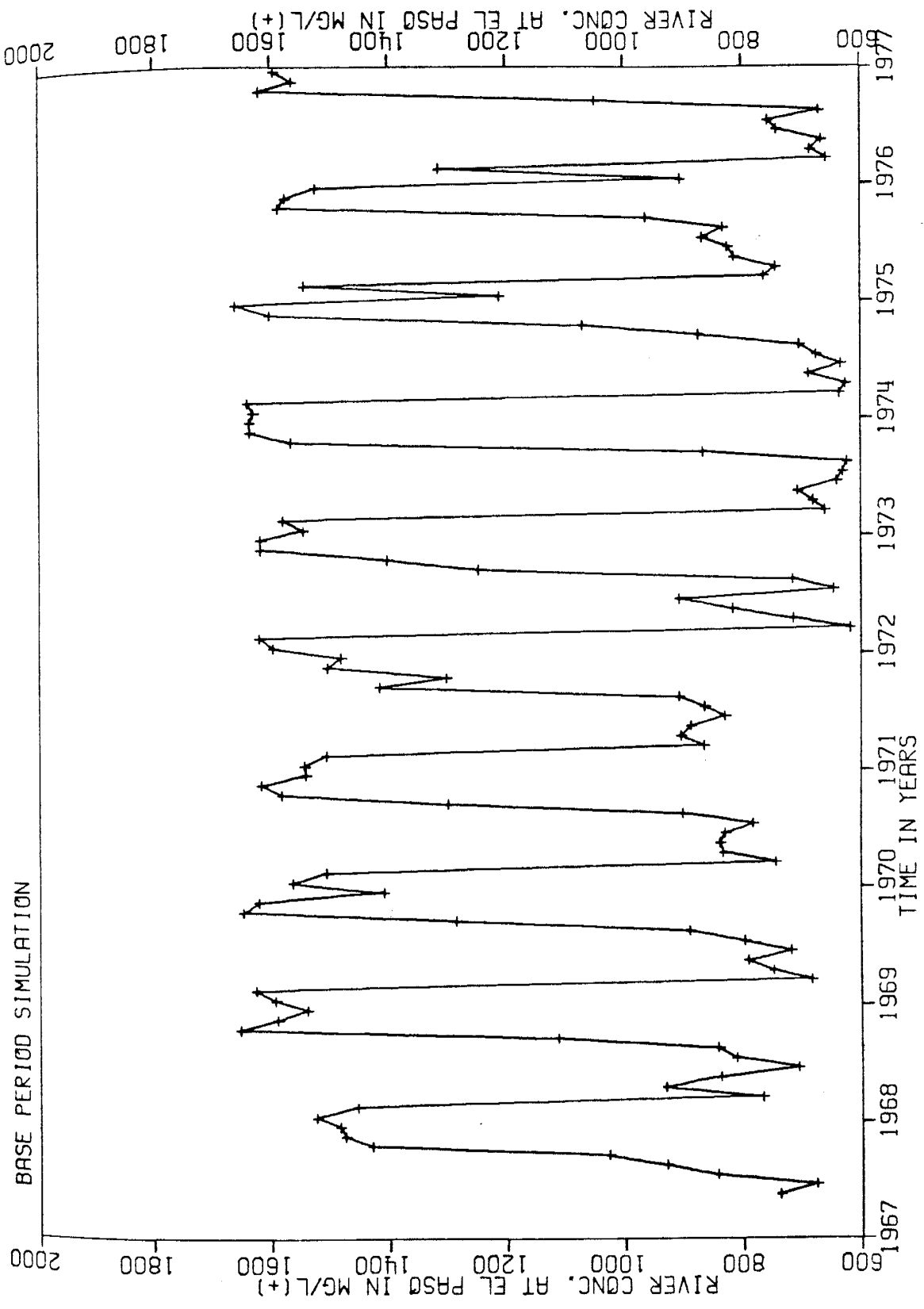


Figure 4.14: Base period simulation of average river TDS at El Paso, Texas, the Mesilla Valley outflow point.

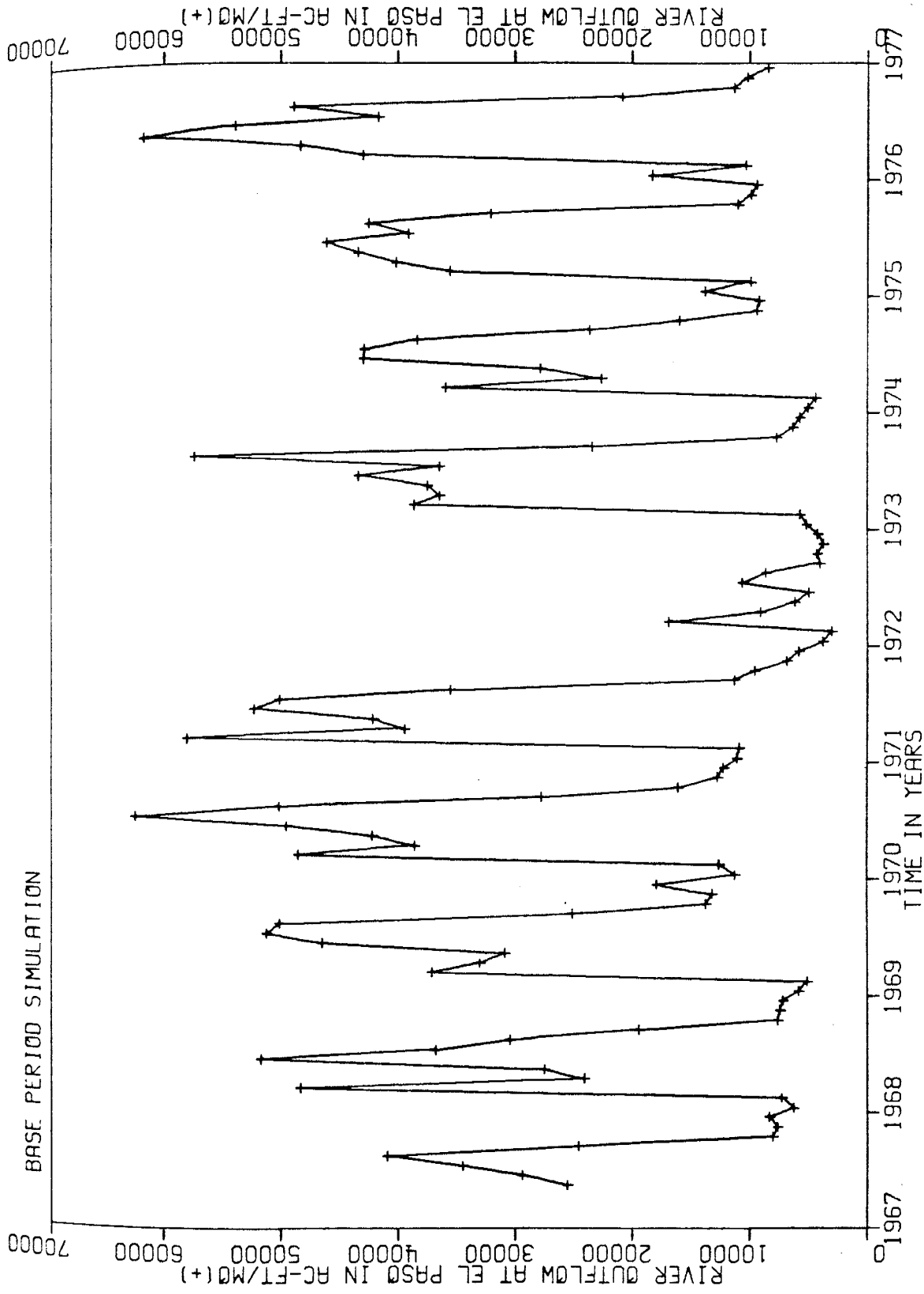


Figure 4.15: Base period simulation of average river outflow at El Paso, Texas, the Mesilla Valley outflow point.

irrigation efficiency (β). The groundwater pumped (q_p) from the valley for irrigation use is not metered; however, it may be estimated if we assume that groundwater will be pumped only after all of the surface diversions are utilized. Hence

$$q_p = q_{et}/\beta - (q_s - q_1) \quad (4.5.1a)$$

$$q_a = q_s - q_1 + q_p - q_{et} \quad (4.5.1b)$$

This last equation is written directly from Figure 4.6. The estimates of $q_1 = 0.45 q_s$ and $\beta = 0.50$ were used in this study (see Gelhar and McLin, 1979, for the complete data basis for these estimates). In view of these assumptions, (4.3.3) can be rewritten for each of the four cases mentioned above; in all cases q_r was assumed to remain constant for a given recharge-recession period and was unchanged from the original simulation. A constant q_r for a given recharge-recession period, which covers several Δt time increments, produced somewhat inconsistent results over a portion of the simulation period for several of the management options examined. To correct these inconsistent results, a given q_r was subsequently assumed to be distributed over individual Δt increments within the same recharge-recession period in proportion to the observed river inflow below the Leasburg dam for that same Δt increment. All results that follow have utilized this later approach; Appendix G summarizes the computer code and output results of these management simulations.

The effect of this q_r distribution over individual Δt increments slightly altered the resulting model predictions for average aquifer water levels, drain flow, aquifer concentration, river outflow, and river

water quality. That is, Figures 4.7 through 4.11 are slightly different than Figures 4.12 through 4.15, except for net aquifer recharge. These small differences can be computed from the listed output contained in Appendix E (the nonlinear simulation output utilizing the constant q_r terms for a given recharge-recession period), and Appendix F (the water management base period simulation output using the appropriate q_r term which is distributed over the corresponding recharge-recession period and identified as q_R for a given Δt increment; however, the sum of all q_R values over a given recharge-recession period still totals the q_r value computed earlier.

The first case again addressed the question of evaluating the hydrologic effects on the system assuming all water demands were met by groundwater pumpage. Such an extreme situation would probably never actually occur (that is, a ten year long period with no surface diversions) but it will serve to illustrate the effects of one rather extreme management option. Again it was assumed that both irrigated acreage and total applied water would not change, and that a sufficient number of wells existed to supply the irrigation demand entirely from groundwater pumpage. In other words, the system averaged stresses remained identical to the original simulation; however, the source and chemical quality of applied water was changed to reflect this increase in pumpage and decrease in applied surface waters. In equation form this is represented by substituting (4.5.1b) into (4.3.1), assuming that $S = n$, and simplifying to obtain

$$n \frac{dh}{dt} + q_d = E + q_R \quad (4.5.2)$$

where $E = q_s + q_n - q_{et}$ and where q_n was assumed to be zero since $q_n \ll q_s$ and $q_n \ll q_{et}$. The evapotranspiration term (q_{et}) was

previously determined (Gelhar and McLin, 1979) for the USBR-EPA simulation study using the Blaney-Criddle method. Therefore in order to simulate the effects of no surface water diversions for case (1), all we need to do is set q_s equal to zero in (4.5.2) and solve for h , using the initial h at time zero and (4.3.2) for q_d along with the parameters previously determined.

The mass balance equation corresponding to (4.5.2) does not change form since only the water source terms and the associated water quality alters. Thus (4.3.5) was used for this and all subsequent water management options that were simulated. Actually this mass balance equation is two separate equations corresponding to the sign on q_R . If q_R is positive we have river to aquifer water transfers and (4.3.5) simply becomes,

$$nh \frac{dc}{dt} + c(E+q_R) = c_S (q_S+q_R) + q_n c_n; q_R > 0 \quad (4.5.3a)$$

where $E = q_S+q_n-q_{et}$, and where $q_n \approx 0 \approx c_n$. If q_R is negative, we have aquifer to river water transfers and (4.3.5) becomes

$$nh \frac{dc}{dt} + cE = q_S c_S + q_n c_n; q_R < 0 \quad (4.5.3b)$$

where E is defined above. For case (1), q_s is set to zero in both (4.5.3a) and (4.5.3b). The results of this first case are presented in Figure 4.16 through 4.19; they indicated the following:

1. The water table elevation throughout the valley decreased by an average of 2.9 (ft.) during the first year, and 43.9 ft. during the tenth year. Corresponding to these declines was an average reduction in groundwater recharge of 0.183 feet per month (ft/mo) during the first

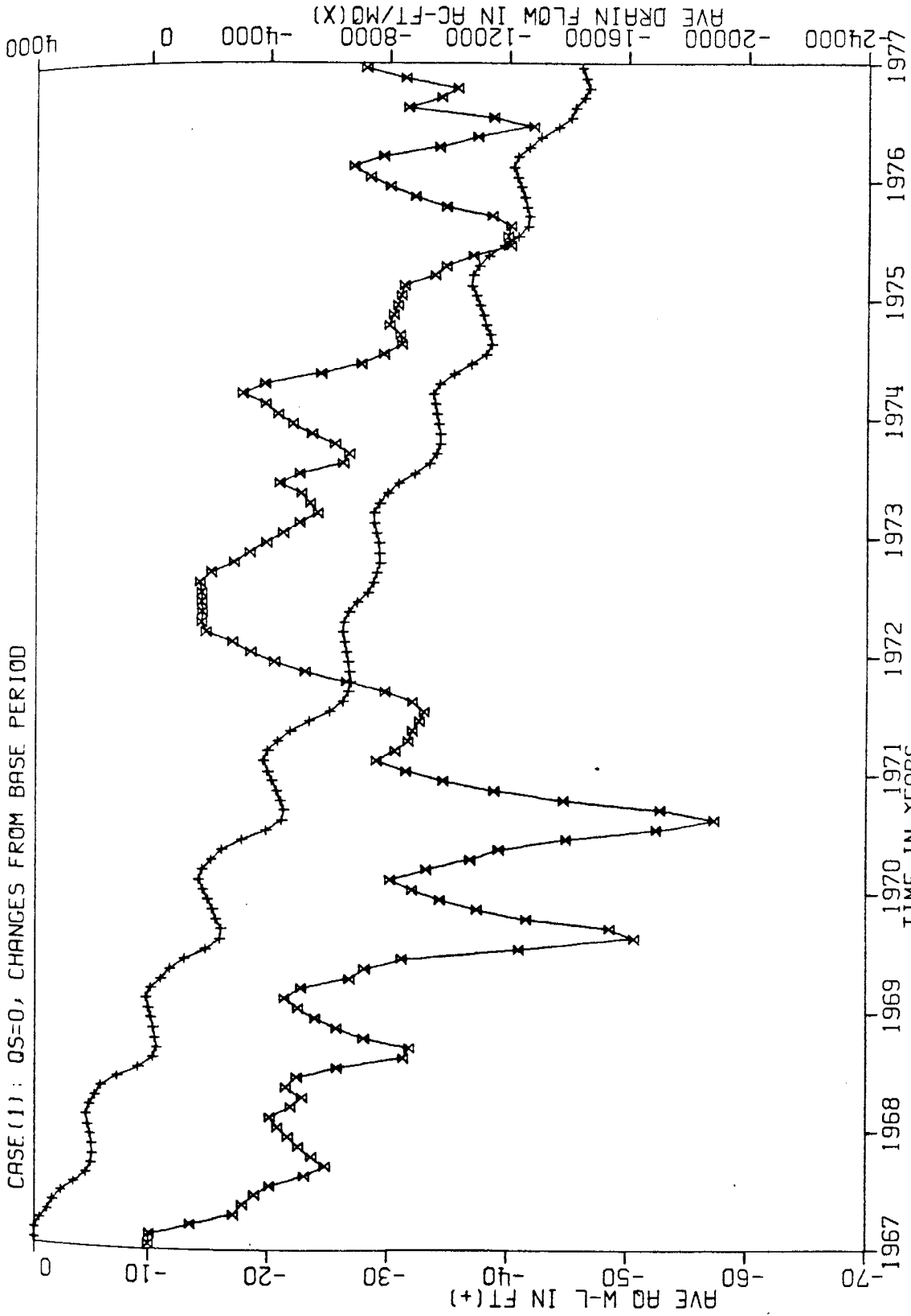


Figure 4.16 Case (1) changes from the base simulation in average aquifer water levels and valley drain flow.

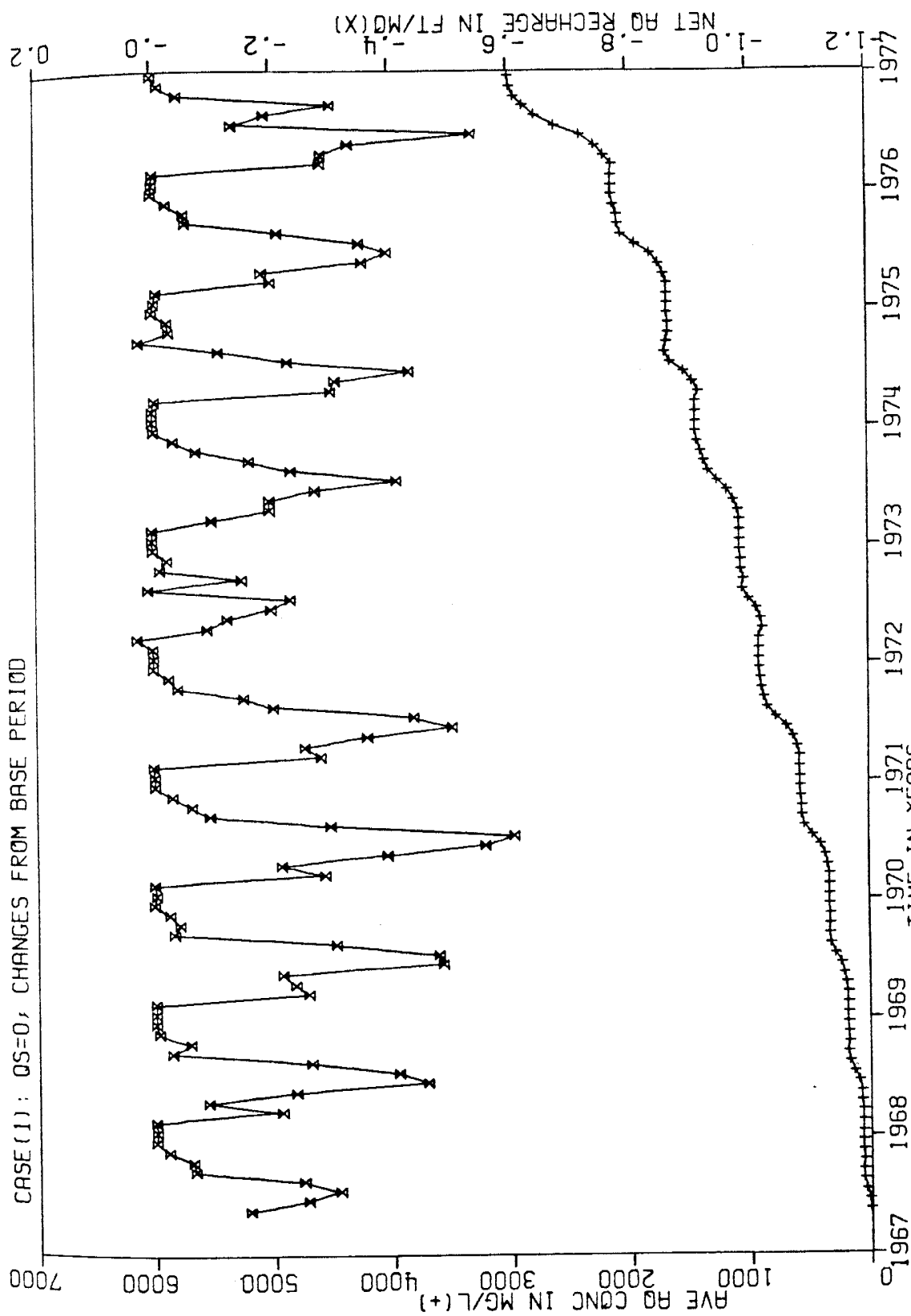


Figure 4.17 Case (1) changes from the base simulation in average aquifer TDS and net average aquifer recharge.

CASE(1): QS=0; CHANGES FROM BASE PERIOD

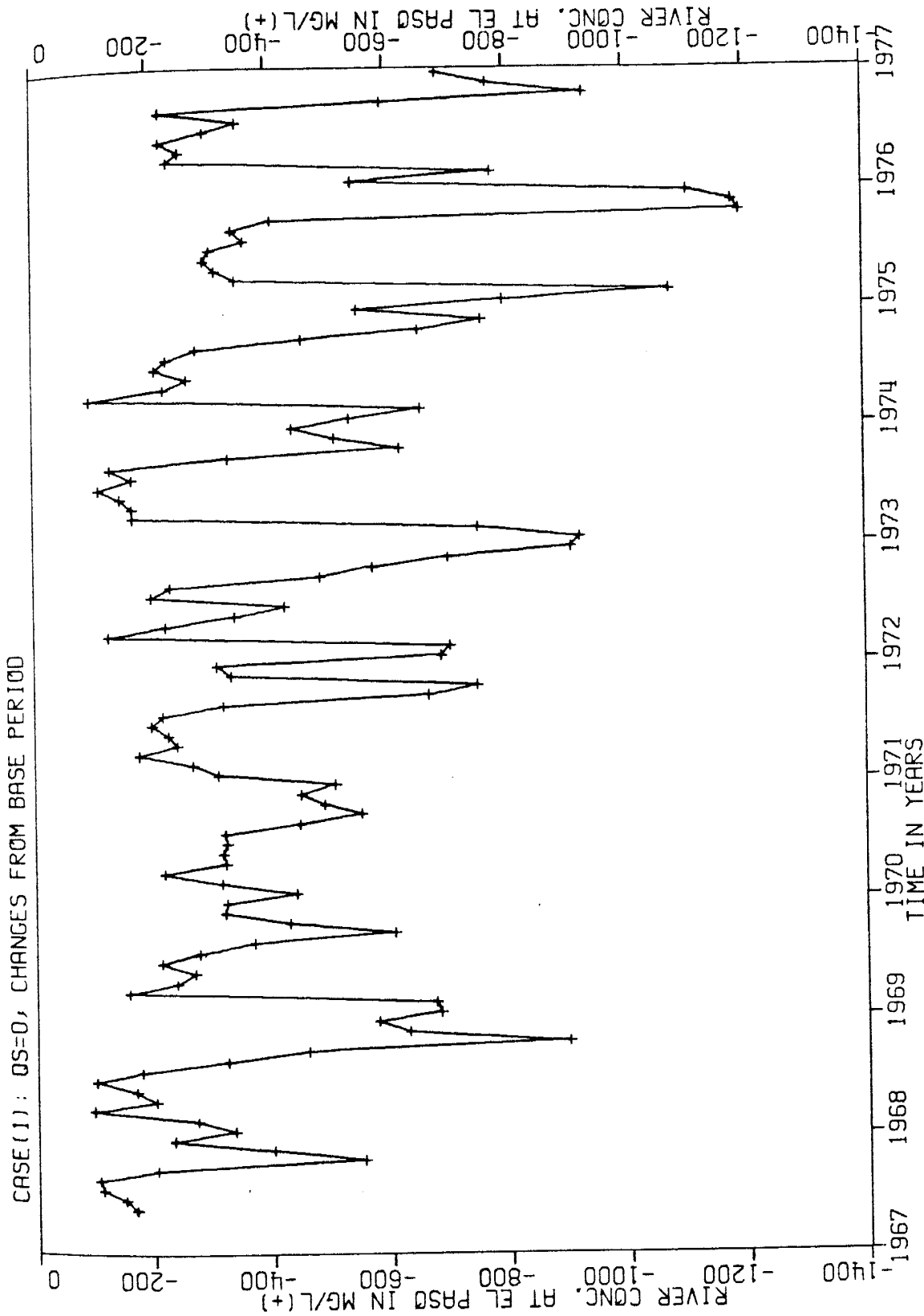


Figure 4.18 Case (1) changes from the base simulation in average river TDS at El Paso, Texas.

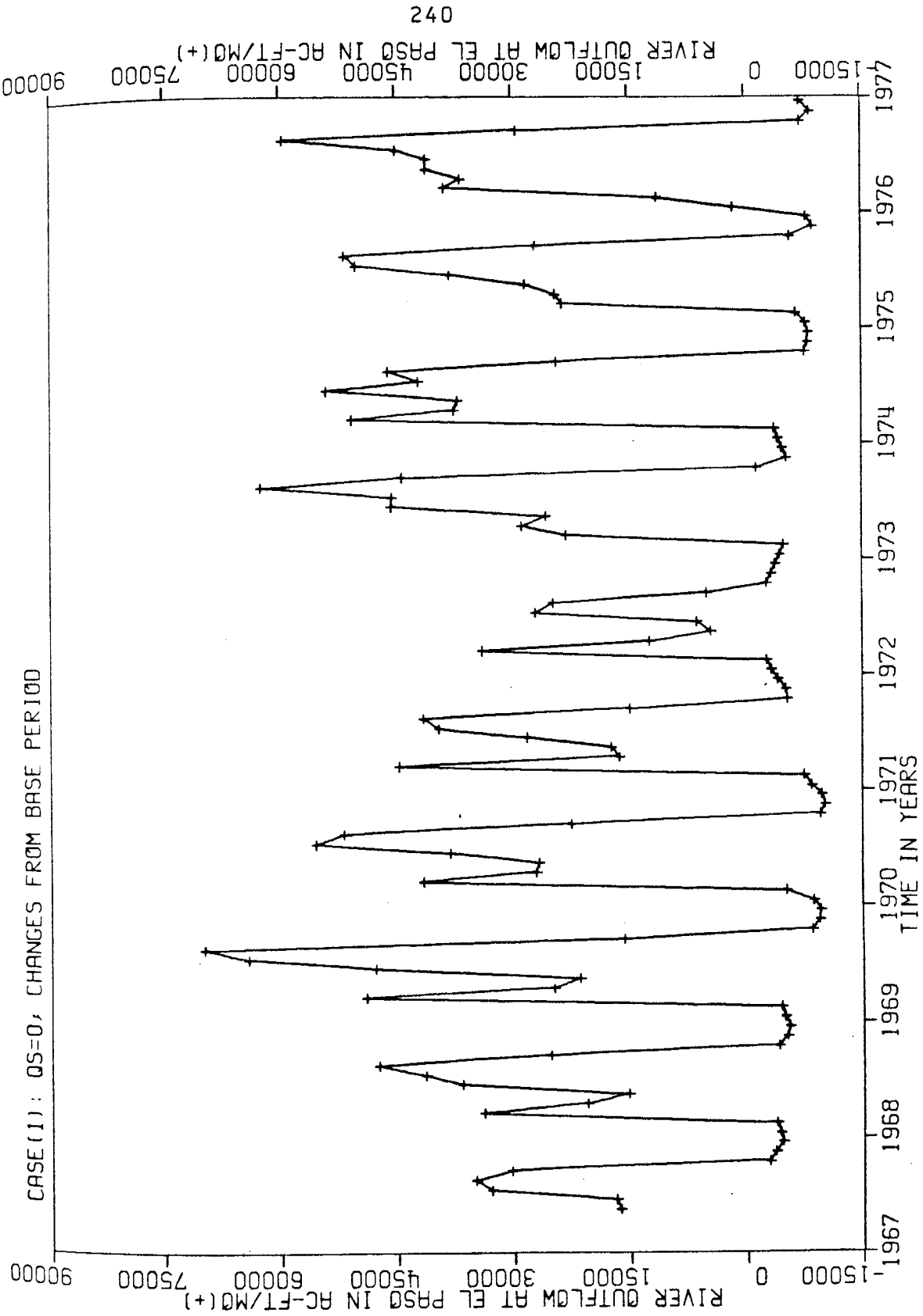


Figure 4.19 Case (1) changes from the base simulation in average river outflow at El Paso, Texas.

year, and 0.179 ft/mo during the tenth year; the average decrease over the ten year period was 0.160 ft/mo.

2. The groundwater quality showed an average TDS increase of 41 mg/l during the first year, and 2527 mg/l during the tenth year.

3. Streamflow gains in the Rio Grande at El Paso, Texas, resulting from decreased surface diversions, increased from an average of 14,967 acre-feet per month (ac-ft/mo) during the first year, to 23,351 ac-ft/mo during the tenth year; the ten year average increase was 17,609 ac-ft/mo. Likewise, there was a decrease in average drain flow of 3463 ac-ft/mo during the first year, and 9273 ac-ft/mo during the tenth year; the study period average drain flow decrease was 7284 ac-ft/mo. These drain flow decreases are obviously tied to the decrease in average aquifer water levels and recharge. In this case (1) all drain flow actually ceased after only a few months into the simulation.

4. The dissolved solids concentration in the Rio Grande at El Paso, Texas, decreased by an average of 243 mg/l during the first year, and 493 mg/l during the tenth year.

These results are graphically presented in Figures 4.16 to 4.19. While this extreme water management option does tend to improve downstream water flows and quality, these benefits are offset by drastic aquifer changes. These changes are a result of no surface water diversions into the valley, but rather simply conveying these supplies through the system. Furthermore, the dramatic water level declines essentially eliminate any drain flow, thereby reducing this source of low quality return flow back to the river. Instead these low quality waters remain in the valley aquifer to control the rapid decline in water quality there. Thus irrigators not only would have higher pumping costs

associated with lower water levels, but the water they would obtain would be of questionable long term value due to its poor quality. This management option does demonstrate, however, the "closed cell" effect on the aquifer mass balance. In other words all contaminant mass from groundwater and drain flow is confined to the valley aquifer, while these same groundwaters are continually recirculated through the system

The second test also addressed an extreme question: What would be the effect on the stream-aquifer system if all irrigation demand were supplied from surface diversions, and no irrigation wells were pumped? Again it was assumed that the overall system stresses remained identical to the base simulation period. Only the source of irrigation waters, and the associated quality, were allowed to change to reflect the new inflow surface diversions. In other words, the original surface diversions were increased by an amount equal to the computed irrigation pumpage, as defined in (4.5.1a). The irrigation pumpage was then set equal to zero. In this simulation an additional constraint was imposed; the procedure used in the constraint was first to check for sufficient river flow that would be available to meet the new surface diversion requirements. If not, the model allowed only a diversion equal to the amount of surface water that was physically observed to be present in the river; a crop water deficit would then occur for that time interval. This situation only rarely occurred in the ten year simulation however. Again all original system stresses remained identical to the original simulation, except for the increased surface diversions. While the mass balance equation did not change form; its predicted output changed to reflect increased use of the better quality surface waters. Figures 4.20 to 4.23 summarize these results; they indicated the following:

1. The water table elevation throughout the valley increased by an average of 0.7 ft. during the first year, and 3.2 ft. during the tenth year. Groundwater recharge also increased from 0.090 ft/mo during the first year, to 0.183 ft/mo during the tenth year; the ten year average increase was 0.146 ft/mo.

2. The shallow groundwater quality showed a TDS decrease of 48 mg/l during the year, and 777 mg/l during the tenth year.

3. Initially streamflow in the Rio Grande decreased by some 4238 ac-ft/mo during the first year, but somewhat surprisingly increased by an average of 13,279 ac-ft/mo during the tenth year. A ten year average increase of 7485 ac-ft/mo was predicted for this second test; these results may have been affected by the relatively large RMS error obtained in the original simulation of river outflow (see Figure 4.10). This initial streamflow reduction may also have resulted in part from the diversion and utilization of all available surface water necessary to meet crop requirements; however, the river flow decline may have been later offset in part by increases in drain flow since there was no aquifer pumpage. The first year average increase in this drain flow was 3467 ac-ft/mo. By the tenth year, the average drain flow increased by 19,963 ac-ft/mo thereby apparently adding greatly to the net increase in river flow. The ten year average increase was 15,179 ac-ft/mo.

4. The dissolved solids in the Rio Grande predictably increased by 394 mg/l during the first year, but somewhat surprisingly decreased by 226 mg/l during the tenth year.

While many water managers would never advocate such an extreme water policy, this technique of diverting all available surface water does evidently show some promise in the Mesilla Valley. This technique does

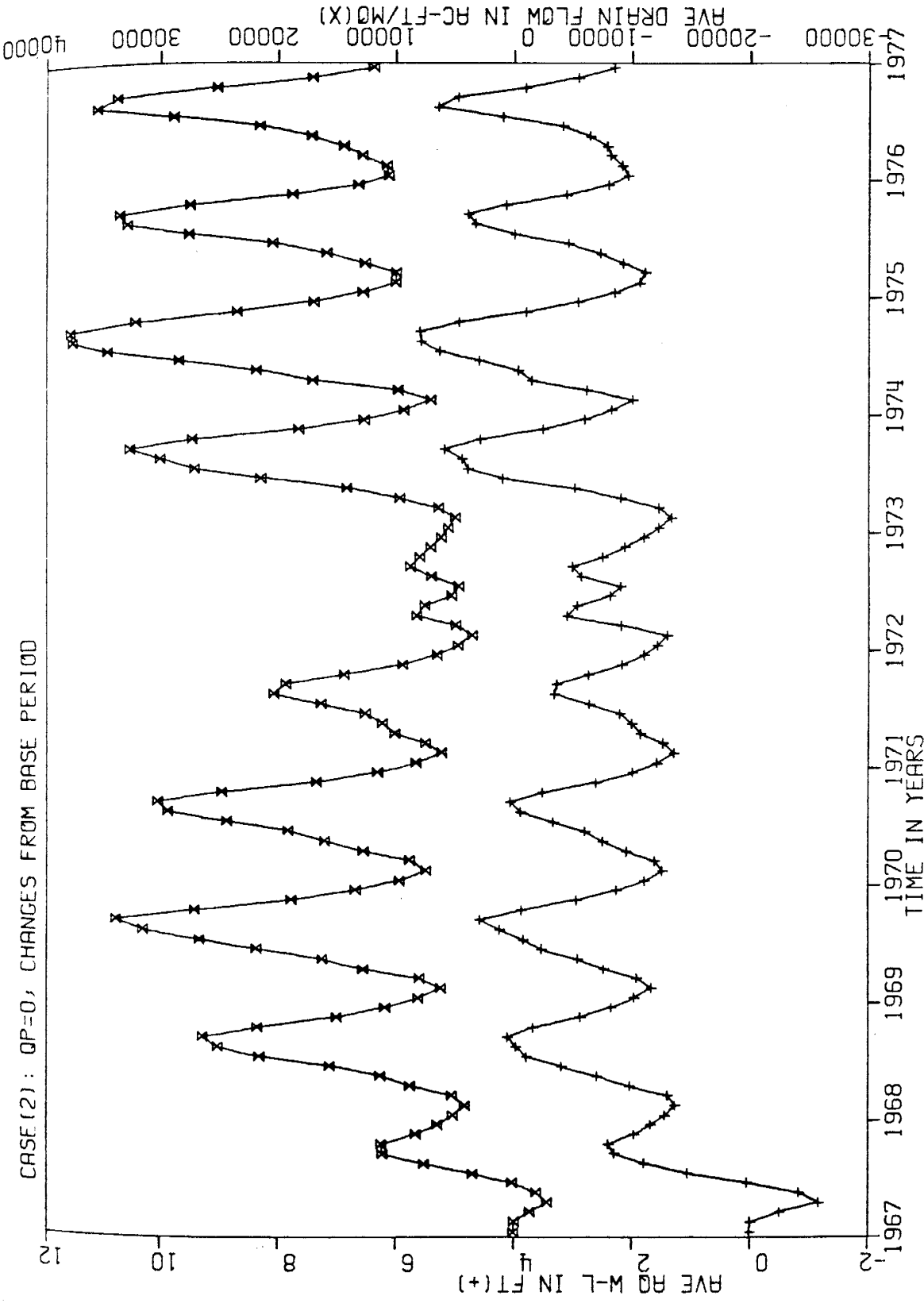


Figure 4.20 Case (2) changes from the base simulation in average aquifer water levels and average valley drain flow.

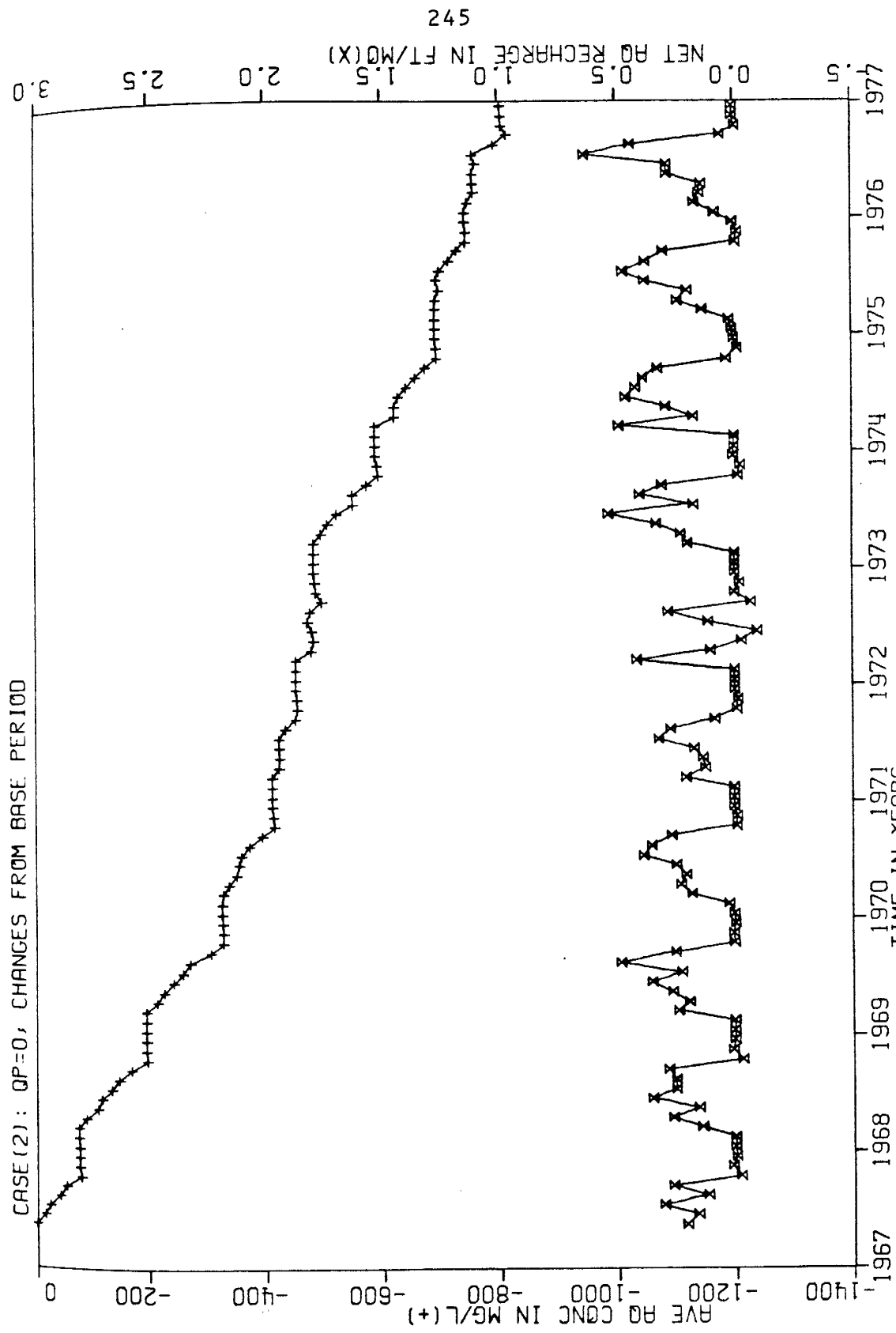
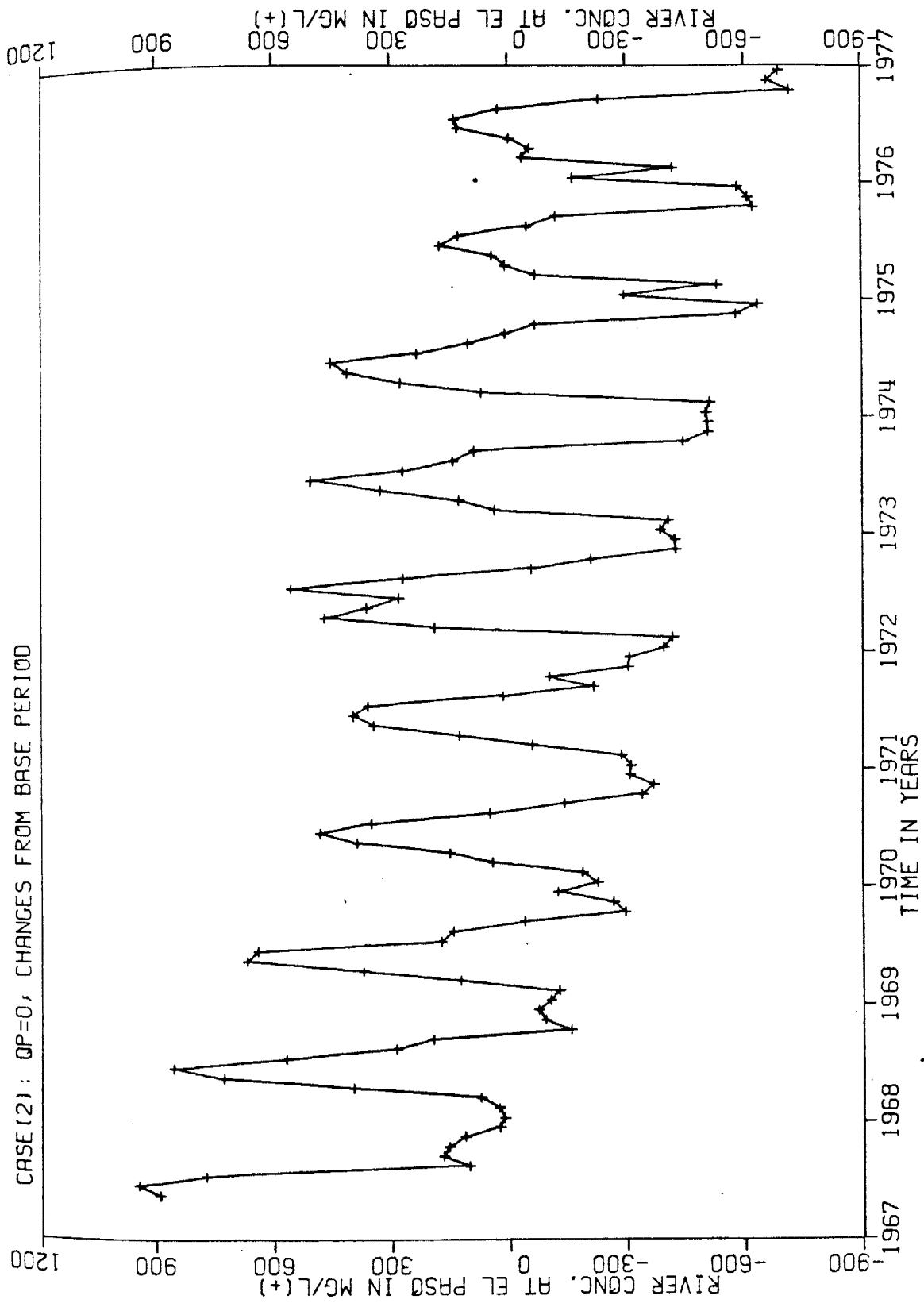


Figure 4.21 Case (2) changes from the base simulation in average aquifer TDS and net average aquifer recharge.



CASE (2): QP=0, CHANGES FROM BASE PERIOD

Figure 4.22 Case (2) changes from the base simulation in average river TDS at El Paso, Texas.

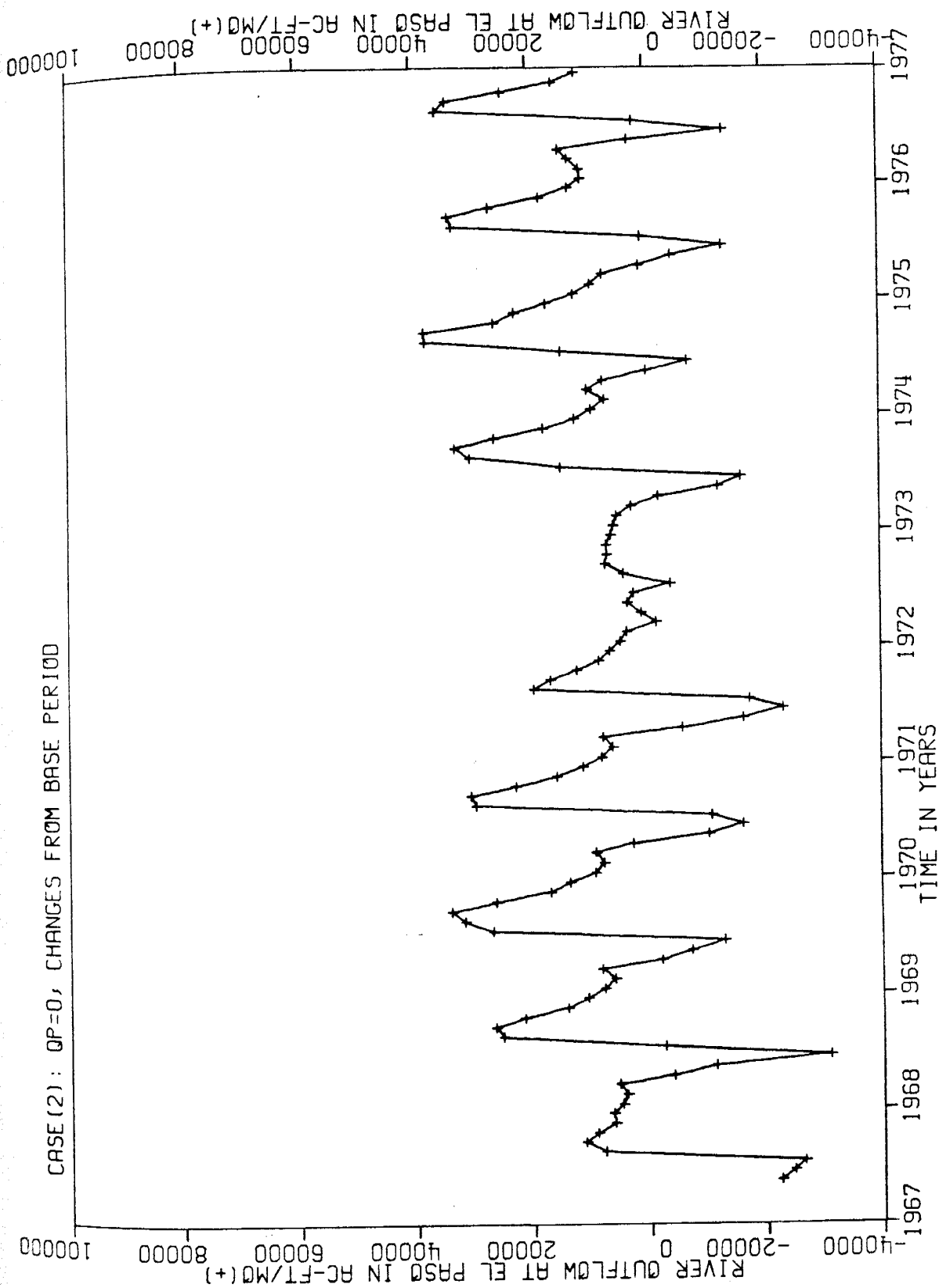


Figure 4.23 Case (2) changes from the base simulation in average river outflow at El Paso, Texas.

assume, however, that sufficient surface waters are available to meet all valley irrigation demands, as it was for most of this ten year period. While downstream users would initially receive lower quality water, the long term results apparently indicate an attractive decline in river contamination and shallow groundwaters. The major drawback to this water management option is rather predictable however. Downstream users would receive lower volumes of river water precisely during the growing season when their requirements are highest, and higher volumes during the off season. A second disadvantage is that the long term improvement in groundwater quality and aquifer water levels would only serve to improve drain flow quality, since no pumpage occurs. Thus for all practical purposes, these improvements are essentially wasted as far as valley irrigators are concerned. This case does imply something we already should know: that a shallow groundwater circulation system exists in the uppermost alluvial aquifer. Thus the water levels and quality fluctuate here according to the quantity and quality of applied waters, with some influence of attenuating contaminant migration through the aquifer.

The third test addressed the question: What would be the hydrologic effects on the stream-aquifer system as a result of improving irrigation efficiency by 25 percent (that is, 25 percent less irrigation water is required to produce the same crop yields from the same irrigated acreage)? Again as with the other management options examined, only the water balance equation needs to be modified. Thus in (4.3.3), we need estimate only changes in E resulting from changes in irrigation efficiency. The form of the mass balance equation does not change, but since the water balance changes, the mass output naturally alters.

We may proceed to estimate these irrigation efficiency related water balance changes by referring back to (4.5.1a), and the basic definitions contained there. Thus these changes may be computed by recalling that the irrigation demand is simply the total crop consumptive use (q_{et}) divided by the average system irrigation efficiency (β). If we assume that we can improve β over some given time interval, then the resulting water savings may be expressed as

$$W = q_{et} (\beta_2 - \beta_1) / \beta_1 \beta_2 \quad (4.5.4)$$

where W is the water savings, β_1 , is the original system irrigation efficiency, and β_2 is the improved efficiency. This net water savings to the system may be taken as either a reduction in groundwater pumpage (q_p) or surface diversion (q_s), or a combination of both. At first thought, it would seem logical to take advantage of these savings as reductions in q_p . If we do so we see that the individual components of net recharge ($E = q_s + q_n - q_{et}$) do not alter in view of (4.5.2); however, changes in soil-water moisture have not been taken into account. With this formulation, we may immediately conclude that any water savings resulting from improvements in irrigation efficiency which are taken in the form of reduced pumpage amount to no net system changes. This conclusion applies to both the water and mass balance relationships, given by (4.3.3) and (4.3.5), respectively, as will become more obvious below.

It is also obvious that when these same savings are taken as reductions in q_s , then system water and mass alterations do occur; naturally the maximum changes occur when all of the savings are in the

form of reduced q_s . We may express the range of these changes in q_s simply as,

$$q_{s2} = q_{s1} - \gamma W \quad (4.5.5)$$

where q_{s2} and q_{s1} are the new and old values for surface diversion corresponding to β_2 and β_1 , respectively, and γ is the fraction of the water savings taken as a reduction in q_s . At the improved β_2 value, the net aquifer recharge is simply $E_2 = q_{s1} + q_n - q_{et}$, where the natural recharge (q_n) and crop consumptive use (q_{et}) do not change. Therefore in light of equation (4.5.5), we see that

$$E_2 = E_1 - \gamma W \quad (4.5.6)$$

If no water savings are taken in the form of reduced q_s but rather all in q_p , then $\gamma = 0$ and $E_2 = E_1$. Likewise if no savings are taken as reductions in q_p but rather in q_s , then $\gamma = 1$ and maximum system changes occur.

In these simulations β_1 was taken as the original irrigation efficiency of 0.50, while β_2 was assumed to be 0.75; the resulting water savings is therefore given by (4.5.4). Since we already know that reductions in q_p result in no system changes, one limiting boundary is already known in the solution (see base period simulations presented earlier). The other boundary will occur at $\gamma = 1$, or when all water savings are taken as reductions in q_s . For $0 \leq \gamma \leq 1$, the net system changes will fall between these two limits. Hence in this simulation, we

CASE (3): IMPR IRR EFF, CHANGES FROM BASE PERIOD

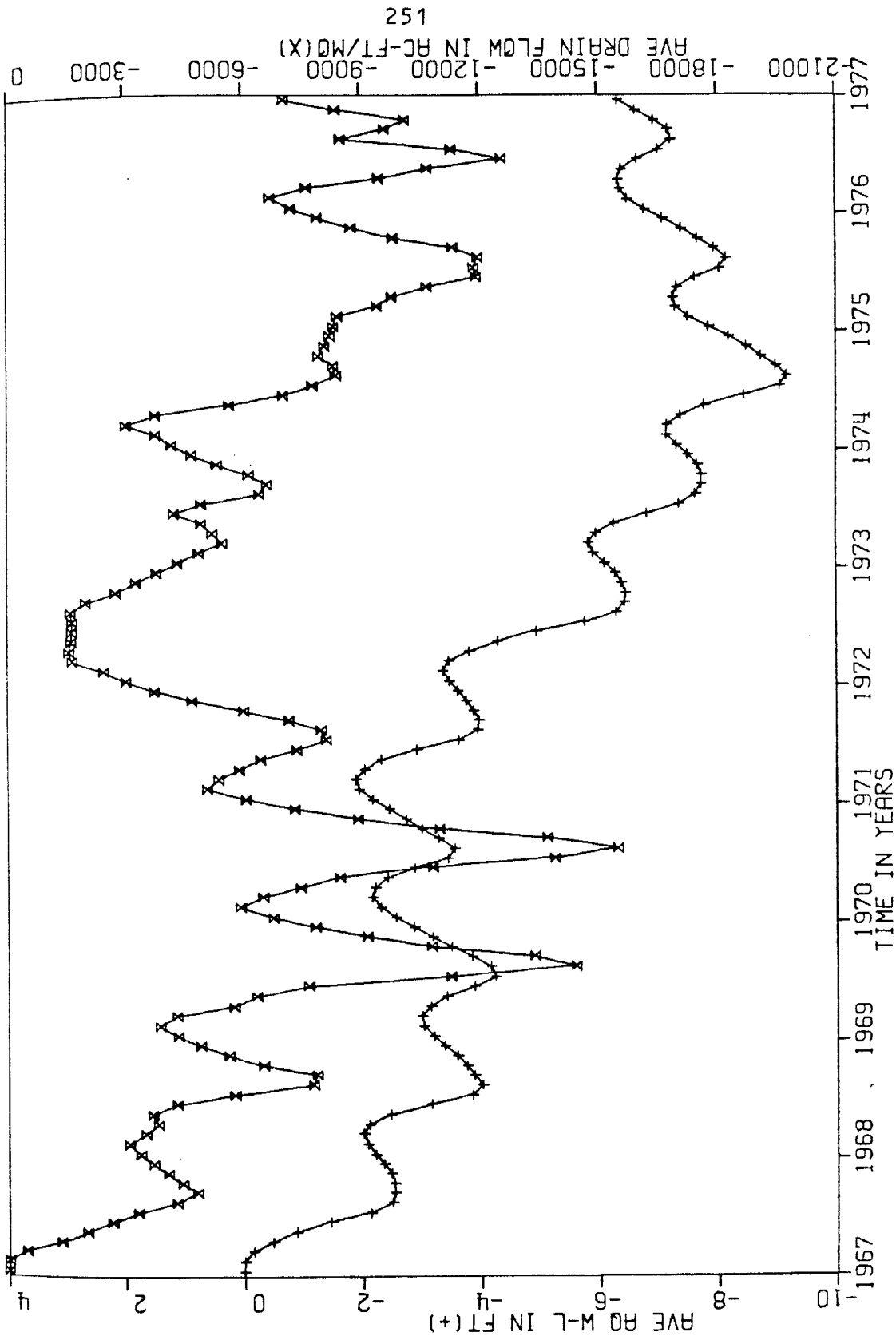


Figure 4.24 Case (3) changes from the base simulation in average aquifer water levels and valley drain flow.

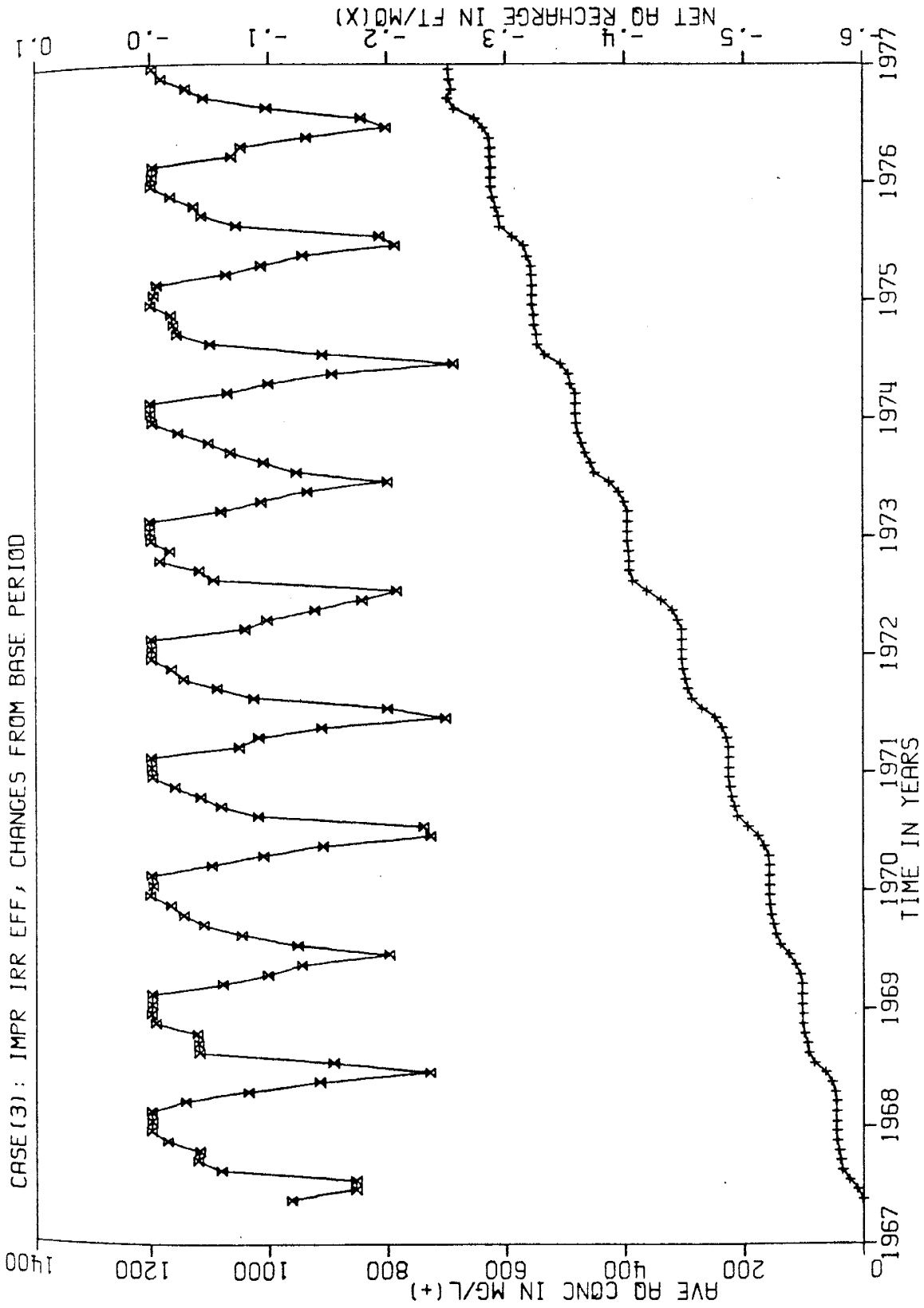


Figure 4.25 Case (3) changes from the base simulation in average aquifer TDS at El Paso, Texas.

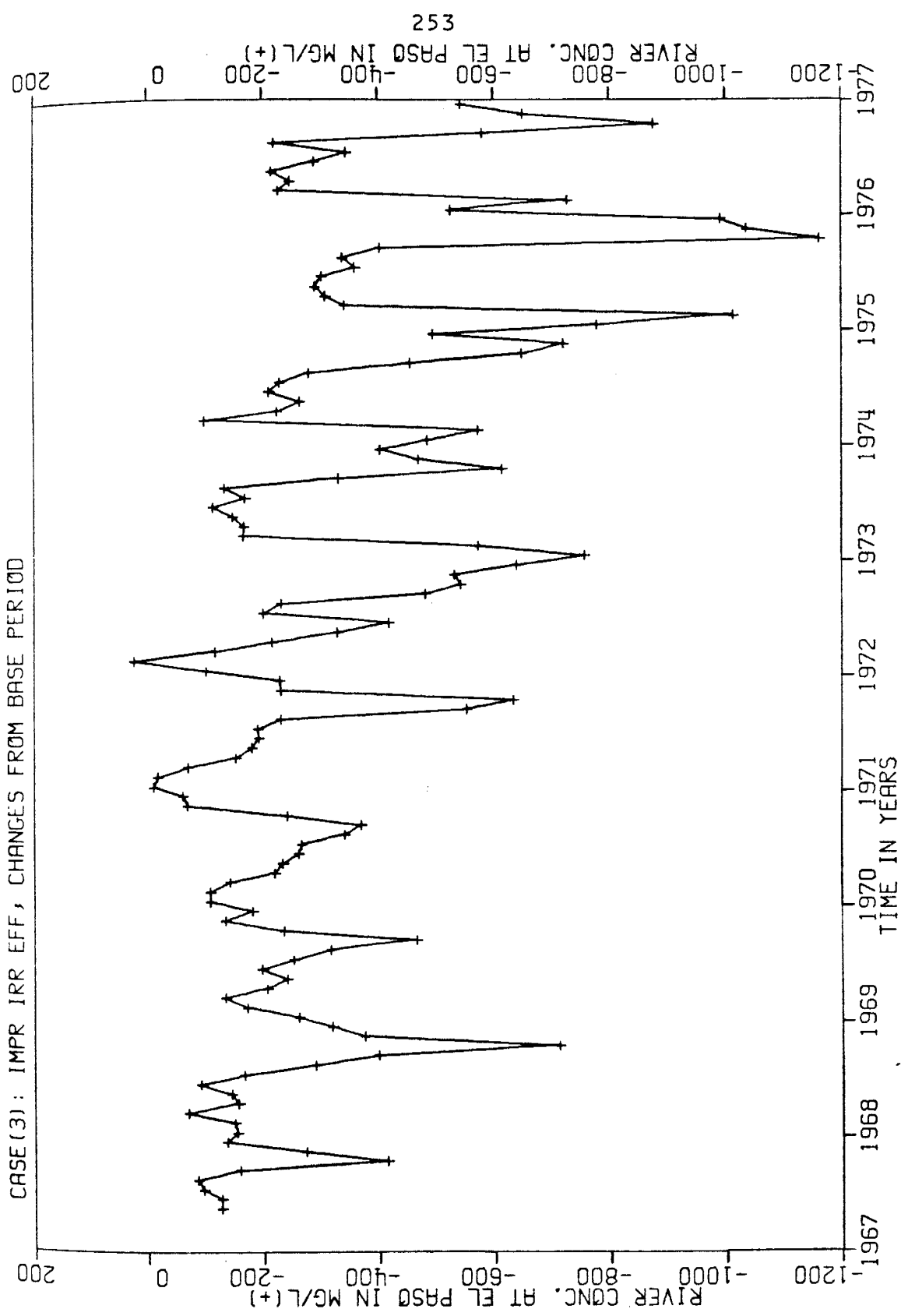


Figure 4.26 Case (3) changes from the base simulation in average river TDS at El Paso, Texas.

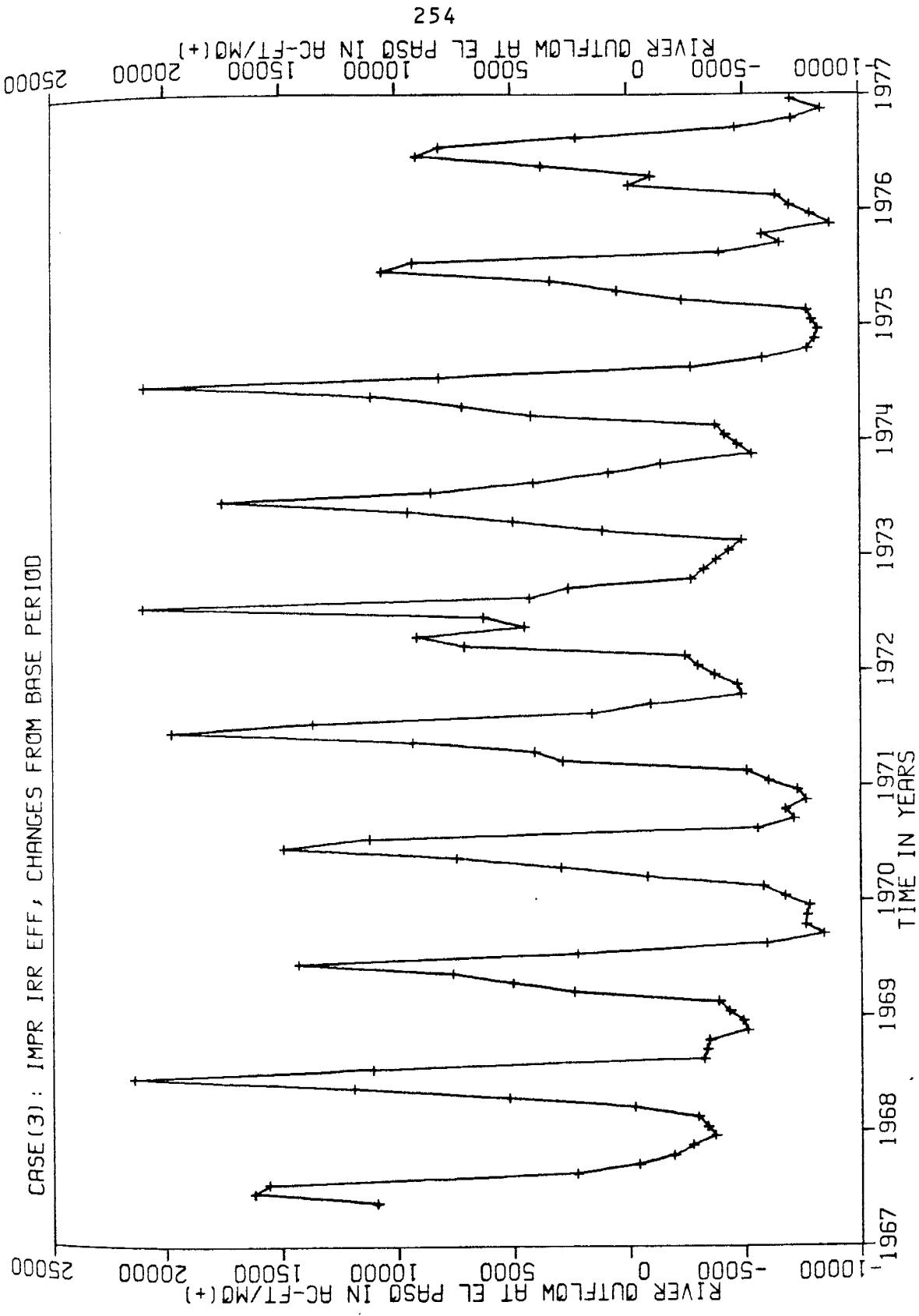


Figure 4.27 Case (3) changes from the base simulation in average river outflow at El Paso, Texas.

have assumed that $\gamma = 1$. These results are shown in Figures 4.24 through 4.27, and indicated the following:

1. The water table elevation throughout the valley decreased by an average of 1.5 ft. during the first year, and 6.7 ft. during the tenth season. These water level declines were also reflected in reduced aquifer recharge. During the first year, recharge declined by 0.099 ft/mo; whereas, during the tenth season these declines averaged 0.071 ft/mo. The ten year average decrease was 0.074 ft/mo.

2. The groundwater quality showed an average TDS increase of 28 mg/l during the first year, and 637 mg/l during the tenth year.

3. Streamflow fluctuations in the Rio Grande at El Paso, Texas, resulting from decreased surface diversions, increased by an average of 4788 ac-ft/mo during the first year, but decreased by an average of 1467 ac-ft/mo during the tenth year. A ten year average increase of 1252 ac-ft/mo was predicted by this option. As would be expected, there was a decrease in average drainflow of 2618 ac-ft/mo during the first year (or 49 percent of that originally available), and 9127 ac-ft/mo during the tenth year (or 98 percent of that available); the study period average drainflow decrease was 6492 ac-ft/mo (or 87 percent of that originally available). Obviously these reductions in drainflow are related to declines in aquifer water levels and recharge, and also influence the total river outflow volume.

4. The dissolved solids concentration in the Rio Grande at El Paso, Texas, decreased by an average of 186 mg/l during the first year, and 461 mg/l during the tenth year.

It is important to recall the conditions which led to the above conclusions. The water savings resulting from improvements in irrigation

efficiency were taken only as reductions in surface diversions. If these same savings were taken as reductions in groundwater pumpage, then no changes from the original base simulation would occur. Some disadvantages of improved irrigation efficiency are obvious, especially when the resulting water savings are taken as reductions in surface diversions. Average aquifer water quality tends to suffer because of reductions in artificial aquifer recharge and drain flow; these reductions tend to restrict subsurface contaminant circulation by confining the dissolved solids in the shallow aquifer. Benefits are seen primarily by downstream users; these include improved downstream surface water quality and initial increases in river flow. Of course valley users see the benefit of reduced crop water requirements and pumpage if water savings are taken here. The downstream improvements are the direct result of slowly eliminating drain flow back to the river. However, the river volume also declines as drain flow is reduced, but this is partially offset by reduced surface diversion if water savings are applied here.

Some advocates of improving irrigation efficiency maintain that declines in aquifer water quality can also be eliminated, thereby making this option even more attractive. This certainly appears to be true in this work, but only if water savings are taken in the form of reduced pumpage. The proponents of this option maintain that the unsaturated zone below the crop roots can be utilized to store infiltrating contamination. However, the model used here does not directly address this question. Whether this contaminant storage in the unsaturated zone would be temporary or permanent remains unclear. Certainly if uncontrolled water applications (i.e., via excess rainfall or system

mismanagement) enter the unsaturated zone in sufficient quantity, then some of the stored contamination would enter the aquifer. Evidently the best option, according to the results obtained here for the Mesilla Valley, would include taking the water savings resulting from an improved irrigation efficiency as reductions in groundwater pumpage.

The fourth and final case addressed the following question: What would be the hydrological and chemical effects on the stream-aquifer system as a result of lining all conveyance canals in the Mesilla Valley to prevent leakage losses. As pointed out earlier, this technique really is a method of improving irrigation efficiency, but it was decided to explore these effects separately. As was the situation in the three previous cases examined, only the water balance equation needs to be modified; however, the mass balance equation naturally alters since water input is changed. The results of this simulation are shown in Figures 4.28 through 4.31, and indicated the following:

1. The water table elevation throughout the valley decreased by an average of 0.5 ft. during the first year, and 0.2 ft. during the tenth season. These water level declines were also reflected in reduced aquifer recharge. During the first year, recharge declined by 0.037 ft/mo; whereas during the tenth season, they averaged 0.014 ft/mo. The ten year average decrease was 0.026 ft/mo.
2. The groundwater quality showed an average TDS increase of 9 mg/l during the first year, and a decrease of 72 mg/l during the tenth year.
3. Streamflow fluctuations in the Rio Grande at El Paso, Texas, resulting from increase surface flows, increased 7,001 ac-ft/mo during the first year, and 13,603 ac-ft/mo during the tenth season. The ten year average increase was 8,626 ac-ft/mo. As would be expected, there

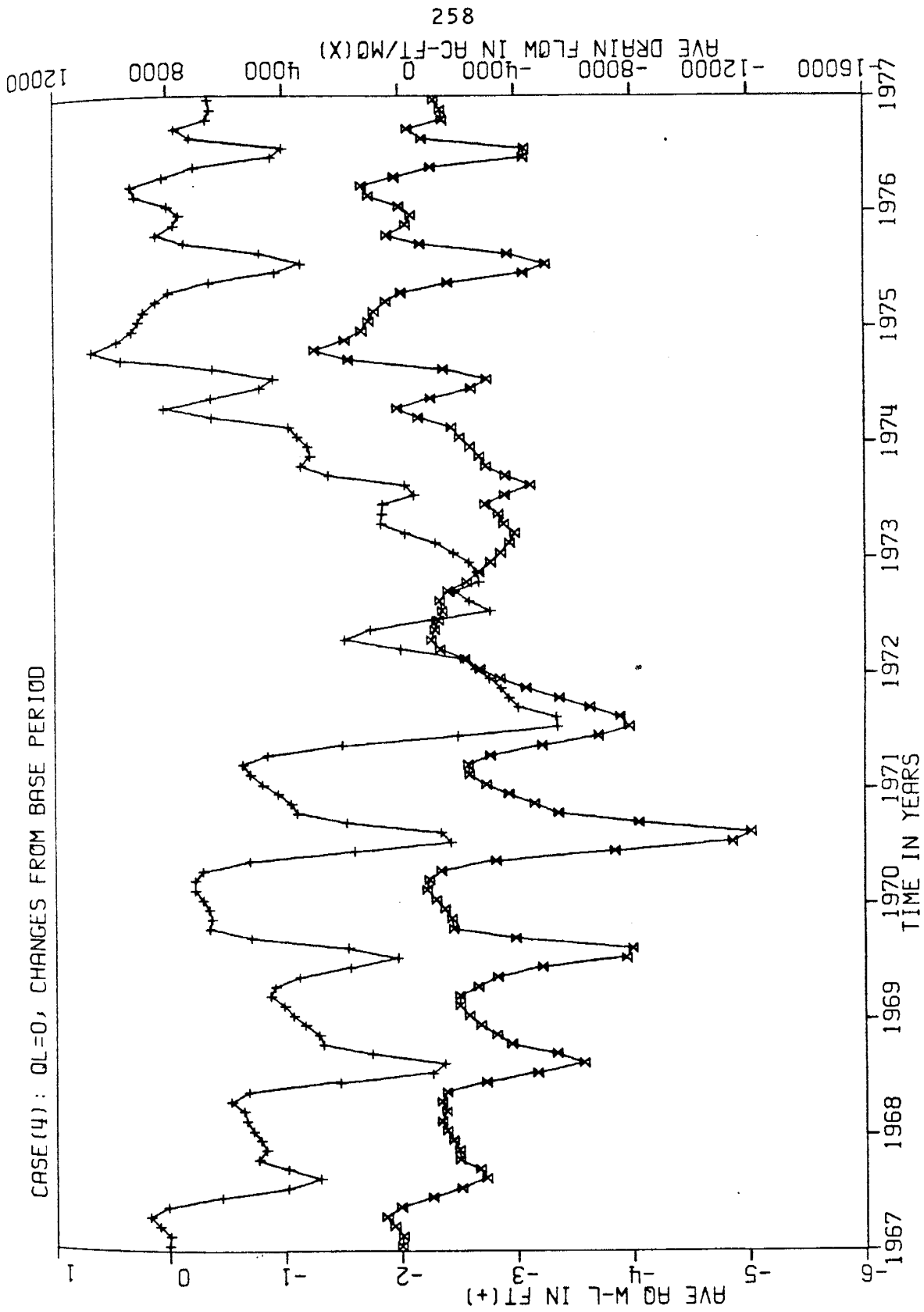


Figure 4.28 Case (4) changes from the base simulation in average aquifer water levels and valley drain flow.

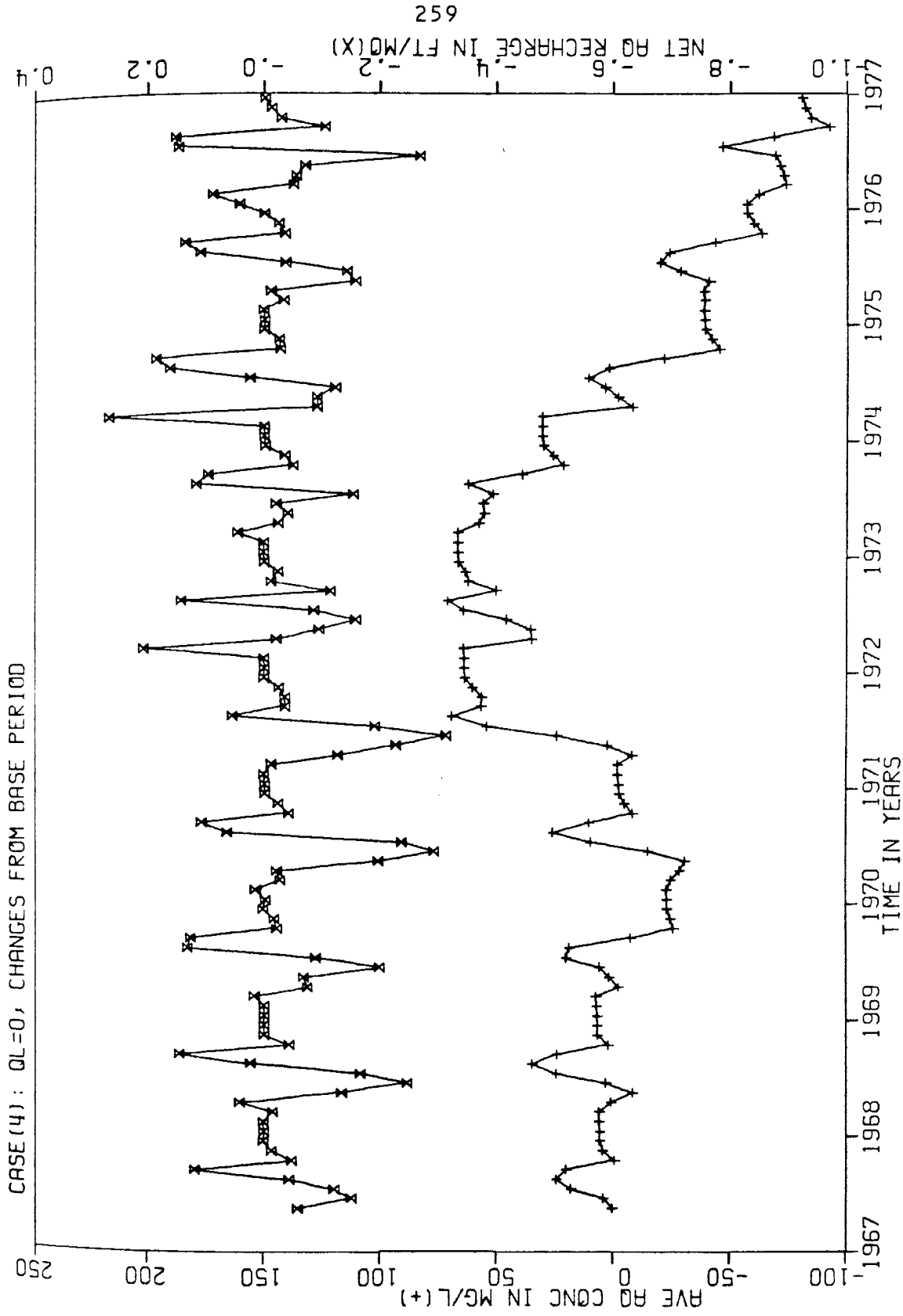


Figure 4.29 Case (4) changes from the base simulation in average aquifer TDS and net average aquifer recharge.

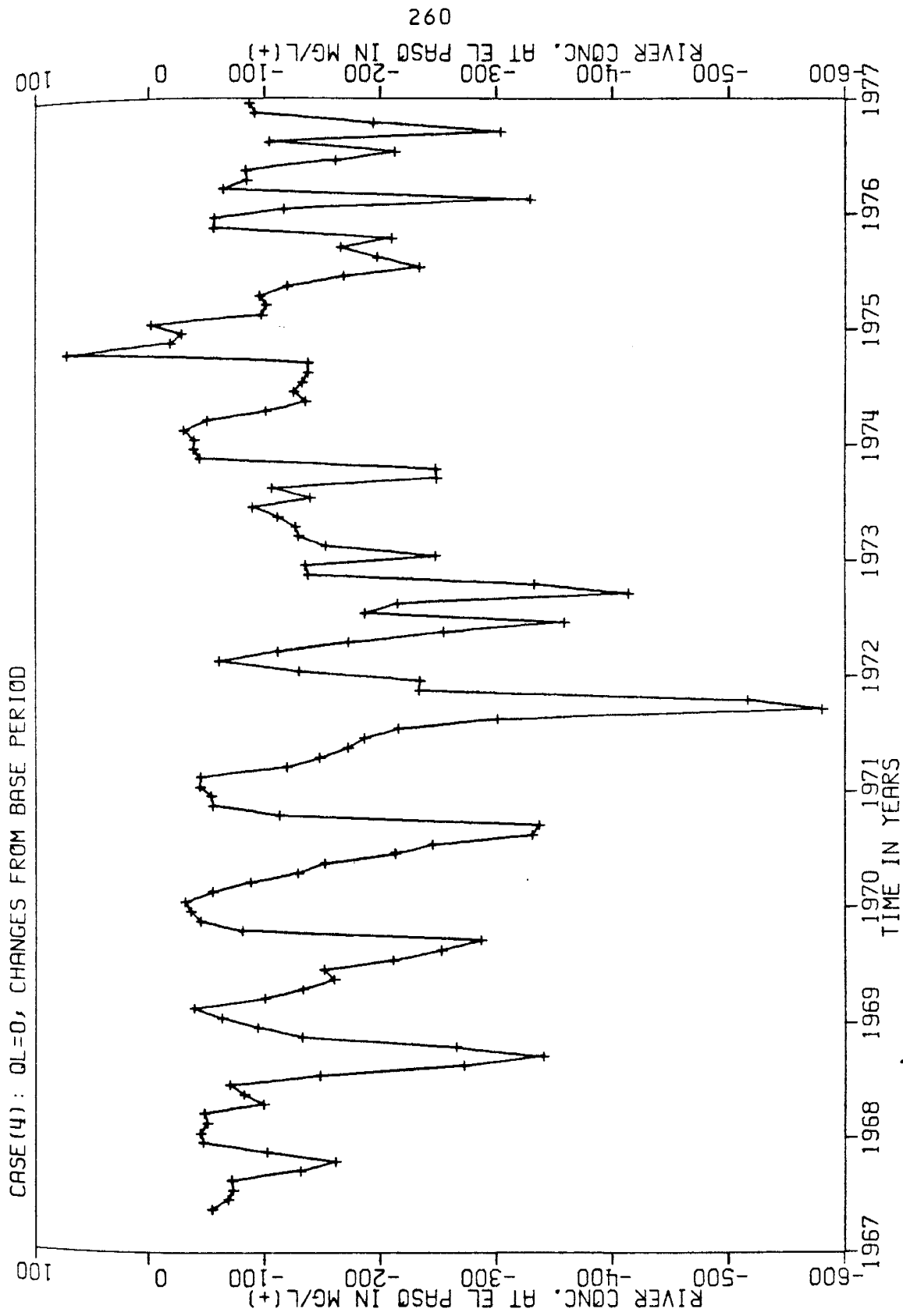


Figure 4.30 Case (4) changes from the base simulation in average river TDS at El Paso, Texas.

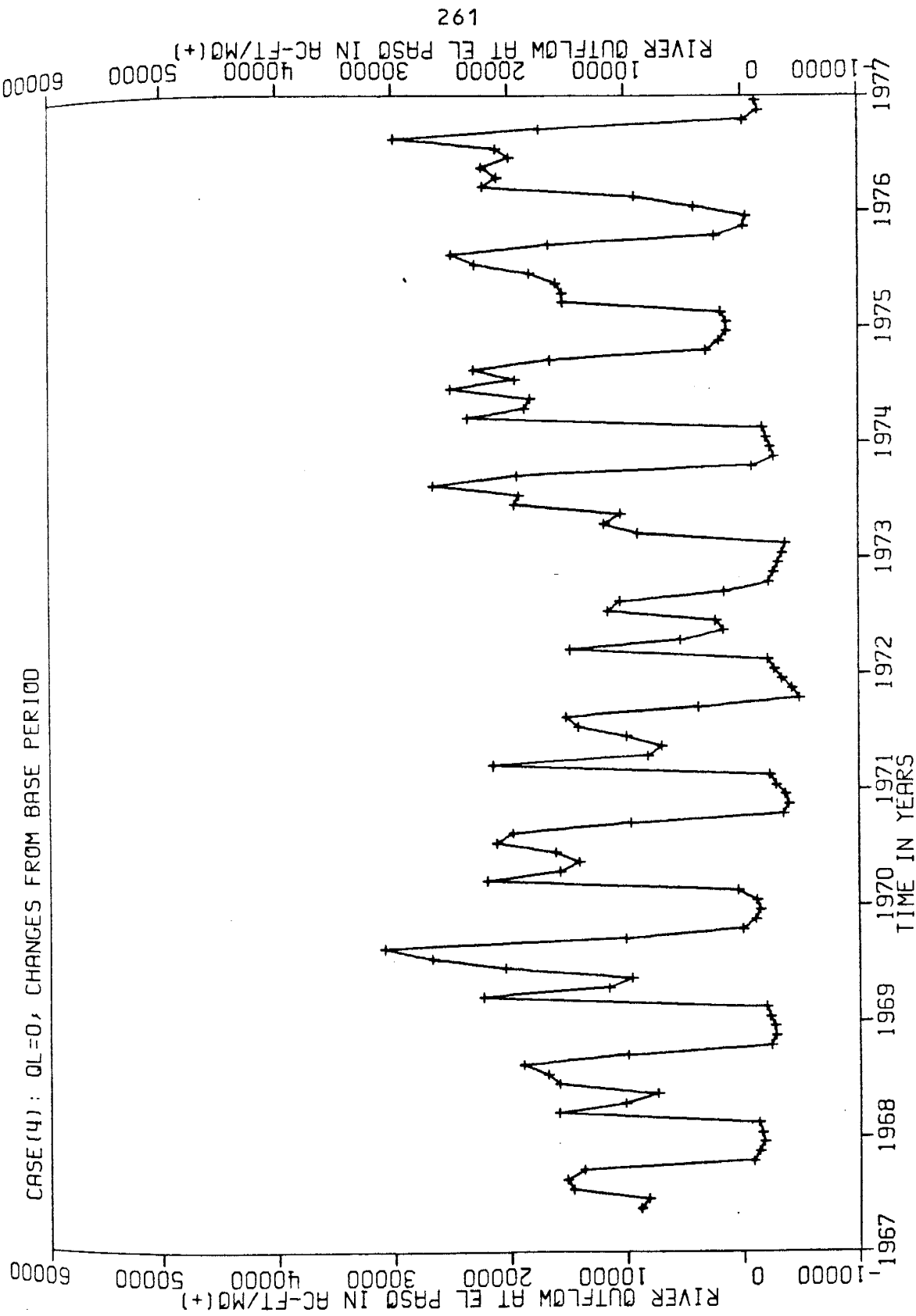


Figure 4.31 Case (4) changes from the base simulation in average river outflow at El Paso, Texas.

was a decrease in average drain flow of 1,156 ac-ft/mo during the first year (or 21 percent of that originally available) and 1,078 ac-ft/mo during the tenth year (or 12 percent of that available); the study period average drain flow decrease was 2,562 ac-ft/mo (or 34 percent of that originally available). Obviously these reductions in drain flow are related to declines in aquifer water levels and recharge, and have some impact on total river outflow volume.

4. The dissolved solids concentration in the Rio Grande at El Paso, Texas, decreased by an average of 90 mg/l during the first year and 153 mg/l during the tenth year.

The reduction in average aquifer TDS after 1973 (see Figure 4.29) came as somewhat of a surprise. However, an explanation may be found in the fact that there was a reduction in groundwater pumpage with a corresponding increase in surface water diversion after 1973 for the original base period simulation (see Figure 4.28 for the corresponding aquifer water level rise after 1973). This simulation indicates that many overall beneficial system changes could reasonably be expected to occur by lining conveyance canals in the Mesilla Valley. However, no cost-benefit analyses were performed for this or any other hypothetical water management scheme envisioned here. Such analyses would obviously need to be performed before any specific recommendations could be made. Results from all four cases are summarized in Table 4.5.

4.6 Conclusions

The results of the ten year simulation in the Mesilla Valley demonstrate that both the USBR-EPA hydrosalinity and nonlinear lumped

Table 4.5: Summary of water management options simulated, showing changes from base period simulation.

	Average aquifer water level (ft)	Average Drain flow (ac-ft/month)	Average aquifer TDS (mg/l)	Average aquifer recharge (ft/month)	Average River TDS at El Paso (mg/l)	Average River flow at El Paso (ac-ft/month)
Original Base Simulation	3824.16 3825.20 3824.62	5393 9273 7486	1648 1706 1681	0.0351 0.0735 0.0453	1074 1017 1105	22,289 31,357 24,608
Case (1) No Surface Diversion	-2.90 -43.93 -24.52	-3463 -9273 -7284	+41 +2527 +977	-0.1830 -0.1793 -0.1600	-243 -493 -412	+14,967 +23,351 +17,609
Case (2) No GW Pumpage	+0.72 +3.18 +2.70	+3467 +19,963 +15,179	-48 -777 -439	+0.0895 +0.1826 +0.1464	+394 -226 +12	-4,238 +13,279 +7,485
Case (3) Improved Irr. Eff.	-1.47 -6.71 -4.83	-2618 -9127 -6492	+28 +637 +313	-0.0986 -0.0705 -0.0743	-186 -461 -334	+4,788 -1,467 +1,252
Case (4) Lining Canals	-0.50 -0.24 -1.09	-1156 -1078 -2562	+9 -72 +3	-0.0369 -0.0141 -0.0255	-90 -153 -145	+7,001 +13,603 +8,626

parameter models can simulate pre-existing conditions quite well if adequate data are available to describe the overall water and mass balance conditions of the aquifer. The experience with the Mesilla Valley system indicates that the data base is reliable and consistent. However, groundwater quality does show a high degree of horizontal and vertical spatial variability; furthermore, the vertical extent of this higher saline groundwater is not well defined. The actual predictive capabilities of the USBR-EPA model are less clear than for the nonlinear lumped parameter model. In order to predict the hydrochemical effects of changes in water application, one must be able to synthesize the resulting changes of flow into and out of the system, because the USBR-EPA model completely ignores aquifer flow dynamics. In other words, the aquifer hydraulic coefficients and changes in flow resulting from changes in water level are not used.

The nonlinear lumped parameter model, on the other hand, does incorporate the mixing cell structure plus the flow dynamics of the aquifer. This model can also be operated with a conceptually much simpler computer program; calculations can even be carried out on a programmable pocket calculator, though somewhat inconveniently for long simulation periods. The nonlinear formulation utilized here has the advantage of avoiding the calculation or measurement of evapotranspiration and groundwater pumpage; these terms probably introduce the greatest sources of error into the typical water balance approach of systems modeling. Furthermore, the nonlinear representation appears to more accurately simulate low drain flow periods than does the linear approach. Finally, the lumped parameter model, in both the linear and nonlinear forms, may be easily utilized in evaluating various system

water management options once certain physical parameters are found which characterize the system response. However in such simulations, the individual components comprising the water balance equation must be estimated; hence, errors which were previously avoided may be inadvertently introduced into these management simulations.

The major disadvantage inherent to the lumped parameter format is that it does not provide a system wide spatial distribution for simulated aquifer water levels and contamination, nor does this model consider hydrodynamic dispersion effects.

CHAPTER 5

SUMMARY, CONCLUSIONS AND RECOMMENDATIONS

This work has considered the problem of developing a simple method of predicting future changes in downstream water quality resulting from various alternative management practices in irrigated agriculture. The technique employed here is based on the generalized lumped parameter model which utilizes either a linear or nonlinear subsurface reservoir. Since this method represents the system only in a spatially averaged sense, a minimum of input data is required; however, several complicating hydraulic and mass transport phenomena are not explicitly incorporated into this lumped approach. Hence, a major facet of this work has attempted to clarify some of these differences between the lumped and spatially distributed models. These considerations included:

1. Vertical flow effects on the system outflow concentration history resulting from fully or partially penetrating streams under steady, saturated, convective transport flow conditions, assuming steady, uniformly distributed, conservative contaminant inputs;
2. Extending the approach of Step (1) to include nonhomogeneous anisotropic aquifers with and without an impermeable basement at some finite depth;
3. Incorporation of pronounced vertical flow effects into the lumped parameter format utilizing simple parallel and/or series configurations; and
4. Field demonstrations of the predictive capabilities of the linear and nonlinear lumped parameter model where direct comparisons to previous modeling efforts and observed data are possible.

The following conclusions can be drawn from this research:

1. A hydraulic and mass equivalence between the linear lumped parameter and linearized Dupuit models was established for a homogeneous, isotropic aquifer system. This analysis produced identical convective outflow concentration histories for each model; furthermore, the instantaneous average concentration in the two-dimensional Dupuit aquifer model analysis with a fully penetrating stream was shown to be identical to this system outflow concentration. Hence, the well-mixed reservoir assumption inherent to the lumped approach is justified from a distributed systems approach. The hydraulic and mass similitude was extended to include multiple aquifer systems utilizing several lumped models in a parallel configuration, and a multi-layered Dupuit model analysis. These results indicated that the total mass outflow from each multiple system was identical, while the respective concentration outflow histories were approximately equivalent.

2. Vertical flow effects on convective outflow concentration resulting from partial stream penetration were examined from a two-dimensional Laplacian point of view. These aquifer geometry effects were combined into an aspect ratio defined as $\eta = L/D (K_z/K_x)^{1/2}$; thus incorporates the drain half-spacing (L), the aquifer depth (D) and the vertical to horizontal hydraulic conductivity contrast (K_z and K_x , respectively). Furthermore, when the impermeable Laplacian basement was taken at some infinite depth below the stream or tile drain, the resulting concentration break-through curve was nearly identical to that for $\eta = 1$. For values of η greater than about five (and hence, most practical applications), the lumped parameter model approximates the

Laplacian concentration break-through curves quite well. When n is less than five, however, pronounced vertical flow effects cause significant differences in outflow histories. Vertical flow effects resulting from horizontal aquifer stratification were also shown to have pronounced effects on the Laplacian outflow history. Both of these effects were approximately incorporated into the combined lumped parameter parallel configuration by utilizing two empirical correction factors. Thus, even the combined vertical flow effects resulting from aquifer layering and partial stream penetration were successfully incorporated into the seemingly crude lumped parameter format with surprisingly good results.

3. The lumped parameter model in the form of a linear reservoir was field tested in an irrigated area along the Arkansas River in southeastern Colorado over a one-year period. Results favorably compared with observed data and with a more complex convective-dispersive model (Konikow and Bredehoeft, 1974a, 1974b) previously tested in the same area under similar conditions. A second application of the model with a nonlinear phreatic aquifer was tested in the Mesilla Valley of southcentral New Mexico over a ten year period. Results showed good to excellent agreement with observed data and with a model (USBR, 1977) previously tested (Gelhar and McLin, 1979) in the same area under similar conditions. Both of these field applications have also demonstrated that the lumped modeling technique can easily be utilized to evaluate the effects of various real and/or hypothetical system stress patterns, including the individual effects of improvements in irrigation efficiency and lining of conveyance canals. These simulations require only that we first determine the system's hydraulic and solute response times from known data, and then estimate the quantity and quality of applied waters

from the various real or hypothetical sources. Hence, the generalized lumped parameter model is a truly predictive management tool.

4. In both the linear and nonlinear lumped parameter formulations, the requirement for showing or independently computing groundwater pumpage and evapotranspiration rates may be avoided in the water and mass balance relationships. This advantage is of major importance since one may avoid precisely those terms that tend to introduce large errors into the typical water balance approach to modeling. Thus, the lumped model format not only realistically incorporates many simplifying physical phenomena, the corresponding data base required for practical application is likewise reduced. These advantages are acquired, however, at the cost of a complete spatial description of hydraulic and mass transport mechanisms.

It is recommended that future research be conducted in the following:

1. An obvious extension of this work is to include the chemical exchange effects between the soil in the vadose zone and associated percolating recharge waters; several such models currently exist which could easily be incorporated (i.e., Dutt, et al., 1972; Bresler, 1977).

2. A question concerning aquifer parameter values from different hydrogeological settings might also be of interest. It is apparent that values for the parameter a from the Arkansas Valley and the Mesilla Valley are nearly identical, but both of these settings are actually hydrogeologically similar. However, irrigated coastal aquifer systems (as found along the northern coast of Lybia or the western shore of Isreal) might yield different values for a .

3. The Theissen polygon weighing technique was utilized to evaluate average aquifer water levels and initial contamination levels. This technique easily assigns appropriate weighing factors to account for horizontal variations in these variables. Such a procedure may not adequately account for vertical changes in aquifer water quality. Hence, other techniques may be more appropriate in obtaining these averages while still retaining the overall advantage of low input data requirements.

4. Many linear and/or nonlinear lumped models could be linked in a series and/or parallel combination in order to simulate regional water sheds where irrigation activity is widespread. The Colorado River Basin and the Rio Grande Basin are examples where water management options might be optimized so as to achieve minimal concentration outflow histories while maintaining economic agricultural practices. Although such an undertaking would be a massive effort, a sufficient data base probably already exists.

5. The nonlinear outflow expression given in (4.3.2) might be altered to produce even better results than those obtained for the Mesilla Valley. Furthermore, the precise form of such an expression may or may not be a site specific characteristic of the system in question.

6. The procedure of incorporating a retardation factor into the mass balance equation, which conceivably could be similar in some respects to the contamination degradation factor, might adequately account for reactive effects in a systems averaged sense.

7. Using a constant river leakage term for a given recharge-recession period may create problems in certain situations. This value is taken as a constant and may cover several time intervals in

the basic simulation. In some applications it might be more appropriate to obtain specific river leakage values over each of the given time intervals used in the water balance equation.

8. In the derivation of the generalized lumped parameter model, it was assumed that the aquifer specific yield was identical to the effective porosity (i.e., that $S = n$). In certain applications, this assumption may introduce errors into the mass balance equation. Hence, the affect of setting $S = n$ should be evaluated.

9. Throughout this work a steady flow situation was employed in constructing idealized concentration break-through curves for the lumped parameter, Dupuit, and Laplace models. A logical extension of this work would be to extend the analysis to the unsteady case as well.

REFERENCES CITED

- Bear, J., 1972, Dynamics of Fluids in Porous Media: American Elsevier, New York, 764 p.
- Biggar, G. W., and Nielsen, D. R., 1963, Miscible displacement, 5, Exchange processes: Soil Sci. Soc. Amer. Proc., v. 27, p. 623-627.
- Blaney, H. F., and Hanson, E. G., 1965, Consumptive use and water requirements in New Mexico: New Mexico State Engr., Tech. Bull. 32, Santa Fe, New Mexico.
- Bresler, E., 1967, A model for tracing salt distribution in the soil profile and estimating the efficient combination of water quantity and quality under varying field conditions: Soil Sci. Soc. of Amer., v. 104, p. 227-233.
- _____, 1973a, Anion exclusion and coupling effects in nonsteady transport through unsaturated soils, Part I, Theory: Soil Sci. Soc. Amer., v. 37, p. 663-669.
- _____, 1973b, Simultaneous transport of solutes and water under transient unsaturated flow conditions: Water Resources Research, v. 9, n. 4, p. 975-986.
- _____, 1977, Models for predicting the impact of irrigation on soil salinity: in Proc. Inter. Salinity Conf. on Managing Saline Water for Irrigation, Texas Tech Univ., Lubbock Texas, p. 299-315.
- Bresler, E., and Hanks, R. J., 1969, Numerical method for estimating simultaneous flow of water and salt in unsaturated soils: Soil Sci. Soc. Amer., v. 33, p. 827-832.
- Carslaw, H. S., and Jaeger, J. C., 1959, Conduction of Heat in Solids: Oxford Press, London, 510 p.
- Collins, M. A., 1976, The extended Boussinesq problem: Water Resources Research, v. 12, n. 1, p. 54-56.
- Cooley, R. L., 1970, Finite element solution of steady state potential flow problems: Hydrologic Engineering Center, generalized computer program 723-G2-L2240, U. S. Army Corps of Engineers, 48 p. plus addendums.
- _____, 1974, Finite element solutions for the equations of ground-water flow: Hydrology and Water Resources Pub. No. 18, Desert Research Institute, Reno, Nevada, 134 p.
- Dooge, J. C. I., 1960, The routing of groundwater recharge through typical elements of linear storage: IAHS Helsinki Publ. 52, p. 286-300.

- _____, 1973, Linear theory of hydrologic systems: U. S. Dept. Agric. Tech. Bull. 1468, Washington, D.C., 327 p.
- Dutt, G. R., 1962, Prediction of the concentration of solutes in soil solutions for soil systems containing gypsum and exchangeable Ca and Mg: *Soil Sci. Soc. Amer.*, v. 26, p. 341-343.
- Dutt, G. R., Shaffer, M. J., and Moore, W. J., 1972, Computer simulation model of dynamic bio-physicochemical processes in soils: Agricultural Experiment Station, Tech. Bull. 196, Univ. of Arizona, Tucson, 101 p.
- Dutt, G. R., Tucker, T. C., Shaffer, M. J., and Moore, W. J., 1970, Predicting nitrate content of agricultural drainage water: Final report on Contract No. 14-06-0-6464 to U. S. Bureau of Reclamation, Dept. of Agriculture Chemistry and Soils, Univ. of Arizona, Tucson.
- Elbakhbekhi, M. A., 1976, Validity and limitations of horizontal flow models in phreatic aquifers: unpub. Ph.D. dissertation, New Mexico Institute Mining and Technology, Socorro, 177 p.
- Eldor, M., and Dagan, G., 1972, Solutions of hydrodynamic dispersion in porous media: *Water Resources Research*, v. 8, No. 5, p. 1316-1331.
- Eliasson, J., 1971, Mechanism of groundwater reservoirs: *J. Nord. Hydrol.*, v. 2, p. 266-277.
- Eriksson, E., 1970, Groundwater time series—an exercise in stochastic hydrology: *J. Nord. Hydrol.*, v. 1, n. 3, p. 181-205.
- Federal Water Pollution Control Act Amendments, 1972, Admendments on 1972 Public Law 92-500: *Fed. W. P. Cont. Act*, Washington, D. C., October 18, 1972.
- Freeze, R. A., 1969a, The mechanism of natural groundwater recharge and discharge, 1. One-dimensional, vertical, unsteady, unsaturated flow above a recharging or discharging groundwater flow system: *Water Resources Research*, v. 5, n. 1, p. 153-171.
- _____, 1969b, Theoretical analysis of regional groundwater flow: Scientific Series no. 3, Inland Waters Branch, Ottawa, Canada, 147 p. plus computer listings.
- _____, 1971, Three-dimensional, transient saturated-unsaturated flow in a groundwater basin: *Water Resources Research*, v. 7, n. 2, p. 347-366.
- _____, 1972, Subsurface hydrology at waste disposal sites: *IBM Jour. of Res. and Develop.*, v. 16, n. 2, p. 117-129.
- Flores, W., A. E. Z., and Gelhar, L. W., 1976, A stochastic management model for the operation of a stream-aquifer system: Report No. 075, New Mexico Water Resources Resarch Institute, Las Cruces, 209 p.

- Flores, W., A.E.Z, Gutjahr, A. L., and Gelhar, L. W., 1978, A stochastic model of the operation of a stream-aquifer system: *Water Resources Research*, v. 14, no. 1, p. 30-38.
- Fried, J. J., 1975, Groundwater Pollution: Elsevier, New York, 330 p.
- Fried, J. J., and Combarous, M. A., 1971, Dispersion in porous media: in Advances in Hydrosience, V. T. Chow (ed), Academic Press, New York, p. 169-282.
- Gardner, W. R., and Brooks, R. H., 1957, A descriptive theory of leaching: *Soil Sci. Soc. Amer.*, v. 83, p. 295-304.
- Gelhar, L. W., 1976, A comparison of groundwater quality modeling techniques: Water Resources Center, Conc. on Groundwater Quality, Univ. of Reading, England, Sept. 6-8, 1976.
- Gelhar, L. W., and Collins, M. A., 1971, General analysis of longitudinal dispersion in nonuniform flows: *Water Resources Research*, v. 7, n. 6, p. 1511-1521.
- Gelhar, L. W., Ko, P. Y., Kwai, H. H., and Wilson, J. L., 1974, Stochastic modeling of groundwater systems: R. M. Parsons Laboratory, Report No. 189, Massachusetts Inst. of Tech., Cambridge, Mass.
- Gelhar, L. W., and McLin, S. G., 1979, Evaluation of a hydrosalinity model of irrigation return flow water quality in the Mesilla Valley, New Mexico: EPA Tech. Series No. EPA-600/2-79-173, R. S. Kerr Environmental Research Laboratory, Ada, Oklahoma, 192 p.
- Gelhar, L. W., and Wilson, J. L., 1974, Ground-water quality modeling: *Ground Water*, v. 12, n. 6, p. 399-408.
- Glover, R. E., 1974, Transient Ground Water Hydraulics: Dept. of Civil Engr., Colorado State Univ., Fort Collins, Colorado, 413 p.
- Gupta, S. P., and Greenkorn, R. A., 1973, Dispersion during flow in porous media with bilinear adsorption: *Water Resources Research*, v. 9, no. 5, p. 1357-1368.
- Guymon, N. M., 1970, A finite element method of the one-dimensional diffusion-convection equation: *Water Resources Research*, v. 6, no. 1, p. 204-210.
- Guymon, G. L., Scott, V. H., and Herrmann, L. R., 1970, A general solution of the two-dimensional diffusion-convection equation by the finite element method: *Water Resources Research*, v. 6, n. 6, p. 1611-1617.

- Hanks, R. J., Willardson, L. S., and Melamed, D., 1977, Modeling salinity of irrigation return flow where sources and sinks are present: in Proc. Irrigation Return Flow Quality Management Conference, Colorado State Univ., Ft. Collins, p. 99-108.
- Helweg, O. J., and Labadie, J. W., 1976, A salinity management strategy for stream-aquifer systems: Hydrology Paper No. 84, Colorado State Univ, Ft. Collins, 39 p.
- Hiester, N. K., and Vermeulen, T., 1952, Saturation performance of ion-exchange and adsorption columns: Chem. Engr. Progr., v. 48, p. 505-516.
- Hornsby, A. G., 1973, Prediction modeling for salinity control in irrigation return flows: U. S. Environmental Protection Agency, Tech. Series No. EPA-R2-73-168, Corvallis, Oregon, 55 p.
- Huyakorn, P. S., 1977, Solution of steady-state, convective transport equation using an upwind finite element scheme: Appl. Math. Modeling, v. 1, p. 187-195.
- Huyakorn, P. S., and Nilkuha, K., 1979, Solution of transient transport equation using an upstream finite element scheme: Appl. Math. Modeling, v. 3, p. 7-17.
- Jury, W. A., 1975, Solute travel-time estimates for tile-drained fields; I. Theory; II. Applications to experimental studies; Soil Sci. Soc. Amer., v. 39, n. 6, p. 1020-1028.
- Jury, W. A., Fluhler, H., and Stolzy, L. H., 1977, Influence of soil properties, leaching fraction, and plant water uptake on solute concentration distribution: Water Resources Research, v. 13, n. 3, p. 645-650.
- Kirkham, D., and Powers, W. L., 1972, Advanced Soil Physics: Wiley-Interscience, New York, 534 p.
- Kirkham, D., Tohsoz, S., and van der Ploeg, R. R., 1974, Steady flow to drains and wells: in Drainage for Agriculture, J. van Schilfgaarde (ed), Agronomy Monograph No. 17, Amer. Soc. Agron., Madison, Wisc., p. 203-244.
- Kirda, C., Nielsen, D. R., and Biggar, J. W., 1973, Simultaneous transport of chloride and water during infiltration: Soil Sci. Soc. Amer., v. 37, no. 3, p. 339-345.
- Konikow, L. A., and Bredehoeft, J. D., 1974a, Modeling flow and chemical quality changes in an irrigated stream-aquifer-system: Water Resources Research, v. 10, n. 3, p. 546-562.
- _____, 1974b, A water quality model to evaluate water management practices in an irrigated stream-aquifer system: in Salinity in Water Resources, Merriman, Boulder, Colorado, p. 36-59.

- Kraijenhoff van de Leur, D. A., 1958, A study of non-steady groundwater flow with special reference to a reservoir coefficient: *Ingenieur*, v. 70, n. 19, p. 87-94.
- Labadie, J. W., Khan, I. A., and Helweg, O. J., 1976, Salinity management strategies for the lower San Luis Rey River basin: Tech. Comp. Report no. C-6301, Office of Water Research and Technology, Washington, D. C., 168 p.
- Lai, S. H., Juinak, J. J., 1972, Cation adsorption in one-dimensional flow through soils; A numerical solution: *Water Resources Research*, v. 8, n. 1, p. 98-107.
- Lapidus, L., and Amundson, N. R., 1952, Mathematics of adsorption in beds, 6, The effect of longitudinal diffusion in ion exchange and chromatographic columns: *J. Phys. Chem.*, v. 56, p. 984-988.
- Luthin, J. N., Fernandez, P., Moslov, B., Woerner, J., and Robinson, F., 1969, Displacement front under ponded leaching: *ASCE J. Irr. Drain.*, v. 96, n. IR3, p. 257-264.
- Mercado, A., 1976, Nitrate and chloride pollution of aquifers - a regional study with the aid of a single cell model: *Water Resources Research*, v. 12, n. 4, p. 731-747.
- McLin, S. G., and Gelhar, L. W., 1979, A field comparison between the USBR-EPA hydrosalinity and generalized lumped parameter models: in The Hydrology of Areas of Low Precipitation, Proc. of the Canberra Symposium, IAHS Pub. n. 128, p. 339-348.
- Miyamoto, S., and Warrick, A. W., 1974a, Salt displacement into ditch tiles under ponded leaching: *Water Resources Research*, v. 10, n. 2, p. 275-278.
- _____, 1974b, Two-dimensional displacement into or from water filled ditches: *Soil Sci. Soc. Amer.*, v. 38, p. 723-727.
- Murray, W. A., and Monkmeyer, P. L., 1973, Validity of Dupint-Forchheimer equation: *J. Hydr. Div.*, ASCE, HY9, p. 1573-1577.
- Muskat, M., 1946, The Flow of Homogeneous Fluids Through Porous Media: Edwards, Ann Arbor, Michigan, 763 p.
- Nimah, M., and Hanks, R. J., 1973, Model for estimating soil water, plant and atmospheric interrelation. I. Description and sensitivity, II. Field test of the model: *Soil Sci. Soc. Amer.*, v. 37, p. 522-532.
- Neuman, S. P., Feddes, R. A., and Bresler, E., 1974, Finite element simulation of flow in saturated-unsaturated soils considering water uptake by plants: project no. ALO-SWC-77, Hydrodynamics and Hydraulic Engineering Laboratory, Israel Institute of Technology, Haifa, p. 104.

- Ortiz, J., and Luthin, J. N., 1970, Movement of salts in ponded anisotropic soils: *ASCE J. Irr. Drain.*, v. 96, n. IR3, p. 257-264.
- Palciuskas, V. V., and Domenico, P. A., 1976, Solution chemistry, mass transfer, and the approach to chemical equilibrium in porous carbonate rocks and sediments: *Geol. Soc. Amer. Bull.* 87, p. 207-214.
- Perez, A. I., Huber, W. C., Heaney, J. P., and Pyatt, E. E., 1972, A water quality model for a conjunctive surface-groundwater system, an overview: *Water Resources Bull.*, v. 8, p. 900-908.
- Pickens, J. S., and Lennox, W. C., 1976, Numerical simulation of waste movement in steady groundwater flow systems: *Water Resources Research*, v. 12, no. 2, p. 171-180.
- Pikul, M. F., Street, R. L., and Remson, I., 1974, A numerical model based on coupled one-dimensional Richards and Boussinesq equations: *Water Resources Research*, v. 10, n. 2, p. 295-304.
- Pinzon-Lizarraga, S., 1978, A non-linear lumped parameter model for the Mesilla Valley, New Mexico: unpub. MS independent study, New Mexico Institute of Mining and Technology, Socorro, New Mexico.
- Prickett, T. A., 1976, Modeling techniques for groundwater evaluation: in Advances in Hydroscience, Academic Press, New York, v. 10, p. 1-143.
- Prickett, T. A., and Lonquist, C. G., 1971, Selected digital computer techniques for groundwater resource evaluation: *Illinois State Water Survey, Bull. No. 55*, Urbana, 62 p.
- Przewlocki, K., and Yurtsever, Y., 1974, Some conceptual mathematical models and digital simulation approach in the use of tracers in hydrologic systems: in Isotope Techniques in Groundwater Hydrology, Inter. Atomic Energy Assoc., Austria, v.2, p. 425-450.
- Raats, P. A. C., 1977, Convective transport of solutes in and below the root zone: in Inter. Conf. on Managing Saline Water for Irrigation, Texas Tech Univ., Lubbock, Texas, 16-20 August 1976.
- Rabinowitz, D. D., Gross, G. W., and Holmes, C. R., 1976, Environmental tritium as a hydrometeorologic tool in the Roswell Basin, New Mexico; Part I, Tritium input function and precipitation-recharge relation; Part II, Tritium patterns in groundwater; Part III, Hydrologic parameters: *J. Hydrology*, v. 32, p. 3-46.
- Reddell, D. L., Sunada, D. K., 1970, Numerical simulation of dispersion in groundwater aquifers: *Hydrology Paper n. 41*, Colorado State Univ., Fort Collins, 79 p.

- Riley, J. P., Chadwich, D. G., and Bagley, J. M., 1966, Applications of an electric analog computer to solutions of hydrologic and river-basin-planning problems: Utah Simulation Model II, PRWG32-1, Utah Water Research Laboratory, Utah State Univ., Logan.
- Robertson, J. B., 1974, Application of digital modeling to the prediction of radioisotope migration in groundwater: Int. Atomic Energy Agency Symp. on Isotope Techniques in Groundwater Hydr., Vienna, IAEA-SM-182/50.
- Rubin, J., and James, R. V., 1973, Dispersion-affected transport of reacting solutes in saturated porous media; Galerkin method applied to equilibrium controlled exchange in unidirectional steady water flow: Water Resources Research, v. 9, n. 5, p. 1332-1356.
- Schwartz, F. W., and Domenico, P. A., 1973, Simulation of hydrochemical patterns in regional groundwater flow: Water Resources Research, v. 9, no. 3, p. 707-720.
- Shaffer, M. J., Ribbens, R. W., and Huntley, C. W., 1976, Prediction of mineral quality of irrigation return flow, v. 5: Final report for project no. EPA-IAG-D4-0371, Office of Research and Development, U. S. Environmental Protection Agency, Washington, D. C.
- Shamir, U. Y., and Harleman, D. R. F., 1967, Numerical solution for dispersion in porous mediums: Water Resources Research, v. 3, n. 3, p. 557-581.
- Tanji, K. K., and Doneen, L. D., 1966, A computer technique for prediction of CaCO_3 precipitation in HCO_3 salt solutions: Soil Sci. Soc. Amer., v. 30, p. 53-56.
- Tanji, K. K., Dutt, G. R., Paul, J. L., and Doneen, L. D., 1967, Quality of percolating waters, II, A computer method for predicting salt concentrations in soils at variable moisture content: Hilgardia, v. 38, n. 9, p. 307-318.
- Thomas, J. L., Riley, J. P., and Israelson, E. R., 1972, A hybrid computer program for predicting the chemical quality of irrigation return flow: Water Resources Bull., v. 8, n. 5, p. 922-934.
- Trescott, P. C., Pinder, G. F., and Larson, S. P., 1976, Finite difference model for aquifer simulation in two dimensions with results of numerical experiments: Book 7, Chapter C1, Techniques of Water Resources Investigations, U. S. Geological Survey, Washington, D. C., 116 p.
- Updegraff, C. D., Gelhar, L. W., 1978, Parameter estimation for a lumped parameter groundwater model the Mesilla Valley, New Mexico: Report No. 097, New Mexico Water Resource Research Institute, Las Cruces, New Mexico.

- U. S. Bureau of Reclamation, 1977, Prediction of mineral quality of irrigation return flow, vol. III, Simulation model of conjunctive use and water quality for a river system or basin: EPA Tech. Series No. 600/2-77-179C, R. S. Kerr Environmental Research Laboratory, Ada, Oklahoma, 285 p.
- U. S. Corps of Engineers, 1972, Finite element solution of steady state potential flow problems. Generalized Computer Program 723-440: Hydrologic Engineering Center, U. S. Army Corps of Engineers, Davis, California.
- U. S. Soil Conservation Service, 1972, Drainage of Agricultural Land, Water Information Center, New York, 430 p.
- van Schilfgaarde, J., 1965, Transient design of drainage systems: J. Irr. Drain. Div., ASCE, v. 91, n. IR3, p. 9-22.
- _____, 1970, Theory of flow to drains: Advances in Hydrosociences, in V. T. Chow (ed), Academic Press, New York, v. 6, p. 43-106.
- Verma, R. D., and Brutsaert, W., 1971, Unsteady free surface groundwater seepage, J. Hydr. Div., ASCE, v. 97, n. HY8, p. 1213-1228.
- Warrick, A. W., Biggar, J. W., and Nielsen, D. R., 1971, Simultaneous solute and water transfer for an unsaturated soil: Water Resources Research, v. 7, n. 5, p. 1216-1225.
- Warrick, A. W., Kichen, J. H., and Thames, J. L., 1972, Solutions for miscible displacement of soil water with time dependent velocity and dispersion coefficients: Soil Sci. Soc. Amer., v. 36, p. 863-867.
- Warrick, A. W., and Kirkham, D., 1969, Two-dimensional seepage of ponded water to full ditch drains: Water Resource Research, v. 5, no. 3, p. 685-693.
- Willis, R., 1976, Optimal groundwater quality management; well injection of waste waters: Water Resources Research, v. 12, no. 1, p. 47-53.
- Wilson, J. L., and Gelhar, L. W., 1974, Dispersive mixing in a partially saturated porous medium: Ralph M. Parson's Laboratory for Water Resources and Hydrodynamics, Report No. 191, Massachusetts Institute of Technology, Cambridge, Mass., 353 p.

APPENDICES

Appendix A

Dupuit Model Program: General Description

The computer program developed for this portion of the study is written in standard Fortran IV. It computes the dimensionless concentrations and times according to the analytical solutions for the combined lumped parameter and multi-layer Dupuit models derived in Sections 2.2.3, 2.3.3, and 2.3.4. The table below relates the appropriate symbol used in the program to the respective equation in the text.

<u>Program Variable</u>	<u>Symbol</u>	<u>Equation</u>	<u>Remarks</u>
AFAC	a	2.2.23	
BFAC	b	2.2.23	
CL	c/c_{aq1}	2.2.23	
TL	t/t_{c1}	below 2.2.20	TL=HU*T
AMU	μ	2.2.23	
GAMMA	γ	2.2.23	
AM	m	2.2.23	
HU	$1+m$	2.2.26 & 2.3.68	see Table 2.1
KU	K_u		upper layer K
KL	K_L		lower layer K
B	b_L/h_0	2.3.43	
ETA	$\hat{\xi}+1$	2.3.43	computational convenience
XMAX	$(1-B)/(1+\hat{\xi})$	below 2.3.51	
C	c/c_{aq}	2.3.51 & 2.3.61	
T	t/t'_{c1}	2.3.61	relative to upper layer
		2.3.64	relative to lower layer
A	$(1+\hat{\xi}) \ln \{ (B+\hat{\xi}) / (1+\hat{\xi}) \}$	below 2.3.56	computational convenience

C DUPUIT TWO-LAYER AQUIFER MODEL VS. COMBINED LUMPED MODEL

00100
00200
00300
00400
00500
00600
00700
00800
00900
01000
01100
01200
01300
01400
01500
01600
01700
01800
01900
02000
02100
02200
02300
02400
02500
02600
02700
02800
02900
03000
03100
03200
03300
03400
03500
03600
03700
03800
03900
04000
04100
04200
04300
04400
04500
04600
04700
04800
04900
05000
05100
05200
05300
05400
05500
05600
05700
05800
05900
06000

```

C
C
REAL KU, KL
DIMENSION X(41), B(7), KU(7), S(7), ERROR(41)
COMMON C(41), I(41), CL(41), IP(41), ANAME(2)
DATA B/5.0, 7.5, 0.90, 0.95/
DATA KU/20.0, 10.0, 5.0, 2.0, 1.0, 2*10.0/
DATA S/7*10.0/
DATA IR, IW, Z2, Z3/
OPEN(UNIT=22, DEVICE='DSK', ACCESS='SEQIN', FILE='DUPI.DAT')
OPEN(UNIT=23, DEVICE='DSK', ACCESS='SEQOUT', FILE='DUP.DAT')
CALL INITIAL(I)
POR1=1.0
POR2=1.0
KL=1.0
DS=0
DO 1 J=1, 7
  READ(IR, 201)(ANAME(I), II=1:2)
  FORMAT(2A5)
  JK=J
  HU=D/(1.0-B(J))
  S(J)=S(J)/D
  E=AP1.0+B(J)*(KL-KU(J))/KU(J)
  XMAX=1.0-B(J)/ZETA
  AM=B(J)/(1.0-B(J))
  AS=ETA*ALOG(1.0-(1.0-B(J))/ETA)
  X(I)=0.0
  X(I)=1.0-X(I)
  T(I)=0.0
  AM=(AM+KL/KU(J))/(1.0+AM*KL/KU(J))
  GAMMA=(POR1*AMU)/(POR2*AM)
  DO 2 I=2, 40
    KOUNT=1
    X(I)=X(I)
    IF(X(I) .GT. XMAX) GO TO 3
    T(I)=1.0-X(I)
    T(I)=ETA*ALOG(1.0-X(I))
    GO TO 2
  3 XMAX=X(I)
  C(I)=1.0-X(I)
  T(I)=X(I)
  K=KOUNT+1
  X(I)=XMAX
  C(I+1)=1.0-X(I+1)
  T(I+1)=ETA*ALOG(KU(J)/KL)*ALOG(KU(J)*ETA*C(I+1))/(KL*B(J))
  GO TO 4
  2 CONTINUE
  C(41)=C(40)
  T(41)=T(40)
  4 IF(KOUNT.GE.40 .OR. KK.GE.41) GO TO 5
  5 K=KK+1
  X(I)=X(I)+0.025
  C(I)=1.0-X(I)
  T(I)=ETA*ALOG(KU(J)/KL)*ALOG(KU(J)*ETA*C(I))/(KL*B(J))
  6 AREA=0.0
  AREAL=0.0
  BFAC=AMU/(GAMMA-1.0)

```

```

104100 AFAC=1.0+BFAC
104200 CL(I)=AFAC+BFAC
104300 DO 7 I=2,41
104400 CL(I)=AFAC*EXP(-HU*TP(I))-BFAC*EXP(-HU*GAMMA*TP(I))
104500 AREAL=AREAL+0.5*(TP(I)-TP(I-1))*((CL(I)+CL(I-1))
104600 AREAL=AREAL+0.5*(TP(I)-TP(I-1))*((C(I)+C(I-1))
104700 ERROR(I)=CL(I)-C(I)
104800 WRITE(IW,101)(ANAME(I),I=1,2),KU(J),S(J),B(J),AMU,GAMMA
104900 FORMAT('MI//,10X,DUPIUT TWO-LAYER ACQUIFER AS COMPARED TO
< THE COMBINED LUMPED MODEL,/,10X,2A5//,11X,KU/KL =,F4.1,/,
< IIX,S/D =,F4.1//,11X,HL/D =,F6.4//,11X,MU =,F6.4//,11X,
< /,DUPIUT CONC. =,2X,STREAMLINE(2X,DIM,TIME,2X,
< (T PRIME),/,)
10500 DO 8 I=1,41
105100 TP(I)=HU*TP(I)
105200 WRITE(IW,102)X(I),TP(I),C(I),CL(I),ERROR(I)
105300 FORMAT(10X,F7.3,4X,F10.5,4X,F6.3,7X,F7.3,8X,F8.2)
105400 AREAL=AREAL+0.20X*AREA UNDER DUPUIT CURVE; C VS. T =,F10.7//,20X,
105500 AREA UNDER LUMPED CURVE; C VS. T =,F10.7)
105600 DO 10 I=1,41
105700 C(I)=ALOG10(C(I))
105800 CL(I)=ALOG10(CL(I))
105900 CALL GRAFF(JK)
10600 CALL RSTR(0)
106100 CLOSE(UNIT=2,DEVICE='DSK',ACCESS='SEQIN',FILE='DUPI.DAT')
106200 CLOSE(UNIT=23,DEVICE='DSK',ACCESS='SEQOUT',FILE='DUP.DAT')
106300 STOP
106400 END
C
106500 SUBROUTINE GRAFF(JR)
106600 COMMON C(4),T(4),CL(4),TP(4),ANAME(2)
106700 DIMENSION LBX(4),LBY(4),TITLE1(8),TITLE2(8)
106800 DATA LBX/SH DIME,SHMSTION,SHLESS,SH CONC/
106900 DATA LBY/SH DIM,SHENSIION,SHNLESS,SH TIME/
107000 DATA TITLE1/SHSOLID,SH LINE,SH FOR,SHDUPUI,SH TWO,SH-LAYE,
107100 DATA TITLE2/SHS,SHYMOD,SHL ONL,SHY FOR,SH COMB,SHINED,
107200 SHLUMPE,SHD MOD,SHEL
107300 CALL RSTR(0)
107400 J0=0
107500 J1=1
107600 N=41
107700 MD=2
107800 FL1=1.0
107900 FL2=0.1
108000 FL3=0.01
108100 XMIN=0.0
108200 DX=0.4
108300 YMIN=0.0
108400 DY=8.0
108500 IF(JR.EQ.7)DY=12.0
108600 DO 1 I=1,N
108700 IF(CL(I).GT.2)CL(I)=2.0
108800 C(I)=(C(I)-XMIN)/DX
108900 CL(I)=(CL(I)-YMIN)/DY
109000 TP(I)=(TP(I)-YMIN)/DY

```

```

12100 CALL PLOT(3,0,1,25,-3)
12200 CALL PLOT(5,0,0,0,2)
12300 CALL PLOT(5,0,8,0,2)
12400 CALL PLOT(0,0,8,0,2)
12500 CALL PLOT(0,0,0,0,2)
12600 CALL SYMBOL(0,5,2,0,0,14,ANAME,90,0,10)
12700 CALL SYMBOL(0,7,2,0,0,14,TITLE1,90,0,40)
12800 CALL SYMBOL(0,9,2,0,0,14,TITLE2,90,0,40)
12900 CALL AXIS(5,0,0,0,LBY,-20,8,0,90,0,YMIN,DY,-1)
13000 CALL SYMBOL(1,3,-0,8,0,14,LBX,0,0,20)
13100 CALL PLOT(2,5,0,0,3)
13200 CALL PLOT(2,5,8,0,2)
13300 CALL NUMBER(0,0,-0,6,0,14,FLT1,90,0,ND)
13400 CALL NUMBER(2,5,-0,6,0,14,FLT2,90,0,ND)
13500 CALL NUMBER(5,0,-0,6,0,14,FLT3,90,0,ND)
13600 CALL PLOT(0,0,0,0,3)
13700 CALL LINE(C,TP,N,JO,1)
13800 CALL PLOT(0,0,0,0,3)
13900 DO 2 I=1,N
14000 XCORD=CU(I)
14100 YCORD=TP(I)
14200 CALL PLOT(XCORD,YCORD,3)
14300 CALL MARKER(J1)
14400 2 CONTINUE
14500 CALL PLOT(0,0,0,0,3)
14600 RETURN
14700 END

```

DUPUIT TWO-LAYER AQUIFER AS COMPARED TO THE COMBINED LUMPED MODEL
CASE(1):

KU/KL = 20.0
S/D = 10.75
BL/D = 1304
MU = .435
GAMMA = .0435

STREAMLINE DIM TIME DUPUIT CONC. LUMPED CONC. & REL. ERROR

0.000	0.00000	1.000	1.000	0.00
0.005	0.02912	0.975	0.975	0.00
0.010	0.05899	0.950	0.950	0.04
0.015	0.08966	0.925	0.925	0.08
0.100	0.12116	0.901	0.875	0.13
0.125	0.13356	0.875	0.852	0.20
0.150	0.18690	0.825	0.827	0.28
0.200	0.22123	0.800	0.803	0.38
0.225	0.29313	0.775	0.775	0.50
0.250	0.33083	0.750	0.755	0.64
0.275	0.36982	0.725	0.731	0.81
0.300	0.41018	0.700	0.707	1.00
0.325	0.45200	0.675	0.683	1.23
0.350	0.49540	0.650	0.660	1.49
0.375	0.54050	0.625	0.636	1.80
0.400	0.58745	0.600	0.613	2.15
0.425	0.63639	0.575	0.590	2.55
0.450	0.68751	0.550	0.567	3.02
0.475	0.74101	0.525	0.544	3.58
0.500	0.79712	0.500	0.521	4.18
0.525	0.85611	0.475	0.498	4.90
0.550	0.91828	0.450	0.476	5.73
0.575	0.98402	0.425	0.453	6.70
0.600	1.05373	0.400	0.431	7.84
0.625	1.12795	0.375	0.409	9.17
0.650	1.20730	0.350	0.388	10.75
0.675	1.29252	0.325	0.366	12.69
0.700	1.38457	0.300	0.345	14.97
0.725	1.48463	0.275	0.324	17.65
0.750	1.59424	0.250	0.303	21.04
0.775	1.71540	0.225	0.282	25.30
0.800	1.85085	0.200	0.262	30.75
0.825	2.00441	0.175	0.241	37.91
0.850	2.18169	0.150	0.221	47.66
0.870	2.34241	0.130	0.206	58.05
0.875	2.42128	0.125	0.195	69.39
0.900	2.45358	0.100	0.175	82.58
0.925	24.39595	0.075	0.155	98.58
0.950	40.33832	0.050	0.141	118.58
0.975		0.025	0.124	143.58

AREA UNDER DUPUIT CURVE: C VS. T = 0.8656083
AREA UNDER LUMPED CURVE: C VS. T = 0.8866023

DUPIIT TWO-LAYER AQUIFER AS COMPARED TO THE COMBINED LUMPED MODEL
CASE(2):

KU/KL = 10.0
S/D = 10.0
RL/D = .75
MU = .2308
GAMMA = .0769

STREAMLINE DIM. TIME DUPIIT CONC. LUMPED CONC. % REL. ERROR

STREAMLINE	DIM.	TIME	DUPIIT CONC.	LUMPED CONC.	% REL. ERROR
0.000	0.00000	1.000	1.000	1.000	0.00
0.025	0.3291	0.975	0.975	0.975	0.01
0.050	0.6668	0.950	0.950	0.950	0.04
0.075	0.10125	0.925	0.926	0.926	0.08
0.100	0.13697	0.900	0.901	0.901	0.15
0.125	0.17359	0.875	0.877	0.877	0.25
0.150	0.21127	0.850	0.853	0.853	0.37
0.175	0.25008	0.825	0.829	0.829	0.52
0.200	0.29009	0.800	0.806	0.806	0.70
0.225	0.33136	0.775	0.782	0.782	0.92
0.250	0.37399	0.750	0.759	0.759	1.19
0.275	0.41806	0.725	0.736	0.736	1.52
0.300	0.46368	0.700	0.713	0.713	1.85
0.325	0.51096	0.675	0.690	0.690	2.27
0.350	0.56002	0.650	0.668	0.668	2.75
0.375	0.61100	0.625	0.646	0.646	3.30
0.400	0.66407	0.600	0.624	0.624	3.94
0.425	0.71940	0.575	0.602	0.602	4.67
0.450	0.77719	0.550	0.580	0.580	5.50
0.475	0.83766	0.525	0.559	0.559	6.46
0.500	0.90109	0.500	0.538	0.538	7.57
0.525	0.96777	0.475	0.517	0.517	8.84
0.550	1.03806	0.450	0.496	0.496	10.32
0.575	1.11237	0.425	0.476	0.476	12.02
0.600	1.19118	0.400	0.456	0.456	14.00
0.625	1.27508	0.375	0.436	0.436	16.32
0.650	1.36477	0.350	0.417	0.417	19.05
0.675	1.46111	0.325	0.397	0.397	22.28
0.700	1.56516	0.300	0.378	0.378	26.14
0.725	1.67828	0.275	0.360	0.360	30.82
0.750	1.80218	0.250	0.341	0.341	36.54
0.775	1.90624	0.225	0.327	0.327	41.87
0.800	2.23537	0.225	0.291	0.291	29.21
0.825	3.76655	0.200	0.204	0.204	2.23
0.850	5.50245	0.175	0.167	0.167	-4.70
0.875	7.50641	0.150	0.141	0.141	-6.41
0.875	9.87659	0.125	0.125	0.125	-6.44
0.900	12.77746	0.100	0.094	0.094	-6.44
0.925	16.51732	0.075	0.075	0.075	-6.44
0.950	21.78837	0.050	0.047	0.047	-6.44
0.975	30.79927	0.025	0.025	0.025	-6.44

AREA UNDER DUPIIT CURVE: C VS. T = 0.9246392
AREA UNDER LUMPED CURVE: C VS. T = 0.9336334

DUPUIT TWO-LAYER AQUIFER AS COMPARED TO THE COMBINED LUMPED MODEL
CASE(3):

KU/KL = 5.0
S/D = 10.0
BL/D = 0.75
MU = 0.3750
GAMMA = 0.1250

STREAMLINE DIM TIME DUPUIT CONC. LUMPED CONC. % REL. ERROR

0.000	0.00000	1.000	1.000	0.00
0.025	0.04051	0.975	0.975	0.02
0.050	0.08207	0.950	0.951	0.06
0.075	0.12474	0.925	0.926	0.15
0.100	0.16858	0.900	0.902	0.27
0.125	0.21365	0.875	0.879	0.43
0.150	0.26003	0.850	0.855	0.64
0.175	0.30780	0.825	0.832	0.90
0.200	0.35703	0.800	0.810	1.22
0.225	0.40783	0.775	0.787	1.59
0.250	0.46029	0.750	0.765	2.03
0.275	0.51453	0.725	0.743	2.55
0.300	0.57068	0.700	0.722	3.14
0.325	0.62887	0.675	0.701	3.83
0.350	0.68925	0.650	0.680	4.62
0.375	0.75201	0.625	0.660	5.52
0.400	0.81732	0.600	0.639	6.55
0.425	0.88542	0.575	0.619	7.72
0.450	0.95654	0.550	0.600	9.06
0.475	1.03097	0.525	0.581	10.59
0.500	1.10904	0.500	0.562	12.32
0.525	1.19110	0.475	0.543	14.30
0.550	1.27761	0.450	0.525	16.57
0.575	1.36907	0.425	0.507	19.18
0.600	1.46607	0.400	0.489	22.15
0.625	1.56933	0.375	0.471	25.65
0.650	1.68933	0.350	0.471	29.50
0.675	1.82127	0.325	0.397	33.58
0.700	1.96413	0.300	0.343	37.92
0.725	2.11805	0.275	0.302	42.55
0.750	2.28305	0.250	0.268	47.45
0.775	2.45993	0.225	0.239	52.58
0.800	2.64919	0.200	0.213	57.92
0.825	2.85145	0.175	0.189	63.45
0.850	3.06845	0.150	0.165	69.18
0.875	3.30183	0.125	0.141	75.15
0.900	3.55222	0.100	0.117	81.30
0.925	3.81943	0.075	0.094	87.67
0.950	4.10448	0.050	0.070	94.25
0.975	4.40854	0.025	0.047	101.92
0.975	23.23371	0.025	0.023	110.00

AREA UNDER DUPUIT CURVE: C VS. T = 0.9538100
AREA UNDER LUMPED CURVE: C VS. T = 0.9575593

DUPIUIT TWO-LAYER AQUIFER AS COMPARED TO THE COMBINED LUMPED MODEL
CASE(4):

KU/KL = 2.0
S/D = 10.0
BL/D = 0.75
MU = 5000
GAMMA = 0.2000

STREAMLINE DIM, TIME DUPIUIT CONC. LUMPED CONC. & REL. ERROR

Streamline Dim	Time	Duipuit Conc.	Lumped Conc.	Rel. Error
0.000	0.00000	1.000	1.000	0.00
0.025	0.06329	0.975	0.975	0.02
0.050	0.12823	0.950	0.951	0.10
0.075	0.19490	0.925	0.927	0.22
0.100	0.26340	0.900	0.904	0.40
0.125	0.33383	0.875	0.881	0.64
0.150	0.40630	0.850	0.858	0.94
0.175	0.48093	0.825	0.836	1.31
0.200	0.55786	0.800	0.814	1.74
0.225	0.63723	0.775	0.792	2.25
0.250	0.71921	0.750	0.771	2.84
0.275	0.80396	0.725	0.750	3.52
0.300	0.89169	0.700	0.730	4.28
0.325	0.98261	0.675	0.710	5.15
0.350	1.07696	0.650	0.690	6.13
0.375	1.17501	0.625	0.670	7.22
0.400	1.27706	0.600	0.651	8.44
0.425	1.38366	0.600	0.651	9.79
0.450	1.49596	0.575	0.613	11.27
0.475	1.61422	0.550	0.578	12.88
0.500	1.73867	0.525	0.544	14.64
0.525	1.86944	0.500	0.512	16.55
0.550	2.00674	0.475	0.482	18.61
0.575	2.15067	0.450	0.452	20.82
0.600	2.30127	0.425	0.424	23.18
0.625	2.45869	0.400	0.396	25.68
0.650	2.62308	0.375	0.370	28.33
0.675	2.79459	0.350	0.344	31.13
0.700	2.97329	0.325	0.325	34.05
0.725	3.15920	0.300	0.293	37.08
0.750	3.35241	0.275	0.268	40.33
0.775	3.55302	0.250	0.243	43.79
0.800	3.76112	0.225	0.219	47.45
0.825	3.97681	0.200	0.194	51.33
0.850	4.20019	0.175	0.170	55.43
0.875	4.43136	0.150	0.145	59.75
0.900	4.67041	0.125	0.121	64.29
0.925	4.91744	0.100	0.097	69.05
0.950	5.17255	0.075	0.073	74.05
0.975	5.43582	0.050	0.048	79.28

AREA UNDER DUPIUIT CURVE: C VS. T = 0.9711963
AREA UNDER LUMPED CURVE: C VS. T = 0.9722227

DUPUIT TWO-LAYER AQUIFER AS COMPARED TO THE COMBINED LUMPED MODEL
CASE(5):

KU/KL = 1.0
S/D = 10.0
BL/D = .75
MU = .7500
GAMMA = .2500

STREAMLINE DIM TIME DUPUIT CONC. LUMPED CONC. % REL. ERROR

STREAMLINE	DIM	TIME	DUPUIT CONC.	LUMPED CONC.	% REL. ERROR
0.000	0.00000	1.000	1.000	1.000	0.00
0.025	0.01027	0.975	0.975	0.975	-0.00
0.050	0.10517	0.950	0.950	0.950	0.00
0.075	0.31185	0.925	0.925	0.925	0.00
0.100	0.42144	0.900	0.900	0.900	0.00
0.125	0.53413	0.875	0.875	0.875	0.00
0.150	0.65008	0.850	0.850	0.850	0.00
0.175	0.76949	0.825	0.825	0.825	0.00
0.200	0.89257	0.800	0.800	0.800	0.00
0.225	1.01957	0.775	0.775	0.775	0.00
0.250	1.15073	0.750	0.750	0.750	0.00
0.275	1.15073	0.750	0.750	0.750	0.00
0.300	1.28633	0.725	0.725	0.725	0.00
0.325	1.42670	0.700	0.700	0.700	0.00
0.350	1.57213	0.675	0.675	0.675	0.00
0.375	1.72317	0.650	0.650	0.650	0.00
0.400	1.88001	0.625	0.625	0.625	0.00
0.425	2.04330	0.600	0.600	0.600	0.00
0.450	2.21354	0.575	0.575	0.575	0.00
0.475	2.39135	0.550	0.550	0.550	0.00
0.500	2.57743	0.525	0.525	0.525	0.00
0.525	2.77259	0.500	0.500	0.500	0.00
0.550	2.97776	0.475	0.475	0.475	0.00
0.575	3.19403	0.450	0.450	0.450	0.00
0.600	3.42266	0.425	0.425	0.425	0.00
0.625	3.66516	0.400	0.400	0.400	0.00
0.650	3.92332	0.375	0.375	0.375	0.00
0.675	4.19929	0.350	0.350	0.350	0.00
0.700	4.49572	0.325	0.325	0.325	0.00
0.725	4.81589	0.300	0.300	0.300	0.00
0.750	5.16394	0.275	0.275	0.275	0.00
0.775	5.54518	0.250	0.250	0.250	0.00
0.800	5.96662	0.225	0.225	0.225	0.00
0.825	6.43775	0.200	0.200	0.200	0.00
0.850	6.97188	0.175	0.175	0.175	0.00
0.875	7.58848	0.150	0.150	0.150	0.00
0.900	8.31776	0.125	0.125	0.125	0.00
0.925	9.21034	0.100	0.100	0.100	0.00
0.950	10.36107	0.075	0.075	0.075	0.00
0.975	11.98293	0.050	0.050	0.050	0.00
0.975	14.75551	0.025	0.025	0.025	0.00

AREA UNDER DUPUIT CURVE: C VS. T = 0.9769744
AREA UNDER LUMPED CURVE: C VS. T = 0.9769744

DUPUIT TWO-LAYER AQUIFER AS COMPARED TO THE COMBINED LUMPED MODEL
CASE(6):

KU/KL = 10.0
S/D = 10.0
BL/D = .90
MU = .4737
GAMMA = .0526

STREAMLINE DIM TIME DUPUIT CONC. LUMPED CONC. % REL. ERROR

0.000	0.00000	1.000	1.000	0.00
0.025	0.04810	0.975	0.975	0.03
0.050	0.0746	0.950	0.951	0.11
0.075	0.1013	0.925	0.927	0.25
0.100	0.12813	0.900	0.904	0.45
0.125	0.15371	0.875	0.881	0.72
0.150	0.17879	0.850	0.859	1.07
0.175	0.20351	0.825	0.837	1.50
0.200	0.22797	0.800	0.816	2.02
0.225	0.25230	0.775	0.795	2.64
0.250	0.27660	0.750	0.775	3.37
0.275	0.30101	0.725	0.756	4.22
0.300	0.32578	0.700	0.736	5.20
0.325	0.35091	0.675	0.718	6.32
0.350	0.37641	0.650	0.699	7.61
0.375	0.40230	0.625	0.682	9.08
0.400	0.42857	0.600	0.665	10.76
0.425	1.05143	0.575	0.648	12.66
0.450	1.13589	0.550	0.632	14.83
0.475	1.22428	0.525	0.616	17.29
0.500	1.31698	0.500	0.600	20.10
0.525	1.41444	0.475	0.586	23.38
0.550	1.51771	0.450	0.586	27.18
0.575	1.62728	0.425	0.486	31.58
0.600	1.74329	0.400	0.432	36.59
0.625	1.86616	0.375	0.397	42.15
0.650	2.00029	0.350	0.369	48.31
0.675	2.14735	0.325	0.348	55.13
0.700	2.30912	0.300	0.318	62.72
0.725	2.48733	0.275	0.294	71.14
0.750	2.68227	0.250	0.245	80.44
0.775	2.89564	0.225	0.220	90.64
0.800	3.12995	0.200	0.196	102.84
0.825	3.38795	0.175	0.171	117.14
0.850	3.66322	0.150	0.147	133.64
0.875	3.95874	0.125	0.122	152.44
0.900	4.27874	0.100	0.098	173.74
0.925	4.62774	0.075	0.075	207.64
0.950	5.01153	0.050	0.049	255.24
0.975	5.43613	0.025	0.024	317.64

AREA UNDER DUPUIT CURVE: C VS. T = 0.9561513
AREA UNDER LUMPED CURVE: C VS. T = 0.9581057

DUPUIT TWO-LAYER AQUIFER AS COMPARED TO THE COMBINED LUMPED MODEL CASE(7):

KU/ML = 10.0
S/D = 10.0
BL/D = .95
MU = .6552
GAMMA = .0345

STREAMLINE DIM. TIME DUPUIT CONC. LUMPED CONC. % REL. ERROR

0.000	0.00000	1.000	1.000	0.00
0.025	0.07342	0.975	0.976	0.05
0.050	0.14875	0.950	0.952	0.22
0.075	0.22609	0.925	0.930	0.51
0.100	0.30555	0.900	0.908	0.92
0.125	0.38724	0.875	0.888	1.46
0.150	0.47130	0.850	0.868	2.15
0.175	0.55788	0.825	0.850	2.99
0.200	0.64712	0.800	0.832	3.96
0.225	0.73919	0.775	0.815	5.22
0.250	0.83428	0.750	0.799	6.52
0.275	0.93259	0.725	0.784	8.08
0.300	1.03436	0.700	0.769	9.86
0.325	1.13982	0.675	0.755	11.89
0.350	1.22628	0.650	0.745	13.68
0.375	1.45614	0.625	0.720	15.81
0.400	2.59354	0.600	0.645	17.13
0.425	3.77738	0.575	0.573	18.34
0.450	5.01161	0.550	0.547	19.61
0.475	6.30071	0.525	0.521	20.71
0.500	7.64979	0.500	0.496	21.72
0.525	9.06470	0.475	0.472	22.72
0.550	10.55221	0.450	0.447	23.72
0.575	12.12016	0.425	0.422	24.72
0.600	13.77775	0.400	0.397	25.72
0.625	15.53587	0.375	0.372	26.72
0.650	17.40748	0.350	0.347	27.72
0.675	19.40828	0.325	0.323	28.72
0.700	21.55741	0.300	0.298	29.72
0.725	23.87865	0.275	0.273	30.72
0.750	26.40198	0.250	0.248	31.72
0.775	29.16597	0.225	0.223	32.72
0.800	32.22143	0.200	0.190	33.72
0.825	35.63713	0.175	0.174	34.72
0.850	39.50954	0.150	0.149	35.72
0.875	43.97991	0.125	0.124	36.72
0.900	49.26723	0.100	0.099	37.72
0.925	55.73840	0.075	0.074	38.72
0.950	64.08117	0.050	0.050	39.72
0.975	75.83965	0.025	0.025	40.72
0.990	95.94090	0.000	0.000	41.72

AREA UNDER DUPUIT CURVE: C VS. T = 0.9665761
AREA UNDER LUMPED CURVE: C VS. T = 0.9672343

Modified U.S. Corps of Engineers Program

00:00:00
00:00:01

// Q PEEPEE FORTRAN
// FLOW EXEC

```

0001 C C C C C C C C C C C C
0002 C C C C C C C C C C C C
0003 C C C C C C C C C C C C
0004 C C C C C C C C C C C C
0005 C C C C C C C C C C C C
0006 C C C C C C C C C C C C
0007 C C C C C C C C C C C C
0008 C C C C C C C C C C C C
0009 C C C C C C C C C C C C
0010 C C C C C C C C C C C C
0011 C C C C C C C C C C C C
0012 C C C C C C C C C C C C
0013 C C C C C C C C C C C C
0014 C C C C C C C C C C C C
0015 C C C C C C C C C C C C
0016 C C C C C C C C C C C C
0017 C C C C C C C C C C C C
0018 C C C C C C C C C C C C
0019 C C C C C C C C C C C C

FINITE ELEMENT SOLUTION OF STEADY STATE POTENTIAL FLOW PROBLEMS
QUADRILATERAL AND/OR TRIANGULAR ELEMENTS
PROGRAM 723-G2-L2440
NOVEMBER 1970
MODIFIED OCTOBER 1971 BY R. COOLEY
MODIFIED IN FEBRUARY 1977, BY S.G. MCLIN
SUBROUTINE GEES IS CALLED FROM STATEMENT 205.03
SUBROUTINE SOLVE IS CALLED FROM STATEMENTS 47.01 AND 300.01
SUBROUTINE PRTOA IS CALLED FROM STATEMENTS 250.02 AND 265.03
SUBROUTINE PRTOB IS CALLED FROM STATEMENTS 250.02 AND 265.03
SUBROUTINE TITLE(21),XCORD(861),YCORD(861),NODE(861),IZONE(861),
DIMENSION YPERM(5),ALPHA(5),NDA(1),NDB(1),PS(141,21)
1 QWELL(1),IWLEL(1),CORDY(41),CORDX(21),NDS,NHDS,XCORD,YCORD,NODE,IZONE,XPERM,YPERM,
3 PRES(1),CORNED,NELS,OMELL,IWLEL,OBND,NDA,NDB,THICK,NMELS,NQBND,
COMMON R,C,NRED,NEL,NMELS,NELT,NPL,N
1 ALPHA,HEAD,NDHD,MAXM,OMELL,IWLEL,OBND,NDA,NDB,THICK,NMELS,NQBND,
2 RADLR,NBEG,NTELS,NDELS,NELT,NPL,N
1 FORMAT (1H,17,9I8)
2 FORMAT (1X,17,9F8.0)
3 FORMAT (1X,1A1,2A3,18A4)
4 FORMAT (1X,17,9F8.0)
5 FORMAT (2X,16,F8.0,7,4I8,F8.0)
6 FORMAT (1H0,10HINPUT DATA)
7 FORMAT (1H,14,9(3X,F10.2))
8 FORMAT (1H,1)
9 FORMAT (1H0,6HOUTPUT)
10 FORMAT (1H0,15X,18HCoefficient Matrix/1H,5X,3HROW,8X,6HCOLUMN,16X
1,1HC)
11 FORMAT (1H,5X,13,9X,13,13X,E10.3)
12 FORMAT (1H0,29X,12HKNOWN VECTOR/1H,5H ROW,10X,1H8,10X,3HROW,10X
1,1HR,10X,3HROW,10X,1HR)
13 FORMAT (1H,5X,13,2(11X,F11.4))
14 FORMAT (1H0,18X,33HCOMPUTED VALUES OF HYDRAULIC HEAD/1H,6H NODE
1,8X,4HHEAD,8X,4HNODE,8X,4HHEAD,8X,4HNODE)
15 FORMAT (45HKNOWN VALUES OF HYDRAULIC HEAD AT BOUNDARIES/1H,5X,
14HNODE,14X,4HHEAD)
16 FORMAT (10HMAXIMUM OF 15,24H, CORRESPONDING TO NODE,15,68H, IS GRE
1,ATER THAN ORDER OF MATRIX (NRED) OR EXCEEDS DIMENSION LIMITS)
17 FORMAT (41HNO. OF QUADRILATERAL ELEMENTS (NOELS) = ,16/38H NO. OF
1 TRIANGULAR ELEMENTS (NTELS) = ,16/23H NO. OF NODES (NND) = ,16/

```



```

24 IH NJ, OF MATERIAL PROPERTY ZONES (NZNS) = ,16/42H NO. OF SPECIFI
3 ED BOUNDARY HEADS (NHDS) = ,16/35H NO. OF SOURCES OR SINKS (NWELS)
4 = ,16/74H NO. OF BOUNDARY ELEMENT SIDES (NOBND) FOR WHICH DISCH
5 CE IS SPECIFIED = ,16)
18 FORMAT (IHO,8X,5HC,COMPUTED VALUES OF HYDRAULIC HEAD AND FLUID PRES
19 SURE HEAD/1H ,6H NODE,5X,8HHYD. HD.,6X,10HPRESS. HD.,4X,4HNODE
20 5X,8HHYD. HD.,6X,10HPRESS. HD.)
19 FORMAT (IHO,2X,4HNODE,6X,5HXCORD,10X,5HXCORD,7X,4HNODE,6X,5HXCOR
1,10X,5HXCORD)
20 FORMAT (IHO,20X,22HQUADRILATERAL ELEMENTS/9HO ELEMENT,2X,6HNODE 1,
21 2X,6HNODE 2,2X,6HNODE 3,2X,6HNODE 4,4X,4HZONE,4X,9HTHICKNESS)
21 FORMAT (5HOZONE,RX,5HXPERM,RX,5-TYPE M,8X,5HALPHA)
22 FORMAT (IHO,9X,9HSPECIFIED/1H ,4HNODE,6X,8HBOUNDARY/1H ,12X,4HHEAD
1)
23 FORMAT (1X,17,518,4FR,0)
24 FORMAT (2X,16,518,4X,F10.2)
25 FORMAT (63HO SOURCES OR SINKS (+ INDICATES RECHARGE, - INDICATES DI
26 SCHARGE)/1H ,19HELEMENT DISCHARGE)
26 FORMAT (IHO,29HSPECIFIED BOUNDARY DISCHARGES/1H ,6X,28HUNIT B
1 BOUNDARY BOUNDARY/1H ,33H DISCHARGE NODE A, NODE B)
27 FORMAT (1X,17,18,8F8.0)
28 FORMAT (1X,3X,F10.3,4X,14,6X,14)
29 FORMAT (63HKNOWN VALUES OF HYDRAULIC HEAD AND PRESSURE HEAD AT 80
UNDARIES/1H ,5X,4HNODE,12X,8HHYD. HD,13X,10HPRESS. HD)
31 FORMAT (IHO,40H*** AXI-SYMMETRIC RADIAL COORDINATES ***))
33 FORMAT (1H ,14,5X,2E10.4,13X),F8.2)
35 FORMAT (IHO,37HMAXIMUM MATRIX BAND WIDTH (MAXBW) OF ,14,21H CORRES
PONDNS TO NODE ,14)
36 FORMAT (IHO,35HSCALE CHANGE FOR NODAL COORDINATES:1
37 EFORMAT (IHO,18X,19HTRIANGULAR ELEMENTS/9HO ELEMENT,2X,6HNODE 1,2X,
16HNODE 2,2X,6HNODE 3,4X,4HZONE,4X,9HTHICKNESS)
38 FORMAT (1X,17,418,4F8.0)
39 FORMAT (1X,17,418,5X,F10.2)
40 FORMAT (1X,17,0,9F8.0)
444 FORMAT (2110,5F10.0,15)
445 FORMAT (1X,1X,NUMBER OF ROWS = ,13,/,1X,NUMBER OF COLUMNS = ,13
, F7.0,/,1X,AQUIFER DEPTH = ,F6.0,/,1X,D
<DISTANCE BETWEEN COLUMN NODES = ,F5.1,/,1X,DISTANCE BETWEEN ROW
<NODES = ,F6.2,/,1X,AFTER',13, ROWS, DISTANCE BETWEEN ROW NODES
<= ,F7.4)
45 PRINT
CARD ADDED
READ (5,444) MML,NML,S,D,CONSTX,CNSTY,CONSTA,NROWS
CARD ADDED

```

0020

0021

0022

0023

0024

0025

0026

0027

0028

0029

0030

0031

0032

0033

0034

0035

0036

0037

0038

0039

0040

0041

0042

0043

0044

0045

0046

0047

0048

0049

C
C
C
C

C
C

00CARD0A

00CARD0B

READ 3 TITLE CARDS

```

DO 46 I=1,3
READ 3, (TITLE(N),N=1,21)
PRINT 3, (TITLE(N),N=1,21)
CONTINUE
46 PRINT 445,MML,NML,S,D,CONSTX,CNSTY,NROWS,CONSTA

```

READ JOB SPECIFICATION

READ 2,NQELS,NTELS,NNDS,NZNS,NHDS,NWELS,NQBND

00CARD0C

```

IF (NNDS.LE.0) STOP
READ 2,IRADL,IPRGS,IUNIT,IPRES
PRINT 6
PRINT 17,NQELS,NTELS,NNDS,NZNS,NHDS,NWELS,NQBND
IF (IRADL.GT.0) PRINT 31
    
```

READ NODAL COORDINATES

00CARD0D

```

CARD ADDED
PRINT 19
DO 47 J=1,NNDS
  READ 4, I,XCORD(I),YCORD(I)
47 CONTINUE
CALL CORDXY(NN1,MML,CONSTX,CONSTY,CONSTA,S,D,NROWS)
CALL PRTOPTA(XCORD,YCORD,NNDS)
CARD ADDED
    
```

00CARD0E

```

CONVERT SCALE
IF (IUNIT.LF.0) GO TO 51
PRINT 36
READ 3, (TITLE(N),N=1,21)
PRINT 3, (TITLE(N),N=1,21)
READ 40,CONST
DO 50 I=1,NNDS
  XCORD(I)=XCORD(I)*CONST
  YCORD(I)=YCORD(I)*CONST
50 CONTINUE
    
```

00CARD0G

```

READ QUADRILATERAL ELEMENT DATA
NELT=0
IF (NQELS.LE.0) GO TO 53
CARD ADDED
PRINT 20
DO 52 K=1,NQELS
  READ 23, I,(NODE(I,J),J=1,4), IZONE(I),THICK(I)
  PRINT 24, I,(NODE(I,J),J=1,4), IZONE(I),THICK(I)
  IF (IZONE(I).LE.0) IZONE(I)=1
  CALL CORRD(MN1,MML,NROWS)
DO 52 I=1,NQELS
  THICK(I)=0.0
PRINT 24,I,(NODE(I,J),J=1,4), IZONE(I),THICK(I)
52 CONTINUE
CARD ADDED
NELT=1
NPL=5
N=4
NELS=NQELS
NREG=1
    
```

00CARD0H

```

READ TRIANGULAR ELEMENT DATA
IF (NTELS.LE.0) GO TO 55
53 PRINT 37
    
```

0050
0051
0052
0053
0054

0055

0056
0057
0058
0059
0060
0061
0062
0063
0064

0065
0066

0067
0068
0069

0070

0071
0072
0073
0074
0075

0076
0077

00CARD0H

```

00 54 K=1,NTELS
READ 36,I,(NODE(I),J),IZONE(I),THICK(I)
PRINT 39,I,(NODE(I),J),J=1,3; IZONE(I),THICK(I)
IF (IZONE(I).LE.0) IZONE(I)=1
54 CONTINUE
NLT=NELT+1
NPI=4
N=3
NELS=NQELS+NTELS
NBEG=NQELS+1

```

READ MEDIA PROPERTIES

00CARD0I

```

55 PRINT 21
DO 56 J=1,NZNS
READ 4,I,XPERM(I),YPERM(I),ALPHA(I)
PRINT 33,I,XPERM(I),YPERM(I),ALPHA(I)
56 CONTINUE

```

READ BOUNDARY CONDITIONS

00CARD0J

```

IF (NHDS.LE.0) GO TO 65
PRINT 22
CARD ADDED
DO 60 I=1,NHDS
READ 4,NDHD(I),HEAD(I)
PRINT 7,NDHD(I),HEAD(I)
60 CONTINUE
ICON=(NMI-2)*(MMI-2)
INUM=NHDS-MMI
JNUM=INUM+1
KNUM=INUM-1
LNUM=MMI+2
MNUM=LNUM-1
DO 60 I=1,MMI
NDHD(I)=I+ICON
HEAD(I)=I.0
HEAD(I)=0.0

```

0094
0095
0096
0097
0098
0099
0100
0101
0102
0103
0104
0105
0106
0107
0108
0109
0110
0111
0112
0113

```

DO 61 I=LNUM,INUM,2
NDHD(I)=I+ICON
HEAD(I)=I.0
DO 62 I=JNUM,NHDS
NDHD(I)=I+ICON
HEAD(I)=I.0
DO 63 I=MNUM,KNUM,2
NDHD(I)=I+ICON
IX=I+ICON
HEAD(I)=XC ORD(I)/S
DO 64 I=1,NHDS
PRINT 7,NDHD(I),HEAD(I)
64 CARD ADDED
IF (NMELS.LE.0) GO TO 75
65 PRINT 25
DO 70 I=1,NMELS
READ 4,I,IMLEFL(I),OWELL(I)
PRINT 1,I,IMLEFL(I),OWELL(I)

```

C C

0114
0115
0116
0117
0118

00CARDOK

```

0119
0120
0121
0122
0123
0124
0125
0126
0127
0128
0129
0130
0131
0132
0133
0134
0135
0136
0137
0138
0139
0140
0141
0142
0143
0144
0145
0146
0147
0148
0149
0150
0151
0152
0153
0154
0155
0156
0157
0158
0159
0160
0161
0162
0163
0164
0165
00CARDOL
70 CONTINUE
75 IF(NQBND.LE.0) GO TO 81
   PRINT 26
   DO 80 I=1,NQBND
     READ 27,NDA(I),NDR(I),QBND(I)
     PRINT 28,QBND(I),NDA(I),NDR(I)
80 CONTINUE
      QBND FOR AXI-SYMMETRIC RADIAL COORDINATES
81 IF((IRADL.LE.0.OR.NQBND.LE.0) GO TO 85
   DO 83 I=1,NQBND
     QBND(I)=QBND(I)/6.283
83 CONTINUE
      DETERMINE MAXIMUM BAND WIDTH (MAXBW)
85 NRED=NND5-NHDS
   MAXRW=0
   DO 200 I=1,NRED
     MINND=VRED
     MAXND=0
     NT=N
     NRG=NELS
     NRG=NREG
     DO 190 NO=1,NELT
       DO 180 J=NRGI,NTMP
         DO 110 K=1,NT
           IF (N3DE(J,K).EQ.1) GO TO 120
110 CONTINUE
         GO TO 180
120 DO 150 K=1,NT
       NJDE=NDE(J,K)
       IF (NUDE.GT.NRED) GO TO 150
       IF (NUDE.GT.MAXND) MAXND=NUDE
       IF (NUDE.LT.MINND) MINND=NUDE
150 CONTINUE
180 CONTINUE
   NRG=1
   NTMP=NELS
   NT=4
190 CONTINUE
   NTEMP=1-MINND
   IF ((MAXND-I).GT.NTEMP) NTEMP=MAXND-I
   IF (NTEMP.LE.MAXBW) GO TO 200
   MAXRW=NTEMP
   KEYND=1
200 CONTINUE
   MAXBW=MAXBW*2+1
   IF (MAXRW.LE.50 .AND. MAXBW.LE.NRED) GO TO 205
   PRINT 16,MAXBW,KEYND
   GO TO 45
205 CONTINUE
   PRINT 8
   PRINT 9
   PRINT 35,MAXBW,KEYND

```

```

0166      CALL SURROUTINE GEES TO SET UP MATRIX EQUATION
0167      IF (IPRGS.LE.0) GO TO 265
0168      PRINT 10
0169      DO 250 I=1,NRED
0170      DO 210 J=1,MAXBW
0171      NROW=I
0172      IF (G(I,J).EQ.0.) GO TO 210
0173      PRINT 11,(NROW,NCOL,G(I,J))
0174      CONTINUE
0175      CONTINUE
0176      PRINT 12
0177      CALL PRTOTB(R,NRED)
0178
0179      SOLVE MATRIX EQUATION
0180      265 CALL RSOLVE(G,B,NRED,MAXBW)
0181      PRINT HYDRAULIC HEADS OR HYDRAULIC HEADS AND PRESSURE HEADS
0182
0183      IF (IPRES.GT.0) GO TO 290
0184      PRINT 14
0185      CALL PRTOTB(R,NRED)
0186      CARDS ADDED
0187      PRINT 15
0188      DO 280 I=1,NHDS
0189      PRINT 13,NHDI(I),HEAD(I)
0190      CONTINUE
0191      ETA=S/D*SORT(IPERM(1))/XPERM(1)
0192      PRINT 1119,ETA
0193      FORMAT(1H1,10X,'ETA = ',F5.1,'//10X,'STREAMLINE VALUES')
0194      DELX=0.0
0195      CORDY(1)=0.0
0196      CORDY(2)=0.0
0197      DO 312 I=2,NMI
0198      DELX=DELX+CONSTX
0199      CORDX(I)=DELX
0200      IF (J.LF.NROWS) DELY=DELY+CONSTY
0201      IF (J.GT.NROWS) DELY=DELY+CONSTA
0202      CORDY(J)=DELY
0203      PSI(1,I)=0.0
0204      DO 316 I=2,NMI
0205      PSI(I,MMI)=1.0
0206      PSI(1,I)=CORDX(I)/S
0207      DO 317 J=2,MML
0208      PSI(I,J)=1.0
0209      PSI(NMI,J)=1.0
0210      MSUR1=MMI-1
0211      NSUR1=NMI-1
0212      KOUNT=0
0213      DO 318 I=2,NSUR1
0214

```

```

0206 DD 318 J=2,MSUB1
0207 KOUNT=KOUNT+1
0208 PSI(I,J)=R(KOUNT)
0209 DD 319 I=1,NN1
0210 PRINT I,09,(PSI(I,J),J=1,MM1)
0211 FORMAT(1H0,22(1X,F5.3))
0212 AREA=S*0
0213 DD 320 I1=1,NSUB1
0214 AREA=0.0
0215 J=1
0216 I=NN1-I1
0217 IF(I.E0. 1)GO TO 320
0218 WRITE(6,1113)X,STREAMLINE',5X,'X-COORD',5X,'Y-COORD',//)
0219 FORMAT(1H1,10X,STREAMLINE',5X,'X-COORD',5X,'Y-COORD',//)
0220 STRMLN=PSI(I,J)
0221 XI=CORDX(I)
0222 YI=CORDY(I)
0223 WRITE(6,1114)STRMLN,XI,YI
0224 S1=STRMLN
0225 XI=XI
0226 YI=YI
0227 J=J+1
0228 S2=PSI(I,J)
0229 Y2=CORDY(I)
0230 Y2=CORDY(J)
0231 I=I-1
0232 S3=CORDX(I)
0233 X3=CORDX(I)
0234 Y3=CORDY(I)
0235 J=J-1
0236 S4=PSI(I,J)
0237 X4=CORDX(I)
0238 Y4=CORDY(I)
0239 IF(S3-STRMLN)322,321,323
0240 321 Y12=Y3
0241 S1=S3
0242 XI=X3
0243 Y1=Y3
0244 J=J+1
0245 GO TO 325
0246 S12=S2
0247 XI2=X2
0248 Y12=Y2
0249 S21=S3
0250 X21=X3
0251 Y21=Y3
0252 IF(Y2.NE. Y3)GO TO 324
0253 A=S12-S21
0254 BR=S12-X21
0255 C=S12-STRMLN
0256 XI2=X12-BA*C/A
0257 I=I+1
0258 J=J+1
0259 S1=S2
0260 XI=X2

```

```

0262
0263
0264
0265
0266
0267
0268
0269
0270
0271
0272
0273
0274
0275
0276
0277
0278
0279
0280
0281
0282
0283
0284
0285
0286
0287
0288
0289
0290
0291
0292
0293
0294
0295
0296
0297
0298
0299
0300
0301
0302
0303
0304
0305
0306
0307
0308
0309
0310
0311
0312
0313
0314
0315
0316
0317

Y1=Y2
GO TO 325
IF(S4-STRMLN)326,332,327
S12=X4
X12=Y4
Y12=X3
S21=S3
X21=X3
Y21=Y3
IF(X3-NE. X4)GO TO 328
A=S21-S12
RA=Y21-Y12
C=STRMLN-S12
Y12=Y12+BB*C/A
S1=S4
X1=X4
Y1=Y4
WRITE(6,1114)STRMLN,X12,Y12
FORMAT(1H0,15X,F5.3,5X,F10.5,3X,F10.5)
AREA=AREA+0.5*(X1-X12)*(Y1+Y12)
XI=X12
YI=Y12
IF(I-1.EQ. 1 .AND. J .EQ. 1)GO TO 329
GO TO 331
WRITE(6,1115)X12,Y12
FORMAT(1H0,15X,F10.5,2X,'Y = ',F10.5)
WRONG AT X = ',F10.5,2X,'Y = ',F10.5)
GO TO 320
WRITE(6,1116)X12,Y12
FORMAT(1H0,15X,F10.5,2X,'Y-COORD = ',F10.5)
WRONG AT X= ',F10.5,2X,'Y-COORD = ',F10.5)
GO TO 320
S12=S1
X12=X1
Y12=Y1
S21=S4
X21=X4
Y21=Y4
IF(Y1-NE. Y4)GO TO 324
A=S12-S21
RA=X12-X21
C=STRMLN-S12
X12=X12-RR*C/A
S2=S1
X2=X1
Y2=Y1
S3=S4
X3=X4
Y3=Y4
I=I+1
J=J+1
I(I,J)
S1=PSI(I,J)
X1=CORDY(I)
Y1=CORDY(J)
I=I-1
J=J-1
S4=PSI(I,J)
X4=CORDY(I)
Y4=CORDY(J)

```

```

0318 WRITE(6,1114)STRMLN,X12,Y12
0319 AREA=AREA+0.5*(X1-X12)*(Y1+Y12)
0320 XI=X12
0321 YI=Y12
0322 IF(I1.EQ.1.AND.J.EQ.1)GO TO 329
0323 GO TO 330
0324 XI2=0.0
0325 YI2=0.0
0326 WRITE(6,1114)STRMLN,X12,Y12
0327 AREA=AREA+0.5*(X1-X12)*(Y1+Y12)
0328 AREA=AREA/AREAL
0329 WRITE(6,1117)STRMLN,AREA,ETA
0330 FORMAT(//,1H0,'THE AREA UNDER PSI = ',F6.4,2X,'IS ',F10.7,3X,'(ET
1117 <A= ',F5.1,1,1)
0331 WRITE(7,2117)ETA,STRMLN,AREA
0332 FORMAT(F5.1,5X,F6.4,5X,F10.7)
0333 GO TO 320
0334 XI2=X4
0335 YI2=Y4
0336 WRITE(6,1114)STRMLN,X12,Y12
0337 AREA=AREA+0.5*(X1-X12)*(Y1+Y12)
0338 XI=X12
0339 YI=Y12
0340 IF(I1.EQ.1.AND.J.EQ.1)GO TO 329
0341 SI=54
0342 XI=X4
0343 YI=Y4
0344 I=I-1
0345 IF(S4-STRMLN)333,332,327
0346 I=I+1
0347 J=J+1
0348 S2=PSI(I,J)
0349 X2=CORDX(I)
0350 Y2=CORDY(J)
0351 I=I-1
0352 S3=PSI(I,J)
0353 X3=CORDX(I)
0354 Y3=CORDY(J)
0355 J=J-1
0356 GO TO 325
0357 CONTINUE
C
0358 CONTINUE
0359 CARDS ADDES
0360 GO TO 45
0290 PRINT 18
DO 300 I=1,NRED
0361 PRES(I)=B(I)-YCORD(I)
0362 CONTINUE
0300 CALL PRITOTA(B,PRES,NRED)
PRINT 29
DO 310 I=1,NHDS
0363 J=NDHD(I)
0364 PRES=HEAD(I)-YCORD(J)
0365 PRINT 13,NDHD(I),HEAD(I),PRES
0366 CONTINUE
0310 GO TO 45
END

```


TOTAL MEMORY REQUIREMENTS 00392C BYTES
 COMPILER HIGHEST SEVERITY CODE WAS 0

```

0001 C SUBROUTINE GEES TO CALCULATE MATRIX ELEMENTS
0002 C SURROUTINE TITLE (21), XCORD(861), YCORD(861), NODE(861,4), IZONE(861),
0003 C DIMENSION XPERM(5), YPERM(5), ALPHA(5), NDHD(120), HEAD(120), G(861,50), B(861),
0004 C QWELL(1), IWELL(1), ORND(1), NDAL(1), NDB(1), THICK(861)
0005 C DIMENSION XX(5), YY(5), HD(5), GT(5), ATRI(4)
0006 C COMMON B, G, XNRED, NELS, NNDS, NHDS, XCORD, YCORD, NODE, IZONE, XPERM, YPERM,
0007 C ALPHA, HEAD, NDHD, MAXBM, QWELL, IWELL, QBND, NDA, NDB, THICK, NHELS, NQBND,
0008 C LR=(MAXBM+1)/2
0009 C DO 2 I=1, NNDS
0010 C B(I)=0
0011 C DO 2 J=1, MAXBM
0012 C G(I,J)=0.
0013 C 2 CONTINUE
0014 C 2 ADD BOUNDARY DISCHARGES TO B VECTOR
0015 C IF(NQBND,LE,0) GO TO 4
0016 C DO 3 I=1, NQBND
0017 C K=NDB(I)
0018 C L=NDB(I)
0019 C DIST=(XCORD(K)-XCORD(L))*2+(YCORD(K)-YCORD(L))*2**0.5
0020 C TEMP=QBND(I)*DIST/2.
0021 C B(K)=B(K)+TEMP
0022 C B(L)=B(L)+TEMP
0023 C 3 CONTINUE
0024 C KOUNT=0
0025 C *** BEGIN OUTER TRIANGULAR AND/OR QUAD ELEMENT LOOP
0026 C 4 DO 300 ND=1, NELT
0027 C DO 200 I=NREG, NELS
0028 C XX(4)=0.0
0029 C YY(4)=0.0
0030 C ATRI(4)=0.
0031 C GC=0.
0032 C RCN=0.
0033 C DO 5 K=L, NPL
0034 C GT(K)=D.
0035 C J=IZONE(I)
0036 C ANGLE=ALPHA(J)
0037 C XK=L.0/YPERM(J)
0038 C YK=L.0/XPERM(J)
0039 C SET HD ARRAY
0040 C DO 9 JJ=L, N
0041 C ITMP=NODE(I, JJ)

```

```

0036 IF(ITMP.LE.NRED) GO TO 9
0037 DO 8 KK=1,NHDS
0038 JTMP=VDHD(KK)
0039 IF(JTMP.NE.JTMP) GO TO 8
0040 HD(JJ)=HEAD(KK)
0041 GO TO 9
0042 8 CONTINUE
0043 9 CONTINUE
0044 HD(NPL)=HD(1)

0045
0046
0047
0048
0049
0050
0051
0052
0053
0054
0055
0056
0057
0058
0059
0060
0061
0062
0063
0064
0065
0066

0067
0068
0069
0070
0071
0072
0073
0074
0075

0076
0077
0078
0079
0080
0081
0082

IF(ITMP.LE.NRED) GO TO 9
DO 8 KK=1,NHDS
JTMP=VDHD(KK)
IF(JTMP.NE.JTMP) GO TO 8
HD(JJ)=HEAD(KK)
GO TO 9
8 CONTINUE
9 CONTINUE
HD(NPL)=HD(1)

      DETERMINE LOCAL COORDINATES
DO 10 II=1,N
KA=NODE(II,II)
ND(II)=KA
XX(II)=XCORD(KA)
YY(II)=YCORD(KA)
10 ANG=AR5(ANGLE) GO TO 12
IF(ANG.LE.0.) GO TO 12
SNALP=STN(ANGLE)
CSALP=COS(ANGLE)
DO 11 I=1,N
TMP=XX(II)
YY=YY(II)
XX(II)=TEMP*CSALP+TMP*SNALP
11 YY(II)=-TEMP*SNALP+TMP*CSALP
Y4=YY(4)
XX(NPL)=XX(1)
YY(NPL)=YY(1)
ND(NPL)=ND(1)
TMP=N
XCN={XX(1)+XX(2)+XX(3)+X4}/TMP
YCN={YY(1)+YY(2)+YY(3)+Y4}/TMP

      *** BEGIN INNER TRIANGULAR ELEMENT LOOP
DO 100 II=1,N
NA=ND(II)
NR=ND(II+1)
XNA=XX(II)
YNA=YY(II)
XNB=XX(II+1)
YNB=YY(II+1)
XKT=XK
YKT=YK

      DETERMINE TRANSMISSIVITY TERMS
IF(IRADL.LE.0) GO TO 19
TMP=(XNA+XNB+XCN)/3.
GO TO 20
19 IF(THICK(II).LE.0.) GO TO 23
20 XKT=XKT*TMP
YKT=YKT*TMP

```

CC

```

THE MAZE
23 BJ=YNB-YCN
   BK=YCN-YNA
   RL=YNA-YNB
   CJ=XCX-XNR
   CK=XNB-XCN
   CL=XNS-XNA
   TEMP=2*(XNA*BJ+XNB*BK+XCN*BL)
   ATRI(11)=TEMP
   XKT=XKT/TEMP
   YKT=YKT/TEMP
   TPXL=YKT*BL
   TPYL=YKT*CL
   TPXJ=XKT*BJ
   TPYJ=YKT*CJ
   GC=GC+TPXL*BL+TPYL*CL
   IF(NA.LE.NRED) GO TO 35
   BCN=BCN-(TPXL*BJ+TPYL*CJ)*HD(11)
   IF(NR.LE.NRED) GO TO 32
   IF(BCN=BCN-(TPXL*BK+TPYL*CK)*HD(11+1)
31 GO TO 100
32 R(NB)=B(NR)-(TPXJ*RK+TPYJ*CK)*HD(11)
33 GT(11+1)=GT(11)+TPXL*BK+TPYL*CK
   G(NB,LR)=G(NB,LR)+XKT*BK*YKT*CK*CK
   GO TO 100
35 GT(11)=GT(11)+TPXJ*BL+TPYJ*CL
   G(NA,LR)=G(NA,LR)+TPXJ*BJ+TPYJ*CJ
   IF(NB.LE.NRED) GO TO 36
   B(NA)=B(NA)-(TPXJ*RK+TPYJ*CK)*HD(11+1)
   GO TO 31
36 NAT=NA-NB+LR
   NRT=NR-NR+LR
   G(NA,NRT)=G(NA,NRT)+TPXJ*BK+TPYJ*CK
   G(NB,NAT)=G(NA,NRT)
   GO TO 33
100 CONTINUE
   GT(11)=GT(11)+GT(NP1)

```

CC

```

ADD SOURCES OR SINKS TO B VECTOR
IF (NWELS.LF.O.OR.KOUNT.GE.NWELS) GO TO 50
DO 45 JJ=1,NWELS
IF (TWELL(JJ).NE.1) GO TO 45
KOUNT=KOUNT+1
TEMP=QWELL(JJ)/((ATRI(1)+ATRI(2)+ATRI(3)+ATRI(4))*3.)
DO 40 KK=1,N
KA=ND(KK+1)
KB=ND(KK+1)
R(KA)=B(KA)+TEMP
R(KB)=B(KB)+TEMP
RCN=RCN+TEMP
40 GO TO 50
45 CONTINUE

```

0083
0084
0085
0086
0087
0088
0089
0090
0091
0092
0093
0094
0095
0096
0097
0098
0099
0100
0101
0102
0103
0104
0105
0106
0107
0108
0109
0110
0111
0112
0113
0114
0115
0116
0117
0118

0119
0120
0121
0122
0123
0124
0125
0126
0127
0128
0129
0130
0131
0132

ELIMINATE CENTRAL NODE POINT OF EACH OUTER TRIANGULAR OR
 QUADRILATERAL ELEMENT FROM MATRIX EQUATION

C
 C
 C

```

0133 DO 61 J,J=1,N
0134 KA=ND(J,J)
0135 IF(KA.GT.NRED) GO TO 61
0136 TEMP=GT(J,J)/GC
0137 R(KA)=R(KA)-BCN*TEMP
0138 DO 60 KK=1,N
0139 KR=ND(KK)
0140 IF(KB.GT.NRED) GO TO 60
0141 KC=KB-KA*LR
0142 G(KA,KC)=G(KA,KC)-GT(KK)*TEMP
0143 CONTINUE
0144 61 CONTINUE
0145 200 NBEG=1
0146 NELS=N-4
0147 NPL=5
0148 N=4
0149 300 CONTINUE
0150 RETURN
0151 END
0152
    
```

TOTAL MEMORY REQUIREMENTS 000EFC BYTES

COMPILER HIGHEST SEVERITY CODE WAS 0

```

0001 SUBROUTINE RSOLVE(C,V,N,M)
0002 SOLVE MATRIX EQUATION BY GAUSS ELIMINATION
0003 DIMENSION C(861,50),V(861)
0004 LR=(M-1)/2
0005 DO 20 L=1,LR
0006 IM=LR-L+1
0007 DO 20 J=1,IM
0008 C(L,J-1)=C(L,J)
0009 KN=N-L
0010 C(L,M)=0
0011 C(KN+1,KM+1)=0
0012 LR=LR+1
0013 IM=N-1
0014 DO 90 I=1,IM
0015 NPIV=I
0016 LS=I+1
0017 DO 30 L=LS,LR
0018 IF(ABS(C(L,I)).GT.ABS(C(NPIV,I))) NPIV=L
0019 CONTINUE
0020 IF(NPIV.LE.I) GO TO 50
0021 DO 40 J=1,M
0022 TEMP=C(I,J)
    
```

0001
 0002
 0003
 0004
 0005
 0006
 0007
 0008
 0009
 0010
 0011
 0012
 0013
 0014
 0015
 0016
 0017
 0018
 0019
 0020
 0021
 0022
 0023

```

0024 C(I,J)=C(NPIV,J)
0025 C(NPIV,J)=TEMP
0026 TEMP=V(I)
0027 V(I)=V(NPIV)
0028 V(NPIV)=TEMP
0029 V(I)=V(I)/C(I,I)
0030 DO 60 J=2,M
0031 C(I,J)=C(I,J)/C(I,I)
0032 DO 60 L=LS,LR
0033 TEMP=C(L,I)
0034 V(L)=V(L)-TEMP*V(I)
0035 DO 70 J=2,M
0036 C(L,J-1)=C(L,J)-TEMP*C(I,J)
0037 C(L,M)=0
0038 IF(LR.LT.N) LR=LR+1
0039 CONTINUE
0040 V(N)=V(N)/C(N,I)
0041 JM=2
0042 DO 110 I=1,IM
0043 L=N-I
0044 DO 100 J=2,JM
0045 KM=L+J
0046 V(L)=V(L)-C(L,J)*V(KM-1)
0047 IF(JM.LT.M) JM=JM+1
0048 CONTINUE
0049 RETURN
0050 END

```

TOTAL MEMORY REQUIREMENTS 0006D8 BYTES
 COMPILER HIGHEST SEVERITY CODE WAS 0

```

0001 C SURROUTINE PRIOATA(VALA,VALB,NO)
0002 PRINT OUT VALUES IN TWO GROUPS OF THREE COLUMNS
0003 DIMENSION VALA(861),VALB(861)
0004 IEND=NO/2
0005 ITMP=(NO+1)/2
0006 DO 1 I=1,IEND
0007 K=ITMP+1
0008 PRINT 20,I,VALA(I),VALB(I),K,VALA(K),VALB(K)
0009 1 CONTINUE
0010 IF(ITMP.EQ.IEND) GO TO 10
0011 PRINT 20,ITMP,VALA(ITMP),VALB(ITMP)
0012 FORMAT(1H ,2X,2(I4,4X,F6.1,9X))
0013 10 RETURN

```

TOTAL MEMORY REQUIREMENTS 0002DC BYTES
 COMPILER HIGHEST SEVERITY CODE WAS 0

```

0001 SUBROUTINE PRTOB(VAL,NOI)
0002 PRINT OUT VALUES IN THREE GROUPS OF TWO COLUMNS
0003 DIMENSION VAL(861)
0004 IEND=NO/3
0005 ITEMP=NO-IEND-IEND-IEND
0006 IF(IITEMP.EQ.0) GO TO 5
0007 ITEMP=IEND+1
0008 IF(IITEMP.EQ.1) IEND=IEND-1
0009 5 DO 10 I=1,IEND
0010 K=ITEMP+1
0011 L=ITEMP+K
0012 PRINT 30,I,VAL(I),K,VAL(K),L,VAL(L)
0013 CONTINUE
0014 IF(IITEMP.EQ.0) GO TO 25
0015 I=IEND+1
0016 K=K+1
0017 PRINT 30,I,VAL(I),K,VAL(K)
0018 I=I+1
0019 K=K+1
0020 PRINT 30,I,VAL(I),K,VAL(K)
0021 30 FORMAT(1H ,2X,3(14,4X,F6.4,10X))
0022 25 RETURN
0023 END
0024

```

TOTAL MEMORY REQUIREMENTS 0003E0 BYTES
 COMPILER HIGHEST SEVERITY CODE WAS 0

```

0001 SUBROUTINE COORD(NNI,M,NROWS)
0002 GENERATES ELEMENT NUMBERS AND SURROUNDING NODE NUMBERS FOR THE N*M
0003 MATRIX - 0 DATA CARDS
0004 DIMENSION TITLE(21),XCORD(861),YCORD(861),NODE(861,4),IZONE(861),
0005 1 XPERM(5),YPERM(5),ALPHA(5),NDHD(120),HEAD(120),G(861,50),B(861),
0006 2 OMELL(1),TWLEL(1),ORND(1),NDA(1),NDR(1),THICK(861),
0007 COMMON A,B,G,NRED,NELS,NMDS,NMDS,ACCORD,YCORD,NODE,IZONE,XPERM,YPERM,
0008 1ALPHA,HEAD,NDHD,MAXW,OMELL,TWLEL,ORND,NDA,NDB,THICK,NWELS,NORND,
0009 2TRADL,LR,NBEG,NTELS,NOELS,NELT,NPL,N
0010 KM=0
0011 K=0
0012 L=(NN1-2)*(M-2)
0013 KL=NN1-(NN1-2)
0014 K2=NN1-2
0015 K3=M-(M-2)
0016 K4=M-2
0017 NODE(1,1)=K+1
0018 NODE(1,2)=L+2
0019 NODE(1,3)=L+1
0020 NODE(1,4)=L+M+1
0021 IZONE(1)=1

```

```

0016 K=K+1
0017 L=L+1
0018 DO 10 J=K3,K4
0019 KELE=1
0020 KELE=KELE+1
0021 NODE(J,1)=K+1
0022 NODE(J,2)=L+2
0023 NODE(J,3)=L+1
0024 NODE(J,4)=K
0025 IF(KELE.LT.NROWS) IZONE(J)=1
0026 IF(KELE.GE.NROWS) IZONE(J)=2
0027 K=K+1
0028 L=L+1
10
0029 NODE(M-1,1)=L+4
0030 NODE(M-1,2)=L+2
0031 NODE(M-1,3)=L+1
0032 NODE(M-1,4)=K
0033 IZONE(M-1)=2
0034 N4=(M-1)*(NN1-1)-2*(M-1)+1
0035 MMM=M-1
0036 K=K+1
0037 L=L+3
0038 DO 20 I=M,NM,MMM
0039 KELE=1
0040 NODE(I,1)=K
0041 NODE(I,2)=K-K4
0042 NODE(I,3)=L
0043 NODE(I,4)=L+2
0044 IZONE(I)=1
0045 L=L+1
0046 K=K+1
0047 DO 30 J=K3,K4
0048 KELE=KELE+1
0049 NODE(J+1-1,1)=K-K4
0050 NODE(J+1-1,2)=K-K4-1
0051 NODE(J+1-1,3)=K-K4-1
0052 NODE(J+1-1,4)=K-1
0053 IF(KELE.LT.NROWS) IZONE(J+1-1)=1
0054 IF(KELE.GE.NROWS) IZONE(J+1-1)=2
30
0055 K=K+1
0056 L=L+1
0057 NODE(I+K4,1)=L+2
0058 NODE(I+K4,2)=L
0059 NODE(I+K4,3)=K-K4-1
0060 IZONE(I+K4)=2
20
0061 L=L+1
0062 N4=(M-1)*(NN1-1)-(M-2)
0063 NODE(NM,1)=L+3
0064 NODE(NM,2)=K-K4
0065 NODE(NM,3)=L
0066 NODE(NM,4)=L+2
0067 IZONE(NM)=1
0068 KELE=1
0069 K=K+1
0070 L=L+3
0071 DO 40 J=K3,K4

```

```

0072 KFL=KFL+1
0073 MM=MM+J-1
0074 NNODE(MM,1)=L+1
0075 NNODE(MM,2)=K-K4
0076 NNODE(MM,3)=K-K4-1
0077 NNODE(MM,4)=L
0078 IF(.GE.NROWS) IZONE(MM)=1
0079 IF(.GE.NROWS) IZONE(MM)=2
0080 K=K+1
0081 L=L+1
0082 NN=(NN1-1)*(M-1)
0083 NNODE(NN,1)=L+1
0084 NNODE(NN,2)=L-K4
0085 NNODE(NN,3)=K-K4-1
0086 NNODE(NN,4)=L
0087 IZONE(NN)=L
0088 RETURN
0089 END

```

TOTAL MEMORY REQUIREMENTS 00095C BYTES
 COMPILER HIGHEST SEVERITY CODE WAS 0

```

0001 SUBROUTINE CORDXY(NN1,M,CONSTX,CONSTY,CONSTA,S,D,NROWS)
0002 C
0003 C
0004 C
0005 C
0006 C
0007 C
0008 C
0009 C
0010 C
0011 C
0012 C
0013 C
0014 C
0015 C
0016 C
0017 C
0018 C
0019 C
0020 C
0021 C
0022 C
0023 C

```

```

      GFNERATES X AND Y COORDINATES FOR AN N*MM NODAL FIELD-C DATA CARDS
      DIMENSION TITLE(21),XCORDE(861),YCORD(861),NDE(861,6),IZONE(861),
      1 XPERM(5),YPERM(5),ALPHA(5),VDHD(120),HEAD(120),G(861,50),B(861),
      2 COMMON B,G,NRED,NELS,NNDS,NHDS,XCORD,YCORD,NODE,THICK(861)
      3 LALPHA,HEAD,NHDH,MAXBW,OWELL,IWELL,OBND,NDAN,NDR,THICK,NWELS,NQBND,
      4 NMAX=NN1*M
      K=0
      L=(NN1-2)*(M-2)
      K1=NN1-(NN1-2)
      K2=NN1-1
      K3=M-(M-2)
      K4=M-1
      SUMY=0.0
      DO 10 J=1,M
      L=L+1
      XCORDE(L)=0.0
      YCORD(L)=SUMY
      IF(J.LT.NROWS) SUMY=SUMY+CONSTY
      IF(J.GE.NROWS) SUMY=SUMY+CONSTA
      SUMX=0.0
      DO 20 I=K1,K2
      L=L+1
      SUMX=SUMX+CONSTX
      XCORDE(L)=SUMX

```



```

0024 YCORD(L)=SUMY
0025 DO 30 J=K3,K4
0026 K=K+1
0027 IF(J.LE.NROWS)SUMY=SUMY+CONSTY
0028 IF(J.GT.NROWS)SUMY=SUMY+CONSTA
0029 XCORD(K)=SUMX
0030 YCORD(K)=SUMY
0031 CONTINUE
0032 L=L+1
0033 XCORD(L)=SUMX
0034 YCORD(L)=SUMY+CONSTA
0035 SUMY=0.0
0036 CONTINUE
0037 SUMX=0.0
0038 SUMY=SUMX+CONSTX
0039 DO 40 J=1,M
0040 L=L+1
0041 XCORD(L)=SUMX
0042 YCORD(L)=SUMY+CONSTA
0043 IF(J.LE.NROWS)SUMY=SUMY+CONSTY
0044 IF(J.GT.NROWS)SUMY=SUMY+CONSTA
0045 XMAX=XCORD(NMAX)
0046 YMAX=YCORD(NMAX)
0047 IF(XMAX.NE.S.AND.YMAX.NE.DIGO TO 50
0048 RETURN
0049 WRITE(6,45)XMAX,YMAX
0050 45 FORMAT(//',,10X,',X AND Y COORDINATES ARE WRONG: XMAX = 'F7.2,3X,',Y
      < ,',F7.2)
0051 CALL EXIT
0052 END

```

TOTAL MEMORY REQUIREMENTS 000674 BYTES

COMPILER HIGHEST SEVERITY CODE WAS 0

BATCH HIGHEST SEVERITY CODE WAS 0

```

// EXEC LNKEDT
LIST PHASE FLOW,*
LIST INCLUDE FLOW,0001,L
LIST INCLUDE FLOW,0002,L
LIST INCLUDE FLOW,0003,L
LIST INCLUDE FLOW,0004,L
LIST INCLUDE FLOW,0005,L
LIST INCLUDE FLOW,0006,L
LIST TFCOM#
LIST AUTOLINK
LIST AUTOLINK PRXPR#
LIST AUTOLINK SIN
LIST AUTOLINK USEROPT
LIST AUTOLINK FIOCS#
LIST AUTOLINK UNITAB#
LIST AUTOLINK ALOG
LIST AUTOLINK EXP
LIST ENTRY

```

```

00:00:59
CREATED
CREATED
CREATED
CREATED
CREATED
CREATED
CREATED

```

```

FROM SYSREL
FROM SYSREL
FROM SYSREL
FROM SYSREL
FROM SYSREL
FROM SYSREL
FROM SYSREL

```

78/019	PHASE	TRANSFER ADDR.	LOCORE	HICORE	BLOCK NO.	ESD TYPE	LABEL	LOADED	REL-FACTOR
COMMON						COMMON		008200	000000
						CREAT	MASSIVE	000000	000000

FLUR	USMU00	USMU00	USMU00	USMU00	USMU00	USMU00
				MAIN44	03A00R	03A000
* ENTRY				GEESG	03D938	03D938
CSECT ENTRY				B SOLVE	03E838	03E838
CSECT ENTRY				PRTOA	03FF10	03FF10
CSECT ENTRY				PRTOB	03FF10	03FF10
CSECT ENTRY				COORD	03F5D0	03F5D0
CSECT ENTRY				CORDX	03FF30	03FF30
CSECT ENTRY				BOAIRCOM	0405A8	0405A8
* ENTRY				IRSCC#	041881	041881
* ENTRY				FIRSTIM	041A20	041A20
* ENTRY				MSGCC#	041A27	041A27
* ENTRY				ADCON#	040664	040664
CSECT ENTRY				BOASSQRT	0430E0	0430E0
CSECT ENTRY				SQRT	0430E6	0430E6
CSECT ENTRY				BOAFRXP	043190	043190
CSECT ENTRY				FRXPR#	043198	043198
CSECT ENTRY				BOASSCN	043288	043288
CSECT ENTRY				SIN	0432A8	0432A8
CSECT ENTRY				COS	04328C	04328C
CSECT ENTRY				BOAFEXIT	0433A0	0433A0
CSECT ENTRY				EXIT	0433A6	0433A6
CSECT ENTRY				BOAUOPT	0433C0	0433C0
CSECT ENTRY				USEROPT	0433C0	0433C0
CSECT ENTRY				BOAFIOCS	0433C8	0433C8
CSECT ENTRY				RCBORG#	043AC0	043AC0
CSECT ENTRY				RUFORG#	043ABC	043ABC
* ENTRY				FIOCS#	0433C8	0433C8
* ENTRY				FIOCD#	043402	043402
* ENTRY				FIXB	043830	043830
* ENTRY				VARA	04383E	04383E
* ENTRY				VDIOCS#	043AC4	043AC4
CSECT					043858	043858

78/019	PHASE	TRANSFER ADDR.	LOCORE	HICORE	BLOCK NO.	ESD TYPE	LABEL	UNITAB #	LOADED	REL-FACTOR
						ENTRY			043B58	
						CSECT	BOASLOG		043BE0	043BE0
						ENTRY	ALOG		043C02	
						* ENTRY	ALOG10		043BE8	
						CSECT	BOASEXP		043CFO	043CFO
						ENTRY	EXP		043CF4	

1 PHASE USED 57 BLOCKS

LINKAGE EDITOR HIGHEST SEVERITY WAS 0

// EXEC FLOW

00:01:14

DISSERTATION - IRRIGATION RETURN FLOW WATER QUALITY
 FINITE ELEMENT PROGRAM TO CALCULATE STREAMLINES IN VERTICAL SECTION
 CASE#1 - FIRST 1/20 IS HIGH K ZONE) : ETA = 10

NUMBER OF ROWS = 21
 NUMBER OF COLUMNS = 41
 AQUIFER WIDTH = 1000.
 AQUIFER DEPTH = 100.
 DISTANCE BETWEEN COLUMN NODES = 25.0
 DISTANCE BETWEEN ROW NODES = 1.25
 AFTER 5 ROWS, DISTANCE BETWEEN ROW NODES = 5.9375

INPUT DATA

NO. OF QUADRILATERAL ELEMENTS (NQELS) = 800
 NO. OF TRIANGULAR ELEMENTS (NTELS) = 0
 NO. OF NODES (NNDS) = 861
 NO. OF MATERIAL PROPERTY ZONES (NZNS) = 2
 NO. OF SPECIFIED BOUNDARY HEADS (NHDS) = 120
 NO. OF SOURCES OR SINKS (NWELS) = 0
 NO. OF BOUNDARY ELEMENT SIDES (NORND) FOR WHICH DISCHARGE IS SPECIFIED = 0

ZONE	XPERM	YPERM	ALPHA
1	0.1000E 02	0.1000E 02	0.0
2	0.1000E 01	0.1000E 01	0.0

STREAMLINE	X-COORD	Y-COORD
0.450	450.00000	0.0
0.450	425.00000	0.63067
0.450	400.00000	1.20869
0.450	398.12939	1.25000
0.450	375.00000	1.74033
0.450	350.00000	2.23081
0.450	335.39575	2.50000
0.450	325.00000	2.68447
0.450	300.00000	3.10480
0.450	275.00000	3.49440
0.450	257.45142	3.75000
0.450	250.00000	3.85476
0.450	225.00000	4.18581
0.450	200.00000	4.48493
0.450	175.00000	4.74548
0.450	150.00000	4.95393
0.450	141.56207	5.00000
0.450	125.00000	5.86116
0.450	100.00000	6.00837
0.450	85.33429	5.00000
0.450	75.00000	4.93706
0.450	50.00000	4.47176
0.450	25.00000	4.06024
0.450	23.13147	3.75000
0.450	17.73195	2.50000
0.450	14.33233	1.25000
0.450	0.0	0.0

THE AREA UNDER PSI = 0.4500 IS 0.0154729 (ETA = 10.0)

Linear Reservoir Model Program: Arkansas River Valley

LUMBER PARAMETER MODEL FOR ARKANSAS RIVER VALLEY, LA JUNTA, COLORADO

00100
00200
00300
00400
00500
00600
00700
00800
00900
01000
01100
01200
01300
01400
01500
01600
01700
01800
01900
02000
02100
02200
02300
02400
02500
02600
02700
02800
02900
03000
03100
03200
03300
03400
03500
03600
03700
03800
03900
04000
04100
04200
04300
04400
04500
04600
04700
04800
04900
05000
05100
05200
05300
05400
05500
05600
05700
05800
05900
06000

1 DIMENSION (12), QL(12), QB(12), QP(12), QM(12), QPET(12), FACTOR(12),
 2 CLIN(12), QCL(12), QMIN(12), QMAX(12), CLIN(12), CLIN(12), CMIN(12),
 3 RIVMAT(12), CCL(12), GMPUP(12), SURMAT(12), ZMON(12), RIVERA(12),
 COMMON /AI/ ERAT(60), QK(60), FTOT(60), QTOT(60), COUT(60),
 1 ALMASS(60), CAQ(60), H(60), HRAR(60), QOUT(60), TIME(62), TITLE(50)
 COMMON /A2/ LB1(40), LB2(40), LB3(40), LB4(40), LB5(40), LB6(40), LB7(40)
 1 LHR(40), LB9(40), LMO(40), ANM1(40), ANM2(40), ANM3(40)
 COMMON /AVERS/ AVECAU, AVEH, AVCOU, AVETON, AVEQR, AVERIV, HMEAN, ESUM,
 1 S AREA
 DATA DAYS/31, 30, 731, 730, 7231, 730, 7231, 730, 7231, 728,
 DATA QM/12*1.20/
 DATA CLIN/5*0.2, 40, 2, 6*0.2/
 DATA EL/0.297, 33, 10, 28, 40, 3, 38, 21, 70, 29, 50, 26, 40, 0, 00314, 7, 17,
 0.019, 0.012, 0.0012, 2,
 DATA QL/9, 20, 13, 03, 13, 00, 13, 36, 14, 74, 15, 70, 16, 32, 13, 33, 10, 84, 9, 89,
 9, 84, 9, 92/
 DATA QS/6, 651, 8, 637, 9, 749, 19, 505, 21, 907, 13, 434,
 2, 159, 3, 773, 11, 001, 3, 040, 0, 00, 0, 00/
 DATA QM/ 3, 621, 5, 346, 5, 303, 9, 145, 7, 400, 7, 403,
 6, 029, 5, 560, 3, 320, 2, 432, 2, 528, 3, 081/
 DATA QPET/ 2, 501, 4, 104, 7, 696, 10, 254, 11, 872, 10, 564,
 5, 765, 3, 856, 1, 291, 0, 112, 0, 007, 0, 490/
 DATA QP/ 2, 378, 40, 550, 32, 701, 12, 697, 29, 728, 40, 233,
 39, 306, 1, 982, 1, 229, 0, 396, 0, 092, 0, 089/
 DATA QIN/42, 0, 95, 0, 139, 0, 464, 0, 413, 0, 114, 0, 109, 0, 106, 0, 36, 0, 30, 0, 0,
 2, 0, 21, 0/
 DATA QM/0, 9, 2*1, 0, 1, 2, 2*1, 4, 1, 6, 5*1, 0/
 DATA CIN/1398, 1509, 1509, 1231, 748, 1296, 1361, 1352, 1296, .,
 1732, 1872, 1835, /
 DATA CLIN/5*6450, 7, 1037, 7, 6450, /
 DATA CMIN/12*5244, /
 DATA CO/1147, 1240, 1379, 1175, 636, 1147, 1203, 1259, 896, 1194, .,
 1509, 1287, /
 DATA ZMON/5H MAR, 5H APR, 5H MAY, 5H JUN, 5H JUL, 5H AUG, 5H SEP, 5H OCT, 5H NOV, 5H DEC, 5H JAN, 5H FEB, /
 DATA IR, 1M/23, 3/
 DATA NCASES, NYR, IYR/5, 3, 1971/
 OPEN UNIT=23, DEVICE=86K, ACCESS=SEQUENCE, FILE=US66, DATA
 CALL INITIAL(1)
 N=12*NYR
 CONST=1600, *24, /43560.
 AREA=0, 7720
 SEU=20
 HR=4011, 16
 AREA=11950, 0
 HRAH(1)=4200, / (5+AREA)
 THES=VAR
 CKIP=EXP(-1.0/TH)
 CK2P=EXP(-0.5/TH)
 CAG(1)=2024, 3
 H(1)=4011, 80
 BASEH(1)=HBAR(1)
 CK=CONST/735, 0
 NYR=IYR
 DO 61 I=1, N
 AISI

```

06100 61 TIME(I)=(AI+1.5)/12.0 + AYK
06200 READ(IR,69)(LP1(I),I=1,78)
06300 READ(IR,69)(LB2(I),I=1,8)
06400 READ(IR,69)(LB3(I),I=1,8)
06500 READ(IR,69)(LB4(I),I=1,8)
06600 READ(IR,69)(LB5(I),I=1,78)
06700 READ(IR,69)(LB6(I),I=1,8)
06800 READ(IR,69)(LB7(I),I=1,8)
06900 READ(IR,69)(LB8(I),I=1,8)
07000 READ(IR,69)(DB9(I),I=1,8)
07100 READ(IR,69)(LH0(I),I=1,8)
07200 READ(IR,69)(ANM1(I),I=1,8)
07300 READ(IR,69)(ANM2(I),I=1,8)
07400 READ(IR,69)(ANM3(I),I=1,78)
07500 FORMAT(10A5)
07600 DO 77 I=1,12
07700 GWPUMP(I)=QP(I)
07800 RIVLEAT(I)=QL(I)
07900 RIVWAT(I)=QIN(I)
08000 SURWAT(I)=QS(I)
08100 DO 98 JR=1,NCASES
08200 JGR=JR
08300 DO 76 I=1,12
08400 E(I)=E1(I)
08500 OP(I)=GWPUMP(I)
08600 QL(I)=RIVLEAT(I)
08700 QIN(I)=RIVWAT(I)
08800 QS(I)=SURWAT(I)
08900 READ(IR,69)(TITLE(I),I=1,10)
09000 IYR=IG-YR
09100 GO TO (80,81,85,86,83),JR
09200 I1 DO 82 I1=1,12
09300 QS(I1)=QS(I1)+QP(II)
09400 QP(II)=0
09500 GO TO 80
09600 DO 88 I1=1,12
09700 QL(I1)=0
09800 GO TO 80
09900 DO 84 I1=1,12
10000 QP(II)=QP(I1)+QS(II)
10100 QS(II)=0
10200 GO TO 80
10300 DO 87 I1=1,12
10400 E(II)=E(I1)+0.393*E(II)
10500 QS(II)=QS(I1)+0.2*QP(II)
10600 QP(II)=QP(I1)+0.2*QP(II)
10700 DO 1 AYR=IYR-1
10800 DO 1 J=1,NYR
10900 JYR=J
11000 AVEGAG=0
11100 AVEH=0
11200 AVCLUUT=0
11300 AVETON=0
11400 AVEQR=0
11500 AVERIV=0
11600 HMEAN=0

```



```

11700 ESUM=0.0
11800 AWR=AYR+1+0
11900 WRITE(IM,250)(TITLE(I),I#1,10)
12000 FORMAT(1H1,10A10)
12100 250 WRITE(1W,200) IYRBEG
12200 200 FORMAT(1X,5X, ARKANSAS RIVER VALLEY STUDY DATA SUMMARY FOR
12300 <14//, 2X, ENET, 2X, FTOT, 2X, AU, CHEM, 3X, AG, W-L,
12400 <3X, RIVER CHEM, 2X, FTOT, 2X, SALI, 2X, RIVER TRANS, 2X,
12500 <3X, RIVER CHEM, 2X, FTOT, 2X, SALI, 2X, RIVER TRANS, 2X,
12600 <3X, RIVER CHEM, 2X, FTOT, 2X, SALI, 2X, RIVER TRANS, 2X,
12700 IF(AYR EQ 1971.) GO TO 2
12800 DAYS(12)=28.0
12900 GO TO 3
13000 3 DAYS(12)=29.0
13100 DO 112 I=1,12
13200 KI=12*(J-1)+1
13300 FACTOR(I)=CONST*DAYS(I)
13400 ENET(KI)=E(I)+QL(I)+QB(I)+QP(I)+QM(I)+QPET(I)
13500 ENET(KI)=ENET(KI)*FACTOR(I)/AREA
13600 DO 111 I=1,12
13700 KI=12*(J-1)+1
13800 IF(KI LE 1) GO TO 7
13900 CK3P=H(KI-1)-HR-ENET(KI-1)/AR
14000 CK4=HR+ENET(KI)/AR
14100 CK5=(ENET(KI-1)-ENET(KI))/AR
14200 H(KI)=CK3P+CK4P+CK5+CK2P
14300 HBAR(KI)=H(KI)+BASE
14400 QR(KI)=AREA*AR*(H(KI)-HR)/FACTOR(I)
14500 AVEQR=AVEQR+QR(KI)
14600 HMEAN=HMEAN+H(KI)
14700 H(QR(KI) GT 0) GO TO 8
14800 IF(QR(KI) EQ 0) GO TO 8
14900 ETOT(KI)=F(I)+QL(I)+QB(I)+QPET(I)+QR(KI)
15000 QTOT(KI)=QS(I)+QB(I)+QR(KI)
15100 CP(KI)=(QS(I)+QL(I)+QB(I)+QP(I)+QR(KI)*CIN(I))/QTOT(KI)
15200 GO TO 9
15300 8 ETOT(KI)=E(I)+QL(I)+QB(I)+QPET(I)
15400 QTOT(KI)=QS(I)+QB(I)
15500 CP(KI)=C(I)
15600 ETOT(KI)=ETOT(KI)*FACTOR(I)/AREA
15700 QTOT(KI)=QTOT(KI)*FACTOR(I)/AREA
15800 ESUM=ESUM+ETOT(KI)
15900 IF(J EQ 1 AND I EQ 1) GO TO 14
16000 CM1=S*(H(KI)+HBAR(KI))+ENET(KI-1)
16100 CM2=ETOT(KI)+ETOT(KI-1)
16200 CM3=QTOT(KI)+QTOT(KI-1)
16300 CAU(KI)=(CAU(KI-1)+CM1+CP(KI)*CM3)/(CM1+CM2)
16400 QOUT(KI)=QIN(I)+QLIN(I)+QBIN(I)+QMIN(I)+QM(I)+QA(KI)
16500 AVEQIV=AVEQIV+QOUT(I)
16600 IF(QR(KI) EQ 0) GO TO 12
16700 CDUT(KI)=(QIN(I)+QLIN(I)+QBIN(I)+QMIN(I)+QM(I)+QR(KI)*CIN
16800 <(I))/QOUT(KI)
16900 GO TO 13
17000 12 COUT(KI)=(QIN(I)+QBIN(I)+QBIN(I)+QBIN(I)+QMIN(I)+QM(I)+QR(KI)*CIN
17100 <(KI))/QOUT(KI)
17200 13 CDAYS=CK*DAYS(I)
17300 ALMASS(KI)=CDAYS*COUNT(KI)*QOUT(KI)
17400 AVEGAG=AVEGAG+CAU(KI)
17500 AVEQIV=AVEQIV+ALMASS(KI)
17600 AVECOUT=AVECOUT+COUT(KI)

```

```

17700 WRITE(IM,201)ZMON(I),ENET(KI),ETOT(KI),QTOF(KI),CAQCK(I),H(KI),
17800 <COUT(KI),QBTK(KI),ALMSS(KI),QR(KI),HBR(KI)
17900 201 FORMAT(1X,A5,2(F6.3,1X),F5.3,2X,F9.2,3X,F7.2,1X,F11.2,F10.2,
18000 3X,F10.2,2X,F7.2,1X,F7.2)
18100 111 CONTINUE
18200 AVEGAG=AVEGAG/12.0
18300 AVEH=AVEH/12.0
18400 AVCOUT=AVCOUT/12.0
18500 AVETON=AVETON/12.0
18600 AVEGR=AVEGR/12.0
18700 HMEAN=HMEAN/12.0
18800 AVERIV=AVERIV/12.0
18900 TC=S*HMEAN/ESUM
19000 WRITE(IM,301)AVEGAG,AVEH,AVCOUT,AVERIV,AVETON,AVEGR,HMEAN
19100 301 FORMAT(7,5X,2YEARLY AVERAGES: ,6X,F9.2,3X,F7.2,1X,F11.2,F10.2,3X,
19200 < F10.2,2X,F7.2,1X,F7.2)
19300 WRITE(IM,202)TH,TC
19400 202 FORMAT(7,5X,HYDRAULIC RESPONSE TIME = ,F4.1,1X,MONTHS ,73X,
19500 < SOLUTE RESPONSE TIME = ,F6.1,1X,YEARS.)
19600 CALL DELTA(LW,JGR,JYR,IYRREG,ZMON,N,QIN,CIN,NYR)
19700 IYRREG=IYRREG+1
19800 CALL GRAFF(TWJGR,NYR,CIN,QIN,CO)
19900 CLOSE(UNIT=23,DEVICE='DSK',ACCESS='SEGIN',FILE='USGS.DAT')
20000 CALL RSTR(0)
20100 CALL RSTR(2)
20200 STOP
20300 END
20400 C
20500 C
20600 C
20700 DIMENSION HKOM(12),DELY2(5),DELY3(5),YSC3B(5),DELY3B(5),
20800 1 CAQKW(12),Y1(62),Y2(62),YSC1(5),YSC2(5),YSC3(5),DELY1(5),
20900 2 QBKS(12),CKOM(12),CKWM(12),LBY1(40),LBY2(40),
21000 3 HGR(12),GRR(12),K(12),GIN(12),QIN(12),GO(12),
21100 4 UDI(14),OD2(14),OD3(14),NUM1(1),NUM2(1),XI(14),
21200 COMMON /A1/ENET(60),QR(60),ETUI(60),QUT(60),COUT(60),
21300 1 ALMSS(60),CAQ(60),H(60),HBR(60),QUT(60),TIME(62),TITLE(50)
21400 COMMON /A2/HB1(40),LB1(40),HB3(40),LB3(40),LB4(40),LB5(40),LB6(40),LB7(40)
21500 1 LB8(40),LB9(40),LH0(40),ANM1(40),ANM2(40),ANM3(40)
21600 COMMON /A3/LUT/H2(60),CAQ2(60),QOUT2(60),COUT2(60)
21700 1 DATA QBKS/55,99,146,473,419,163,122,677,1,47,35,29,,28./
21800 DATA QBKS/174,77,168,107,170,43,127,12,677,1,153,13,
21900 1 1528,0,1518,7,1667,3,1862,4,2038,9,2141,1/
22000 1 DATA QKW/47,0,98,1,138,5,456,4,406,4,152,9,
22100 1 108,9,108,1,43,6,38,2,34,5,28,9/
22200 DATA CKOM/143,0,161,6,157,0,1,247,7,772,7,1307,9,
22300 1 1459,1,144,8,1617,0,2006,6,2138,1,2168,5/
22400 1 DATA HUB/4011,80,4011,24,4011,13,4010,57,4010,82,
22500 1 4010,50,4010,57,4010,64,4011,21,4011,61,4011,70,4011,65/
22600 1 DATA CAQOM/204,37,205,1,208,4,87,214,3,12216,67,2270,4,
22700 1 2300,2,2278,7,2214,0,2212,6,2209,5,2180,9/
22800 1 DATA HKOM/4013,74,4013,60,4013,64,4013,74,4013,76,
22900 1 4013,45,4013,35,4013,16,4013,64,4013,66,4013,68,4013,68/
23000 1 DATA CAQKW/1880,8,1890,6,1920,0,1948,3,1942,6,
23100 < 1984,7,1998,8,2001,5,1961,9,1958,5,1944,9,1926,7/
23200 DATA NUM1/4HI971/

```



```

29300
29400
29500
29600
29700
29800
29900
30000
30100
30200
30300
30400
30500
30600
30700
30800
30900
31000
31100
31200
31300
31400
31500
31600
31700
31800
31900
32000
32100
32200
32300
32400
32500
32600
32700
32800
32900
33000
33100
33200
33300
33400
33500
33600
33700
33800
33900
34000
34100
34200
34300
34400
34500
34600
34700
34800
34900

451 CALL PLOT(0,0,0,0,0,3)
CALL LINE(X,Y2,N,J2,1)
CALL PLOT(0,0,6,10,3)
CALL SYMBOL(0,6,10,0,14,TITLE,0,0,7,50)
CALL PLOT(0,6,10,0,14,TITLE,0,0,7,50)
KOUNT=KOUNT+1
IF(KOUNT=11,3)GO TO 5
IF(JR,GT,1)RETURN
ICOUNT=1
55 CONTINUE
CALL RSTR(1)
CALL RSTR(1)
GO TO (11,22,65,33,44),ICOUNT
11 DO 112 I=1,12
AI=I
X1(I)=AYR+(AI+1.5)/12.0
CONST=QIN(I)
OD1(I)=OUBS(I)-CONST
OD2(I)=KOW(I)-CONST
112 DO 113 I=1,8
OD3(I)=GOUT(I)-CONST
113 LBY1(I)=LB4(I)
OIMIN=10.0
DOD1=10.0
GO TO 44
GO TO 44
22 DO 220 I=1,12
AI=I
X1(I)=AYR+(AI+1.5)/12.0
CONST=CO(I)
OD1(I)=CORS(I)-CONST
OD2(I)=KOW(I)-CONST
220 OD3(I)=COUT(I)-CONST
DO 221 I=1,8
LBY1(I)=LB5(I)
OIMIN=0.0
DOD1=150.0
GO TO 44
GO TO 44
65 DO 650 I=1,12
AI=I
X1(I)=AYR+(AI+1.5)/12.0
CONST=CIN(I)
OD1(I)=CORS(I)-CONST
OD2(I)=KOW(I)-CONST
650 OD3(I)=COUT(I)-CONST
DO 651 I=1,8
LBY1(I)=LB4(I)
OIMIN=100.0
DOD1=100.0
GO TO 44
GO TO 44
33 DO 330 I=1,12
AI=I
X1(I)=AYR+(AI+1.5)/12.0
OD1(I)=HCH(I)
OD2(I)=HROW(I)
330 OD3(I)=H(I)
DO 331 I=1,8
LBY1(I)=LB9(I)

```

```

35000 OD1MIN=4009.0
35100 DDD1=1.0
35200 GO TO 444
35300 44 DO 440 I=1,12
35400 AI=1
35500 XI(I)=AYR*(AI+1.57)/13.0
35600 OD1(I)=CAJOB(I)
35700 OD2(I)=CAKOW(I)
35800 OD3(I)=CAO(I)
35900 DO 441 I=1,12
36000 441 LRYI(I)=D90(I)
36100 OD1MIN=1500.0
36200 DOD1=200.0
36300 444 OD2MIN=OD1MIN
36400 DOD2=DOD1
36500 OD3MIN=OD1MIN
36600 DOD3=DOD1
36700 XMIN=AYR
36800 DX=1.0/4.0
36900 DO 445 I=1,12
37000 XI(I)=(XI(I)-XMIN)/DX
37100 OD1(I)=(OD1(I)-OD1MIN)/DOD1
37200 OD2(I)=(OD2(I)-OD2MIN)/DOD2
37300 445 OD3(I)=(OD3(I)-OD3MIN)/DOD3
37400 CALL PLOT(1.56,2.5,-3)
37500 CALL PLOT(0.0,0.0,1.3)
37600 CALL PLOT(0.0,0.0,2)
37700 CALL PLOT(6.0,0.0,2)
37800 CALL PLOT(0.0,0.0,3)
37900 DELX=0.0
38000 DO 446 IX=1,18
38100 DELX=DELX+1.0/3.0
38200 CALL PLOT(DELX,0.0,3)
38300 CALL MARKER(J)
38400 CALL PLOT(4.0,0.0,1.3)
38500 CALL PLOT(4.0,0.0,1.2)
38600 CALL SYMBOL(2.0,-0.5,0.14,LB1,0.0,40)
38700 CALL SYMBOL(-0.25,-0.37,0.14,NUM1,0.0,4)
38800 CALL SYMBOL(3.75,-0.37,0.14,NUM2,0.0,4)
38900 CALL PLOT(0.0,6.0,3)
39000 CALL PLOT(6.0,6.0,2)
39100 CALL PLOT(6.0,0.0,2)
39200 CALL PLOT(0.0,0.0,3)
39300 CALL AXIS(0.0,0.0,LRY1,40,6.0,90.0,OD1MIN,DOD1,-1)
39400 CALL PLOT(0.0,0.0,3)
39500 CALL LINE(X1,OD1,I2,J2,I1)
39600 CALL PLOT(0.0,0.0,3)
39700 CALL LINE(X1,OD2,I2,J2,I1)
39800 CALL PLOT(0.0,0.0,3)
39900 CALL LINE(X1,OD3,I2,J2,I1)
40000 CALL SYMBOL(0.6,46,0.14,ANM1,0.0,40)
40100 CALL SYMBOL(0.6,26,0.14,ANM2,0.0,40)
40200 CALL SYMBOL(0.6,06,0.14,ANM3,0.0,40)
40300 CALL PLOT(1.625,6.166,3)
40400 CALL MARKER(J3)
40500 CALL PLOT(0.0,-2.5,-3)
40600 IFCOUNT=ICOUNT+1
40700 IF(ICOUNT)BT,60 TU,55
40800 RETURN
40900 END

```

41000 C SUBROUTINE DELTA(IW,JK,JYR,IYR,ZMON,N,OIN,CIN,NYR)
 41100 C
 41200 DIMENSION CAQI(60),HI(60),QOUTI(60),ALMASI(60),ORI(60),HBI(60),
 41300 ZMON(12),OIN(12),CIN(12),ACAGI(5),AH(5),ACOUT(5),ATON(5),AJR(5),
 41400 ARIV(5),HBAR(5),GINI(12),COUTI(60),DELTA(60)
 41500 COMMON /AL/ENET(60),QH(60),EHT(60),GTOF(60),COUT(60),ALMASS(60),
 41600 CAQI(60),H(60),HBAR(60),QOUT(60),X(62),TITLE(50)
 41700 1 COMMON /AVERS/AVECAQ,AVEH,AVCOUT,AVEION,AVEQR,AVERIV,HMEAN,ESUM,
 41800 1 SA,AREA
 41900 COMMON /APLOT/H2(60),CAG2(60),QOUT2(60),COUT2(60)
 42000 IF(JR.GT.1)GO TO 1
 42100 FACI=SA*AREA
 42200 ACAQ(JYR)=AVECAQ
 42300 AH(JYR)=AVEH
 42400 ACOUT(JYR)=AVCOUT
 42500 ATON(JYR)=AVEION
 42600 AJR(JYR)=AVEQR
 42700 ARIV(JYR)=AVERIV
 42800 HBAR(JYR)=HMEAN
 42900 DO 2 KI=1,12
 43000 I=12*(JYR-1)+KI
 43100 DELTA(I)=CAG(I)
 43200 CAQI(I)=CAQ(I)*HBAR(I)*FACI/735.
 43300 HI(I)=H(I)*FACI
 43400 QOUTI(I)=QOUT(I)*QIN(KI)
 43500 ALMASI(I)=ALMASS(I)
 43600 COUTI(I)=COUT(I)*CIN(KI)
 43700 HBI(I)=HBI(I)
 43800 GINI(KI)=GINI(KI)
 43900 IF(I)=GH(I)
 44000 2 IF(JYR-NYR)6,5,5
 44100 DO 7 I=1,N
 44200 H2(I)=HI(I)*FACI
 44300 QOUT2(I)=QOUTI(I)
 44400 COUT2(I)=COUTI(I)
 44500 CAQ2(I)=CAQI(I)
 44600 GO TO 6
 44700 7
 44800 1 AVEC=AVECAQ-ACAG(JYH)
 44900 AVECAQ=(AVECAQ+HMEAN-ACAG(JYR)*HBAR(JYR))*FACI/735.0
 45000 AVEH=(AVEH-AH(JYR))*FACI
 45100 AVCOUT=AVCOUT-ACOUT(JYR)
 45200 AVEION=AVEION-ATON(JYH)
 45300 AVEQR=AVEQR-AQR(JYR)
 45400 AVERIV=AVERIV-AHAR(JYR)
 45500 HMEAN=HMEAN-AHAR(JYR)
 45600 WRITE(IW,200)IYR
 45700 200 FORMATT(//,2X),ARKANSAS RIVER VALLEY STUDY: CHANGES FROM BASE PERIO
 45800 D FOR I=1,14,16X, AQ. CHEM., 2X, AQ. VOL., 3X,
 45900 RIVER CHEM., 2X, RIVER FLOW, 2X, SALT, 2X, RIVER TRANS, 2X,
 46000 HBAR(IYR), (MG/H), 5X, (TONS), 7X, (CFS), 2X, (FT), //,
 46100 DO 3 KI=1,12
 46200 I=12*(JYR-1) + KI
 46300 EG=CAQ(I)-DELTA(I)
 46400 CAQ2(I)=CAQ(I)*HBAR(I)*FACI/735.0-CAQI(I)
 46500 H2(I)=H(I)*FACI - H3(I)
 46600

```

46700 QOUT2(I)=QOUT(I)-QOUT1(I)-QIM1(KI)
46800 ALMAS2=ALMAS(I)-ALMAS1(I)
46900 COUT2(I)=COUT(I)-COUT1(I)-CIN(KI)
47000 QR2=QR(I)-QR1(I)
47100 HBAR2=HBAR(I)-HBAR1(I)
47200 3 WRITE(IW,201)ZMUR(KI)DC,CAGR(I),H2(I),COUT2(I),QOUT2(I),ALMAS2
47300 < QR2,HBAR2
201 FORNAT(4X,A5,7X,F9.2,2X,F9.2,F10.2,F12.2,F10.2,F13.2,F9.2,F8.2)
47400 WRITE(IM,301)AVEC,AVECAQ,AVEN,AVCOUT,AVERTUN,AVEQUR,UMEAN
47500 301 FORNAT(75X,AVE,CHANGE,71X,F9.2,71X,F9.2,71X,F10.2,71X,F12.2,71X,F10.2,71X,F13.2,
47600 1 F9.2,F8.2)
47700 6 RETURN
47800 6 END
47900

```

Data Summary for Arkansas River Valley Study

Input Data Summary

Output Model Simulation Results

Definition of Terms

The following definitions are used in the Arkansas River valley linear reservoir model program (see Appendix C and Figure 3.1). All definitions apply to the study reach area (information source: L.F. Konikow, personal communication, 1977).

<u>Program Term</u>	<u>Chap. 3 Term</u>	<u>Definition</u>
QB	q_b	Net aquifer boundary inflow
QIN		Arkansas R. streamflow above La Junta
QL	q_L	Ft. Lyon and feeder canal leakage to aquifer
QLIN		Tributary inflows into Arkansas River
QM	q_m	La Junta municipal pumpage from groundwater
QMIN		La Junta municipal sewage effluent to river
QP	q_p	Study reach irrigation pumpage from aquifer
QPET		Phreatophyte evapotranspiration
QS	q_s	Surface water irrigation applications
E1	ϵ	Total aquifer recharge
HBAR	\bar{h}	Average aquifer saturated thickness(17.57 ft.)
CIN		Arkansas R. water quality above La Junta
CLIN		Tributary inflow water quality
CMIN		Municipal sewage water quality
CO	c_o	Ft. Lyon canal water quality above La Junta
QOBS, COBS		Arkansas R. streamflow and water quality at downstream gage
QKOW,CKOW		Konikow-Bredehoeft predicted river water flow and quality at downstream gage
HOB, CAQOB		Average aquifer water levels and water quality from Thiessen polygon weighting technique
HKOW,CAQKOW		Konikow-Bredehoeft predicted average aquifer water levels and water quality
CROP ET	$q_{crop\ et}$	Blaney-Criddle crop consumptive use
QDIV		Ft. Lyon canal diversions
A		Study reach area (11,950 acres)

Arkansas River Data Summary

(Information Source: L.F. Konikow, personal communication, 1977)

Note: All water quality values are for TDS in mg/l.

<u>Month</u>	<u>CIN</u> (mg/l)	<u>CO</u> (mg/l)	<u>CLIN</u> (mg/l)	<u>CMIN</u> (mg/l)	<u>QOBS</u> (cfs)	<u>COBS</u> (mg/l)
Mar 71	1397.9	1147.1	6450	5244	55.0	1741.7
Apr	1509.4	1240.0	6450	5244	99.0	1881.0
May	1509.4	1379.4	6450	5244	146.0	1704.5
Jun	1230.7	1175.0	6450	5244	473.0	1277.2
Jul	747.6	636.2	6450	5244	419.0	877.7
Aug	1295.7	1147.1	1037	5244	163.0	1537.3
Sep	1360.8	1202.8	6450	5244	122.0	1528.0
Oct	1351.5	1258.6	6450	5244	105.0	1518.7
Nov	1295.7	896.3	6450	5244	47.0	1667.3
Dec	1732.4	1193.6	6450	5244	35.0	1862.4
Jan 72	1871.7	1509.4	6450	5244	29.0	2038.9
Feb	1834.6	1286.5	6450	5244	28.0	2142.2

<u>Month</u>	<u>QKOW</u> (cfs)	<u>CKOW</u> (mg/l)	<u>HOB</u> (ft)	<u>CAQOB</u> (mg/l)	<u>HKOW</u> (ft)	<u>CAQKOW</u> (mg/l)
Mar 71	47.0	1643.0	4011.80	2024.3	4013.74	1880.8
Apr	98.1	1613.6	4011.24	2051.1	4013.60	1899.8
May	138.5	1570.7	4011.13	2084.8	4013.64	1920.0
Jun	456.4	1247.7	4010.57	2143.1	4013.74	1948.7
Jul	406.4	772.7	4010.82	2216.6	4013.76	1942.9
Aug	152.9	1387.9	4010.50	2270.4	4013.45	1984.7
Sep	108.9	1459.1	4010.57	2300.2	4013.35	1998.8
Oct	108.1	1444.8	4010.64	2278.7	4013.46	2001.5
Nov	43.6	1617.0	4011.21	2214.0	4013.64	1961.9
Dec	38.2	2006.6	4011.61	2212.6	4013.66	1958.5
Jan 72	34.5	2138.1	4011.70	2209.5	4013.68	1944.9
Feb	28.9	2168.5	4011.65	2180.9	4013.68	1926.7

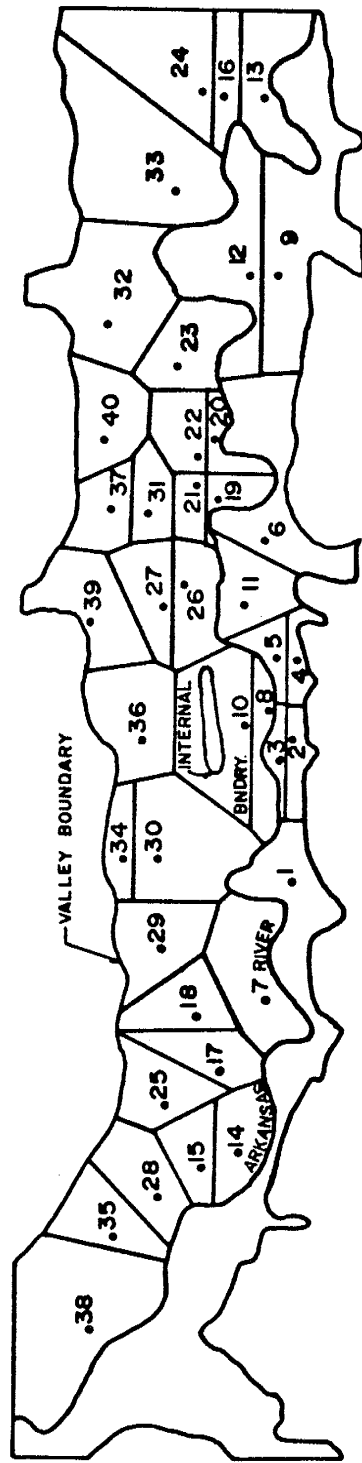


Figure D.1

Arkansas River Valley Study Area Showing Well Location and Thiessen Weighing Polygons

ARIZONA ARKANSAS RIVER VALLEY DATA, L.F. 1, 1979
 REFERENCE: KEN IKOH, L.F. 1, 1979

**** AVERAGE VALUES USING THE ISSCN POLYGON WEIGHTING FACTORS ****

MONTH	AVE H-L (FT)	AVE CHEM (PPM)	AVE CHEM (SP. COND.)
0	4012.41	2047.3	2556.2
1	4011.80	2024.3	2531.5
2	4011.24	2051.1	2560.3
3	4011.13	2084.8	2596.6
4	4010.57	2143.1	2654.2
5	4010.82	2216.6	2738.3
6	4010.50	2270.4	2796.2
7	4010.57	2300.2	2828.2
8	4010.64	2278.7	2805.1
9	4011.21	2214.0	2735.5
10	4011.61	2212.6	2734.0
11	4011.70	2209.5	2730.7
12	4011.65	2180.9	2695.9

**** ARITHMETIC MEAN OF 40 OBSERVATION WELLS ****

MONTH	AVE H-L (FT)	AVE CHEM (PPM)	AVE CHEM (SP. COND.)
0	4010.74	2236.3	2759.5
1	4010.50	2240.8	2764.3
2	4010.27	2276.0	2802.2
3	4010.30	2302.3	2830.4
4	4010.17	2374.3	2907.0
5	4010.29	2417.9	2954.7
6	4009.54	2471.8	3012.7
7	4009.91	2493.0	3035.5
8	4010.09	2464.3	3004.6
9	4010.43	2408.7	2944.8
10	4010.51	2395.8	2931.0
11	4010.52	2395.8	2931.0
12	4010.50	2361.9	2894.5

ARKANSAS RIVER VALLEY DATA SOUTHEASTERN COLORADO
 REFERENCE: KCN IKOM, L.F., 1977

WEIGHT FACTOR = 0.04035

SP. COND.

GW CHEM

W-L ELEV

MONTH

OBS. WELL #

OBS. WELL #	MONTH	W-L ELEV	GW CHEM	SP. COND.
1	0	4022.5	PPM	AT 25 C
	1	4021.9	1713.8	2197.6
	2	4021.4	1730.5	2215.6
	3	4021.5	1859.4	2354.2
	4	4022.1	1919.7	2419.0
	5	4022.1	2021.9	2528.9
	6	4021.6	2144.1	2660.3
	7	4021.4	2319.4	2848.8
	8	4021.9	2497.9	3040.8
	9	4021.9	2412.4	2948.8
	10	4021.9	2295.7	2823.3
	11	4021.9	1903.2	2401.3
12	4021.8	1904.2	2402.4	
		1905.3	2403.5	

WEIGHT FACTOR = 0.01272

SP. COND.

GW CHEM

W-L ELEV

MONTH

OBS. WELL #

OBS. WELL #	MONTH	W-L ELEV	GW CHEM	SP. COND.
2	0	4014.0	PPM	AT 25 C
	1	4013.8	3107.3	3696.0
	2	4013.7	3106.4	3695.1
	3	4013.8	3105.5	3694.1
	4	4014.3	3104.9	3693.4
	5	4014.3	3103.0	3691.4
	6	4014.1	3029.5	3612.4
	7	4014.0	3028.8	3611.6
	8	4013.9	3027.9	3610.6
	9	4013.7	2979.3	3558.4
	10	4013.7	2954.9	3532.2
	11	4013.7	2953.9	3531.1
12	4013.6	2952.3	3529.4	
		2880.1	3451.7	

WEIGHT FACTOR = 0.00603

SP. COND.

GW CHEM

W-L ELEV

MONTH

OBS. WELL #

OBS. WELL #	MONTH	W-L ELEV	GW CHEM	SP. COND.
3	0	4014.0	PPM	AT 25 C
	1	4013.6	3200.2	3795.9
	2	4013.6	3200.4	3756.1
	3	4013.8	3199.9	3795.6
	4	4014.5	3197.6	3793.1
	5	4013.9	3017.8	3599.8
	6	4013.9	3011.6	3593.1
	7	4013.9	3001.5	3582.3
	8	4013.6	3001.7	3546.5
	9	4013.6	2822.5	3389.8
	10	4013.6	2818.3	3383.1
	11	4013.6	2648.4	3202.6
12	4013.6	2645.7	3199.7	
		2484.6	3026.5	

ARKANSAS RIVER VALLEY DATA, SOUTHEASTERN COLORADO
 REFERENCE: KERRICK, L.F., 1977

WEIGHT FACTOR = 0.00577

SP. COND.

GW CHEM

M-L ELEV.

MONTH

OBS. WELL #

OBS. WELL #	MONTH	M-L ELEV.	GW CHEM	SP. COND.	AT 25 C	
					PPM	FT.
4	0	4009.5	3386.0	3995.7		
	1	4008.9	3392.6	3992.0		
	2	4008.8	3379.0	3908.2		
	3	4008.9	3375.4	3904.3		
	4	4009.3	3370.8	3919.4		
	5	4009.5	3372.4	3981.1		
	6	4009.1	3368.8	3977.2		
	7	4009.0	3365.3	3973.4		
	8	4008.9	3362.0	3969.9		
	9	4008.8	3357.4	3964.9		
	10	4008.7	3356.0	3963.4		
	11	4008.7	3351.6	3958.7		
	12	4008.7	3346.9	3953.7		

WEIGHT FACTOR = 0.00623

SP. COND.

GW CHEM

M-L ELEV.

MONTH

OBS. WELL #

OBS. WELL #	MONTH	M-L ELEV.	GW CHEM	SP. COND.	AT 25 C	
					PPM	FT.
5	0	4009.0	3571.8	4195.5		
	1	4008.7	3566.0	4189.2		
	2	4008.6	3558.8	4181.5		
	3	4008.8	3549.4	4171.4		
	4	4009.4	3461.7	4077.1		
	5	4009.5	3419.7	3924.4		
	6	4008.9	3253.5	3853.2		
	7	4008.2	3239.2	3837.8		
	8	4008.8	3246.7	3888.9		
	9	4008.7	2994.9	3575.2		
	10	4009.6	2935.6	3565.2		
	11	4008.6	2975.9	3554.7		
	12	4008.6	2883.0	3454.8		

332

WEIGHT FACTOR = 0.02121

SP. COND.

GW CHEM

M-L ELEV.

MONTH

OBS. WELL #

OBS. WELL #	MONTH	M-L ELEV.	GW CHEM	SP. COND.	AT 25 C	
					PPM	FT.
6	0	4003.5	2364.1	2896.9		
	1	4003.5	2360.4	2892.9		
	2	4003.4	2356.8	2889.0		
	3	4003.5	2353.3	2885.1		
	4	4003.8	2350.3	2882.0		
	5	4003.9	2346.9	2878.4		
	6	4003.7	2342.9	2874.1		
	7	4003.6	2168.5	2686.6		
	8	4003.5	2185.1	2704.4		
	9	4003.4	2183.8	2703.0		
	10	4003.4	2182.2	2701.3		
	11	4003.3	2180.5	2699.5		
	12	4003.3	2166.6	2684.5		

ARKANSAS RIVER VALLEY DATA [1977] SOUTHEASTERN COLORADO
 REFERENCE KEN KONTI, L.P.

WEIGHT FACTOR = 0.03938

OBS. WELL #	MONTH	W-L ELEV	GW CHEM	SP. COND.
7		FT.	PPM	AT 25 C
	0	4029.5	1806.7	2297.5
	1	4029.1	1840.4	2333.8
	2	4029.4	1965.7	2468.5
	3	4029.5	2072.9	2583.8
	4	4029.6	2245.8	2769.7
	5	4029.9	2212.1	2733.4
	6	4029.4	2434.8	2972.9
	7	4029.3	2627.8	3180.4
	8	4029.2	2645.8	3199.8
	9	4029.3	2600.5	3151.1
	10	4029.8	2462.8	3003.0
	11	4029.3	2458.1	2958.0
	12	4029.4	2360.6	2893.1

WEIGHT FACTOR = 0.00590

OBS. WELL #	MONTH	W-L ELEV	GW CHEM	SP. COND.
8		FT.	PPM	AT 25 C
	0	4012.0	3478.9	4095.6
	1	4011.9	3478.1	4094.7
	2	4012.0	3477.9	4094.5
	3	4012.2	3431.6	4044.7
	4	4012.7	3422.4	4034.8
	5	4012.8	3419.7	4031.9
	6	4012.2	3287.8	3890.1
	7	4012.1	3243.1	3842.0
	8	4012.1	3159.6	3752.3
	9	4012.0	3106.2	3694.8
	10	4011.9	3037.0	3620.4
	11	4011.9	3039.8	3623.4
	12	4011.9	3043.0	3626.9

333

WEIGHT FACTOR = 0.03503

OBS. WELL #	MONTH	W-L ELEV	GW CHEM	SP. COND.
9		FT.	PPM	AT 25 C
	0	3978.5	1806.7	2297.5
	1	3979.1	1807.2	2298.1
	2	3978.8	1879.4	2375.7
	3	3978.7	2000.5	2505.9
	4	3978.9	2077.3	2588.5
	5	3978.4	2162.4	2680.0
	6	3978.4	2271.2	2797.0
	7	3978.3	2453.1	2992.6
	8	3978.8	2520.8	3065.4
	9	3979.0	2516.8	3061.1
	10	3979.2	2538.2	3084.1
	11	3979.3	2613.3	3186.3
	12	3979.4	2624.8	3177.2

ARKANSAS RIVER VALLEY DATA, SOUTHEASTERN COLORADO
 REFERENCE: KENIKOH, L.F., 1977

WEIGHT FACTOR = 0.02757

OBS. WELL #	MONTH	W-L ELEV	GW CHEM	SP. COND.
10		FT.	PPM	AT 25 C
	0	4011.5	3571.8	4195.5
	1	4011.8	3643.9	4273.0
	2	4012.0	3731.8	4367.5
	3	4012.2	3746.0	4382.8
	4	4012.4	3993.9	4649.4
	5	4012.8	4091.2	4754.0
	6	4012.6	4322.8	5002.8
	7	4012.1	4291.8	4969.7
	8	4012.0	4148.0	4815.1
	9	4012.0	4268.5	4944.6
	10	4011.9	4271.2	4947.5
11	4011.9	4158.0	4825.8	
12	4011.9	4053.3	4713.2	

WEIGHT FACTOR = 0.01771

OBS. WELL #	MONTH	W-L ELEV	GW CHEM	SP. COND.
11		FT.	PPM	AT 25 C
	0	4005.5	1528.0	197.8
	1	4005.9	1545.4	2018.6
	2	4005.7	1612.9	2085.1
	3	4005.8	1678.2	2159.4
	4	4006.4	2191.6	2711.4
	5	4006.3	2288.6	2815.7
	6	4005.8	2498.0	3042.6
	7	4005.7	2588.0	3137.6
	8	4006.0	2732.1	3292.6
	9	4005.9	2722.0	3281.7
	10	4005.9	2767.5	3330.6
11	4005.9	2873.2	3444.3	
12	4005.9			

WEIGHT FACTOR = 0.03010

OBS. WELL #	MONTH	W-L ELEV	GW CHEM	SP. COND.
12		FT.	PPM	AT 25 C
	0	3979.0	1902.5	2597.3
	1	3979.5	1995.3	2800.3
	2	3979.5	2046.3	2855.2
	3	3979.5	2128.5	2643.5
	4	3979.6	2173.6	2692.0
	5	3979.7	2217.9	2739.7
	6	3979.4	2306.4	2899.4
	7	3979.2	2487.9	3000.0
	8	3979.4	2456.2	2995.9
	9	3979.6	2470.2	3011.0
	10	3979.7	2431.3	2969.1
11	3979.8	2571.7	3120.1	
12	3979.8	2611.5	3162.9	

ARKANSAS RIVER VALLEY DATA, SOUTHEASTERN COLORADO
 REFERENCE: KONIKOW, L.F., 1977

OBS. WELL #	MONTH	W-L ELEV FT.	GW CHEM PPM	SP. COND. AT 25 C	WEIGHT FACTOR =
13	0	3967.0	1435.1	1898.0	
	1	3967.2	1630.8	2108.4	
	2	3967.3	1725.5	2210.2	
	3	3967.4	1891.8	2389.0	
	4	3968.0	1884.8	2381.5	
	5	3968.1	1962.0	2564.5	
	6	3967.9	2014.5	2521.0	
	7	3967.5	2033.7	2541.6	
	8	3967.4	1905.8	2404.1	
	9	3967.3	1665.0	2145.2	
	10	3967.2	1719.9	2634.3	
	11	3967.2	1743.7	2229.8	
	12	3967.2	2164.0	2681.7	

OBS. WELL #	MONTH	W-L ELEV FT.	GW CHEM PPM	SP. COND. AT 25 C	WEIGHT FACTOR =
14	0	4039.5	1528.0	1997.8	
	1	4039.7	1534.5	2004.8	
	2	4039.7	1551.2	2022.8	
	3	4039.7	1562.6	2035.1	
	4	4038.6	1594.3	2069.1	
	5	4039.0	1620.3	2097.1	
	6	4038.5	1583.5	2057.5	
	7	4039.6	1621.1	2098.0	
	8	4039.9	1606.7	2082.5	
	9	4040.1	1592.7	2067.4	
	10	4040.2	1521.8	1991.2	
	11	4040.2	1539.0	2009.7	
	12	4040.3	1536.7	2007.2	

OBS. WELL #	MONTH	W-L ELEV FT.	GW CHEM PPM	SP. COND. AT 25 C	WEIGHT FACTOR =
15	0	4040.0	1620.9	2097.7	
	1	4040.1	1633.1	2110.9	
	2	4040.1	1660.0	2139.8	
	3	4040.2	1632.2	2109.9	
	4	4039.7	1694.9	2177.3	
	5	4039.8	1612.4	2088.6	
	6	4039.5	1581.0	2054.8	
	7	4040.0	1586.3	2060.5	
	8	4040.2	1620.7	2097.5	
	9	4040.3	1607.1	2082.9	
	10	4040.4	1566.6	2039.4	
	11	4040.5	1471.3	1936.9	
	12	4040.5	1471.6	1937.0	

ARIZONA RIVER VALLEY DATA, SOUTHEASTERN COLORADO
 REFERENCE: KCMKCM, L.F., 1977

WEIGHT FACTOR = 0.00947

SP. COND.

16

OBS. WELL #	MONTH	W-L ELEV	GW CHEM	PPM
		FT.		
		3971.0		2549.9
		3971.0		2504.7
		3971.1		3007.6
		3971.2		2434.5
		3971.4		2495.0
		3971.5		2483.4
		3971.9		2417.9
		3971.2		2447.4
		3971.1		2986.5
		3971.0		2918.3
		3970.9		2908.8
		3971.0		2383.8
		3971.0		2646.1
		3971.0		2642.5

WEIGHT FACTOR = 0.01356

SP. COND.

17

OBS. WELL #	MONTH	W-L ELEV	GW CHEM	PPM
		FT.		
		4034.5		2457.0
		4034.6		2471.5
		4034.8		2534.1
		4035.0		2492.3
		4034.8		3072.3
		4034.8		2525.8
		4034.7		2505.2
		4034.9		2577.1
		4034.7		3125.9
		4034.9		3089.2
		4034.9		3059.7
		4034.9		3072.7
		4035.0		2920.0
		4035.0		2921.0

336

WEIGHT FACTOR = 0.01933

SP. COND.

18

OBS. WELL #	MONTH	W-L ELEV	GW CHEM	PPM
		FT.		
		4024.0		3386.0
		4024.5		3312.3
		4024.5		3209.7
		4024.7		3186.8
		4024.8		3172.5
		4024.8		3327.6
		4024.4		3343.7
		4024.3		3950.2
		4024.6		3463.4
		4024.6		3382.5
		4024.8		3049.3
		4024.8		3633.7
		4024.8		3692.7
		4024.8		2917.8
		4024.9		3026.4

ARKANSAS RIVER VALLEY DATA, SOUTHEASTERN COLORADO
 REFERENCE: KCAIKOH, L.F., 1977

OBS. WELL # MONTH W-L ELEV GW CHEM SP. COND. WEIGHT FACTOR = 0.00649

OBS. WELL #	MONTH	W-L ELEV	GW CHEM	SP. COND.	WEIGHT FACTOR
19					
		FT.	PPM	AT 25 C	
	0	4002.0	2271.2	2797.0	
	1	4002.2	2383.8	2518.1	
	2	4002.2	2549.9	3096.7	
	3	4002.3	2400.6	2936.1	
	4	4002.9	2431.2	2569.0	
	5	4002.9	2408.0	2944.1	
	6	4002.4	2504.4	3047.7	
	7	4002.2	2514.6	3058.7	
	8	4002.2	1956.2	2458.3	
	9	4002.3	2352.8	2884.7	
	10	4002.2	2739.4	3300.4	
	11	4002.2	2477.8	3019.1	
	12	4002.2	2704.3	3262.1	

OBS. WELL # MONTH W-L ELEV GW CHEM SP. COND. WEIGHT FACTOR = 0.02057

OBS. WELL #	MONTH	W-L ELEV	GW CHEM	SP. COND.	WEIGHT FACTOR
20					
		FT.	PPM	AT 25 C	
	0	3999.5	2364.1	2896.9	
	1	4000.0	2488.3	3030.4	
	2	4000.0	2611.9	3163.3	
	3	4000.1	2739.5	3300.5	
	4	4000.6	2794.6	3359.8	
	5	4000.6	2799.7	3365.3	
	6	4000.2	2902.2	3475.2	
	7	4000.0	3131.2	3721.7	
	8	4000.0	3149.9	3741.8	
	9	4000.1	3571.9	4195.6	
	10	4000.1	3634.6	4263.0	
	11	4000.1	3533.5	4154.3	
	12	4000.0	3673.5	4304.8	

OBS. WELL # MONTH W-L ELEV GW CHEM SP. COND. WEIGHT FACTOR = 0.00779

OBS. WELL #	MONTH	W-L ELEV	GW CHEM	SP. COND.	WEIGHT FACTOR
21					
		FT.	PPM	AT 25 C	
	0	4002.5	2828.6	3396.3	
	1	4002.7	3148.0	3735.8	
	2	4002.6	3322.8	3927.7	
	3	4002.7	3392.3	4002.5	
	4	4003.0	3606.5	4232.8	
	5	4003.1	3834.0	4477.4	
	6	4002.6	3882.7	4529.8	
	7	4002.4	3835.1	4476.4	
	8	4002.5	3857.1	4502.7	
	9	4002.8	3584.4	4209.0	
	10	4002.7	3710.1	4344.2	
	11	4002.6	3522.6	4142.6	
	12	4002.6	3458.6	4073.8	

ARKANSAS RIVER VALLEY DATA, SOUTHEASTERN COLORADO
 REFERENCE: KCAIKOM, L.F., 1977

OBS. WELL # MONTH W-L ELEV W-G CHEM SP. COND. WEIGHT FACTOR = 0.01155

OBS. WELL #	MONTH	W-L ELEV	W-G CHEM	SP. COND.	WEIGHT FACTOR
22		FT.	PPM	AT 25 C	
	0	4000.0	2828.6	3396.3	
	1	4000.6	3090.6	3667.3	
	2	4000.5	3408.4	4019.8	
	3	4000.6	3636.1	4264.6	
	4	4000.8	3817.3	4455.5	
	5	4000.2	3924.4	4574.6	
	6	4000.6	4019.8	4677.2	
	7	4000.3	3899.1	4547.4	
	8	4000.4	3950.6	4554.4	
	9	4000.8	3640.0	4268.8	
	10	4000.7	3657.9	4288.1	
	11	4000.6	3692.7	4325.5	
	12	4000.6	3695.9	4328.9	

OBS. WELL # MONTH W-L ELEV W-G CHEM SP. COND. WEIGHT FACTOR = 0.04055

OBS. WELL #	MONTH	W-L ELEV	W-G CHEM	SP. COND.	WEIGHT FACTOR
23		FY.	PPM	AT 25 C	
	C	3990.0	3571.8	4195.5	
	1	3990.7	3529.1	4149.6	
	2	3990.8	3478.8	4095.5	
	3	3990.9	3519.4	4139.1	
	4	3991.3	3570.8	4194.4	
	5	3991.4	3620.0	4247.3	
	6	3991.0	3624.0	4251.6	
	7	3990.8	3622.0	4240.5	
	8	3990.8	3804.6	4445.8	
	9	3990.9	3782.5	4422.0	
	10	3990.9	3993.3	4646.7	
	11	3990.9	4008.5	4665.1	
	12	3990.9	3906.5	4555.4	

OBS. WELL # MONTH W-L ELEV W-G CHEM SP. COND. WEIGHT FACTOR = 0.03776

OBS. WELL #	MONTH	W-L ELEV	W-G CHEM	SP. COND.	WEIGHT FACTOR
24		FT.	PPM	AT 25 C	
	0	3972.5	2457.0	2696.8	
	1	3972.7	2391.7	2926.6	
	2	3972.8	2340.6	2871.6	
	3	3972.8	2266.1	2791.5	
	4	3972.8	2321.4	2851.0	
	5	3972.9	2384.9	2919.2	
	6	3972.7	2348.0	2879.6	
	7	3972.5	2267.9	2793.3	
	8	3972.4	2418.3	2955.2	
	9	3972.5	2261.9	2787.9	
	10	3972.6	2414.4	2951.0	
	11	3972.6	2437.8	2976.1	
	12	3972.6	2870.1	3441.0	

ARKANSAS RIVER VALLEY DATA, SOUTHEASTERN COLORADO
 REFERENCE: KGNIKOW, L.F., 1977

WEIGHT FACTOR = 0.01901

SP. CCND.

GW CHEM

W-L ELEV

OBS. WELL # MONTH

OBS. WELL #	MONTH	W-L ELEV	GW CHEM	SP. CCND.
25		FT.	PPM	AT 25 C
	0	4036.5	2178.3	2697.1
	1	4036.9	2180.9	2699.9
	2	4037.1	2195.8	2715.9
	3	4037.2	2147.0	2663.4
	4	4037.2	2151.5	2668.3
	5	4037.2	2108.1	2621.6
	6	4037.1	2089.8	2601.9
	7	4037.1	2023.9	2531.1
	8	4037.1	2006.7	2512.6
	9	4037.1	2004.9	2510.6
	10	4037.2	1946.8	2448.2
	11	4037.3	1927.3	2427.2
	12	4037.3	1913.1	2411.9

WEIGHT FACTOR = 0.01804

SP. CCND.

GW CHEM

W-L ELEV

OBS. WELL # MONTH

OBS. WELL #	MONTH	W-L ELEV	GW CHEM	SP. CCND.
26		FT.	PPM	AT 25 C
	0	4008.5	3664.7	4295.4
	1	4008.3	3503.6	4122.1
	2	4008.1	3443.0	4057.6
	3	4008.0	3170.0	3763.4
	4	4008.2	3262.8	3863.2
	5	4008.3	3219.8	3817.0
	6	4007.9	3259.1	3859.2
	7	4007.7	2957.0	3520.4
	8	4008.2	2944.0	3520.4
	9	4008.1	2587.7	3137.3
	10	4008.1	2572.8	3121.1
	11	4008.0	3630.5	4312.4
	12	4008.0	2704.8	3263.2

WEIGHT FACTOR = 0.01499

SP. CCND.

GW CHEM

W-L ELEV

OBS. WELL # MONTH

OBS. WELL #	MONTH	W-L ELEV	GW CHEM	SP. CCND.
27		FT.	PPM	AT 25 C
	0	4009.5	3200.2	3795.9
	1	4009.4	3224.8	3822.4
	2	4009.1	2860.5	3430.6
	3	4008.9	2706.5	3265.1
	4	4008.9	2523.9	3068.7
	5	4008.9	2524.9	3068.7
	6	4008.7	2608.1	3159.2
	7	4008.4	2673.6	3229.7
	8	4008.5	2638.3	3191.7
	9	4009.2	2543.6	3098.9
	10	4009.1	2551.6	3098.9
	11	4009.0	2524.0	3068.8
	12	4009.0	2539.2	3085.2

ARKANSAS RIVER VALLEY DATA, SOUTHEASTERN COLORADO
 REFERENCE: KENTON, L. P., 1977

WEIGHT FACTOR = 0.02374

OBS. WELL #	MONTH	W-L ELEV	GW CHEM	SP. COND.	
				AT 25 C	PPM
28	0	4043.0	1620.9	2097.7	
	1	4043.1	1624.4	2101.5	
	2	4042.9	1643.5	2122.0	
	3	4042.8	1683.9	2165.5	
	4	4042.2	1724.1	2208.7	
	5	4042.0	1685.2	2166.9	
	6	4041.8	1633.6	2165.2	
	7	4041.9	1664.3	2144.3	
	8	4042.1	1673.0	2153.8	
	9	4042.6	1659.8	2139.6	
	10	4042.9	1654.6	2134.0	
	11	4043.0	1556.0	2028.0	
	12	4043.0	1554.2	2026.0	

WEIGHT FACTOR = 0.02277

OBS. WELL #	MONTH	W-L ELEV	GW CHEM	SP. COND.	
				AT 25 C	PPM
29	0	4031.0	1806.7	2297.5	
	1	4030.1	1736.4	2221.9	
	2	4029.7	1864.2	2359.4	
	3	4029.7	1931.4	2431.6	
	4	4029.8	1963.2	2465.8	
	5	4029.6	2235.1	2758.2	
	6	4029.3	2367.1	2900.1	
	7	4029.3	2521.9	3066.6	
	8	4030.1	2474.9	3079.0	
	9	4030.3	2346.2	2877.6	
	10	4030.3	2237.5	2760.8	
	11	4030.4	2006.7	2512.6	
	12	4030.4	1584.2	2058.3	

340

WEIGHT FACTOR = 0.02615

OBS. WELL #	MONTH	W-L ELEV	GW CHEM	SP. COND.	
				AT 25 C	PPM
30	0	4026.0	1063.5	1498.4	
	1	4025.2	1040.3	1471.4	
	2	4024.7	1126.1	1565.7	
	3	4024.9	1228.4	1675.7	
	4	4025.0	1334.1	1843.1	
	5	4024.8	1479.2	1945.4	
	6	4024.5	1583.6	2057.6	
	7	4024.5	1679.1	2160.3	
	8	4025.5	1629.4	2138.2	
	9	4025.5	1610.6	2128.4	
	10	4025.5	1610.6	2086.7	
	11	4025.5	1523.3	1992.8	
	12	4025.6	1450.1	1914.1	

ARKANSAS RIVER VALLEY DATA, SOUTHEASTERN COLORADO
 REFERENCE: KENIKOW, L.F., 1977

OBS. WELL #	MONTH	M-L ELEV FT.	GW CHEM PPM	SP. COND. AT 25 C	WEIGHT FACTOR =
31	0	4005.0	3014.4	3596.1	
	1	4004.6	2927.0	3503.2	
	2	4004.1	2585.0	3556.3	
	3	4004.1	3039.8	3623.4	
	4	4003.9	3294.4	3851.2	
	5	4004.0	3265.4	3973.5	
	6	4003.7	3250.5	3850.0	
	7	4003.3	3558.2	4180.9	
	8	4003.9	3507.1	4125.9	
	9	4004.7	3352.2	4174.4	
	10	4004.6	3288.5	3883.3	
	11	4004.3	3090.0	3677.4	
	12	4004.5	2646.5	3200.5	

OBS. WELL #	MONTH	M-L ELEV FT.	GW CHEM PPM	SP. COND. AT 25 C	WEIGHT FACTOR =
32	0	3988.0	1899.6	2397.4	
	1	3989.1	1682.2	2163.7	
	2	3989.8	1598.8	2074.0	
	3	3989.1	1635.4	2113.3	
	4	3989.3	1675.5	2160.8	
	5	3989.4	1870.9	2366.6	
	6	3988.8	1992.5	2497.3	
	7	3988.8	2010.9	2517.1	
	8	3989.0	2075.4	2586.5	
	9	3989.3	2019.6	2526.5	
	10	3980.3	1943.1	2444.2	
	11	3989.2	1857.7	2352.4	
	12	3989.2	1939.0	2439.8	

OBS. WELL #	MONTH	M-L ELEV FT.	GW CHEM PPM	SP. COND. AT 25 C	WEIGHT FACTOR =
33	0	3982.0	1778.3	2697.1	
	1	3981.7	1945.3	2446.6	
	2	3981.4	1753.5	2240.3	
	3	3981.1	1826.6	2336.1	
	4	3980.5	1784.0	2273.1	
	5	3980.6	2079.9	2591.3	
	6	3980.4	2272.1	2758.0	
	7	3979.9	2303.6	2831.8	
	8	3980.2	2266.9	2792.4	
	9	3981.3	1858.5	2353.2	
	10	3981.3	1590.8	2065.4	
	11	3981.2	1629.8	2107.3	
	12	3981.2	1466.4	1931.6	

ARKANSAS RIVER VALLEY DATA [1977] SOUTHEASTERN COLORADO
 REFERENCE MONITORING, L.P.

OBS. WELL # MONTH M-L ELEV GW CHEM SP. COND. WEIGHT FACTOR = 0.00947

34

OBS. WELL #	MONTH	M-L ELEV	GW CHEM	SP. COND.	WEIGHT FACTOR
34	0	4027.0	PPM	AT 25 C	0.00947
	1	4026.2	784.8	1198.7	
	2	4026.3	838.7	1256.7	
	3	4026.3	984.3	1411.2	
	4	4026.0	161.8	1604.1	
	5	4026.1	1357.2	1814.2	
	6	4026.0	1284.9	1736.5	
	7	4026.0	1386.2	1933.5	
	8	4026.1	1216.8	1845.4	
	9	4026.3	1214.7	1663.2	
	10	4026.3	1172.1	1661.0	
	11	4026.3	1148.7	1615.2	
12	4026.3	1242.6	1590.0		
				1691.0	

OBS. WELL # MONTH M-L ELEV GW CHEM SP. COND. WEIGHT FACTOR = 0.02361

35

OBS. WELL #	MONTH	M-L ELEV	GW CHEM	SP. COND.	WEIGHT FACTOR
35	0	4045.5	PPM	AT 25 C	0.02361
	1	4042.9	1156.4	1558.3	
	2	4040.5	1120.9	1560.1	
	3	4040.3	1269.4	1719.8	
	4	4035.3	1218.2	1664.7	
	5	4035.9	1366.6	1824.3	
	6	4036.1	1338.9	1744.5	
	7	4036.8	1430.9	1893.4	
	8	4039.2	1414.3	1875.6	
	9	4042.8	1356.2	1813.1	
	10	4043.1	1361.7	1819.0	
	11	4043.2	1358.6	1815.7	
12	4042.6	1353.4	1810.1		
				1808.2	

OBS. WELL # MONTH M-L ELEV GW CHEM SP. COND. WEIGHT FACTOR = 0.02615

36

OBS. WELL #	MONTH	M-L ELEV	GW CHEM	SP. COND.	WEIGHT FACTOR
36	0	4019.0	PPM	AT 25 C	0.02615
	1	4018.1	784.8	1198.7	
	2	4018.1	821.7	1238.4	
	3	4017.8	921.2	1345.4	
	4	4017.9	1010.2	1441.5	
	5	4017.7	1103.7	1627.6	
	6	4017.8	1261.4	1711.2	
	7	4017.7	1375.5	1833.9	
	8	4017.9	1525.3	1994.9	
	9	4018.2	1475.9	1941.8	
	10	4018.2	1422.3	1884.2	
	11	4018.3	1468.9	1934.3	
12	4018.3	1432.4	1895.1		
				1882.9	

342

ARKANSAS RIVER VALLEY DATA, SOUTHEASTERN COLORADO
 REFERENCE: KERIKOW, L.F., 1977

WEIGHT FACTOR = 0.01401

OBS. WELL #	MONTH	W-L ELEV FT.	GW CHEM PPM	SP. CGND. AT 25 C
37	0	4007.0	1435.1	1898.0
	1	4006.0	1354.7	1911.5
	2	4005.8	1335.7	1791.1
	3	4005.7	1448.4	1912.3
	4	4005.3	1549.1	2020.5
	5	4005.5	1619.0	2095.7
	6	4005.4	1665.4	2145.6
	7	4005.0	1722.6	2207.1
	8	4005.3	1790.1	2279.7
	9	4006.1	1728.6	2213.5
	10	4006.0	1633.3	2175.6
	11	4005.9	1737.7	2233.3
	12	4005.8	1708.5	2191.9

WEIGHT FACTOR = 0.13546

OBS. WELL #	MONTH	W-L ELEV FT.	GW CHEM PPM	SP. CGND. AT 25 C
38	0	4050.0	1713.8	2177.9
	1	4046.0	1631.8	2109.5
	2	4043.0	1680.6	2161.9
	3	4042.0	1667.6	2148.0
	4	4038.5	1615.1	2091.5
	5	4039.7	1726.2	2211.0
	6	4039.4	1807.6	2283.8
	7	4040.3	1987.0	2681.3
	8	4039.5	1499.2	1966.9
	9	4041.5	1518.0	1987.1
	10	4044.3	1609.2	2085.2
	11	4045.0	1678.6	2159.8
	12	4044.7	1557.9	2030.0

343

WEIGHT FACTOR = 0.02478

OBS. WELL #	MONTH	W-L ELEV FT.	GW CHEM PPM	SP. CGND. AT 25 C
39	0	4015.0	1156.4	1598.3
	1	4014.0	1215.7	1682.0
	2	4013.8	1293.6	1692.0
	3	4013.6	1293.5	1745.7
	4	4013.0	1450.9	1914.9
	5	4013.3	1510.5	1985.5
	6	4013.2	1572.9	2066.1
	7	4012.9	1372.4	1827.7
	8	4013.0	1279.8	1731.0
	9	4013.8	1207.5	1653.2
	10	4013.7	1243.6	1692.0
	11	4013.6	1235.4	1683.2
	12	4013.6	1255.8	1705.2

ARKANSAS RIVER VALLEY DATA, SOUTHEASTERN COLORADO
 REFERENCE: KENIKOM, L.F., 1977

WEIGHT FACTOR = 0.01765

OBS. WELL #	MONTH	M-L ELEV FT.	GW CHEM PPM	SP. CCND. AT 25 C
40				
	0	4005.5	970.6	1398.5
	1	4004.9	1018.2	1449.7
	2	4004.3	1103.1	1541.0
	3	4004.3	1194.5	1631.2
	4	4004.2	1342.9	1758.8
	5	4004.0	1286.7	1738.4
	6	4003.7	1366.8	1818.1
	7	4003.4	1501.5	1969.4
	8	4003.0	1464.0	1929.0
	9	4004.9	1376.7	1835.2
	10	4004.9	1394.7	1900.3
	11	4004.8	1437.3	1900.3
	12	4004.8	1371.0	1829.0

BASE PERIOD SIMULATION (MODEL CAL:3-71 TO 2-72)

ARKANSAS RIVER VALLEY STUDY, DATA SUMMARY FOR 1971

	EMFT (FT)	ETOT (FT)	QTOT (FT)	AQ, CHEM. (MG/LB)	AQ, W-L (FT)	RIVER CHEM. (MG/B)	RIVER FLOW (CFS)	SALT RIVER TRANS (CFS)	HBAK (FT)
MAR	0.011	0.024	.082	2024.30	4011.80	1591.65	52.06	6931.28	17.57
APR	-0.013	0.007	.108	2048.79	4011.60	1588.39	102.53	13184.43	17.37
MAY	-0.016	0.005	.117	2069.50	4011.40	1555.67	143.61	10689.38	17.18
JUN	-0.071	-0.029	.165	2146.19	4011.14	1243.60	465.08	46824.23	16.91
JUL	-0.058	-0.006	.209	2193.85	4010.88	766.11	410.62	26317.16	16.65
AUG	-0.061	0.003	.182	2264.82	4010.71	1263.29	149.32	15780.86	16.48
SEP	-0.036	0.025	.129	2312.86	4010.65	1430.92	101.41	11979.98	16.42
OCT	0.016	0.066	.116	2328.19	4010.77	1400.26	101.77	11921.87	16.55
NOV	0.067	0.083	.115	2308.13	4011.07	1434.55	35.94	4173.56	16.85
DEC	0.046	0.052	.067	2294.14	4011.33	1904.81	33.60	5354.91	17.11
JAN	0.043	0.050	.051	2285.69	4011.47	2064.60	31.50	5440.85	17.24
FEB	0.036	0.045	.048	2272.70	4011.54	2079.54	27.88	4537.08	17.31

YEARLY AVERAGES: 2214.07 4011.20 1526.95 .138.11 14261.30 0.54 16.97

HYDRAULIC RESPONSE TIME = 2.5 MONTHS SOLUTE RESPONSE TIME = 10.4 YEARS

BASE PERIOD SIMULATION (MODEL CAL:3-71 TO 2-72)

ARKANSAS RIVER VALLEY STUDY, DATA SUMMARY FOR 1972

	ENET (FT)	ETOT (FT)	QTOT (FT)	AQ (MG/L)	CHEM. (MG/L)	AQ. W-L (FT)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT (TONS)	RIVER TRANS (GFS)	HBAH (FT)
MAR	0.011	0.024	0.082	2271.57	1582.18	4011.52	1582.18	48.13	6370.27	5.03	17.29
APR	-0.013	0.007	0.108	2295.29	1583.96	4011.46	1583.96	99.70	12384.63	3.50	17.18
MAY	-0.016	0.005	0.117	2336.39	1551.07	4011.27	1551.07	141.70	18345.96	1.50	17.04
JUN	-0.071	-0.022	0.172	2392.76	1243.64	4011.04	1243.64	463.71	46686.89	-1.69	16.82
JUL	-0.058	-0.001	0.214	2438.96	1766.15	4010.81	1766.15	409.70	26258.94	-4.50	16.58
AUG	-0.061	0.007	0.186	2507.95	1263.15	4010.67	1263.15	148.67	15710.44	-6.93	16.44
SEP	-0.036	0.027	0.131	2554.72	1431.24	4010.62	1431.24	102.95	11928.39	-7.85	16.39
OCT	0.016	0.068	0.118	2566.14	1400.41	4010.75	1400.41	101.46	11386.13	-5.74	16.52
NOV	0.067	0.085	0.116	2540.29	1435.43	4011.00	1435.43	35.71	4149.67	-1.49	16.83
DEC	0.046	0.052	0.067	2521.52	1918.32	4011.32	1918.32	33.45	5388.22	2.25	17.09
JAN	0.043	0.050	0.051	2509.71	2093.77	4011.46	2093.77	31.39	5498.96	4.19	17.23
FEB	0.035	0.043	0.046	2493.92	2125.60	4011.53	2125.60	27.95	4489.72	5.75	17.30
YEARLY AVERAGES:				2452.45	1532.91	4011.12	1532.91	137.04	14126.52	-0.53	16.89

HYDRAULIC RESPONSE TIME = 2.6 MONTHS SOLUTE RESPONSE TIME = 9.8 YEARS

BASE PERIOD SIMULATION (MODEL CAL:3-71 TO 2-72)

ARKANSAS RIVER VALLEY STUDY, DATA SUMMARY FOR 1973

	ENET (FT)	ETOT (FT)	QTOT (FT)	AQ. CHEM. (MG/L)	AQ. W-L (FT)	RIVER CHEM. (MG/L)	RIVER FLOW (CF5)	SALT RIVER (TON5)	RIVER TRANS (CF5)	HBAR (FT)
MAR	0.011	0.024	0.082	2490.90	4011.51	1603.00	48.01	6438.86	4.91	17.28
APR	-0.013	0.007	0.10R	2513.64	4011.40	1590.86	99.62	12829.75	3.42	17.17
MAY	-0.016	-0.005	0.117	2554.91	4011.26	1552.97	141.64	18401.37	1.44	17.04
JUN	-0.071	-0.022	0.172	2611.26	4011.04	1243.64	463.66	46682.89	-1.74	16.81
JUL	-0.058	-0.001	0.214	2658.15	4010.81	766.15	409.67	26257.25	-4.93	16.58
AUG	-0.061	0.007	0.186	2726.91	4010.66	1263.14	148.66	15708.41	-6.94	16.44
SEP	-0.036	0.027	0.131	2772.71	4010.62	1431.25	102.93	11928.64	-7.87	16.39
OCT	0.016	0.068	0.118	2780.79	4010.75	1400.41	101.75	11865.07	-5.75	16.52
NOV	0.067	0.085	0.116	2750.12	4011.06	1435.46	35.70	4148.93	-1.50	16.83
DEC	0.046	0.052	0.067	2727.18	4011.32	1932.05	33.45	5405.87	2.25	17.09
JAN	0.043	0.056	0.051	2712.34	4011.46	2120.78	31.39	5569.29	4.19	17.23
FEB	0.035	0.043	0.046	2693.84	4011.53	2166.70	27.95	4576.15	5.75	17.30
YEARLY AVERAGES:				2666.01	4011.12	1542.20	137.01	14152.56	-0.56	16.89

HYDRAULIC RESPONSE TIME = 2.8 MONTHS SOLUTE RESPONSE TIME = 9.8 YEARS

BASE PERIOD SIMULATION (MODEL CAL:3-71 TO 2-72)
 ARKANSAS RIVER VALLEY STUDY, DATA SUMMARY FOR 1974

	ENET (FT)	ETOT (FT)	QTOT (FT)	AQ, CHEM. (MG/L)	AQ, W-L (FT)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT RIVER TRANS (TUNGS)	RIVER TRANS (CFS)	HBAR (FT)
MAR	0.011	0.024	.082	2688.90	4011.51	1623.24	48.01	6519.98	4.91	17.28
APR	-0.013	0.007	.108	2710.75	4011.40	1597.61	99.61	12884.96	3.41	17.17
MAY	-0.016	-0.005	.117	2751.06	4011.26	1554.97	141.64	19424.97	1.44	17.04
JUN	-0.071	-0.022	.172	2808.50	4011.04	1243.64	463.89	46682.89	-1.74	16.81
JUL	-0.058	-0.001	.214	2856.09	4010.81	766.15	409.87	26257.25	-4.93	16.58
AUG	-0.061	0.007	.186	2924.68	4010.69	1263.14	148.66	15708.41	-6.94	16.44
SEP	-0.036	0.027	.131	2969.46	4010.62	1431.23	102.93	11926.89	-7.87	16.39
OCT	0.016	0.068	.118	2974.71	4010.75	1400.41	101.45	11885.07	-5.75	16.52
NOV	0.067	0.085	.116	2939.70	4011.06	1435.46	35.70	4148.93	-1.50	16.83
DEC	0.046	0.052	.067	2913.01	4011.32	1944.53	33.45	5440.79	2.25	17.09
JAN	0.043	-0.056	.051	2895.43	4011.48	2145.22	31.39	5633.48	4.19	17.23
FEB	0.035	0.043	.046	2874.48	4011.53	2203.87	27.95	4654.66	5.75	17.30
YEARLY AVERAGES:				2858.90	4011.12	1550.79	137.01	14180.61	-0.56	16.89

HYDRAULIC RESPONSE TIME = 2.8 MONTHS SOLUTE RESPONSE TIME = 9.8 YEARS

BASE PERIOD SIMULATION (MODEL CAL:3-71 TO 2-72)

ARKANSAS RIVER VALLEY STUDY, DATA SUMMARY FOR 1975

	ENET (ft)	ETOT (ft)	QTOT (ft)	AG. CHEM. (MG/L)	AO. W-L (ft)	RIVER CHEM. (MG/L)	RIVER FLOW (CF6)	SALT (TONS)	NIVER TRANS (CF6)	HBAR (ft)
MAR	0.011	0.024	.082	2867.40	4011.51	1641.55	48.01	6593.51	4.91	17.28
APR	-0.013	0.007	.108	2888.85	4011.40	1503.71	99.61	12933.29	3.41	17.17
MAY	-0.016	0.005	.117	2928.83	4011.20	1556.70	141.04	18446.38	1.44	17.04
JUN	-0.071	-0.022	.172	2986.73	4011.04	1243.64	463.69	46682.89	-1.74	16.81
JUL	-0.058	-0.001	.214	3034.94	4010.81	766.15	409.67	26257.25	-4.93	16.58
AUG	-0.061	0.007	.186	3103.38	4010.66	1263.14	148.66	15708.41	-6.94	16.44
SEP	-0.036	0.027	.131	3147.22	4010.62	1431.25	102.93	11926.89	-7.87	16.39
OCT	0.016	0.028	.118	3149.93	4010.75	1400.41	101.45	11885.07	-5.75	16.52
NOV	0.067	0.085	.116	3111.00	4011.06	1435.46	35.70	4148.93	-1.50	16.83
DEC	0.046	0.052	.067	3080.91	4011.32	1955.81	33.45	5472.34	-2.25	17.09
JAN	0.043	0.050	.051	3050.86	4011.46	2167.31	31.39	5691.48	4.10	17.23
FEB	0.035	0.043	.046	3037.69	4011.53	2237.46	27.95	4725.59	5.75	17.30
YEARLY AVERAGES:				3033.18	4011.12	1558.55	137.01	14206.00	-0.56	16.89

HYDRAULIC RESPONSE TIME = 2.8 MONTHS SOLUTE RESPONSE TIME = 9.8 YEARS

CASE(1): QS=0; CHANGES FROM BASE PERIOD

ARKANSAS RIVER VALLEY STUDY, DATA SUMMARY FOR 1971

	ENET (FT)	ETOT (FT)	QTOT (FT)	AQ. CHEM. (MG/L)	AQ. W-L (FT)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT (TONS)	RIVER TRANS (CFS)	HBAR (FT)
MAR	-0.023	-0.010	0.047	2024.30	4011.80	1591.65	52.06	6931.28	8.96	17.57
APR	-0.056	-0.036	0.055	2057.95	4011.43	1577.86	100.16	12794.54	3.96	17.21
MAY	-0.066	-0.046	0.072	2109.23	4011.09	1542.92	139.25	17973.99	-0.95	16.44
JUN	-0.168	-0.088	0.107	2183.39	4010.60	1243.81	437.34	46032.97	-8.06	16.38
JUL	-0.171	-0.060	0.155	2263.55	4010.08	1266.62	399.20	25601.88	-15.40	15.83
AUG	-0.130	0.002	0.181	2351.03	4009.77	1260.12	136.12	14349.45	-19.48	15.54
SEP	-0.047	0.073	0.177	2396.05	4009.63	1439.91	91.63	10981.56	-17.17	15.61
OCT	-0.003	0.093	0.142	2359.11	4010.14	1404.86	92.91	10919.18	-14.29	15.91
NOV	0.012	0.072	0.104	2379.71	4010.47	1479.07	27.20	3256.52	-10.00	16.24
DEC	0.031	0.065	0.079	2368.79	4010.77	1905.12	25.74	4101.91	-5.46	16.54
JAN	0.043	0.059	0.059	2357.41	4011.05	2034.52	25.59	4366.96	-1.61	16.82
FEB	0.036	0.045	0.048	2341.88	4011.25	2047.04	23.47	3760.41	1.27	17.02
YEARLY AVERAGES:				2269.37	4010.69	1524.88	130.89	13399.22	-6.69	16.46

350

HYDRAULIC RESPONSE TIME = 2.8 MONTHS SOLUTE RESPONSE TIME = 18.9 YEARS

ARKANSAS RIVER VALLEY STUDY: CHANGES FROM BASE PERIOD FOR 1971

	AQ. CHEM. (MG/L)	AQ. CHEM. (TONS)	AQ. VOL (AC-FT)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT (TONS)	RIVER TRANS (CFS)	HBAR (FT)
MAR	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
APR	9.16	-578.89	-391.50	-10.52	-2.37	-389.88	-2.37	-0.16
MAY	19.33	-1054.99	-743.88	-12.75	-4.36	-715.39	-4.36	-0.31
JUN	37.26	-1754.57	-1279.13	0.21	-7.74	-771.26	-7.74	-0.54
JUL	69.69	-2239.51	-1952.38	0.52	-11.43	-715.29	-11.43	-0.82
AUG	87.01	-2548.91	-2255.00	-3.17	-13.20	-1431.41	-13.20	-0.94
SEP	83.09	-1912.50	-1947.75	8.99	-11.78	-1238.42	-11.78	-0.81
OCT	70.22	-1126.92	-1514.25	4.69	-4.87	-1002.69	-4.87	-0.63
NOV	71.59	-756.13	-1444.63	44.53	-8.74	-917.04	-8.74	-0.60
DEC	74.66	-177.99	-1343.75	0.30	-7.67	-1252.94	-7.67	-0.56
JAN	71.72	785.09	-1008.88	-25.08	-5.91	-1073.89	-5.91	-0.42
FEB	69.19	1652.12	-703.88	-32.50	-4.41	-776.67	-4.41	-0.29
AVE CHANGE:	55.30	-701.39	-1215.35	-2.07	-7.22	-862.08	-7.22	-0.51

CASE(1): QS=0; CHANGES FROM BASE PERIOD

ARKANSAS RIVER VALLEY STUDY, DATA SUMMARY FOR 1972

	ENET (ft)	ETOT (ft)	QTOT (ft)	AQ (MG/L)	CHEM. (MG/L)	AQ. H-L (ft)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT (TONS)	RIVER TRANS (CFS)	HBAR (ft)
MAR	-0.023	-0.010	0.047	2346.01	4011.24	1521.97	44.16	5622.24	-1.06	17.01	
APR	-0.056	-0.027	0.074	2381.93	4011.04	1559.00	94.47	11922.91	-1.73	16.81	
MAY	-0.066	-0.026	0.092	2434.13	4010.82	1543.88	135.41	17488.66	-4.79	16.59	
JUN	-0.168	-0.074	0.120	2509.10	4010.41	1243.89	454.57	45776.75	-10.83	16.18	
JUL	-0.171	-0.050	0.165	2590.14	4009.93	766.71	397.33	25484.78	-17.27	15.70	
AUG	-0.130	0.009	0.188	2678.33	4009.67	1259.77	134.82	14207.91	-20.78	15.45	
SEP	-0.047	0.078	0.181	2717.16	4009.77	1440.73	90.69	10577.86	-20.11	15.54	
OCT	0.003	0.096	0.145	2710.30	4010.09	1405.23	92.27	10877.31	-14.93	15.87	
NOV	0.012	0.074	0.106	2681.89	4010.44	1482.21	26.74	3208.46	-10.46	16.52	
DEC	0.031	0.067	0.081	2684.20	4010.75	1907.22	25.43	457.11	-5.77	16.52	
JAN	0.043	0.060	0.060	2647.12	4011.03	2040.94	25.30	4935.19	-1.82	17.80	
FEB	0.035	0.043	0.046	2627.28	4011.23	2058.52	25.31	3625.36	-1.11	17.00	

YEARLY AVERAGES:

	2582.36	4010.53	1519.17	128.71	13096.05	-8.86	16.31
--	---------	---------	---------	--------	----------	-------	-------

HYDRAULIC RESPONSE TIME = 2.8 MONTHS SOLUTE RESPONSE TIME = 13.3 YEARS

ARKANSAS RIVER VALLEY STUDY - CHANGES FROM BASE PERIOD FOR 1972

	AQ (MG/L)	CHEM. (TUNS)	AQ CHEM. (AC-ft)	AQ VOL (MG/L)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT (TONS)	RIVER TRANS (CFS)	HBAR (ft)
MAR	74.45	2021.23	-678.25	-60.21	-3.97	-748.03	-0.28		
APR	86.94	2036.77	-864.63	-24.96	-5.23	-861.72	-0.36		
MAY	97.78	1861.10	-1074.00	-7.18	-9.29	-897.30	-0.45		
JUN	116.34	1208.46	-1509.50	0.25	-9.13	-910.14	-0.63		
JUL	151.88	741.41	-2113.00	0.57	-12.37	-774.16	-0.88		
AUG	170.37	481.50	-2367.13	-3.34	-13.86	-1502.53	-0.95		
SEP	162.23	1156.69	-2026.00	9.49	-12.26	-150.54	-0.85		
OCT	144.16	1460.43	-1568.75	4.82	-9.18	-1038.82	-0.66		
NOV	141.59	2338.61	-1482.75	46.78	-8.97	-941.20	-0.62		
DEC	142.69	2964.38	-1370.25	-11.10	-8.02	-1311.10	-0.57		
JAN	137.41	3999.78	-1027.38	-52.83	-6.02	-1165.77	-0.43		
FEB	133.36	4942.28	-716.75	-67.08	-4.65	-864.36	-0.30		
AVE CHANGE:	129.91	2217.88	-1399.88	-13.74	-8.33	-1030.47	-0.59		

CASE(1): QS=0; CHANGES FROM BASE PERIOD
 ARKANSAS RIVER VALLEY STUDY, DATA SUMMARY FOR 1973

	ENET (ft)	ETOT (ft)	QTOT (ft)	AQ. CHEM. (MG/L)	AQ. W-L (ft)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT (TUNS)	RIVER TRANS (CFS)	HBAR (ft)
MAR	-0.023	0.010	0.047	2630.40	4011.22	1524.63	43.99	5610.92	0.89	17.00
APR	-0.056	0.026	0.074	2667.73	4011.03	1559.06	94.35	11908.35	-1.85	16.91
MAY	-0.066	0.020	0.092	2721.40	4010.81	1543.90	135.33	17478.53	-4.87	16.78
JUN	-0.198	0.074	0.121	2800.84	4010.41	1543.90	454.51	45770.99	-10.89	15.18
JUL	-0.171	0.050	0.163	2888.24	4009.92	1766.71	397.29	25482.32	-17.31	15.70
AUG	-0.130	0.009	0.188	2977.63	4009.67	1259.76	134.79	14204.95	-20.81	15.45
SEP	-0.047	0.078	0.182	3012.27	4009.77	1440.75	90.67	10575.72	-20.13	15.54
OCT	0.003	0.096	0.145	2997.43	4010.09	1405.24	92.26	10845.81	-14.94	15.87
NOV	0.012	0.074	0.106	2961.56	4010.44	1482.27	26.73	3207.44	-10.47	16.21
DEC	0.031	0.067	0.081	2937.96	4010.75	1907.27	25.42	4056.19	-5.78	16.52
JAN	0.043	0.060	0.060	2915.49	4011.03	2040.97	25.37	4332.45	-11.83	16.88
FEB	0.035	0.043	0.046	2891.91	4011.23	2070.96	23.30	3646.68	1.10	17.00
YEARLY AVERAGES:										
				2866.95	4010.53	1520.45	128.67	13093.36	-8.91	16.30

HYDRAULIC RESPONSE TIME = 2.8 MONTHS SOLUTE RESPONSE TIME = 13.2 YEARS

ARKANSAS RIVER VALLEY STUDY - CHANGES FROM BASE PERIOD FOR 1973

	AQ. CHEM. (MG/L)	AQ. CHEM. (TONS)	AQ. VOL (AC-FT)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT (TUNS)	RIVER TRANS (CFS)	HBAR (ft)
MAR	139.40	5375.77	-687.13	-78.37	-4.02	-827.93	-4.02	-0.29
APR	154.09	5442.34	-870.88	-31.80	-5.27	-921.39	-5.27	-0.30
MAY	167.49	5285.57	-1078.25	-9.07	-6.31	-922.84	-6.31	-0.45
JUN	189.58	4601.28	-1512.38	0.25	-9.15	-911.90	-9.15	-0.63
JUL	230.09	4094.03	-2115.25	0.57	-12.38	-774.93	-12.38	-0.89
AUG	250.72	3805.29	-2368.50	-3.38	-13.87	-1503.46	-13.87	-0.99
SEP	239.56	4450.73	-2026.88	-9.50	-12.26	-1351.17	-12.26	-0.85
OCT	216.64	5239.57	-1569.38	4.84	-9.19	-1039.25	-9.19	-0.69
NOV	211.44	5594.72	-1483.25	46.82	-8.97	-941.69	-8.97	-0.62
DEC	210.78	6237.73	-1370.50	-24.78	-8.02	-1349.69	-8.02	-0.57
JAN	203.44	7323.48	-1027.63	-79.81	-6.02	-1236.84	-6.02	-0.43
FEB	198.07	8324.23	-717.00	-95.74	-4.65	-929.47	-4.65	-0.30
AVE CHANGE:	200.94	5567.13	-1402.14	-21.75	-8.34	-1059.20	-8.34	-0.59

CASE(1): QS=0; CHANGES FROM BASE PERIOD

ARKANSAS RIVER VALLEY STUDY, DATA SUMMARY FOR 1974

	ENET (PT)	ETOT (PT)	QTOT (PT)	AQ, CHEM. (MG/L)	AQ, W-L (FT)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT RIVER FLOW (TONS)	RIVER TRANS (CFS)	HBAR (FT)
MAR	-0.023	-0.010	.047	2893.65	4011.22	1529.89	43.99	5629.95	0.89	17.00
APR	-0.056	-0.026	.074	2932.53	4011.03	1559.06	94.34	11908.09	-1.86	16.80
MAY	-0.066	-0.020	.092	2988.44	4010.81	1543.90	135.33	17478.31	-4.87	16.58
JUN	-0.168	-0.074	.121	3071.34	4010.41	1243.89	454.51	45770.86	-10.89	16.18
JUL	-0.171	-0.050	.155	3164.11	4009.92	1266.71	397.29	25482.27	-17.31	15.70
AUG	-0.130	-0.009	.188	3255.33	4009.67	1259.76	134.79	14204.86	-20.81	15.45
SEP	-0.047	0.078	.162	3286.13	4009.77	1440.75	105.67	10575.62	-20.13	15.54
OCT	-0.003	0.096	.145	3263.91	4010.09	1305.24	92.26	10845.76	-14.94	15.87
NOV	0.012	0.074	.106	3221.14	4010.44	1482.27	26.73	3207.44	-10.47	16.21
DEC	0.031	0.067	.081	3192.07	4010.75	1907.27	25.42	4056.19	-5.78	16.52
JAN	0.043	0.060	.060	3165.17	4011.03	2040.97	25.37	4332.45	-1.83	16.80
FEB	0.035	0.043	.046	3137.54	4011.23	2082.60	23.30	3667.17	1.10	17.00
YEARLY AVERAGES:				3130.95	4010.53	1521.86	128.67	13096.58	-8.91	16.30

HYDRAULIC RESPONSE TIME = 2.8 MONTHS SOLUTE RESPONSE TIME = 13.2 YEARS

ARKANSAS RIVER VALLEY STUDY: CHANGES FROM BASE PERIOD FOR 1974

	AQ, CHEM. (MG/L)	AQ, CHEM. (TONS)	AQ, VOL (AC-FT)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT RIVER FLOW (TONS)	RIVER TRANS (CFS)	HBAR (FT)
MAR	204.75	8801.71	-687.38	-93.35	-4.02	-890.03	-4.02	-0.29
APR	221.78	8906.58	-811.00	-38.54	-5.27	-975.97	-5.27	-0.36
MAY	237.38	8765.29	-1078.38	-11.07	-6.31	-946.66	-6.31	-0.45
JUN	262.84	8449.02	-1312.63	0.25	-7.15	-912.03	-7.15	-0.63
JUL	308.02	7501.14	-2115.38	0.57	-12.38	-774.98	-12.38	-0.89
AUG	330.65	7181.73	-2368.63	-3.38	-13.87	-1503.55	-13.87	-0.99
SEP	316.67	7813.46	-2027.00	9.50	-12.26	-1351.27	-12.26	-0.85
OCT	289.20	8507.78	-1569.50	4.82	-8.97	-1039.30	-8.97	-0.68
NOV	241.43	8901.45	-1431.25	46.82	-8.97	-941.48	-8.97	-0.62
DEC	279.06	9558.86	-1370.50	-37.26	-8.02	-1384.60	-8.02	-0.57
JAN	269.73	10689.51	-1027.63	-104.25	-6.02	-1301.03	-6.02	-0.43
FEB	263.07	11741.93	-717.60	-121.28	-6.65	-987.48	-6.65	-0.40
AVE CHANGE:	272.05	8968.48	-1402.21	-28.93	-8.34	-1084.03	-8.34	-0.59

CASE(1) QS=0; CHANGES FROM BASE PERIOD

ARKANSAS RIVER VALLEY STUDY, DATA SUMMARY FOR 1975

	ENET (FT)	ETOT (FT)	QTOT (FT)	AQ, CHEM. (MG/L)	AQ, W-L (FT)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT (TONS)	RIVER TRANS (CFS)	HBAR (FT)
MAR	-0.023	0.010	0.47	3138.10	4011.22	1534.83	43.99	5648.13	0.89	17.00
APR	-0.056	0.026	0.74	3178.32	4011.03	1559.06	94.34	11908.09	-1.86	16.80
MAY	-0.066	0.020	0.92	3255.42	4010.41	1543.90	135.33	17478.31	-4.87	16.58
JUN	-0.168	0.074	1.21	3322.45	4010.41	1243.89	454.51	45770.86	-10.89	15.18
JUL	-0.171	0.050	1.65	3420.20	4009.92	1766.71	397.29	25482.27	-17.31	15.45
AUG	-0.130	0.009	1.88	3513.11	4009.67	1259.76	134.79	14204.86	-20.81	15.45
SEP	-0.047	0.078	1.82	3540.34	4009.77	1440.75	190.67	10575.62	-20.13	15.54
OCT	0.003	0.096	1.45	3511.28	4010.09	1405.24	92.26	10845.76	-14.94	15.87
NOV	0.012	0.074	1.06	3462.11	4010.44	1482.27	26.73	3207.44	-10.47	16.21
DEC	0.031	0.067	0.81	3427.95	4010.75	1907.27	25.42	4056.19	-5.78	16.52
JAN	0.043	0.060	0.60	3398.66	4011.03	2040.97	25.37	4332.45	-1.83	16.80
FEB	0.035	0.043	0.46	3365.57	4011.23	2093.40	23.30	3686.19	1.10	17.00

YEARLY AVERAGES: 3376.00 4010.53 1523.17 128.67 13099.68 -8.91 16.30

HYDRAULIC RESPONSE TIME = 2.8 MONTHS SOLUTE RESPONSE TIME = 13.2 YEARS

ARKANSAS RIVER VALLEY STUDY - CHANGES FROM BASE PERIOD FOR 1975

	AQ, CHEM. (MG/L)	AQ, CHEM. (TONS)	AQ VOL (AC-FT)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT (TONS)	RIVER TRANS (CFS)	HBAR (FT)
MAR	270.30	12257.33	-687.38	-106.72	-4.02	-945.38	-4.02	-0.29
APR	289.47	12394.68	-871.00	-44.65	-5.27	-1025.20	-5.27	-0.36
MAY	307.12	12265.92	-1078.38	-12.87	-6.31	-968.07	-6.31	-0.45
JUN	314.72	11516.47	-1412.63	0.25	-9.15	-912.03	-9.15	-0.63
JUL	385.25	10928.26	-2115.38	0.57	-12.38	-774.98	-12.38	-0.99
AUG	409.73	10577.98	-2368.64	-3.38	-13.87	-1503.55	-13.87	-0.85
SEP	393.12	11186.00	-2027.00	9.50	-12.26	-1351.27	-12.26	-0.66
OCT	361.35	11415.41	-1569.60	4.82	-9.19	-1039.30	-9.19	-0.62
NOV	351.10	12227.81	-1483.25	46.82	-8.97	-941.48	-8.97	-0.62
DEC	347.04	12897.46	-1370.50	48.54	-6.02	-1416.16	-6.02	-0.57
JAN	335.80	14068.07	-1027.63	-126.34	-8.02	-1359.03	-8.02	-0.57
FEB	327.87	15166.11	-717.00	-144.06	-4.65	-1039.40	-4.65	-0.50
AVE CHANGE:	342.82	12388.24	-1402.21	-35.38	-8.34	-1106.32	-8.34	-0.59

CASE(2): QP=0; CHANGES FROM BASE PERIOD

ARKANSAS RIVER VALLEY STUDY, DATA SUMMARY FOR 1971

	EMET (FT)	ETOT (FT)	QTOT (FT)	AO, CHEM. (MG/L)	AO, W-L (FT)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT (TONS)	RIVER TRANS (CFS)	HBAR (FT)
MAR	0.024	0.036	0.094	2024.30	4011.80	1591.65	52.06	6931.24	8.96	17.57
APR	0.189	0.209	0.310	2024.84	4012.08	1614.96	109.55	14322.52	13.35	17.86
MAY	0.152	0.173	0.285	2030.46	4012.51	1600.76	159.13	21309.92	18.93	18.29
JUN	0.007	0.032	0.227	2055.42	4012.34	1273.20	483.02	49787.06	17.62	18.15
JUL	0.095	0.126	0.342	2060.20	4012.21	810.30	429.30	29103.32	14.72	17.99
AUG	0.146	0.178	0.357	2083.22	4012.41	1347.30	173.08	19508.32	17.48	18.18
SEP	0.160	0.184	0.288	2085.74	4012.68	1535.29	132.74	16498.73	21.94	18.45
OCT	0.026	0.049	0.098	2085.98	4012.58	1505.71	127.13	16014.06	19.93	18.36
NOV	0.073	0.083	0.115	2074.62	4012.37	1636.14	54.70	7245.83	17.50	18.14
DEC	0.048	0.055	0.069	2065.56	4012.25	1937.58	46.50	7537.64	15.30	18.03
JAN	0.044	0.051	0.051	2060.51	4012.12	2039.81	40.57	6923.68	13.37	17.89
FEB	0.036	0.045	0.048	2050.74	4011.99	2037.56	34.67	5528.39	12.47	17.77
YEARLY AVERAGES:				2058.52	4012.28	1577.52	153.54	16725.86	15.96	18.06

355

HYDRAULIC RESPONSE TIME = 2.6 MONTHS SOLUTE RESPONSE TIME = 3.0 YEARS

ARKANSAS RIVER VALLEY STUDY - CHANGES FROM BASE PERIOD FOR 1971

	AO, CHEM. (MG/L)	AO, CHEM. (TONS)	AO, VOL (AC-FT)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT (TONS)	RIVER TRANS (CFS)	HBAR (FT)
MAR	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
APR	-23.95	1842.62	1159.88	26.58	7.02	1138.09	7.02	0.49
MAY	-59.04	4028.71	2651.38	45.09	15.52	2620.54	15.52	1.11
JUN	-90.77	3297.79	2964.00	29.60	17.93	2952.81	17.93	1.24
JUL	-133.66	1714.06	3193.00	44.21	18.69	2756.15	18.69	1.34
AUG	-180.80	1810.69	4058.25	84.01	23.76	3727.46	23.76	1.70
SEP	-227.02	1633.09	4847.13	104.37	29.32	4519.75	29.32	2.03
OCT	-242.21	738.14	431.38	105.45	25.36	4092.19	25.36	1.81
NOV	-233.20	-4035.45	3101.88	201.67	18.77	3072.27	18.77	1.30
DEC	-228.58	-6522.10	2203.13	32.77	12.90	2182.94	12.90	0.92
JAN	-225.18	-8281.46	1549.00	-24.79	9.07	1482.33	9.07	0.65
FEB	-221.91	-9465.18	1085.13	-41.98	6.79	991.51	6.79	0.45
AVE CHANGE:	-155.55	-1314.69	2595.39	50.58	15.43	2464.56	15.43	1.09

CASE(2): QP=0; CHANGES FROM BASE PERIOD

ARKANSAS RIVER VALLEY STUDY, DATA SUMMARY FOR 1972

EMET (FT)	ETOT (FT)	QTOT (FT)	AO, CHEM. (MG/L)	AO, W-L (FT)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT RIVER TRANS (CFS)	HBAR (FT)
MAR 0.024	0.036	0.094	2050.22	4011.86	1603.90	52.96	7106.55	17.64
APR 0.189	0.209	0.310	2049.89	4012.13	1620.58	110.20	14458.33	17.90
MAY 0.152	0.172	0.285	2054.84	4012.54	1604.84	159.57	21423.63	18.32
JUN -0.007	0.032	0.227	2078.51	4012.40	1274.57	483.33	49873.68	18.17
JUL 0.095	0.126	0.342	2082.79	4012.23	811.73	429.54	29168.72	18.00
AUG 0.146	0.178	0.357	2104.88	4012.42	1350.14	174.23	19566.44	18.19
SEP 0.160	0.184	0.288	2106.58	4012.68	1539.17	132.55	16553.97	18.48
OCT 0.026	0.049	0.098	2105.99	4012.59	1509.20	127.21	16060.37	18.39
NOV 0.073	0.083	0.115	2094.28	4012.37	1642.87	54.76	7282.66	18.15
DEC 0.048	0.055	0.099	2084.85	4012.26	1944.04	46.54	7568.72	18.03
JAN 0.044	0.051	0.051	2079.52	4012.12	2048.09	40.60	6948.66	17.89
FEB 0.035	0.044	0.047	2069.70	4011.99	2044.67	35.09	5421.23	17.77

YEARLY AVERAGES:

2080.16	4012.30	1582.65	153.82	16786.08	16.25	18.07
---------	---------	---------	--------	----------	-------	-------

HYDRAULIC RESPONSE TIME = 2.8 MONTHS SOLUTE RESPONSE TIME = 3.0 YEARS

ARKANSAS RIVER VALLEY STUDY: CHANGES FROM BASE PERIOD FOR 1972

	AO, CHEM. (MG/L)	AO, CHEM. (TONS)	AO, VOL (AC-FT)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT RIVER TRANS (CFS)	HBAR (FT)
MAR	-221.35	-10142.84	825.88	21.72	4.84	4.84	0.35
APR	-245.39	-8862.70	1736.25	36.62	10.50	10.50	0.73
MAY	-281.71	-7073.32	3053.50	53.76	17.88	17.88	1.28
JUN	-314.25	-8408.05	3244.50	30.93	19.63	19.63	1.36
JUL	-356.17	-9602.83	3388.88	45.59	19.84	19.84	1.42
AUG	-403.07	-9532.10	4194.75	87.00	24.50	24.50	1.76
SEP	-448.34	-9726.32	4942.50	107.93	29.90	29.90	2.07
OCT	-460.14	-12421.11	4397.88	108.79	25.75	25.75	1.84
NOV	-446.01	-15438.20	3148.25	207.44	19.05	19.05	1.32
DEC	-436.66	-17930.95	2235.38	25.71	13.09	13.09	0.94
JAN	-430.19	-19659.94	1571.50	-47.68	9.20	9.20	0.66
FEB	-424.22	-20770.61	1100.89	-80.93	7.14	7.14	0.48
AVE CHANGE:	-372.29	-12469.76	2820.04	49.74	16.78	16.78	1.18

CASE(2): QP=0; CHANGES FROM BASE PERIOD

ARKANSAS RIVER VALLEY STUDY, DATA SUMMARY FOR 1973

	EMET (FT)	ETOT (FT)	QTOT (FT)	AQ (MG/L)	AQ (MG/L)	W-L (FT)	RIVER CHEM. (MG/L)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT (TONS)	RIVER TRANS (CFS)	HBAR (FT)
MAR	0.024	0.036	0.094	2069.10	4011.80		1606.98	52.91	7113.49	9.81	17.63	
APR	0.189	0.209	0.310	2068.15	4012.13		1622.76	110.17	14472.97	13.97	17.90	
MAY	0.152	0.173	0.285	2071.28	4012.54		1606.88	159.55	21447.51	19.35	18.32	
JUN	0.007	0.032	0.227	2095.38	4012.40		1275.17	483.32	49895.18	17.82	18.17	
JUL	0.095	0.126	0.342	2099.30	4012.23		812.27	429.25	29187.25	14.93	18.00	
AUG	0.146	0.178	0.357	2120.72	4012.42		1351.72	173.23	19588.40	17.63	18.19	
SEP	0.160	0.184	0.288	2121.87	4012.68		1541.65	132.84	16579.87	22.04	18.46	
OCT	0.026	0.049	0.098	2120.83	4012.59		1511.48	127.20	16084.18	20.00	18.36	
NOV	0.073	0.083	0.115	2108.85	4012.37		1647.45	54.75	7302.63	17.55	18.03	
DEC	0.048	0.055	0.069	2098.95	4012.26		1948.68	46.54	7586.52	15.34	18.03	
JAN	0.054	0.051	0.051	2093.42	4012.12		2050.68	40.59	6994.01	13.39	17.88	
FEB	0.035	0.044	0.047	2083.42	4011.99		2049.71	35.09	5434.44	12.89	17.77	
YEARLY AVERAGES:				2095.95	4012.30		1585.45	153.81	16804.70	16.23	18.07	

HYDRAULIC RESPONSE TIME = 2.8 MONTHS SOLUTE RESPONSE TIME = 3.0 YEARS

ARKANSAS RIVER VALLEY STUDY: CHANGES FROM BASE PERIOD FOR 1973

	AQ (MG/L)	AQ (TONS)	RIVER CHEM. (MG/L)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT (TONS)	RIVER TRANS (CFS)	HBAR (FT)
MAR	-421.79	-21350.74	836.75	3.98	4.90	674.63	4.90	0.35
APR	-445.49	-19964.97	1743.75	31.90	10.55	1643.22	10.55	0.73
MAY	-482.33	-18096.45	3058.75	53.90	17.91	3046.15	17.91	1.28
JUN	-515.88	-18943.45	3248.25	31.53	19.65	3212.29	19.65	1.36
JUL	-558.85	-20445.02	3391.38	46.12	19.86	2930.00	19.86	1.42
AUG	-606.19	-20290.92	4196.63	88.58	24.57	3880.00	24.57	1.76
SEP	-651.04	-20424.83	4943.75	110.40	29.91	4652.98	29.91	2.07
OCT	-660.16	-22775.88	4398.88	111.06	25.73	4193.11	25.73	1.84
NOV	-641.47	-26070.18	3149.00	212.00	19.05	3153.71	19.05	1.32
DEC	-628.23	-28533.91	2236.00	16.63	13.09	2180.65	13.09	0.94
JAN	-618.92	-30204.88	1571.88	-70.10	9.20	1394.72	9.20	0.66
FEB	-610.42	-31226.79	1101.00	-117.00	7.14	858.29	7.14	0.46
AVE CHANGE:	-570.06	-23260.49	2823.03	43.25	16.80	2652.14	16.80	1.18

CASE(2) QP#01 CHANGES FROM BASE PERIOD
 ARKANSAS RIVER VALLEY STUDY, DATA SUMMARY FOR 1974

	ENET (ft)	ETOT (ft)	QTOT (ft)	AQ. CHEM. (MG/L)	AQ. W-L (ft)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT (TONS)	RIVER TRANS (CFS)	HBAR (ft)
MAR	0.024	0.036	0.094	2082.67	4011.86	1609.49	52.91	7124.47	9.81	17.63
APR	0.189	0.209	0.310	2081.26	4012.13	1624.42	110.16	14487.72	13.96	17.90
MAY	0.152	0.173	0.283	2084.43	4012.54	1608.79	159.55	21467.60	19.35	18.32
JUN	0.095	0.032	0.227	2107.49	4012.40	1275.62	483.32	49912.75	17.92	18.17
JUL	0.146	0.178	0.357	2132.09	4012.23	1872.68	429.53	29202.05	14.93	18.00
AUG	0.140	0.184	0.357	2132.51	4012.42	1352.88	173.23	19605.17	17.63	18.19
SEP	0.026	0.049	0.088	2131.13	4012.59	1543.45	132.84	16599.71	22.04	18.46
OCT	0.073	0.083	0.115	2114.97	4012.37	1513.13	127.20	16101.76	20.00	18.36
NOV	0.048	0.055	0.069	2109.08	4012.26	1650.76	54.75	7317.30	17.55	18.15
DEC	0.044	0.051	0.051	2103.40	4012.12	1952.02	46.54	7599.51	15.34	18.03
JAN	0.044	0.051	0.051	2093.27	4011.99	2053.97	40.59	6975.19	13.39	17.89
FEB	0.035	0.044	0.047	2093.27	4011.99	2053.32	35.09	5444.04	12.89	17.77
YEARLY AVERAGES:				2107.29	4012.30	1587.51	153.81	16819.73	16.23	18.07

HYDRAULIC RESPONSE TIME * 2.4 MONTHS SOLUTE RESPONSE TIME = 3.0 YEARS

ARKANSAS RIVER VALLEY STUDY: CHANGES FROM BASE PERIOD FOR 1974

	AQ. CHEM. (MG/L)	AQ. CHEM. (TONS)	AQ. VOL (AC-FT)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT (TONS)	RIVER TRANS (CFS)	HBAR (FT)
MAR	-606.23	-31700.87	836.88	-13.75	4.90	604.50	4.90	0.35
APR	-629.49	-30206.10	1743.88	26.81	10.55	1603.66	10.55	0.73
MAY	-666.63	-28253.89	3058.88	53.42	17.91	3042.63	17.91	1.28
JUN	-701.01	-24411.63	3248.25	31.98	17.91	2229.88	17.91	1.46
JUL	-744.94	-30423.80	3391.38	46.54	19.65	2944.80	19.65	1.42
AUG	-792.59	-30188.64	4196.63	89.73	24.57	3896.77	24.57	1.76
SEP	-836.94	-30259.22	4043.75	112.20	29.91	4672.32	29.91	2.07
OCT	-843.58	-32547.72	4348.68	112.72	25.75	4218.69	25.75	1.84
NOV	-820.73	-35836.16	3149.00	215.31	19.09	3168.37	19.09	1.32
DEC	-803.93	-34269.29	2236.00	7.49	13.09	2158.72	13.09	0.94
JAN	-792.03	-39883.59	1571.88	-91.25	19.20	1341.71	19.20	0.66
FEB	-781.21	-40822.20	1101.00	-150.55	7.14	789.38	7.14	0.46
AVE CHANGE:	-751.61	-33188.20	2823.17	36.72	16.80	2639.12	16.80	1.18

CASE(2): OP=0; CHANGES FROM BASE PERIOD

ARKANSAS RIVER VALLEY STUDY, DATA SUMMARY FOR 1975

EMET (FT)	ETOT (FT)	QTOT (FT)	AQ, CHEM. (MG/L)	AQ, M-L (FT)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT RIVER (TONS)	RIVER TRANS (CFS)	HBAR (FT)
MAR 0.024	0.036	0.094	2092.41	4011.86	1611.30	524.91	7132.47	9.81	17.63
APR 0.189	0.209	0.310	2090.68	4012.13	1625.61	110.16	14498.36	13.96	17.90
MAY 0.152	0.173	0.265	2093.98	4012.54	1609.47	159.55	21482.97	17.35	18.32
JUN 0.007	0.032	0.227	2116.19	4012.40	1275.94	483.32	49925.36	17.92	18.17
JUL 0.095	0.176	0.342	2119.16	4012.23	812.98	429.53	29212.68	14.93	18.00
AUG 0.146	0.178	0.357	2140.28	4012.42	1353.71	173.23	19617.21	17.63	18.19
SEP 0.160	0.184	0.288	2140.29	4012.68	1544.74	127.84	16713.89	22.04	18.46
OCT 0.026	0.049	0.098	2138.67	4012.59	1514.32	127.70	16114.38	20.00	18.36
NOV 0.073	0.083	0.115	2126.38	4012.37	1653.14	54.75	7327.83	17.55	18.15
DEC 0.048	0.055	0.069	2116.35	4012.26	1954.41	46.54	7608.84	15.34	18.03
JAN 0.044	0.051	0.051	2110.57	4012.12	2056.34	46.59	6983.72	13.39	17.89
FEB 0.035	0.044	0.047	2100.34	4011.99	2055.92	35.09	5450.92	12.89	17.77

YEARLY AVERAGES:

2115.43	4012.30	1588.99	153.81	16830.54	16.23	18.07
---------	---------	---------	--------	----------	-------	-------

HYDRAULIC RESPONSE TIME = 2.0 MONTHS SOLUTE RESPONSE TIME = 3.0 YEARS

ARKANSAS RIVER VALLEY STUDY - CHANGES FROM BASE PERIOD FOR 1975

MO	AQ, CHEM. (MG/L)	AQ, CHEM. (TONS)	AQ, VOL (AC-FT)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT RIVER (TONS)	RIVER TRANS (CFS)	HBAR (FT)
MAR	-775.39	-41197.31	836.88	-30.25	4.90	538.96	4.90	0.35
APR	-798.17	-39601.28	1743.88	21.90	10.55	1565.07	10.55	0.73
MAY	-835.46	-37569.26	1028.88	52.69	17.91	3035.69	17.91	1.28
JUN	-870.54	-36241.48	3249.25	32.30	10.65	3242.47	10.65	1.36
JUL	-915.29	-39568.60	3391.38	46.83	19.89	2955.42	19.89	1.42
AUG	-963.12	-39265.72	4196.63	90.56	24.57	3908.81	24.57	1.76
SEP	-1006.93	-4530.52	4943.75	113.49	29.91	4686.21	29.91	2.07
OCT	-984.22	-4773.14	3348.88	113.90	25.75	4229.31	25.75	1.84
NOV	-964.56	-47175.81	3149.00	217.68	19.05	3178.90	19.05	1.32
DEC	-950.29	-48736.65	2236.00	-110.27	13.20	2191.74	13.20	0.66
JAN	-937.36	-49597.65	11014.00	-181.53	7.14	725.34	7.14	0.46
FEB								
AVE CHANGE:	-917.75	-42282.04	2823.17	30.43	16.80	2624.54	16.80	1.18

CASE(3): IMPR IRR EFF; CHANGES FROM BASE PERIOD

ARKANSAS RIVER VALLEY STUDY, DATA SUMMARY FOR 1971

	ENET (FT)	ETOT (FT)	QTOT (FT)	AO (MG/L)	AO (FT)	W-L (FT)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT (TONS)	RIVER TRANS (CFS)	HBAR (FT)
MAR	0.013	0.026	0.075	2024.30	4011.80		1591.65	52.06	6931.28	8.96	17.57
APR	-0.038	-0.017	0.099	2052.70	4011.55		1585.18	101.77	13060.81	5.57	17.32
MAY	-0.040	-0.019	0.107	2105.02	4011.27		1548.57	141.68	18354.76	1.48	17.04
JUN	-0.064	-0.014	0.155	2161.15	4011.01		1243.65	463.24	46640.36	-2.16	16.78
JUL	-0.072	-0.012	0.194	2203.84	4010.77		1766.17	409.12	26222.85	-5.48	16.54
AUG	-0.079	-0.004	0.179	2276.47	4010.57		1262.85	147.33	15564.70	-8.27	16.34
SEP	-0.048	-0.024	0.138	2330.57	4010.44		1432.52	101.06	11220.39	-9.74	16.26
OCT	0.018	0.078	0.122	2345.07	4010.64		1401.16	99.54	11711.43	-7.29	16.41
NOV	0.054	0.079	0.113	2323.48	4010.95		1441.47	34.23	3994.20	-2.97	16.73
DEC	0.045	0.051	0.063	2309.99	4011.22		1886.59	32.07	5061.18	0.87	17.00
JAN	0.043	0.050	0.051	2300.94	4011.39		2058.18	30.40	5234.63	3.20	17.16
FEB	0.036	0.045	0.048	2287.68	4011.48		2076.39	27.06	4397.12	4.86	17.26
YEARLY AVERAGES:				2226.77	4011.09		1524.53	136.66	14074.41	-0.91	16.87

HYDRAULIC RESPONSE TIME = 2.8 MONTHS SOLUTE RESPONSE TIME = 11.7 YEARS

ARKANSAS RIVER VALLEY STUDY; CHANGES FROM BASE PERIOD FOR 1971

	AO (MG/L)	AO (TONS)	AO VOL (AC-FT)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT (TONS)	RIVER TRANS (CFS)	HBAR (FT)
MAR	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
APR	3.91	-128.30	-125.00	-3.21	-0.76	-123.62	-0.76	-0.05
MAY	15.12	-196.86	-328.75	-7.10	-1.97	-334.62	-1.92	-0.14
JUN	14.96	-74.88	-504.88	0.05	-1.84	-183.87	-1.84	-0.11
JUL	9.99	-232.58	-258.00	0.07	-1.51	-94.51	-1.51	-0.11
AUG	12.45	-387.40	-340.50	-0.44	-1.99	-216.17	-1.99	-0.14
SEP	17.62	-293.87	-349.38	1.63	-2.36	-259.59	-2.36	-0.16
OCT	16.88	-105.64	-317.88	-0.90	-1.86	-110.44	-1.86	-0.13
NOV	15.35	52.12	-282.50	6.92	-1.71	-179.36	-1.71	-0.12
DEC	15.85	56.98	-262.38	-18.22	-1.54	-293.73	-1.54	-0.11
JAN	15.29	267.19	-198.88	-9.41	-1.10	-206.82	-1.10	-0.08
FEB	14.98	436.12	-130.88	-3.16	-0.92	-139.96	-0.82	-0.08
AVE CHANGE:	12.70	-38.55	-243.97	-2.41	-1.45	-186.89	-1.45	-0.10

CASE(3): IMPR IRR EFF; CHANGES FROM BASE PERIOD

ARKANSAS RIVER VALLEY STUDY, DATA SUMMARY FOR 1972

	ENET (FT)	ETOT (FT)	GTOT (FT)	AQ. CHEM. (MG/L)	AQ. W-L (FT)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT (TONS)	RIVER TRANS (CFS)	HBAR (FT)
MAR	0.013	0.026	.075	2284.64	4011.49	1576.59	47.66	6285.42	4.56	17.26
APR	-0.038	-0.017	.099	2313.16	4011.33	1576.48	98.60	12584.34	2.40	17.10
MAY	-0.040	-0.045	.110	2367.66	4011.11	1542.85	139.54	18010.49	-0.66	16.89
JUN	-0.064	-0.006	.163	2422.94	4010.90	1743.69	461.69	46486.49	-3.71	16.68
JUL	-0.072	-0.006	.200	2463.46	4010.69	1766.22	408.08	26157.42	-9.52	16.47
AUG	-0.079	-0.001	.182	2535.25	4010.52	1262.68	146.60	15485.85	-9.00	16.29
SEP	-0.048	0.027	.141	2587.52	4010.45	1432.92	100.53	11662.63	-10.27	16.22
OCT	0.018	0.080	.123	2597.24	4010.61	1401.33	99.56	11671.39	-7.64	16.39
NOV	0.054	0.080	.114	2569.16	4010.94	1442.56	31.97	3967.44	-3.23	16.51
DEC	0.045	0.051	.063	2550.75	4011.21	1489.55	31.90	5041.85	0.70	16.98
JAN	0.043	0.050	.051	2538.20	4011.18	2001.33	30.28	5271.93	3.08	17.15
FEB	0.035	0.043	.046	2521.91	4011.48	2119.00	27.10	4338.55	4.90	17.25
YEARLY AVERAGES:				2479.27	4011.01	1527.93	135.46	13913.05	-2.12	16.78

361

HYDRAULIC RESPONSE TIME = 2.8 MONTHS SOLUTE RESPONSE TIME = 10.8 YEARS

ARKANSAS RIVER VALLEY STUDY; CHANGES FROM BASE PERIOD FOR 1972

	AQ. CHEM. (MG/L)	AQ. CHEM. (TONS)	AQ. VOL (AC-FT)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT (TONS)	RIVER TRANS (CFS)	HBAR (FT)
MAR	13.07	484.05	-80.75	-5.59	-0.47	-84.85	-0.47	-0.03
APR	17.87	427.48	-181.25	-7.48	-1.10	-200.29	-1.10	-0.08
MAY	30.71	516.76	-368.00	-8.22	-2.15	-375.47	-2.15	-0.15
JUN	50.18	554.63	-332.38	-0.06	-2.01	-209.40	-2.01	-0.14
JUL	24.49	392.08	-277.00	0.07	-1.62	-101.52	-1.62	-0.12
AUG	27.30	238.68	-353.88	-0.46	-2.07	-221.59	-2.07	-0.15
SEP	32.60	333.71	-398.63	1.69	-2.41	-265.76	-2.41	-0.17
OCT	31.10	524.84	-324.25	-0.93	-2.41	-214.74	-2.41	-0.17
NOV	28.87	576.47	-287.13	7.13	-1.74	-182.23	-1.74	-0.12
DEC	29.23	702.90	-265.63	-28.77	-1.56	-326.37	-1.56	-0.11
JAN	28.49	937.90	-190.63	-12.44	-1.12	-227.03	-1.12	-0.08
FEB	27.99	1121.99	-132.13	-6.59	-0.86	-151.17	-0.86	-0.06
AVE CHANGE:	26.82	576.33	-266.07	-4.97	-1.58	-212.87	-1.58	-0.11

CASE(3): IMPR IRR EFF; CHANGES FROM BASE PERIOD
 ARKANSAS RIVER VALLEY STUDY, DATA SUMMARY FOR 1973

	ENET (FT)	FTOT (FT)	QTOT (FT)	AQ. CHEM. (MG/L)	AQ. W-L (FT)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT (TONS)	RIVER TRANS (CFS)	HBAR (FT)
MAR	0.013	0.026	0.075	2516.79	4011.48	1596.48	47.54	6348.71	4.44	17.25
APR	-0.038	-0.017	0.099	2545.03	4011.32	1581.29	98.51	12611.68	2.31	17.09
MAY	-0.040	-0.015	0.111	2600.01	4011.11	1542.86	139.48	18093.16	-0.72	16.88
JUN	-0.054	-0.006	0.163	2656.55	4010.90	1243.69	461.65	46482.31	-3.75	16.67
JUL	-0.072	-0.006	0.200	2697.41	4010.69	766.22	408.05	26155.66	-6.55	16.47
AUG	-0.079	-0.001	0.182	2769.41	4010.52	1262.68	146.58	15483.72	-9.02	16.29
SEP	-0.048	-0.027	0.141	2820.71	4010.45	1432.93	100.52	11061.08	-10.28	16.26
OCT	0.018	0.080	0.123	2826.04	4010.61	1401.34	99.55	11670.33	-7.65	16.16
NOV	0.054	0.080	0.114	2793.12	4010.94	1442.59	33.97	3966.74	-3.23	16.71
DEC	0.045	0.051	0.063	2770.41	4011.21	1894.24	31.89	5053.64	0.69	16.98
JAN	0.043	0.050	0.051	2754.64	4011.38	2103.27	30.29	5326.98	3.08	17.15
FEB	0.035	0.043	0.046	2735.44	4011.48	2157.55	27.09	4417.16	4.89	17.25
YEARLY AVERAGES:				2707.18	4011.01	1535.43	135.43	13931.76	-2.15	16.78

HYDRAULIC RESPONSE TIME = 2.8 MONTHS SOLUTE RESPONSE TIME = 10.7 YEARS

ARKANSAS RIVER VALLEY STUDY - CHANGES FROM BASE PERIOD FOR 1973

	AQ. CHEM. (MG/L)	AQ. CHEM. (TONS)	AQ. VOL (AC-FT)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT (TONS)	RIVER TRANS (CFS)	HBAR (FT)
MAR	25.89	1175.20	-81.75	-6.53	-0.48	-90.14	-0.48	-0.03
APR	31.39	1122.44	-182.00	-9.57	-1.10	-218.07	-1.10	-0.08
MAY	45.70	1228.29	-368.34	-10.11	-2.16	-398.21	-2.16	-0.15
JUN	45.29	1273.84	-332.64	0.06	-2.01	-249.58	-2.01	-0.14
JUL	39.26	1099.50	-277.38	0.07	-1.62	-101.60	-1.62	-0.12
AUG	42.50	937.73	-354.00	-0.46	-2.07	-224.69	-2.07	-0.15
SEP	47.99	1027.56	-398.75	1.69	-2.41	-265.81	-2.41	-0.17
OCT	45.45	1216.14	-324.25	0.92	-1.74	-214.74	-1.74	-0.14
NOV	43.00	1262.53	-287.00	7.13	-1.74	-182.19	-1.74	-0.12
DEC	43.22	1401.59	-265.50	-37.83	-1.55	-352.23	-1.55	-0.11
JAN	42.29	1655.69	-190.63	-17.51	-1.12	-242.31	-1.12	-0.08
FEB	41.60	1048.88	-132.25	-9.15	-0.86	-156.99	-0.86	-0.06
AVE CHANGE:	41.17	1280.56	-200.22	-6.77	-1.58	-220.80	-1.58	-0.11

CASE(3): IMPR IRR EFF; CHANGES FROM BASE PERIOD
 ARKANSAS RIVER VALLEY STUDY, DATA SUMMARY FOR 1974

	ENET (FT)	ETOT (FT)	WTOT (FT)	AQ. CHEM. (MG/L)	AQ. W-L (FT)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT (TONS)	RIVER TRANS (CFS)	HBAR (FT)
MAR	0.013	0.026	0.075	2728.70	4011.48	1616.24	47.53	6426.88	4.43	17.25
APR	-0.038	-0.017	0.099	2756.17	4011.32	1586.24	98.51	12651.05	2.31	16.88
MAY	-0.040	-0.015	0.111	2812.15	4011.41	1542.86	139.48	14043.10	-0.75	16.67
JUN	-0.064	-0.006	0.163	2869.36	4010.90	1243.69	461.65	46482.27	-3.55	16.46
JUL	-0.072	-0.006	0.200	2910.62	4010.69	766.22	408.05	28155.63	-0.22	16.29
AUG	-0.079	-0.001	0.182	2982.84	4010.52	1262.68	148.58	15493.67	-0.02	16.22
SEP	-0.048	0.027	0.141	3033.78	4010.45	1432.93	100.52	11661.03	-10.28	16.39
OCT	0.018	0.080	0.123	3035.78	4010.61	1401.34	99.55	11670.28	-7.65	16.71
NOV	0.054	0.080	0.114	2997.31	4010.94	1442.39	33.96	3966.69	-3.24	16.98
DEC	0.045	0.051	0.063	2970.68	4011.21	1898.55	31.89	5085.12	0.69	17.15
JAN	0.043	0.050	0.051	2951.98	4011.38	2123.70	26.27	5327.64	3.07	17.25
FEB	0.035	0.043	0.046	2930.13	4011.48	2192.71	27.09	4489.06	4.89	17.25
YEARLY AVERAGES:				2914.88	4011.01	1542.44	135.43	13952.70	-2.15	16.78

HYDRAULIC RESPONSE TIME = 2.8 MONTHS SOLUTE RESPONSE TIME = 10.7 YEARS

ARKANSAS RIVER VALLEY STUDY; CHANGES FROM BASE PERIOD FOR 1974

	AQ. CHEM. (MG/L)	AQ. CHEM. (TONS)	AQ. VOL (AC-FT)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT (TONS)	RIVER TRANS (CFS)	HBAR (FT)
MAR	39.30	1905.49	-81.75	-7.07	-0.48	-93.10	-0.48	-0.03
APR	45.43	1853.96	-181.88	-11.37	-1.10	-233.07	-1.10	-0.08
MAY	61.09	1974.85	-368.38	-12.11	-2.16	-421.87	-2.16	-0.15
JUN	60.46	2028.32	-332.75	0.06	-2.01	-200.62	-2.01	-0.14
JUL	54.53	1841.75	-277.38	0.07	-1.62	-101.62	-1.62	-0.15
AUG	58.16	1671.62	-354.00	-0.48	-2.07	-224.73	-2.07	-0.17
SEP	63.82	1755.32	-398.75	0.69	-2.41	-265.86	-2.41	-0.15
OCT	61.07	1941.08	-324.38	0.92	-1.70	-214.79	-1.70	-0.12
NOV	57.61	1981.97	-287.13	7.13	-1.74	-182.23	-1.74	-0.11
DEC	57.67	2132.09	-265.63	-45.98	-1.56	-375.67	-1.56	-0.11
JAN	56.55	2402.96	-190.75	-21.92	-1.12	-265.84	-1.12	-0.08
FEB	55.65	2604.11	-132.25	-11.16	-0.86	-165.59	-0.86	-0.08
AVE CHANGE:	55.98	2018.69	-266.15	-8.35	-1.59	-227.91	-1.59	-0.11

CASE(3): IMPR IRR EFF; CHANGES FROM BASE PERIOD

ARKANSAS RIVER VALLEY STUDY, DATA SUMMARY FOR 1975

	ENFT (FT)	ETOT (FT)	QTOT (FT)	AQ, CHEM. (MG/L)	AG, W-L (FT)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT RIVER (TONS)	RIVER TRANS (CFS)	HBAR (FT)
MAR	0.013	0.026	0.075	2920.96	4011.48	1634.16	47.53	6498.39	4.43	17.25
APR	-0.038	-0.017	0.099	2948.09	4011.32	1590.76	94.51	12687.12	2.31	17.09
MAY	-0.040	-0.015	0.111	3005.58	4011.11	1542.86	139.48	18083.10	-0.72	16.88
JUN	-0.064	-0.006	0.163	3063.40	4010.90	1243.69	461.65	4682.27	-3.75	16.67
JUL	-0.072	-0.006	0.200	3105.02	4010.69	1766.22	408.05	26155.63	-6.55	16.46
AUG	-0.079	-0.001	0.182	3177.44	4010.52	1262.68	146.58	15483.67	-9.02	16.29
SEP	-0.058	0.027	0.141	3227.10	4010.43	1432.93	199.52	11661.03	-10.28	16.22
OCT	0.018	0.080	0.123	3226.47	4010.61	1401.34	199.55	11670.28	-17.65	16.39
NOV	0.054	0.080	0.114	3183.50	4010.94	1442.59	31.96	3966.69	-3.24	16.71
DEC	0.045	0.051	0.063	3153.50	4011.21	1902.51	31.89	5025.68	0.69	16.98
JAN	0.043	0.050	0.051	3131.92	4011.38	2141.58	30.27	5423.92	3.83	17.15
FEB	0.035	0.043	0.046	3107.64	4011.48	2224.78	27.09	4554.71	4.89	17.25
YEARLY AVERAGES:				3104.25	4011.01	1548.84	135.43	13971.87	-2.15	16.78

364

HYDRAULIC RESPONSE TIME = 2.8 MONTHS SOLUTE RESPONSE TIME = 10.7 YEARS

ARKANSAS RIVER VALLEY STUDY: CHANGES FROM BASE PERIOD FOR 1975

	AQ, CHEM. (MG/L)	AQ, CHEM. (TONS)	AG, VOL (AC-FT)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT RIVER (TONS)	RIVER TRANS (CFS)	HBAR (FT)
MAR	53.16	2662.80	-81.75	-7.39	-0.88	-95.12	-0.48	-0.03
APR	59.85	2611.37	-181.88	-12.95	-1.10	-246.17	-1.10	-0.08
MAY	76.75	2745.13	-368.38	-13.91	-2.16	-443.27	-2.16	-0.15
JUN	76.68	2805.15	-337.75	0.06	-2.01	-200.62	-2.01	-0.14
JUL	70.08	2606.77	-277.38	0.07	-1.62	-101.62	-1.62	-0.12
AUG	74.06	2428.03	-354.00	-0.46	-2.07	-224.73	-2.07	-0.15
SEP	79.87	2505.67	-394.75	1.69	-2.41	-265.86	-2.41	-0.17
OCT	76.54	2688.19	-324.38	0.94	-1.90	-214.73	-1.90	-0.14
NOV	72.49	2723.75	-287.13	7.13	-1.74	-182.23	-1.74	-0.12
DEC	72.38	2883.60	-265.63	-5.33	-1.56	-396.67	-1.56	-0.11
JAN	71.05	3169.19	-190.75	-25.73	-1.12	-267.56	-1.12	-0.08
FEB	69.95	3376.62	-132.25	-12.68	-0.86	-170.88	-0.86	-0.06
AVERAGE:	71.07	2779.11	-266.15	-9.71	-1.59	-234.13	-1.59	-0.11

CASE(4): QLE0; CHANGES FROM BASE PERIOD
 ARKANSAS RIVER VALLEY STUDY, DATA SUMMARY FOR 1971

	ENET (ft)	ETOT (ft)	QTOT (ft)	AG, CHEM. (MG/L)	AG, W-L (ft)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT (TONS)	RIVER TRANS (CFS)	MBAK (ft)
MAR	-0.036	-0.024	.034	2024.30	4011.80	1591.65	52.06	6931.28	8.96	17.57
APR	-0.078	-0.058	.043	2062.19	4011.36	1572.72	99.07	12614.60	2.87	17.13
MAY	-0.063	-0.047	.065	2117.14	4010.56	1543.48	137.40	17739.92	-2.80	16.73
JUN	-0.137	-0.053	.141	2187.78	4010.55	1243.83	456.56	45974.56	-8.84	16.12
JUL	-0.134	-0.031	.184	2251.43	4010.16	1766.36	400.97	25693.69	-13.83	15.94
AUG	-0.141	-0.018	.161	2336.68	4009.88	1260.54	137.76	14526.82	-17.84	15.66
SEP	-0.117	0.010	.114	2399.02	4009.73	1441.20	190.15	10518.98	-20.65	15.51
OCT	-0.052	0.066	.116	2418.66	4009.82	1407.53	88.45	10415.18	-18.75	15.59
NOV	-0.013	0.096	.127	2490.30	4010.15	1515.81	22.65	2779.64	-14.55	15.93
DEC	-0.005	0.051	.065	2393.64	4010.47	1938.64	21.56	3496.90	-9.64	16.24
JAN	-0.007	0.036	.036	2300.84	4010.65	2085.13	20.12	3509.21	-7.08	16.43
FEB	-0.012	0.026	.028	2386.57	4010.77	2100.66	16.31	2680.71	-5.89	16.54
YEARLY AVERAGES:										
				2280.73	4010.53	1538.97	128.56	13073.46	-9.01	16.30

HYDRAULIC RESPONSE TIME = 2.8 MONTHS SOLUTE RESPONSE TIME = 59.9 YEARS

ARKANSAS RIVER VALLEY STUDY: CHANGES FROM BASE PERIOD FOR 1971

	AG, CHEM. (MG/L)	AG, CHEM. (TONS)	AG, VOL (AC-FT)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT (TONS)	RIVER TRANS (CFS)	MBAK (ft)
MAR	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
APR	13.40	-844.95	-571.00	-15.66	-3.45	-569.83	-3.45	-0.24
MAY	27.43	-1522.98	-1060.63	-12.30	-6.21	-949.46	-6.21	-0.44
JUN	41.59	-1907.45	-1404.13	0.24	-8.53	-849.67	-8.53	-0.59
JUL	57.58	-2095.73	-1701.75	0.45	-9.96	-623.47	-9.96	-0.71
AUG	72.66	-2385.65	-1975.63	-2.75	-11.57	-1254.04	-11.57	-0.83
SEP	86.06	-2557.85	-2191.63	10.28	-13.26	-1461.00	-13.26	-0.92
OCT	90.47	-2615.71	-2275.23	17.27	-13.52	-1506.69	-13.52	-0.95
NOV	92.17	-2122.18	-2195.88	81.26	-13.29	-1393.91	-13.29	-0.92
DEC	99.50	-1164.18	-2056.88	33.83	-12.04	-1858.00	-12.04	-0.86
JAN	105.15	-429.56	-1944.38	20.53	-11.38	-1931.64	-11.38	-0.81
FEB	113.87	406.63	-1849.00	21.12	-11.57	-1856.38	-11.57	-0.77
AVE CHANGE:								
	66.66	-1294.66	-1602.57	12.02	-9.55	-1187.84	-9.55	-0.67

CASE(4): QL=0; CHANGES FROM BASE PERIOD

ARKANSAS RIVER VALLEY STUDY, DATA SUMMARY FOR 1972

	ENET (FT)	ETOT (FT)	QTOT (FT)	AQ, CHEM. (MG/L)	AQ, W-L (FT)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT RIVER FLOW (TONS)	RIVER TRANS (CFS)	HBAR (FT)
MAR	-0.036	0.004	0.062	2393.05	4010.78	1516.35	37.78	4792.94	-5.32	16.55
APR	-0.078	-0.021	0.080	2428.52	4010.65	1562.20	88.79	11228.90	-7.41	16.42
MAY	-0.083	-0.012	0.100	2440.50	4010.40	1545.21	130.45	16863.01	-7.75	16.24
JUN	-0.137	-0.028	0.166	2545.98	4010.20	1243.98	451.55	45475.42	-13.85	15.98
JUL	-0.134	-0.013	0.202	2502.96	4009.92	1766.71	397.92	25482.11	-17.32	15.70
AUG	-0.141	-0.008	0.173	2685.83	4009.72	1259.92	135.40	14271.07	-20.20	15.49
SEP	-0.117	0.048	0.122	2744.17	4009.64	1442.74	88.45	10331.59	-22.35	15.39
OCT	-0.052	0.072	0.122	2756.50	4009.74	1408.26	87.30	10285.30	-16.90	15.51
NOV	0.005	0.100	0.131	2727.27	4010.10	1524.14	21.82	2692.79	-15.38	15.87
DEC	-0.005	0.054	0.068	2712.26	4010.43	1944.14	21.00	3415.91	-10.20	16.20
JAN	-0.007	0.038	0.038	2704.64	4010.63	2089.34	19.73	3448.15	-7.47	16.48
FEB	-0.011	0.027	0.030	2696.97	4010.75	2109.02	15.81	2519.36	-6.39	16.52
YEARLY AVERAGES:				2623.20	4010.25	1534.33	124.61	12567.21	-12.96	16.02

366

HYDRAULIC RESPONSE TIME = 2.8 MONTHS SOLUTE RESPONSE TIME = 13.7 YEARS

ARKANSAS RIVER VALLEY STUDY: CHANGES FROM BASE PERIOD FOR 1972

	AQ, CHEM. (MG/L)	AQ, CHEM. (TONS)	AQ, VOL (AC-FT)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT RIVER FLOW (TONS)	RIVER TRANS (CFS)	HBAR (FT)
MAR	121.48	1078.28	-1766.88	-65.83	-10.34	-1577.33	-10.34	-0.74
APR	133.24	1481.52	-1803.63	-21.76	-10.91	-1525.73	-10.91	-0.75
MAY	144.25	1511.18	-1920.50	-5.86	-11.24	-1522.96	-11.24	-0.80
JUN	153.22	1418.02	-2009.25	0.44	-12.16	-1211.47	-12.16	-0.84
JUL	164.00	1334.14	-2120.25	0.57	-12.41	-775.83	-12.41	-0.89
AUG	177.48	1201.80	-2267.63	-3.22	-13.28	-1439.38	-13.28	-0.95
SEP	189.25	1142.75	-2395.38	11.51	-14.49	-1596.80	-14.49	-1.00
OCT	190.16	1161.38	-2417.38	7.85	-14.15	-1600.83	-14.15	-1.01
NOV	186.94	1716.52	-2295.13	88.71	-13.89	-1456.88	-13.89	-0.96
DEC	190.74	2756.81	-2126.13	25.82	-12.45	-1952.31	-12.45	-0.89
JAN	194.93	3590.44	-1922.63	-4.43	-11.67	-2050.81	-11.67	-0.83
FEB	203.05	4551.10	-1873.50	-16.58	-12.14	-1970.36	-12.14	-0.78
AVE CHANGE:	170.75	1947.59	-2082.42	1.43	-12.43	-1559.31	-12.43	-0.87

CASE(4): QL=0; CHANGES FROM BASE PERIOD

ARKANSAS RIVER VALLEY STUDY, DATA SUMMARY FOR 1973

	ENET (FT)	FTOT (FT)	QTOT (FT)	AQ, CHEM. (MG/L)	AQ, W-L (FT)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT (TUNGS)	RIVER TRANS (CFS)	HBAR (FT)
MAR	-0.036	0.005	0.062	2701.60	4010.77	1516.89	37.61	4773.06	-5.49	16.54
APR	-0.078	-0.020	0.081	2737.60	4010.64	1562.87	88.66	11213.97	-7.54	16.41
MAY	-0.043	-0.011	0.011	2748.42	4010.46	1545.23	130.37	16852.55	-9.83	16.23
JUN	-0.137	-0.028	0.165	2858.18	4010.20	1243.98	451.49	45469.49	-13.91	15.97
JUL	-0.134	-0.013	0.202	2817.07	4009.92	1766.72	397.24	25479.57	-17.36	15.69
AUG	-0.141	-0.005	0.173	3000.41	4009.71	1259.92	135.37	14268.01	-20.23	15.49
SEP	-0.117	0.018	0.122	3058.45	4009.61	1442.76	88.43	10329.36	-22.37	15.34
OCT	-0.052	0.072	0.131	3066.19	4009.74	1408.27	87.24	10283.75	-19.51	15.51
NOV	0.013	0.100	0.131	3028.66	4010.10	1524.25	21.81	2691.77	-15.39	15.87
DEC	-0.005	0.054	0.068	3009.58	4010.43	1944.21	21.00	3414.98	-10.20	16.20
JAN	-0.007	0.038	0.038	2994.94	4010.63	2089.39	19.72	3447.42	-7.48	16.40
FEB	-0.011	0.027	0.030	2984.53	4010.75	2109.07	15.81	2518.90	-6.39	16.52
YEARLY AVERAGES:				2928.74	4010.25	1534.41	124.57	12561.90	-13.01	16.02

HYDRAULIC RESPONSE TIME = 2.8 MONTHS SOLUTE RESPONSE TIME = 13.5 YEARS

ARKANSAS RIVER VALLEY STUDY: CHANGES FROM BASE PERIOD FOR 1973

	AQ, CHEM. (MG/L)	AQ, CHEM. (TUNGS)	AQ, VOL (AC-FT)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT (TUNGS)	RIVER TRANS (CFS)	HBAR (FT)
MAR	210.70	5312.48	-1776.50	-86.11	-10.40	-1685.79	-10.40	-0.74
APR	223.96	5760.93	-1810.38	-28.59	-10.95	-1615.78	-10.95	-0.76
MAY	236.71	5802.27	-1925.25	-7.74	-11.27	-1518.82	-11.27	-0.81
JUN	246.92	5673.66	-2012.50	0.34	-12.18	-1213.40	-12.18	-0.84
JUL	258.92	5535.73	-2122.75	0.57	-12.43	-777.68	-12.43	-0.89
AUG	273.50	5255.43	-2269.13	-3.22	-13.29	-1440.40	-13.29	-1.00
SEP	285.74	5255.43	-2396.50	11.52	-14.50	-1597.53	-14.50	-1.00
OCT	285.40	5245.22	-2418.13	7.85	-14.18	-1601.31	-14.18	-1.01
NOV	278.54	5783.44	-2295.63	88.79	-13.89	-1457.16	-13.89	-0.96
DEC	279.40	6831.59	-2126.38	12.16	-12.45	-1990.89	-12.45	-0.89
JAN	282.50	7709.83	-1992.88	-31.39	-11.67	-2121.87	-11.67	-0.83
FEB	290.49	8737.08	-1873.75	-57.63	-12.15	-2057.26	-12.15	-0.78
AVE CHANGE:	262.73	6122.55	-2085.05	-7.79	-12.44	-1590.06	-12.44	-0.87

CASE(4): QL=0; CHANGES FROM BASE PERIOD

ARKANSAS RIVER VALLEY STUDY, DATA SUMMARY FOR 1974

	ENET (FT)	ETOT (FT)	QTOT (FT)	AQ (MG/L)	CHEM. (MG/L)	AG (MG/L)	W-L (FT)	RIVER FLOW (CFS)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT (TONS)	RIVER TRANS (CFS)	HBAR (FT)
MAR	-0.036	0.005	0.062	2987.58	4010.77	1516.89	37.61	4772.86	-5.49	16.54	11213.81	-7.54	16.41
APR	-0.078	-0.020	0.081	3024.24	4010.64	1562.27	88.66	16852.44	-9.83	16.23	16852.44	-9.83	16.23
MAY	-0.083	-0.041	0.101	3079.04	4010.46	1545.23	130.37	45469.44	-11.91	15.97	45469.44	-11.91	15.97
JUN	-0.137	-0.028	0.166	3147.96	4010.20	1743.98	451.49	25479.57	-17.36	15.69	25479.57	-17.36	15.69
JUL	-0.134	-0.013	0.202	3208.76	4009.92	1766.72	397.27	14268.01	-20.23	15.49	14268.01	-20.23	15.49
AUG	-0.141	-0.005	0.173	3292.97	4009.71	1259.92	135.37	10329.36	-22.37	15.39	10329.36	-22.37	15.39
SEP	-0.117	0.018	0.122	3350.40	4009.51	1442.79	88.43	10283.75	-19.91	15.51	10283.75	-19.91	15.51
OCT	-0.052	0.072	0.131	3353.90	4010.10	1408.27	21.81	2691.77	-15.39	15.87	2691.77	-15.39	15.87
NOV	-0.005	0.100	0.131	3308.70	4010.43	1524.25	21.00	3414.98	-10.20	16.20	3414.98	-10.20	16.20
DEC	-0.005	0.054	0.068	3280.05	4010.43	1944.21	19.72	1447.42	-7.48	16.40	1447.42	-7.48	16.40
JAN	-0.007	0.038	0.038	3264.50	4010.43	2009.59	15.81	2518.90	-6.39	16.52	2518.90	-6.39	16.52
FEB	-0.011	0.027	0.030	3251.55	4010.75	2109.07	15.81	12561.86	-13.01	16.02	12561.86	-13.01	16.02

368

YEARLY AVERAGES:

YEARLY AVERAGES:	3212.45	4010.25	1534.41	124.57	12561.86	-13.01	16.02
------------------	---------	---------	---------	--------	----------	--------	-------

HYDRAULIC RESPONSE TIME * 2.8 MONTHS SOLUTE RESPONSE TIME * 13.5 YEARS

ARKANSAS RIVER VALLEY STUDY CHANGES FROM BASE PERIOD FOR 1974

	AQ (MG/L)	AQ (TONS)	AQ CHEM. (TONS)	AG VOL (AC-FT)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT (TONS)	RIVER TRANS (CFS)	HBAR (FT)
MAR	298.69	9565.88	-1776.63	-106.35	-10.40	-1747.11	-10.40	-0.74	
APR	313.50	10053.43	-1810.38	-35.34	-10.95	-1670.25	-10.95	-0.76	
MAY	327.98	10103.50	-1925.23	-9.74	-11.27	-1572.53	-11.27	-0.81	
JUN	349.46	9999.07	-2012.50	0.34	-12.18	-1213.45	-12.18	-0.84	
JUL	352.66	9747.58	-2122.75	0.57	-12.43	-777.68	-12.43	-0.89	
AUG	368.29	9517.91	-2269.13	3.22	-13.29	-1440.40	-13.29	-0.95	
SEP	380.94	9377.25	-2395.50	11.52	-14.50	-1597.53	-14.50	-1.00	
OCT	370.10	9437.76	-2418.13	11.85	-14.50	-1601.31	-14.50	-1.01	
NOV	368.99	9859.56	-2295.53	88.79	-13.89	-1457.16	-13.89	-0.96	
DEC	367.04	10912.20	-1926.38	0.32	-12.45	-2025.81	-12.45	-0.89	
JAN	369.07	11829.95	-1992.88	-55.83	-11.67	-2186.06	-11.67	-0.83	
FEB	376.87	12917.88	-1873.75	-94.80	-12.15	-2135.76	-12.15	-0.78	
AVE CHANGE:	353.56	10306.24	-2084.90	-16.38	-12.44	-1618.75	-12.44	-0.87	

CASE(4): QL=0; CHANGES FROM BASE PERIOD
 ARKANSAS RIVER VALLEY STUDY, DATA SUMMARY FOR 1975

	ENRT (FT)	ETOT (FT)	QTOT (FT)	AQ, CHEM. (MG/L)	AQ, W-L (FT)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT (TONS)	RIVER TRANS (CFS)	HBAR (FT)
MAR	-0.036	0.005	0.062	3253.33	4010.77	1516.89	37.61	4772.86	-5.49	16.54
APR	-0.078	-0.020	0.081	3290.61	4010.64	1562.27	88.66	11213.81	-7.54	16.41
MAY	-0.083	-0.011	0.101	3346.68	4010.46	1545.23	130.37	16852.44	-9.83	16.23
JUN	-0.137	-0.028	0.166	3417.25	4010.20	1243.98	451.49	45469.44	-17.91	15.97
JUL	-0.134	-0.013	0.202	3479.61	4009.92	786.72	397.24	25479.57	-17.36	15.69
AUG	-0.141	-0.005	0.173	3564.84	4009.71	1259.92	135.37	14268.01	-20.23	15.49
SEP	-0.117	0.018	0.122	3621.70	4009.61	1442.76	88.43	10329.36	-22.37	15.39
OCT	-0.052	0.072	0.122	3621.70	4009.74	1408.27	87.29	10283.75	-19.91	15.51
NOV	-0.013	0.100	0.131	3568.93	4010.10	1524.25	21.81	2691.77	-15.39	15.87
DEC	-0.005	0.054	0.068	3534.19	4010.43	1944.21	21.00	3414.98	-10.20	16.20
JAN	-0.007	0.034	0.038	3515.09	4010.63	2089.53	19.72	2447.42	-7.48	16.40
FEB	-0.011	0.027	0.030	3499.49	4010.75	2109.07	15.81	2518.90	-6.39	16.52
YEARLY AVERAGES:										
				3476.10	4010.25	1534.41	124.57	12561.86	-13.01	16.02

HYDRAULIC RESPONSE TIME = 2.8 MONTHS SOLUTE RESPONSE TIME = 13.5 YEARS

ARKANSAS RIVER VALLEY STUDY: CHANGES FROM BASE PERIOD FOR 1975

	AQ, CHEM. (MG/L)	AQ, CHEM. (TONS)	AQ, VOL (AC-FT)	RIVER CHEM. (MG/L)	RIVER FLOW (CFS)	SALT (TONS)	RIVER TRANS (CFS)	HBAR (FT)
MAR	385.53	13804.64	-1776.63	-124.65	-10.40	-1820.65	-10.40	-0.74
APR	401.77	14325.47	-1810.38	-41.44	-10.95	-1719.48	-10.95	-0.76
MAY	417.85	14381.03	-1925.25	-11.55	-11.27	-1593.94	-11.27	-0.81
JUN	430.52	14190.41	-2012.50	0.34	-12.18	-1213.45	-12.18	-0.84
JUL	444.87	13936.06	-2122.75	0.57	-12.43	-777.68	-12.43	-0.89
AUG	461.46	13658.26	-2269.13	-3.22	-13.29	-1440.40	-13.29	-0.95
SEP	474.47	13477.78	-2396.50	11.52	-14.50	-1597.53	-14.50	-1.00
OCT	471.34	13406.45	-2418.13	17.85	-14.16	-1501.31	-14.16	-1.01
NOV	457.93	13913.41	-2295.63	88.79	-13.89	-1457.16	-13.89	-0.96
DEC	453.27	14969.87	-2126.38	-11.60	-12.45	-2057.36	-12.45	-0.89
JAN	454.23	15927.57	-1992.88	-77.92	-11.67	-2244.06	-11.67	-0.83
FEB	461.79	17063.70	-1873.75	-28.38	-12.15	-2206.69	-12.15	-0.78
AVE CHANGE:								
	442.92	14466.58	-2084.90	-24.14	-12.44	-1644.14	-12.44	-0.87

Non-linear Reservoir Model Program for the Mesilla Valley with
Output Results

```

*****
PROGRAM TO ESTIMATE MONTHLY VALUES OF
RECHARGE, PREDICT WATER LEVELS AND
DRAIN FLOWS USING A NON-LINEAR
LUMPED-PARAMETER MODEL.
*****
*****
INTEGER DATE
DIMENSION ABSE(32,14),SMSE(32,14),ABSEX(32),SMSEX(32),RMEX(32)
< I(COUNT(32,14)
DIMENSION C(14),D(32,14),E(32,14),F(32,14),G(32,14)
COMMON DATE(32,14),H(32,14),DRAIN(32,14),GR(32),RECH(32,14),
< HPRD(32,14),DRED(32,14)
DATA IR,IV,25,26/
OPEN(UNIT=25,DEVICE='DISK',ACCESS='SEQUIN',FILE='OBSV.DAT')
OPEN(UNIT=26,DEVICE='DISK',ACCESS='SEQUIN',FILE='APNZ.DAT')
OPEN(UNIT=27,DEVICE='DISK',ACCESS='SEQUIN',FILE='AMES.DAT')
*****
*****
MODEL PARAMETERS
*****
*****
DEFINITION OF VARIABLES:
ALPHA:CONSTANT FOR THE NON-LINEAR TERM
HO:CONSTANT FOR THE NON-LINEAR TERM
NREC: # OF RECHARGE-RECESSION PERIODS
POR: POROSITY OF THE AQUIFER
Z:CONSTANT FOR THE NON-LINEAR TERM
ALPHA=0.0755
HO=3825.0
NREC=31
PUR=0.218
Z=9.05
WRITE(1M,6)
WRITE(1M,16)
*****
*****
READ THE DATA
*****
*****
DEFINITION OF VARIABLES:
J: PERIOD NUMBER
I: MONTHLY VALUES (12 IN EACH J)
DATE(J,I):DATE OF WATER LEVEL AND DRAIN FLOW OBSERVATION
HC(J,I): WATER LEVEL OBSERVATION
DRAIN(J,I):DRAIN FLOW OBSERVATION
GR(J): DRAIN-RIVER LEAKAGE RECHARGE TERM
DO 2 J=1,NREC
READ(IR,15) GR(J)
WRITE(1W,26) J,GR(J)
DO 1 I=1,12
READ(IR,15) DATE(J,I),H(J,I),DRAIN(J,I)
G(J,I)=(1./Z)*ALOG(EXP(ALPHA*Z*(H(J,I)-HO))+1.0)

```

0000000000

0000000000

0000000000

```

00100
00200
00300
00400
00500
00600
00700
00800
00900
01000
01100
01200
01300
01400
01500
01600
01700
01800
01900
02000
02100
02200
02300
02400
02500
02600
02700
02800
02900
03000
03100
03200
03300
03400
03500
03600
03700
03800
03900
04000
04100
04200
04300
04400
04500
04600
04700
04800
04900
05000
05100
05200
05300
05400
05500
05600
05700
05800
05900
06000

```

WRITE(IW,36)DATE(J,I),H(J,I),DRAIN(J,I),G(J,I)
CONTINUE
CONTINUE

***** SOLUTION OF THE TRIANGULAR MATRIX TO ESTIMATE RECHARGE *****

DEFINITION OF VARIABLES:
A: CONSTANT COEFFICIENT
B: CONSTANT COEFFICIENT
C(I): COEFFICIENT
D(J,I): SOLUTION OF THE MODEL BY SIMPSON'S RULE
E(J,I): COEFFICIENT
F(J,I): COEFFICIENT

N: # OF MONTHS OF POSSIBLE RECHARGE
DRIN: INITIAL DRAIN FLOW VALUE (FEB. 1946)
HIN: INITIAL WATER LEVEL VALUE (FEB. 1946)
GIN: INITIAL VALUE OF THE NON-LINEAR TERM (FEB. 1946)
RECH(J,I): RECHARGE MONTHLY ESTIMATES

A=1.0
B=4.0
DRIN=0.0898
HIN=3826.12
GIN=(1./Z)*ALOG(EXP(ALPHA*Z*(HIN-H0))+1.0)
DO 8 J=1,NREC
WRITE(IW,46) J
D(J,1)=3.0*POR*(H(J,2)-HIN)-6.0*QR(J)+G(J,2)+4.0*G(J,1)+GIN
F(J,1)=C(1)/B
F(J,1)=D(J,1)/B
WRITE(IW,56)
WRITE(IW,66) D(J,1),E(J,1),F(J,1)
DRIN=DRAIN(J,12)
HIN=H(J,12)
GIN=(1./Z)*ALOG(EXP(ALPHA*Z*(HIN-H0))+1.0)
N=7

DO 3 I=2,N
C(I)=1.0
C(N)=0.0
D(J,I)=3.0*POR*(H(J,I+1)-H(J,I-1))-6.0*QR(J)+G(J,I+1)+
#4.0*G(J,I-1)+G(J,I-1)
E(J,I)=C(I)/(B-E(J,I-1)*A)
F(J,I)=D(J,I)-A*F(J,I-1)/(B-A*E(J,I-1))
WRITE(IW,76) D(J,I),E(J,I),F(J,I)
CONTINUE

RECH(J,N)=F(J,N)
RECH(J,N-1)=F(J,N-1)*RECH(J,N)
N=N-1
IF(N EQ. 1)GO TO 8
CU TO 4

CONTINUE
DO 12 J=1,NREC
WRITE(IW,R6) J
DO 11 I=1,12
WRITE(IW,106) DATE(J,I),RECH(J,I)

06100
06200
06300
06400
06500
06600
06700
06800
06900
07000
07100
07200
07300
07400
07500
07600
07700
07800
07900
08000
08100
08200
08300
08400
08500
08600
08700
08800
08900
09000
09100
09200
09300
09400
09500
09600
09700
09800
09900
10000
10100
10200
10300
10400
10500
10600
10700
10800
10900
11000
11100
11200
11300
11400
11500
11600
11700
11800
11900

1 2

CCCCCCCCCCCCCCCCCCCC

3

4

8

11
12

CONTINUE
CONTINUE

SIMULATION

AN ITERATIVE PROCESS TO PREDICT WATER LEVELS AND DRAIN FLOWS

DEFINITION OF VARIABLES:

AVREC=AVERAGE OF TWO ADJACENT RECHARGE VALUES
HBEG=BEGINNING ESTIMATION OF WATER LEVEL
HNEN=NEW ESTIMATION OF WATER LEVEL
HOLD=OLD ESTIMATION OF WATER LEVEL
GBEG=NON-LINEAR FCN. EVALUATED AT HBEG
GOLD=NON-LINEAR FCN. EVALUATED AT HOLD
AVG=AVERAGE OF TWO ADJACENT NON-LINEAR FCN VALUES
DIFF=ABSOLUTE ERROR BETWEEN A NEW AND OLD ESTIMATE
TOL=TOLERANCE IN ITERATION
HPRED(J,I):PREDICTED WATER LEVEL
DPRED(J,I):PREDICTED DRAIN FLOW

TOL=0.01
HPRED(1,1)=3826.38
DO 23 J=1,NREC
DO 22 I=1,12
ICOUNT(J,I)=0
DRED(J,I)=(1.0/Z)*ALOG(EXP(ALPHA*Z*(HPRED(J,I)-H0))+1.0)
IF(J.GT.1 AND.1.GT.I) GO TO 13
HOLD=H(I)
HBEG=HOLD
GO TO 14

13
14
18

HBEG=HNEN
GO TO 14
GO TO 14
HOLD=HNEN

GREG=(1.0/Z)*ALOG(EXP(ALPHA*Z*(HBEG-H0))+1.0)
GOLD=(1.0/Z)*ALOG(EXP(ALPHA*Z*(HOLD-H0))+1.0)
AVG=(GOLD+GREG)/2.0
ICOUNT(J,I)=ICOUNT(J,I)+1
IF(I.EQ.12) GO TO 19
AVREC=(GREG(J,I)+GREG(J,I))/2.0
GO TO 20

AVREC=(GREG(J,I)+GREG(J,I))/2.0
HNEN=HBEG+(AVREC+GR(J)-AVG)/POR
DIFF=ABS(HNEW-HOLD)
IF(DIFF.GT.TOL) GO TO 14
IF(I.LT.12) GO TO 21
HPRED(J,I)=HNEN
GO TO 22

19
20
21
22
23

AVREC=(GREG(J,I)+GREG(J,I))/2.0
HNEN=HBEG+(AVREC+GR(J)-AVG)/POR
DIFF=ABS(HNEW-HOLD)
IF(DIFF.GT.TOL) GO TO 14
IF(I.LT.12) GO TO 21
HPRED(J,I)=HNEN
GO TO 22
HPRED(J,I+1)=HNEN
CONTINUE
CONTINUE

COMPUTATION OF ERRORS IN SIMULATION

DEFINITION OF VARIABLES:

COMPUTATION OF ERRORS IN SIMULATION

12000
12100
12200
12300
12400
12500
12600
12700
12800
12900
13000
13100
13200
13300
13400
13500
13600
13700
13800
13900
14000
14100
14200
14300
14400
14500
14600
14700
14800
14900
15000
15100
15200
15300
15400
15500
15600
15700
15800
15900
16000
16100
16200
16300
16400
16500
16600
16700
16800
16900
17000
17100
17200
17300
17400
17500
17600
17700
17800
17900

CCCCCCCCCCCCCCCC

CCCCCCCC


```

ARSE(J,I) ABSOLUTE ERROR BETWEEN EACH PRED. & OBS. VALUE
SMSE(J,I) MEAN SQUARE ERROR BETWEEN EACH PRED. & OBS. VALUE
ABSEX(J,I) ABSOLUTE ERROR FOR J PERIOD
SMSEX(J,I) MEAN SQUARE ERROR FOR J PERIOD
RMEX(J,I) ROOT MEAN SQUARE ERROR FOR J PERIOD
SABS: ABSOLUTE ERROR FOR SIMULATION
SSQ: MEAN SQUARE ERROR FOR THE SIMULATION
RMS: ROOT MEAN SQUARE ERROR FOR THE SIMULATION

```

CCCCCCCC

```

18000
18100
18200
18300
18400
18500
18600
18700
18800
18900
19000
19100
19200
19300
19400
19500
19600
19700
19800
19900
20000
20100
20200
20300
20400
20500
20600
20700
20800
20900
21000
21100
21200
21300
21400
21500
21600
21700
21800
21900
22000
22100
22200
22300
22400
22500
22600
22700
22800
22900
23000
23100
23200
23300
23400
23500
23600
23700
23800

KOUNT=0
NPIS=IREC*12
SABS=0.0
SSQ=0.0
WRITE(IW,116)
IF(KOUNT.EQ.1) GO TO 25
WRITE(IW,126)
GO TO 255
25 WRITE(IW,136)
DO 32 J=1,NREC
ABSEX(J)=0.0
SMSEX(J)=0.0
DO 31 I=1,12
IF(KOUNT.EQ.1) GO TO 27
ARSE(J,I)=HPRED(J,I)-H(J,I)
GO TO 28
27 ABSE(J,I)=DRED(J,I)-DRAIN(J,I)
SMSE(J,I)=ABSE(J,I)+ARSE(J,I)
IF(KOUNT.EQ.1) GO TO 29
WRITE(IW,146) DATE(J,I),HPRED(J,I),H(J,I),ABSE(J,I),SMSE(J,I)
ACOUNT(J,I)
GO TO 30
29 WRITE(IW,156) DATE(J,I),DRED(J,I),DRAIN(J,I),ABSE(J,I),SMSE(J,I)
ABSEX(J)=ABSEX(J)+ABSE(J,I)
SMSEX(J)=SMSEX(J)+SMSE(J,I)
CONTINUE
SABS=SABS+ABSEX(J)
SSQ=SSQ+SMSEX(J)
RMEX(J)=SQRT(SMSEX(J)/12.0)
WRITE(IW,166) J,ABSEX(J),SMSEX(J),RMEX(J)
CONTINUE
AMPTS=NPIS
RMS=SQRT(SSQ/AMPTS)
WRITE(IW,176) SABS,SSQ,RMS
KOUNT=KOUNT+1
IF(KOUNT.EQ.1) GO TO 24
CALL MASHAL(IR,IW,NREC,ALPHA,PUR)
CLOSE UNIT=25,DEVICE='DSK',ACCESS='SEQOUT',FILE='APN2.DAT')
CLOSE UNIT=26,DEVICE='DSK',ACCESS='SEQOUT',FILE='AMES.DAT')
CLOSE UNIT=27,DEVICE='DSK',ACCESS='SEQOUT',FILE='AMES.DAT')
STOP
FORMAT(14,F10.2,F10.4)
FORMAT(45X,DATA)
FORMAT(14,F10.2,F10.2,F10.4,F10.4)
FORMAT(2X,F10.4)
FORMAT(20X,DATE,AX,WATER LEVEL',3X,'DRAIN FLOW',3X,
#NOM-U NEAR FCN,/)
FORMAT(//,20X,12,2A,PERIOD',2X,GR=',F10.4)
FORMAT(20X,14,3X,F10.2,3X,F10.4,3X,F10.4)
FORMAT(//,5X,TRIDIAGONAL MATRIX COEFFS FOR THE',1X,I2,1X,

```

```

23900 * PERIOD'/'
24000 FORMAT(5X,'D',13X,'E',13X,'F'/'
24100 FORMAT(3(3X,F10.2))
24200 FORMAT(3(3X,F10.2))
24300 FORMAT(20X,DATE,3X,RECHARGE ESTIMATES FOR THE,2X,12,1X,'TH PERIOD'/'
24400 FORMAT(20X,DATE,3X,RECHARGE ESTIMATES FOR THE,2X,12,1X,'TH PERIOD'/'
24500 FORMAT(20X,14,3X,F10.4)
24600 FORMAT(1H1,5X,'MODEL PREDICTIONS AND COMPUTATION OF ERRORS'/'
24700 FORMAT(2X,DATE,5X,'PRED W.L.',6X,'OBS. W.L.',4X,'ABS. ERROR',
24800 #4X,'SQ OF ABS. ERR.'/'
24900 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
25000 #ABS. ERROR',4X,'SQ OF ABS. ERR.'/'
25100 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
25200 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
25300 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
25400 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
25500 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
25600 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
25700 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
25800 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
25900 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
26000 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
26100 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
26200 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
26300 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
26400 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
26500 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
26600 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
26700 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
26800 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
26900 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
27000 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
27100 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
27200 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
27300 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
27400 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
27500 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
27600 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
27700 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
27800 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
27900 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
28000 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
28100 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
28200 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
28300 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
28400 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
28500 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
28600 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
28700 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
28800 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
28900 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
29000 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
29100 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
29200 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
29300 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
29400 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
29500 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
29600 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
29700 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,
29800 #4X,'DATE',5X,'PRED. DRAIN',4X,'OBS. DRAIN',4X,

```

```

29900
30000
30100
30200
30300
30400
30500
30600
30700
30800
30900
31000
31100
31200
31300
31400
31500
31600
31700
31800
31900
32000
32100
32200
32300
32400
32500
32600
32700
32800
32900
33000
33100
33200
33300
33400
33500
33600
33700
33800
33900
34000
34100
34200
34300
34400
34500
34600
34700
34800
34900
35000
35100
35200
35300
35400
35500
35600
35700
35800

100 FORMAT(I4)
   READ(IR,101)((QS(J,I),I=1,12)
   READ(IR,101)((QRIN(J,I),I=1,12)
   READ(IR,101)((CSUR(J,I),I=1,12)
   READ(IR,101)((CHRS(J,I),I=1,12)
   READ(IR,101)((CONUSE(J,I),I=1,12)
   CONTINUE
101 FORMAT(8F10.0,/,4F10.0)
   CALL RIVLEA(QL,US,OP,CONUSE,QRIN,QLR,MON,IYR,AREA,IW)
   STUT1=0.0
   STUT2=0.0
   STUT3=0.0
   DO 3 J=1,10
     A1=0.0
     A2=0.0
     A3=0.0
     A4=0.0
     A5=0.0
     A6=0.0
     A7=0.0
     TC(J)=0.0
     ETOT=0.0
     DO 4 I=1,12
       TC(J)=TC(J)+WL(J,I)
       IF(QLR(J,I).LT.0.0)GO TO 44
       QLEAK=QLR(J,I)
       GU TO 43
44 QLEAK=0.0
43 TOTREC=ENET(J,I)+QLEAK/AREA
   ETOT=ETOT+TOTREC
   IF(J.EQ.1)AND(I.LT.5)GO TO 4
   AK1=TOTREC/(POR*(J,I))
   AK2=(QS(J,I)+QLEAK)/(AREA*POR*WL(J,I))
   IF(I.EQ.12)GO TO 5
   CAQ(J,I+1)=CSUR(J,I)*2.*AK2/(AK1+2.)-CAQ(J,I)*(AK1-2.)/(AK1+2.)
   GO TO 6
5 CAQ(J+1,I)=CSUR(J,I)*2.*AK2/(AK1+2.)-CAQ(J,I)*(AK1-2.)/(AK1+2.)
6 QROUT(J,I)=QD(J,I)+QRIN(J,I)-QOUT
   IF(QLR(J,I).LT.0.0)GO TO 45
   COU(J,I)=((QRIN(J,I)-QOUT)*CSUR(J,I)+QD(J,I)+CAQ(J,I))/QROUT(J,I)
   GO TO 4
45 COU(J,I)=((QRIN(J,I)-QS(J,I))*CSUR(J,I)+CAQ(J,I)*(QD(J,I)-QLR(J,I)
   <J)/QROUT(J,I)
4 CONTINUE
277 WRITE(27,277)((ENET(J,I),I=1,12)
   FORMAT(8F10.7,/,4F10.7)
278 WRITE(27,278)(QLR(J,I),I=1,12)
   FORMAT(8F10.2,/,4F10.2)
   TC(J)=POR*TC(J)/(12.0*ETOT)
200 WRITE(IW,200)IYR(J)
   FORMAT(IH,/,/,5X,'MPSILLA VALLEY, N.MEX;',HYDROSALINITY MODEL,RE
   <SULTS FOR THE WATER YEAR,15,/,5X,'MONTH,2X, AO HEAD,2X, DRA
   <IN FLOW,2X, LEAKS, GM,2X, LEAS, GM,2X, LEAS, CHEM,2X, EL DPA
   <SU GR,2X, EL PASO GR,2X, EL PA CHEM,2X, EL PA CHEM,2X, EL DPA
   <D; FT),2X, (PRD:AC-FT),1X, (OBS:AC-FT),2X, (PRD:MG/L),3X, (PRD:
   <G/L),2X, (PD:AC-FT),2X, (OBS:AC-FT),2X, (PRD:MG/L),2X, (OBS:MG/
   <L),/,)
   SD15G=0.0

```

```

35900 SD2SQ=0.0
36000 SD3SQ=0.0
36100 DO 7 I=1,12
36200 D1(I)=UROUT(J,I)-QOBS(J,I)
36300 D2(I)=COUT(J,I)-COBS(J,I)
36400 TUNOB(I)=QOBS(J,I)*COBS(J,I)/735.0
36500 TUMPR(I)=QROUT(J,I)*COUT(J,I)/735.0
36600 D3(I)=TUMPR(I)-TUNOB(I)
36700 SD1SQ=SD1SQ+D1(I)*D1(I)
36800 SD2SQ=SD2SQ+D2(I)*D2(I)
36900 SD3SQ=SD3SQ+D3(I)*D3(I)
37000 WLEVEL=WL(J,I)+HO
37100 HPD(J,I)=WLEVEL
37200 A1=A1+WLEVEL
37300 A2=A2+UD(J,I)
37400 A3=A3+CAQ(J,I)
37500 A4=A4+UROUT(J,I)
37600 A5=A5+QOBS(J,I)
37700 A6=A6+COUT(J,I)
37800 A7=A7+COBS(J,I)
37900 WRITE(IW,201)MGN(I),WLEVEL,QD(J,I),QRIN(J,I),CSUR(J,I),CAQ(J,I),
38000 < QROUT(J,I),QOBS(J,I),COUT(J,I),COBS(J,I)
38100 FORMAT(5X,A5,9(2X,F10.2))
38200 CONTINUE
38300 STOT1=STOT1+SD1SQ
38400 STOT2=STOT2+SD2SQ
38500 STOT3=STOT3+SD3SQ
38600 IF(J.EQ.1)GO TO 33
38700 RMS1=SQRT(SD1SQ/12.0)
38800 RMS2=SQRT(SD2SQ/12.0)
38900 RMS3=SQRT(SD3SQ/12.0)
39000 A1=A1/12.0
39100 A2=A2/12.0
39200 A3=A3/12.0
39300 A4=A4/12.0
39400 A5=A5/12.0
39500 A6=A6/12.0
39600 A7=A7/12.0
39700 GO TO 34
39800 RMS1=SQRT(SD1SQ/8.0)
39900 RMS2=SQRT(SD2SQ/8.0)
40000 RMS3=SQRT(SD3SQ/8.0)
40100 A1=A1/12.0
40200 A2=A2/12.0
40300 A3=A3/8.0
40400 A4=A4/8.0
40500 A5=A5/8.0
40600 A6=A6/8.0
40700 A7=A7/8.0
40800 CONTINUE
40900 WRITE(IW,501)A1,A2,A3,A4,A5,A6,A7
41000 FORMAT(/,5X, 'YEARLY AVERAGES: ',/,10X,2(2X,F10.2),24X,5(2X,F10.2))
41100 TH,TC(J)
41200 WRITE(IW,300)TH,TC(J)
41300 CONTINUE
41400 < RESPONSE TIME = F4.1, MO.,/,5X, 'SOLUTE
41500 < CIED AND OBSERVED',/,13X, 'RIVER FLOW & TDS AT EL PASO',21X, 'TONS
41600 < OF SALT AT EL PASO',/,5X, 'MONTH',5X, 'OPRED-QOBS',5X, 'CMPRED-CMOBS
41700 < (MG/L)',/,16X, 'DIFF',/,16X, '(AC-FT)',10X,
41800 DO 77 I=1,12

```

```

41900
42000
42100
42200
42300
42400
42500
42600
42700
42800
42900
43000
43100
43200
43300
43400
43500
43600
43700
43800
43900
44000
44100
44200
44300
44400
44500
44600
44700
44800
44900
45000
45100
45200
45300
45400
45500
45600
45700
45800
45900
46000
46100
46200
46300
46400
46500
46600
46700
46800
46900
47000
47100
47200
47300
47400
47500
47600
47700
47800

77 WRITE(IW,301)MON(I),D1(I),D2(I),TONPR(I),TUNUH(I),D3(I)
301 FORMAT(5X,A5,5X,F10.1,6X,F10.1,19X,F11.1,4X,F11.1,F10.2)
302 FURMAT(//,4X,RMS ERRORS: ,F10.1,6X,F10.1,44X,F10.1)
CONTINUE
TRMS1=SQRT(STOT1/119.0)
TRMS2=SQRT(STOT2/119.0)
TRMS3=SQRT(STOT3/119.0)
303 WRITE(IW,303)TRMS1,TRMS2,TRMS3
FORMAT(//,5X,TOTAL RMS ERRORS: ,F10.1,/,5X,RMS ERROR IN RIVER FLOW = ,
F10.1,/,5X,RMS ERROR IN TDS = ,F10.1,/,5X,RMS ERROR IN TONS =
F10.1)
CALL CRAFF(IR,IW)
RETURN
END
SUBROUTINE RIVLEA(QL,QS,QP,CONUSE,QRIN,QLR,MON,IYR,AREA,IW)
DIMENSION QL(10,12),QPIN(10,12),QGR(10,12),MUN(12),IYR(10),
PERC(10),QS(10,12),PEKS(10),QP(10,12),CONUSE(10,12)
BETA=0.50
FLEAK=1.0-0.45
DO 1 J=1,10
SRFLO=0.0
SRFLW=0.0
SOL=0.0
SQS=0.0
DO 2 I=1,12
OP(J,I)=CONUSE(J,I)/BETA-FLEAK*QS(J,I)
IF(QP(J,I).LT.0.0)QP(J,I)=0.0
IF(J.EQ.1.AND.I.LT.5)GO TO 2
SRFLO=SRFLO+QRIN(J,I)
SRFLW=SRFLW+QRIN(J,I)
SOL=SQS+AREA*OL(J,I)
SQS=SQS+QS(J,I)
2 CONTINUE
PERC(J)=SOL/SRFLW
PERS(J)=100.0*SQS/SRFLW
DO 3 I=1,12
QLR(J,I)=PERC(J)*(QPIN(J,I)-QS(J,I))
IF(QGR(I).AND.I.LT.5)QGR(J,I)=6.0
1 PERC(J)=100.0*PERC(J)
WRITE(IW,101)(IYR(J),J=1,10)
101 FORMAT(1H1,/,/,5X,MESILLA VALLEY, N MEX. ,/,5X,CALC. RIVER LE
AC-FT/MO),/,12X,10(3X,I4,3X))
DO 4 I=1,12
MUN(I)=(QGR(J,I),J=1,10)
4 WRITE(IW,102)MUN(I),QGR(J,I),J=1,10)
102 FORMAT(5X,A5,10F10.1)
WRITE(IW,103)(PERC(J),J=1,10)
103 FURMAT(//,5X,PERCENTAGE OF YEARLY RIVER LEAKAGE INTO ALLUVIAL AQU
IFER: ,/,13X,10(3X,F4.1,3X))
WRITE(IW,201)(IYR(J),J=1,10)
201 FURMAT(//,5X,ORS. SURFACE DIVERSION (AC-FT/MO),/,12X,
10(3X,I4,3X))
DO 5 I=1,12
MUN(I)=(QS(J,I),J=1,10)
5 WRITE(IW,102)MUN(I),(QS(J,I),J=1,10)
303 FURMAT(//,5X,PERCENTAGE OF YEARLY RIVER FLOW THAT IS DIVERTED: ,
/,13X,10(3X,F4.1,3X))
WRITE(IW,301)(IYR(J),J=1,10)
301 <UNSUMPTIVE USE (AC-FT/MO),/,12X,10(3X,I4,3X) C

```

```

47900 DO 6 I=1,12
48000 6 WRITE(IM,102)MON(I),(CONUSE(J,I),J=1,10)
48100 WRITE(IM,401)(IYR(J),J=1,10)
48200 401 FORMAT(//,/,5X,'CALC. GROUNDWATER PUMPAGE (AC-FT/MO)',/,/,12X,
48300 10(3X,14,3X))
48400 DO 7 I=1,12
48500 7 WRITE(IM,102)MON(I),(QP(J,I),J=1,10)
48600 RETURN
48700 END
48800 SUBROUTINE GRAFF(IR,1W)
48900 COMMON/GRAPH/CAO(11,12),HOB(10,12),HPD(10,12),QOOR(10,12),COUT(10
49000 1,12),GROUT(10,12),CUBS(10,12),QD(10,12),QOBS(10,12)
49100 DIMENSION X(122),Y1(122),Y2(122),LBX(35),LBI(35),LB2(35),LB3(35),
49200 LB4(35),LB5(35),LB6(35),LB7(35),LB8(35),LBY1(35),LBY2(35),LBY3(35)
49300 CALL INITIAL(1)
49400 KOUNT=1
49500 J1=1
49600 J2=5
49700 N=120
49800 READ(IR,101)(LBX(I),I=1,7)
49900 READ(IR,101)(LBI(I),I=1,7)
50000 READ(IR,101)(LB2(I),I=1,7)
50100 READ(IR,101)(LB3(I),I=1,7)
50200 READ(IR,101)(LB4(I),I=1,7)
50300 READ(IR,101)(LB5(I),I=1,7)
50400 READ(IR,101)(LB6(I),I=1,7)
50500 READ(IR,101)(LB7(I),I=1,7)
50600 READ(IR,101)(LB8(I),I=1,7)
50700 READ(IR,101)(LB9(I),I=1,7)
50800 101 XMIN=1967.0
50900 DX=1.0
51000 CONTINUE
51100 IN=5
51200 CALL RSTR(0)
51300 CALL RSTR(1)
51400 GO TO (1,44,3,2,55),KOUNT
51500 1 DO 12 J=1,12
51600 DO 12 I=1,12
51700 K=12*(J-1)+I
51800 AK=K
51900 X(K,0)=1967.0+(AK-0.5)/12.0
52000 Y1(K)=HPD(J,I)
52100 Y2(K)=HOB(J,I)
52200 DO 11 I=1,7
52300 LBY1(I)=LBI(I)
52400 LBY2(I)=LB2(I)
52500 LBY3(I)=LB3(I)
52600 YMIN=3622.0
52700 DY1=1.0
52800 GO TO 4
52900 2 DO 23 J=1,10
53000 IF(.5*GT.1)IN=1
53100 DO 23 I=IN,12
53200 K=12*(J-1)+I
53300 AK=K
53400 X(K,0)=1967.0+(AK-0.5)/12.0
53500 Y1(K,0)=GROUT(J,I)/1000.0
53600 23 Y2(K,0)=QOBS(J,I)/1000.0
53700 DO 22 I=1,7
53800 LBY1(I)=LB3(I)

```

```

53900
54000
54100
54200
54300
54400
54500
54600
54700
54800
54900
55000
55100
55200
55300
55400
55500
55600
55700
55800
55900
56000
56100
56200
56300
56400
56500
56600
56700
56800
56900
57000
57100
57200
57300
57400
57500
57600
57700
57800
57900
58000
58100
58200
58300
58400
58500
58600
58700
58800
58900
59000
59100
59200
59300
59400
59500
59600
59700
59800

22 LBY2(I)=LB4(I)
   Y1MIN=0.0
   GO TO 4
3   DO 34 J=1,10
   IF(J.GT.1)IN=1
   DO 34 I=IN,12
   K=12*(J-1)+1
   AK=K
   X(K-4)=1967.0+(AK-0.5)/12.0
   Y1(K-4)=COUT(J,I)/100.0
   Y2(K-4)=CURS(J,I)/100.0
34 DO 33 I=1,7
   LBY1(I)=LB5(I)
   LBY2(I)=LB6(I)
33 N=116
   Y1MIN=0.0
   DY1=4.0
   GO TO 4
44 DO 45 J=1,10
   DO 45 I=1,12
   K=12*(J-1)+1
   AK=K
   X(K)=1967.0+(AK-0.5)/12.0
   Y1(K)=COUT(J,I)/100.0
   Y2(K)=CURS(J,I)/100.0
45 DO 46 I=1,7
   LBY1(I)=LB2(I)
   LBY2(I)=LB8(I)
46 Y1MIN=0.0
   DY1=4.0
   GO TO 4
55 DO 56 J=1,10
   IF(J.GT.1)IN=1
   DO 56 I=IN,12
   K=12*(J-1)+1
   AK=K
   X(K-4)=1967.0+(AK-0.5)/12.0
   Y1(K-4)=CAQ(J,I)
   Y2(K-4)=CAQ(J,I)
56 DO 57 I=1,7
   LBY1(I)=LB9(I)
   LBY2(I)=LB9(I)
57 Y1MIN=1500.0
   DY1=50.0
   Y2MIN=Y1MIN
   DY2=DY1
4   DO 444 I=1,N
   X(I)=(X(I)-XMIN)/DX
   Y1(I)=(Y1(I)-Y1MIN)/DY1
   Y2(I)=(Y2(I)-Y2MIN)/DY2
444 WRITE(LW,500)KOUNT
500 FORMAT(1H1,/,/,/,5X,'PLOTTED VALUES FOR KOUNT = ',I4,/,/,5X,
   < 'X',10Y,1,10X,12,/,/)
   DO 502 I=1,N
502 WRITE(LW,503)I,X(I),Y1(I),Y2(I)
503 FORMAT(1X,13,3(X,F10.2))
59500 CALL PLOT(0.35,1.0,-3)
59600 CALL AXIS(0.0,0,LBX,-35,10.0,0,XMIN,DX,-1)
59700 CALL PLOT(0.0,7.0,3)
59800

```

```

50900 CALL PLOT(10,0,7,0,2)
60000 CALL PLOT(10,0,0,0,2)
60100 CALL PLOT(0,0,0,0,3)
60200 CALL AXIS(0,0,0,UB1,35,7,90,,Y1MIN,DY1,-1)
60300 CALL PLOT(0,0,0,3)
60400 CALL LINE(X,Y1,N,J1,1)
60500 CALL PLOT(10,0,0,3)
60600 CALL AXIS(10,0,0,UB2,-35,7,90,,Y2MIN,DY2,-1)
60700 CALL PLOT(0,0,0,3)
60800 CALL LINE(X,Y2,N,J2,1)
60900 CALL PLOT(15,-1,-3)
61000 KOUNT=KOUNT+1
61100 IF(KOUNT.LT.6)GO TO 5
61200 CALL RSTR(0)
61300 CALL RSTR(2)
61400 RETURN
61500 END

```


0.1294
0.1767
0.2027
0.2439
0.2443
0.2149
0.1682
0.1330
0.1177

0.1180
0.2188
0.2280
0.2755
0.2839
0.2226
0.1478
0.1282
0.1104
0.0775

3826.17
3827.01
3827.27
3827.43
3828.06
3828.11
3826.87
3826.52
3826.24
3826.02
3825.94

349
449
549
649
749
849
949
1049
1149
1249
150

0.1309
0.2008
0.2266
0.2395
0.1578
0.1262
0.1123

0.1384
0.2039
0.2289
0.2696
0.1676
0.1077
0.0658

3826.38
3827.59
3827.80
3827.99
3826.63
3826.33
3826.19
3825.83

5
550
550
750
850
950
1050
1150
151
151

0.1168
0.1299
0.1322
0.1327
0.1180
0.0652

0.0769
0.1277
0.1181
0.1388
0.1471
0.0714
0.0531
0.0460

3826.93
3826.99
3826.18
3826.23
3826.28
3826.82
3825.76
3824.69
3824.59

6
651
451
551
651
751
851
951
1051
1151
152

0.0417
0.0542
0.0652
0.0652
0.0949
0.0757
0.0658
0.0652

0.0309
0.0607
0.0770
0.0837
0.0832
0.0824
0.0689
0.0527
0.0471
0.0306

3824.86
3824.33
3824.56
3824.77
3825.45
3824.96
3824.70
3824.71
3824.50

7
352
452
552
652
752
852
952
1052
1152
153

DATA

NON-LINEAR FCN.

DATE WATER LEVEL DRAIN FLOW

PERIOD 0.0776
3826.38
3827.34
3827.53
3828.00
3828.36
3828.42
3827.68
3826.84
3826.52
3826.17
3825.95
3825.89

0.1268
0.2031
0.2308
0.2485
0.2855
0.3010
0.2441
0.1613
0.1254
0.1122
0.0977

0.1405
0.1970
0.2099
0.2399
0.2684
0.2188
0.1666
0.1291
0.1182
0.1152

PERIOD 0.0723

3826.18
3827.50
3827.92
3828.32
3826.84
3826.38
3825.18
3825.99

0.1827
0.1997
0.2098
0.2636
0.2179
0.1959
0.1013
0.0829
0.0718

0.1399
0.1872
0.2066
0.2261
0.2466
0.2144
0.1409
0.1138
0.1104

PERIOD 0.0699

3825.74
3827.21
3828.05
3827.24
3826.95
3826.25
3825.89

0.0743
0.1622
0.1940
0.2270
0.2477
0.1904
0.1334
0.1044
0.0821

0.1080
0.1631
0.1889
0.1995
0.2432
0.2594
0.2173
0.1516
0.1158

PERIOD 0.0833

3826.18
3827.50
3827.92
3828.32
3826.84
3826.38
3825.18
3825.99

0.1827
0.1997
0.2098
0.2636
0.2179
0.1959
0.1013
0.0829
0.0718

0.1399
0.1872
0.2066
0.2261
0.2466
0.2144
0.1409
0.1138
0.1104

1156
1256
1357
257

3819.59
3818.73
3818.59
3819.00

0.0016
0.0021
0.0023
0.0019

0.0014
0.0015
0.0018
0.0018

12 57
457
557
657
757
857
1057
1157
1258

PERIOD
QR= 13
3817.43
3817.53
3817.73
3817.72
3818.03
3819.03
3819.34
3819.84
3820.07

0.0483
0.0015
0.0013
0.0012
0.0037
0.0088
0.0144
0.0111
0.0063
0.0043

0.0010
0.0007
0.0006
0.0006
0.0019
0.0029
0.0031
0.0040
0.0037

13 58
458
558
658
758
858
1058
1158
1259

PERIOD
QR= 14
3821.44
3822.02
3822.37
3822.45
3824.10
3824.46
3824.46
3824.46
3823.96

0.0090
0.0055
0.0094
0.0080
0.0090
0.0053
0.0047
0.0059
0.0033
0.0032

0.0048
0.0038
0.0262
0.0408
0.0684
0.0740
0.0645
0.0581
0.0342

14 59
459
559
659
759
859
1059
1159
1260

PERIOD
QR= 15
3824.52
3825.02
3825.38
3825.80
3825.88
3825.18
3824.56
3824.42

0.0273
0.0560
0.0887
0.1197
0.1158
0.1157
0.0830
0.0730
0.0671
0.0428

0.0033
0.0773
0.0919
0.1127
0.1136
0.0826
0.0722
0.0612
0.0369

15 60
460
560
660

PERIOD
QR= 16
3824.82
3825.71
3825.07

0.0286
0.0534
0.1073
0.1139

0.0700
0.0997
0.1066
0.1242

8 53
453
553
653
753
853
1053
1153
1254

PERIOD
QR= 65
3824.65
3825.01
3824.79
3824.48
3824.72
3824.91
3824.46
3824.40
3824.33
3823.95

0.0327
0.0284
0.0798
0.0774
0.0766
0.1005
0.1163
0.0553
0.0440
0.0381
0.0339
0.0233

0.0642
0.0622
0.0687
0.0581
0.0652
0.0581
0.0562
0.0542
0.0439

9 54
454
554
654
754
854
1054
1154
1255

PERIOD
QR= 33
3823.33
3823.30
3823.13
3823.17
3823.20
3822.20
3822.13
3822.38
3822.51
3822.41

0.0322
0.0164
0.0394
0.0323
0.0307
0.0279
0.0184
0.0171
0.0131
0.0110
0.0073

0.0306
0.0272
0.0152
0.0145
0.0176
0.0173
0.0190
0.0174

10 55
455
555
655
755
855
1055
1155
1256

PERIOD
QR= 16
3821.16
3820.71
3820.16
3820.99
3820.94
3820.56
3820.77
3820.66

0.0249
0.0106
0.0193
0.0029
0.0089
0.0093
0.0112
0.0036
0.0032
0.0029

0.0088
0.0079
0.0057
0.0041
0.0040
0.0035
0.0045
0.0028
0.0060
0.0056

11 56
456
556
656
756
856
1056
1257

PERIOD
QR= 66
3820.66
3820.98
3820.50
3820.14
3820.14
3820.27
3820.37

0.0390
0.0014
0.0077
0.0058
0.0037
0.0050
0.0035
0.0029
0.0020

0.0028
0.0035
0.0025
0.0012
0.0010
0.0011
0.0012

0.0351
0.0233
0.0245
0.0115
0.0073
0.0075
0.0074
0.0087
0.0095
0.0096

0.0283
0.0314
0.0192
0.0133
0.0150
0.0127
0.0137
0.0076
0.0053
0.0049
0.0047
0.0034

3823.56
3822.47
3822.96
3822.54
3821.76
3821.08
3821.11
3821.34
3821.48
3821.49
3821.45

364
464
564
664
764
864
964
1064
1164
1265
165

0.0078
0.0041
0.0038
0.0059
0.0117
0.0154
0.0167
0.0161

0.0194
0.0097
0.0059
0.0151
0.0325
0.0460
0.0266
0.0202
0.0159
0.0174
0.0144

20 PERIOD GR= 8
3821.10
3820.96
3820.59
3820.59
3821.84
3822.23
3822.35
3822.45
3822.29
3822.32

0.0185
0.0383
0.0553
0.0718
0.1011
0.0812
0.0704
0.0575
0.0480

0.0103
0.0193
0.0554
0.0728
0.0918
0.1308
0.1062
0.0480
0.0457
0.0377
0.0288

21 PERIOD GR= 13
3822.53
3822.37
3822.37
3822.44
3822.59
3825.17
3825.83
3822.54
3822.44
3822.11

0.0598
0.0483
0.0425
0.0522
0.0679
0.0486
0.0442
0.0417
0.0350
0.0353

0.0175
0.0398
0.0508
0.0447
0.0472
0.0573
0.0631
0.0679
0.0486
0.0333
0.0350
0.0295

22 PERIOD GR= 52
3824.52
3824.45
3822.98
3822.89
3822.27
3822.41
3822.62
3822.41
3822.36
3822.87
3822.57

0.1395
0.1660
0.1455
0.1172
0.1987
0.0877
0.0770
0.0693

0.1700
0.2072
0.1730
0.1089
0.0793
0.0647
0.0589
0.0504

3826.36
3826.83
3826.47
3825.93
3825.28
3825.01
3824.80

0.0590
0.0877
0.0840
0.1002
0.1094
0.1182
0.0868
0.0766
0.0718
0.0662

0.0370
0.0565
0.0888
0.0967
0.1077
0.1314
0.1404
0.0875
0.0669
0.0623
0.0589
0.0423

16 PERIOD GR= 49
3824.49
3825.28
3825.19
3825.57
3825.77
3825.95
3825.26
3824.87
3824.71
3824.59

0.0647
0.0997
0.1123
0.1189
0.1749
0.1637
0.1357
0.1142
0.1031
0.0793

0.0427
0.0669
0.1065
0.1155
0.1378
0.1722
0.1820
0.0885
0.1094
0.0835
0.0760
0.0591
0.0519

17 PERIOD GR= 55
3824.65
3825.83
3826.98
3827.27
3827.27
3826.79
3826.39
3825.62
3825.47
3825.07

0.1034
0.1212
0.1160
0.1090
0.0942
0.0890
0.0816
0.0722
0.0653

0.0397
0.0739
0.1029
0.1027
0.0979
0.1070
0.1205
0.1054
0.0481
0.0471
0.0417
0.0303

18 PERIOD GR= 64
3825.64
3826.01
3825.77
3825.77
3825.52
3825.13
3824.98
3824.74
3824.50

0.0274

19 PERIOD GR=

1171 0.0403
 1271 0.0413
 1772 0.0360
 2773 0.0302

27 PERIOD QR= 0.0890
 3772 0.0461
 4772 0.0513
 5772 0.0338
 6772 0.0323
 7772 0.0372
 8772 0.0371
 9772 0.0251
 10772 0.0349
 11772 0.0188
 12773 0.0159
 1773 0.0119

28 PERIOD QR= -0.0016
 3773 0.0191
 4773 0.0323
 5773 0.0477
 6773 0.0735
 7773 0.0839
 8773 0.0527
 9773 1.0055
 10773 0.0512
 11774 0.0444
 12774 0.0347

29 PERIOD QR= 0.0830
 3774 0.0336
 4774 0.0725
 5774 0.0783
 6774 0.0873
 7774 0.1050
 8774 0.0773
 9774 0.1168
 10774 0.1089
 11774 0.1045
 12775 0.0509
 1775 0.0532
 2775 0.0408

30 PERIOD QR= 0.0337
 3775 0.0577
 4775 0.0751
 5775 0.0790
 6775 0.0860

23 PERIOD QR= 0.0065
 3776 0.0375
 4776 0.0558
 5776 0.0584
 6776 0.0603
 7776 0.0807
 8776 0.0994
 9776 0.0986
 10776 0.0817
 11776 0.0887
 12776 0.0462
 1777 0.0309

24 PERIOD QR= 0.0389
 3777 0.0451
 4777 0.0645
 5777 0.0782
 6777 0.0770
 7777 0.1193
 8777 0.1123
 9777 0.1465
 10777 0.0948
 11777 0.0767
 12777 0.0704
 1778 0.0598
 2778 0.0507

25 PERIOD QR= 0.0221
 3778 0.0744
 4778 0.1035
 5778 0.1019
 6778 0.1269
 7778 0.1154
 8778 0.1556
 9778 0.0949
 10778 0.0712
 11778 0.0598
 12778 0.0430

26 PERIOD QR= -0.0416
 3779 0.0577
 4779 0.0749
 5779 0.0661
 6779 0.0668
 7779 0.0847
 8779 0.0975
 9779 0.0830
 10779 0.0588

27 PERIOD QR= 0.0516
 3780 0.0516
 4780 0.0700
 5780 0.1212
 6780 0.1544
 7780 0.1080
 8780 0.0919
 9780 0.0856
 10780 0.0804
 11780 0.0665

0.0578
 0.0413
 0.0302
 0.0268

0.0212
 0.0230
 0.0203
 0.0189
 0.0237
 0.0345
 0.0337
 0.0383
 0.0449

0.0590
 0.0454
 0.0351
 0.0496
 0.0523
 0.0495
 0.0472
 0.0444
 0.0347

0.0366
 0.0379
 0.0545
 0.0676
 0.0773
 0.0816
 0.0632
 0.0619
 0.0676

0.0800
 0.0785
 0.0902
 0.1071

775 3825.57 0.1022
 875 3825.74 0.1143
 975 3825.92 0.1431
 1075 3825.24 0.1418
 1175 3824.88 0.1022
 1275 3824.48 0.0631
 176 3824.65 0.0709
 276 3824.59 0.0732

31 PERIOD CR= -0.0001
 376 3824.63 0.0784
 476 3825.34 0.0983
 576 3825.38 0.1158
 676 3825.92 0.1150
 776 3825.84 0.1109
 876 3824.60 0.1131
 976 3825.57 0.1114
 1076 3824.45 0.1073
 1176 3824.43 0.0630
 177 3824.22 0.0572
 277 3823.94 0.0509
 377 3823.94 0.0436

TRIAGONAL MATRIX COEFFS FOR THE 1 PERIOD

D 1.22 0.25
 1.24 0.27
 1.31 0.27
 1.37 0.27
 0.65 0.27
 -0.19 0.00

E 0.30
 0.25
 0.24
 0.28
 0.10
 -0.08

F 0.25
 0.27
 0.27
 0.27
 0.27
 0.00

TRIAGONAL MATRIX COEFFS FOR THE 2 PERIOD

D 1.23 0.25
 1.51 0.27
 1.48 0.27
 1.44 0.27
 0.74 0.27
 -0.08 0.00

E 0.31
 0.24
 0.23
 0.30
 0.12
 -0.05

F 0.25
 0.27
 0.27
 0.27
 0.27
 0.00

TRIAGONAL MATRIX COEFFS FOR THE 3 PERIOD

D -0.16
 0.34
 0.21
 0.32

E 0.25
 0.27
 0.27

F -0.04
 0.10
 0.03
 0.08

0.93 0.25
 1.49 0.27
 1.09 0.27
 1.26 0.27
 1.27 0.27
 0.83 0.27
 0.04 0.00

TRIAGONAL MATRIX COEFFS FOR THE 4 PERIOD

D 1.06 0.25
 1.25 0.27
 0.92 0.27
 1.37 0.27
 1.66 0.27
 -0.04 0.00

E 0.25
 0.27
 0.27
 0.27
 0.27
 0.00

F 0.27
 0.26
 0.18
 0.29
 0.10
 -0.04

TRIAGONAL MATRIX COEFFS FOR THE 5 PERIOD

D 1.01 0.25
 1.33 0.27
 1.11 0.27
 1.14 0.27
 1.63 0.27
 -0.15 0.00

E 0.25
 0.27
 0.27
 0.27
 0.27
 0.00

F 0.25
 0.29
 0.22
 0.24
 0.10
 -0.07

TRIAGONAL MATRIX COEFFS FOR THE 6 PERIOD

D 1.05 0.25
 0.84 0.27
 0.68 0.27
 0.94 0.27
 0.87 0.27
 0.66 0.00

E 0.25
 0.27
 0.27
 0.27
 0.27
 0.00

F 0.26
 0.15
 0.14
 0.21
 0.18
 0.08

TRIAGONAL MATRIX COEFFS FOR THE 7 PERIOD

D -0.16
 0.34
 0.21
 0.32

E 0.25
 0.27
 0.27

F -0.04
 0.10
 0.03
 0.08

0.53
0.71
0.12

0.27
0.27
0.00

0.12
0.16
-0.01

TRIDIAGONAL MATRIX COEFFS FOR THE 8 PERIOD

D
0.53
0.20
0.05
0.13
0.48
0.05

E
0.25
0.27
0.27
0.27
0.27
0.00

F
0.13
0.02
0.01
0.03
0.12
-0.02

TRIDIAGONAL MATRIX COEFFS FOR THE 9 PERIOD

D
-0.37
-0.31
-0.21
-0.43
-0.09
0.10

E
0.25
0.27
0.27
0.27
0.27
0.00

F
-0.09
0.08
-0.14
-0.08
-0.00
0.03

TRIDIAGONAL MATRIX COEFFS FOR THE 10 PERIOD

D
-0.89
-0.53
-0.77
-0.48
-0.31
-0.24
0.15

E
0.25
0.27
0.27
0.27
0.27
0.00

F
-0.22
-0.08
-0.18
-0.08
-0.08
-0.05
0.05

TRIDIAGONAL MATRIX COEFFS FOR THE 11 PERIOD

D
-0.66
-0.32
-0.74
-0.88
-0.33
-0.08

E
0.25
0.27
0.27
0.27
0.27
0.00

F
-0.16
-0.19
-0.20
-0.18
-0.04
-0.01

TRIDIAGONAL MATRIX COEFFS FOR THE 12 PERIOD

D
-1.12
-0.68
-0.53
-0.43
0.28
0.85
0.45

E
0.25
0.27
0.27
0.27
0.27
0.00

F
-0.28
-0.11
-0.11
-0.08
0.10
0.20
0.07

TRIDIAGONAL MATRIX COEFFS FOR THE 13 PERIOD

D
0.81
1.06
1.17
1.26
1.33
1.31
0.51

E
0.25
0.27
0.27
0.27
0.27
0.00

F
0.20
0.23
0.25
0.27
0.28
0.25
0.07

TRIDIAGONAL MATRIX COEFFS FOR THE 14 PERIOD

D
0.84
0.59
0.60
0.97
0.99
0.55
-0.10

E
0.25
0.27
0.27
0.27
0.27
0.00

F
0.21
0.10
0.13
0.22
0.20
0.09
-0.05

TRIDIAGONAL MATRIX COEFFS FOR THE 15 PERIOD

D
1.01
0.99
0.81
1.00
1.17
1.15
0.10

E
0.25
0.27
0.27
0.27
0.27
0.00

F
0.25
0.20
0.17
0.22
0.25
0.16
-0.01

TRIDIAGONAL MATRIX COEFFS FOR THE 16 PERIOD

D	E	F
0.48	0.25	0.12
0.73	0.27	0.16
0.49	0.27	0.09
0.75	0.27	0.18
0.60	0.27	0.14
0.00	0.00	0.12
		-0.03

TRIDIAGONAL MATRIX COEFFS FOR THE 17 PERIOD

D	E	F
0.80	0.25	0.20
1.02	0.27	0.24
0.95	0.27	0.19
1.34	0.27	0.31
1.18	0.27	0.23
0.66	0.27	0.11
0.25	0.00	0.04

TRIDIAGONAL MATRIX COEFFS FOR THE 18 PERIOD

D	E	F
0.99	0.25	0.25
0.64	0.27	0.11
0.29	0.27	0.05
0.33	0.27	0.08
0.36	0.27	0.08
0.23	0.27	0.04
-0.02	0.00	-0.02

TRIDIAGONAL MATRIX COEFFS FOR THE 19 PERIOD

D	E	F
-0.61	0.25	-0.15
-0.36	0.27	-0.06
-0.62	0.27	-0.15
-0.84	0.27	-0.18
-1.05	0.27	-0.23
-0.54	0.27	-0.08
-0.11	0.00	-0.01

TRIDIAGONAL MATRIX COEFFS FOR THE 20 PERIOD

D	E	F
-0.89	0.25	-0.22
-0.89	0.27	-0.16

D	E	F
-0.16	0.27	0.08
0.32	0.27	0.08
1.05	0.27	0.26
1.16	0.27	0.24
0.00	0.00	0.00

TRIDIAGONAL MATRIX COEFFS FOR THE 21 PERIOD

D	E	F
0.74	0.25	0.18
0.90	0.27	0.19
0.86	0.27	0.18
1.01	0.27	0.22
1.22	0.27	0.27
1.01	0.27	0.27
0.22	0.00	0.00

TRIDIAGONAL MATRIX COEFFS FOR THE 22 PERIOD

D	E	F
0.46	0.25	0.12
-0.03	0.27	-0.04
-0.12	0.27	-0.02
-0.01	0.27	-0.01
0.35	0.27	0.09
0.69	0.27	0.16
0.18	0.00	0.01

TRIDIAGONAL MATRIX COEFFS FOR THE 23 PERIOD

D	E	F
0.63	0.25	0.16
0.05	0.27	-0.03
0.15	0.27	0.05
0.61	0.27	0.15
1.01	0.27	0.23
0.94	0.27	0.19
0.07	0.00	-0.03

TRIDIAGONAL MATRIX COEFFS FOR THE 24 PERIOD

D	E	F
0.75	0.25	0.19
0.54	0.27	0.09
0.36	0.27	0.07
1.04	0.27	0.26
1.54	0.27	0.34
1.02	0.27	0.18

TRIDIAGONAL MATRIX COEFFS FOR THE 29 PERIOD

D	-0.23	E	0.25	F	-0.06
	0.59		0.27		0.17
	0.51		0.27		0.09
	0.28		0.27		0.05
	0.12		0.27		0.02
	0.10		0.27		-0.05
	-0.18		0.00		

TRIDIAGONAL MATRIX COEFFS FOR THE 30 PERIOD

D	0.46	E	0.25	F	0.12
	0.45		0.27		0.09
	0.78		0.27		0.19
	0.57		0.27		0.10
	0.43		0.27		0.09
	0.47		0.27		0.10
	0.07		0.00		-0.01

TRIDIAGONAL MATRIX COEFFS FOR THE 31 PERIOD

D	0.90	E	0.25	F	0.22
	1.01		0.27		0.21
	0.95		0.27		0.21
	0.23		0.27		0.12
	-0.29		0.27		-0.11
	0.99		0.00		0.24

RECHARGE ESTIMATES FOR THE 1 TH PERIOD

DATE	RECHARGE (FT/MO)
346	0.2419
446	0.2507
546	0.1793
646	0.2712
746	0.2460
846	0.1194
946	-0.0771
1046	0.0000
1146	0.0000
1246	0.0000
147	0.0000
247	0.0000

RECHARGE ESTIMATES FOR THE 2 TH PERIOD

0.05 0.00 -0.04

TRIDIAGONAL MATRIX COEFFS FOR THE 25 PERIOD

D	0.90	E	0.25	F	0.23
	0.67		0.27		0.23
	0.97		0.27		0.27
	1.24		0.27		0.30
	1.40		0.27		0.16
	0.90		0.27		0.00
	0.18		0.00		

TRIDIAGONAL MATRIX COEFFS FOR THE 26 PERIOD

D	0.79	E	0.25	F	0.20
	0.82		0.27		0.16
	0.80		0.27		0.18
	0.75		0.27		0.17
	0.49		0.27		0.16
	0.34		0.27		0.09
			0.00		0.07

TRIDIAGONAL MATRIX COEFFS FOR THE 27 PERIOD

D	-0.64	E	0.25	F	0.16
	-0.45		0.27		-0.08
	-0.44		0.27		-0.10
	-0.48		0.27		-0.09
	-0.35		0.27		-0.10
	0.26		0.00		-0.07
					0.09

TRIDIAGONAL MATRIX COEFFS FOR THE 28 PERIOD

D	0.10	E	0.25	F	0.02
	0.02		0.27		0.06
	0.12		0.27		-0.01
	0.37		0.27		0.03
	0.94		0.27		0.25
	0.33		0.27		0.11
			0.00		0.06

650	0.1677
750	0.2124
850	0.1158
950	-0.0662
1050	0.0000
1150	0.0000
1250	0.0000
1351	0.0000
1451	0.0000

RECHARGE ESTIMATES FOR THE 6 TH PERIOD

DATE	RECHARGE (FT/MO)
351	0.2292
451	0.1286
551	0.0929
651	0.1755
751	0.1481
851	0.1062
951	0.0841
1051	0.0000
1151	0.0000
1251	0.0000
1352	0.0000
1452	0.0000

RECHARGE ESTIMATES FOR THE 7 TH PERIOD

DATE	RECHARGE (FT/MO)
352	-0.0655
452	0.0993
552	0.0121
652	0.0586
752	0.0773
852	0.1615
952	-0.0093
1052	0.0000
1152	0.0000
1252	0.0000
1353	0.0000
1453	0.0000

RECHARGE ESTIMATES FOR THE 8 TH PERIOD

DATE	RECHARGE (FT/MO)
353	0.1282
453	0.0175
553	0.0008
653	0.0340
753	-0.0083
853	0.1276
953	-0.0191
1053	0.0000
1153	0.0000

RECHARGE (FT/MO)

DATE	RECHARGE (FT/MO)
347	0.2387
447	0.2768
547	0.1688
647	0.2618
747	0.2616
847	0.1323
947	-0.0518
1047	0.0000
1147	0.0000
1247	0.0000
1348	0.0000
1448	0.0000

RECHARGE ESTIMATES FOR THE 3 TH PERIOD

RECHARGE (FT/MO)

DATE	RECHARGE (FT/MO)
348	0.1589
448	0.2278
548	0.1409
648	0.2296
748	0.3008
848	0.1388
948	-0.0237
1048	0.0000
1148	0.0000
1248	0.0000
1349	0.0000
1449	0.0000

RECHARGE ESTIMATES FOR THE 4 TH PERIOD

RECHARGE (FT/MO)

DATE	RECHARGE (FT/MO)
349	0.2074
449	0.2307
549	0.1181
649	0.2215
749	0.2600
849	0.1091
949	-0.0362
1049	0.0000
1149	0.0000
1249	0.0000
1350	0.0000
1450	0.0000

RECHARGE ESTIMATES FOR THE 5 TH PERIOD

RECHARGE (FT/MO)

DATE	RECHARGE (FT/MO)
350	0.1930
450	0.2414
550	0.1745

RECHARGE ESTIMATES FOR THE 9 TH PERIOD

DATE	RECHARGE (FT/MO)
1253	0.0000
154	0.0000
254	0.0000
RECHARGE ESTIMATES FOR THE 9 TH PERIOD	
DATE	RECHARGE (FT/MO)
354	-0.0953
454	0.0135
554	-0.0505
654	-0.1234
754	-0.0727
854	-0.0126
954	0.0287
1054	0.0000
1154	0.0000
1254	0.0000
155	0.0000
255	0.0000

RECHARGE ESTIMATES FOR THE 10 TH PERIOD

DATE	RECHARGE (FT/MO)
355	-0.2123
455	-0.0376
555	-0.1657
655	-0.0682
755	-0.0447
855	-0.0622
955	0.0540
1055	0.0000
1155	0.0000
1255	0.0000
156	0.0000
256	0.0000

RECHARGE ESTIMATES FOR THE 11 TH PERIOD

DATE	RECHARGE (FT/MO)
356	-0.1645
456	-0.0003
556	-0.1533
656	-0.1539
756	-0.1726
856	-0.0357
956	-0.0103
1056	0.0000
1156	0.0000
1256	0.0000
157	0.0000
257	0.0000

RECHARGE ESTIMATES FOR THE 12 TH PERIOD

DATE	RECHARGE (FT/MO)
360	0.2118
460	0.1634
560	0.1207

RECHARGE ESTIMATES FOR THE 13 TH PERIOD

DATE	RECHARGE (FT/MO)
358	0.1579
458	0.1765
558	0.1949
658	0.2116
758	0.2217
858	0.2308
958	0.0698
1058	0.0000
1158	0.0000
1258	0.0000
159	0.0000
259	0.0000

RECHARGE ESTIMATES FOR THE 14 TH PERIOD

DATE	RECHARGE (FT/MO)
359	0.1894
459	0.0790
559	0.0862
659	0.1758
759	0.1759
859	0.1061
959	-0.0518
1059	0.0000
1159	0.0000
1259	0.0000
160	0.0000
260	0.0000

RECHARGE ESTIMATES FOR THE 15 TH PERIOD

DATE	RECHARGE (FT/MO)
360	0.2118
460	0.1634
560	0.1207

1263	0.0000
164	0.0000
264	0.0000

RECHARGE ESTIMATES FOR THE 19 TH PERIOD

DATE RECHARGE (FT/MO)

364	-0.1449
464	-0.0255
564	-0.1173
664	-0.1277
764	-0.2095
864	-0.0813
964	-0.0063
1064	0.0000
1164	0.0000
1264	0.0000
165	0.0000
265	0.0000

RECHARGE ESTIMATES FOR THE 20 TH PERIOD

DATE RECHARGE (FT/MO)

365	-0.1783
465	-0.1760
565	-0.0042
665	0.0329
765	0.1924
865	0.2429
965	0.0001
1065	0.0000
1165	0.0000
1265	0.0000
166	0.0000
266	0.0000

RECHARGE ESTIMATES FOR THE 21 TH PERIOD

DATE RECHARGE (FT/MO)

366	0.1459
466	0.1557
566	0.1338
666	0.1668
766	0.2135
866	0.1979
966	0.0044
1066	0.0000
1166	0.0000
1266	0.0000
167	0.0000
267	0.0000

RECHARGE ESTIMATES FOR THE 22 TH PERIOD

660	0.1662
760	0.2110
860	0.1632
960	-0.0146
1060	0.0000
1160	0.0000
1260	0.0000
161	0.0000
261	0.0000

RECHARGE ESTIMATES FOR THE 16 TH PERIOD

DATE RECHARGE (FT/MO)

361	0.0839
461	0.1495
561	0.0478
661	0.1510
761	0.0999
861	0.1320
961	-0.0319
1061	0.0000
1161	0.0000
1261	0.0000
162	0.0000
262	0.0000

RECHARGE ESTIMATES FOR THE 17 TH PERIOD

DATE RECHARGE (FT/MO)

362	0.1479
462	0.2051
562	0.1227
662	0.2524
762	0.2055
862	0.1052
962	0.0352
1062	0.0000
1162	0.0000
1262	0.0000
163	0.0000
263	0.0000

RECHARGE ESTIMATES FOR THE 18 TH PERIOD

DATE RECHARGE (FT/MO)

363	0.2236
463	0.0961
563	0.0348
663	0.0581
763	0.0647
863	0.0444
963	-0.0159
1063	0.0000
1163	0.0000

670	0.2025
770	0.2587
870	0.1601
970	0.0050
1070	0.0000
1170	0.0000
1270	0.0000
171	0.0000
271	0.0000

RECHARGE ESTIMATES FOR THE 26 TH PERIOD
DATE RECHARGE (FT/MO)

371	0.1695
471	0.1175
571	0.1422
671	0.1293
771	0.1384
871	0.0708
971	0.0679
1071	0.0000
1171	0.0000
1271	0.0000
172	0.0000
272	0.0000

RECHARGE ESTIMATES FOR THE 27 TH PERIOD
DATE RECHARGE (FT/MO)

372	-0.1458
472	-0.0650
572	-0.0820
672	-0.0697
772	-0.0798
872	-0.0887
972	0.0876
1072	0.0000
1172	0.0000
1272	0.0000
173	0.0000
273	0.0000

RECHARGE ESTIMATES FOR THE 28 TH PERIOD
DATE RECHARGE (FT/MO)

373	0.0105
473	0.0566
573	-0.0030
673	-0.0262
773	0.2257
873	0.0889
973	0.0811
1073	0.0000
1173	0.0000

DATE RECHARGE (FT/MO)

367	0.1239
467	-0.0328
567	-0.0200
667	-0.0038
767	0.0494
867	0.1580
967	0.0066
1067	0.0000
1167	0.0000
1267	0.0000
168	0.0000
268	0.0000

RECHARGE ESTIMATES FOR THE 23 TH PERIOD
DATE RECHARGE (FT/MO)

368	0.1669
468	-0.0342
568	0.0208
668	0.1042
768	0.1769
868	0.2002
968	-0.0331
1068	0.0000
1168	0.0000
1268	0.0000
169	0.0000
269	0.0000

RECHARGE ESTIMATES FOR THE 24 TH PERIOD
DATE RECHARGE (FT/MO)

369	0.1664
469	0.0871
569	0.0230
669	0.1821
769	0.2926
869	0.1896
969	-0.0352
1069	0.0000
1169	0.0000
1269	0.0000
170	0.0000
270	0.0000

RECHARGE ESTIMATES FOR THE 25 TH PERIOD
DATE RECHARGE (FT/MO)

370	0.2084
470	0.0714
570	0.1729

1273 0.0000
 174 0.0000
 274 0.0000

RECHARGE ESTIMATES FOR THE 29 TH PERIOD

DATE RECHARGE (FT/MO)

374 -0.0942
 474 0.1516
 574 0.0781
 674 0.0476
 774 0.0098
 874 0.0373
 974 -0.0550
 1074 0.0000
 1174 0.0000
 1274 0.0000
 175 0.0000
 275 0.0000

RECHARGE ESTIMATES FOR THE 30 TH PERIOD

DATE RECHARGE (FT/MO)

375 0.1039
 475 0.0447
 575 0.1628
 675 0.0862
 775 0.0593
 875 0.1033
 975 -0.0071
 1075 0.0000
 1175 0.0000
 1275 0.0000
 176 0.0000
 276 0.0000

RECHARGE ESTIMATES FOR THE 31 TH PERIOD

DATE RECHARGE (FT/MO)

376 0.1412
 476 0.1728
 576 0.1349
 676 0.2420
 776 -0.1299
 876 0.0452
 976 0.2360
 1076 0.0000
 1176 0.0000
 1276 0.0000
 177 0.0000
 277 0.0000

MODEL PREDICTIONS AND COMPUTATION OF ERRORS

DATE	PRED W.L.	OBS. W.L.	ABS. ERROR	SO. OF ABS. ERR.	RMS#
346	3826.38	3826.38	0.00	0.0000	4
446	3827.34	3827.34	-0.22	0.0474	3
546	3827.51	3827.51	-0.19	0.0375	3
946	3827.51	3828.00	-0.09	0.0089	4
846	3828.34	3828.42	-0.08	0.0057	4
1046	3827.75	3827.68	0.07	0.0044	4
1146	3826.48	3826.84	-0.12	0.0143	3
1146	3826.23	3826.57	-0.06	0.0017	3
146	3825.56	3826.17	-0.01	0.0038	3
246	3825.78	3825.89	-0.11	0.0000	3
	ERRORS FOR 1 TH PERIOD	ABS#	-0.07	SO(ABS)#	0.1358
					RMS#
347	3826.33	3826.18	0.05	0.0022	4
447	3826.59	3827.50	-0.19	0.0350	4
547	3827.47	3827.80	-0.03	0.0008	3
647	3828.23	3828.32	-0.09	0.0000	4
847	3828.28	3827.77	0.05	0.0084	2
947	3827.70	3827.77	-0.07	0.0014	4
1047	3826.00	3826.84	-0.09	0.0019	4
1147	3826.58	3826.84	-0.16	0.0269	3
1147	3826.26	3826.38	0.20	0.0409	3
148	3826.20	3826.18	0.08	0.0061	3
148	3825.80	3825.86	-0.14	0.0209	3
248	3825.80	3825.79	0.01	0.0002	3
	ERRORS FOR 2 TH PERIOD	ABS#	0.24	SO(ABS)#	0.1475
					RMS#
348	3825.97	3825.74	0.23	0.0538	4
448	3826.52	3826.78	-0.26	0.0675	4
548	3827.09	3827.21	-0.12	0.0143	3
648	3827.39	3827.38	0.01	0.0000	4
748	3827.93	3828.05	-0.12	0.0153	3
848	3828.14	3828.24	-0.10	0.0001	4
948	3827.66	3827.69	-0.03	0.0010	4
1048	3827.02	3826.95	0.07	0.0009	3
1148	3826.58	3826.58	0.00	0.0000	3
1148	3826.25	3826.25	0.00	0.0000	3
149	3825.99	3825.89	-0.10	0.0096	3
249	3825.78	3825.86	-0.08	0.0063	3
	ERRORS FOR 3 TH PERIOD	ABS#	-0.30	SO(ABS)#	0.1818
					RMS#
349	3826.04	3826.17	-0.13	0.0160	4
449	3826.87	3827.01	-0.14	0.0187	4
549	3827.23	3827.23	-0.04	0.0013	3
749	3827.49	3827.43	0.06	0.0031	4
849	3828.06	3828.06	-0.01	0.0009	3
949	3828.11	3827.98	0.03	0.0010	4
1049	3827.58	3827.67	-0.04	0.0016	4
1149	3826.62	3826.52	0.10	0.0043	3
					0.12

1249	3826.33	4 TH PERIOD	ABS#	0.09	0.0079	0.0915	RMS#	0.09
150	3826.11			-0.01	0.0073			
250	3825.94				0.0001			
		ERRORS FOR			SQ(ABS)#			
350	3826.19			-0.04	0.0002			
450	3826.68			-0.01	0.0104			
550	3827.36			-0.04	0.0017			
650	3827.57			-0.02	0.0003			
750	3827.81			-0.01	0.0001			
850	3827.89			-0.10	0.0108			
950	3827.77			0.01	0.0000			
1050	3826.44			0.09	0.0086			
1150	3826.18			0.11	0.0118			
1250	3825.98			0.07	0.0049			
1351	3825.82			-0.09	0.0075			
1451				-0.01	0.0002			
		ERRORS FOR			SQ(ABS)#			
351	3826.16			0.08	0.0565			0.07
451	3826.93			0.23	0.0528			
551	3826.99			-0.14	0.0097			
651	3826.18			-0.06	0.0041			
751	3826.22			-0.02	0.0004			
851	3826.00			-0.06	0.0041			
951	3825.74			-0.11	0.0031			
1051	3825.88			-0.48	0.0121			
1151	3824.76			-0.12	0.0234			
1252	3824.69			-0.13	0.0146			
1352	3824.59			-0.30	0.0159			
		ERRORS FOR			SQ(ABS)#			
352	3823.86			-0.76	0.4591			0.20
452	3824.33			0.05	0.0025			
552	3824.23			-0.38	0.1431			
652	3824.26			-0.13	0.0165			
752	3824.29			-0.73	0.0511			
852	3824.53			-0.10	0.0100			
952	3825.45			-0.14	0.0201			
1052	3824.96			-0.23	0.0550			
1152	3824.84			0.07	0.0051			
1253	3824.78			0.08	0.0033			
1353	3824.71			-0.08	0.0070			
1453	3824.50			-0.02	0.0005			
		ERRORS FOR			SQ(ABS)#			
353	3824.61			-0.86	0.3252			0.16
453	3824.81			0.16	0.0247			
553	3824.01			-0.18	0.0339			
653	3824.74			-0.12	0.0133			
753	3824.48			-0.15	0.0232			
853	3824.56			-0.08	0.0065			
953	3824.72			-0.03	0.0007			
1053	3824.91			-0.13	0.0167			
1153	3824.46			0.13	0.0167			
1253	3824.46			0.06	0.0041			

1253	3824.36	8 TH PERIOD	ABS#	0.09	50(ABS)#	0.0072	RMS#	0.13
154	3824.26			-0.07		0.0049		
254	3823.95			0.23		0.0513		
		ERRORS FOR	ABS#	0.30	0.2025			
354	3823.90	3823.33		0.57	0.3202			
454	3823.12	3823.13		-0.18	0.0335			
554	3823.19	3823.19		-0.11	0.0234			
654	3822.62	3822.20		-0.01	0.0281			
754	3822.20	3822.20		-0.00	0.0000			
854	3822.04	3822.13		-0.16	0.0247			
954	3822.12	3822.43		-0.01	0.0001			
1054	3822.23	3822.38		-0.20	0.0420			
1154	3822.29	3822.40		-0.11	0.0145			
1255	3822.32	3822.55		-0.23	0.0115			
1355	3822.35	3822.41		-0.06	0.0508			
					0.0031			
		ERRORS FOR	ABS#	-0.78	0.5407			0.21
355	3821.91	3821.36		0.55	0.2992			
455	3820.49	3820.71		-0.22	0.1048			
555	3820.05	3820.16		-0.11	0.0491			
655	3819.89	3820.16		-0.27	0.0705			
755	3819.74	3819.99		-0.17	0.0288			
855	3819.82	3819.99		-0.17	0.0284			
955	3820.04	3820.56		-0.43	0.0877			
1055	3820.14	3820.77		-0.49	0.1759			
1155	3820.33	3820.77		-0.44	0.1430			
1256	3820.42	3820.66		-0.24	0.0558			
		ERRORS FOR	ABS#	-2.64	1.3625			0.34
356	3820.14	3819.66		0.48	0.2312			
456	3819.45	3819.50		-0.53	0.1054			
556	3819.26	3819.14		-0.12	0.0579			
656	3818.73	3818.40		-0.33	0.1679			
756	3818.15	3818.25		-0.10	0.0604			
856	3817.92	3818.37		-0.45	0.0834			
956	3818.07	3818.25		-0.18	0.1085			
1056	3818.25	3818.59		-0.34	0.1891			
1156	3818.42	3818.73		-0.31	0.1185			
1257	3818.57	3818.93		-0.36	0.0965			
1357	3818.76	3819.00		-0.24	0.1582			
					0.0555			
		ERRORS FOR	ABS#	-3.15	1.5069			0.35
357	3818.34	3818.13		0.21	0.0451			
457	3817.56	3817.53		-0.03	0.0247			
557	3817.39	3817.35		-0.04	0.0189			
657	3817.19	3817.35		-0.16	0.0264			
757	3817.29	3817.31		-0.02	0.0003			
857	3818.05	3818.22		-0.17	0.0292			
957	3818.84	3819.03		-0.19	0.0443			
1057	3819.20	3819.37		-0.17	0.0165			
1157	3819.41	3819.67		-0.26	0.0655			

1257	3819.62	3819.81	-0.19	0.0347	0.4209	RMS#	0.19
1258	3819.83	3820.17	-0.34	0.1145		2	
1259	3820.04	3820.07	-0.03	0.0011		2	
	ERRORS FOR 12 TH PERIOD	ABS#		SO(ABS)#		RMS#	
358	3820.60	3820.43	0.17	0.0291		2	
458	3821.21	3821.34	-0.10	0.0172		3	
558	3822.05	3822.07	-0.02	0.0000		3	
658	3823.82	3823.82	0.00	0.0173		3	
758	3823.82	3823.82	0.00	0.0000		3	
858	3824.66	3824.79	-0.13	0.0164		3	
958	3824.64	3825.15	-0.09	0.0074		3	
1058	3824.64	3824.93	-0.01	0.0002		3	
1158	3824.41	3824.66	-0.02	0.0003		3	
1259	3824.02	3824.47	-0.05	0.0028		3	
1359	3824.02	3825.96	-0.06	0.0045		3	
	ERRORS FOR 13 TH PERIOD	ABS#		SO(ABS)#		RMS#	0.09
359	3824.27	3824.54	-0.26	0.0658		3	
459	3825.13	3825.02	0.09	0.0004		3	
559	3825.13	3825.38	-0.05	0.0077		3	
659	3825.13	3825.89	-0.05	0.0028		3	
759	3825.13	3825.89	-0.05	0.0000		3	
859	3825.13	3825.10	0.04	0.0000		3	
959	3825.13	3825.86	-0.05	0.0025		3	
1059	3825.13	3825.18	0.02	0.0298		3	
1159	3824.88	3824.68	0.19	0.0426		3	
1260	3824.88	3824.56	0.12	0.0334		3	
1360	3824.51	3824.56	-0.09	0.0134		3	
	ERRORS FOR 14 TH PERIOD	ABS#		SO(ABS)#		RMS#	0.13
360	3825.84	3824.82	0.63	0.0001		3	
460	3825.84	3825.57	0.12	0.0155		3	
560	3825.84	3825.74	-0.05	0.0019		3	
660	3826.41	3826.36	0.05	0.0030		3	
760	3826.41	3826.83	-0.05	0.0023		3	
860	3826.41	3826.47	0.01	0.0149		3	
960	3825.97	3825.93	0.04	0.0015		3	
1060	3825.97	3825.52	0.07	0.0055		3	
1160	3825.97	3825.28	0.04	0.0001		3	
1261	3824.83	3825.01	-0.03	0.0001		3	
1361	3824.83	3824.80	0.03	0.0010		1	
	ERRORS FOR 15 TH PERIOD	ABS#		SO(ABS)#		RMS#	0.06
361	3824.83	3824.49	0.34	0.1165		3	
461	3825.16	3825.28	-0.39	0.1496		3	
561	3825.16	3825.19	-0.03	0.0010		3	
661	3825.68	3825.19	-0.19	0.0346		3	
761	3825.68	3825.77	-0.09	0.0086		3	
861	3825.76	3825.95	-0.07	0.0055		3	
961	3825.76	3825.88	-0.19	0.0354		3	
1061	3825.76	3825.95	-0.14	0.0184		3	
1161	3825.76	3825.00	0.16	0.0265		3	

1261	3824.97	3824.87	0.10	0.0099	3	0.19
162	3824.81	3824.71	0.10	0.0093	3	
262	3824.67	3824.59	0.08	0.0061	3	
ERRORS FOR 16 TH PERIOD ABS=						
362	3824.86	3824.65	-0.05	SQ(ABS)=	0.4215	RMS=
462	3825.31	3825.56	0.21			
562	3825.80	3825.83	-0.03			
762	3826.29	3826.35	-0.06			
862	3826.97	3827.98	-0.14			
962	3827.72	3827.00	0.03			
1062	3828.72	3826.79	0.07			
1162	3829.33	3826.29	0.06			
1262	3829.62	3825.87	0.00			
1362	3829.37	3825.62	0.04			
1462	3829.17	3825.41	-0.04			
1562	3828.17	3825.07	0.10			
ERRORS FOR 17 TH PERIOD ABS=						
363	3825.47	3825.64	-0.22	SQ(ABS)=	0.1548	RMS=
463	3825.04	3826.01	0.03			0.11
563	3825.97	3826.91	0.06			
763	3826.83	3827.76	0.07			
863	3827.79	3828.77	0.02			
963	3828.72	3829.68	0.04			
1063	3829.33	3829.52	-0.02			
1163	3829.62	3829.13	0.10			
1263	3829.37	3829.99	0.04			
1363	3829.17	3829.88	-0.01			
1463	3828.17	3829.74	0.00			
1563	3827.17	3829.50	0.12			
ERRORS FOR 18 TH PERIOD ABS=						
364	3823.21	3823.56	0.26	SQ(ABS)=	0.0670	RMS=
464	3823.83	3823.47	0.65			0.07
564	3824.51	3822.96	-0.13			
764	3825.21	3822.54	-0.23			
864	3826.02	3821.08	0.16			
964	3826.99	3821.11	0.20			
1064	3827.08	3821.10	-0.11			
1164	3827.18	3821.34	-0.25			
1264	3827.49	3821.48	-0.30			
1364	3827.35	3821.49	-0.23			
1464	3827.17	3821.45	-0.10			
1564	3826.17	3821.45	-0.10			
ERRORS FOR 19 TH PERIOD ABS=						
365	3821.03	3821.18	-1.44	SQ(ABS)=	0.8892	RMS=
465	3820.08	3821.20	0.13			0.27
565	3820.82	3820.96	0.12			
765	3820.20	3820.10	0.12			
865	3821.85	3820.59	0.21			
965	3822.43	3821.82	0.03			
1065	3822.44	3821.44	-0.01			
1165	3822.22	3822.22	0.00			
1265	3822.22	3822.35	0.09			

1265	3822.45	20 TH PERIOD	ABS	0.04	50(ABS)	0.2477	RMS	0.14
166	3822.46			0.11				
166	3822.29			0.17				
ERRORS FOR 20 TH PERIOD								
366	3822.80			1.19				
466	3822.14			0.29				
566	3822.70			-0.19				
666	3822.74			-0.01				
766	3822.85			-0.11				
866	3822.85			-0.02				
966	3822.48			-0.17				
1066	3822.54			-0.05				
1166	3822.19			0.01				
1166	3822.87			0.04				
1167	3822.44			0.07				
1267	3822.11			-0.06				
1267	3822.19			-0.08				
ERRORS FOR 21 TH PERIOD								
367	3824.28		ABS	-0.09	50(ABS)	0.1822	RMS	0.12
467	3824.25			-0.24				
567	3824.24			0.08				
667	3824.12			0.12				
767	3822.98			0.06				
867	3822.89			0.12				
967	3822.44			0.09				
1067	3822.62			-0.09				
1167	3822.19			0.24				
1167	3822.96			0.20				
1168	3822.86			0.06				
1268	3822.87			0.25				
1268	3822.97							
ERRORS FOR 22 TH PERIOD								
368	3824.08		ABS	1.07	50(ABS)	0.3091	RMS	0.16
468	3824.15			-0.18				
568	3824.11			0.20				
668	3822.95			0.16				
768	3822.03			0.17				
868	3822.51			0.11				
968	3822.08			0.10				
1068	3822.30			-0.10				
1168	3822.49			0.31				
1168	3822.45			0.09				
1168	3822.20			0.10				
1269	3822.20			0.10				
1269	3822.92			0.20				
ERRORS FOR 23 TH PERIOD								
369	3824.13		ABS	1.26	50(ABS)	0.3242	RMS	0.16
469	3824.75			-0.12				
569	3824.82			-0.02				
669	3822.14			0.13				
769	3822.95			-0.06				
869	3822.61			-0.02				
969	3822.66			-0.10				
1069	3822.74			0.21				
1169	3822.58			0.24				
1169	3822.53							

1269	3825.36	24	PERIOD	ABS#	0.13	0.0164	3	0.14	RMS#
170	3825.14				0.04	0.0016	3		
270	3824.72				0.24	0.0566	3		
ERRORS FOR 24 TH PERIOD ABS#									
370	3825.48				-0.24	0.0553	3		
470	3825.47				0.28	0.0804	3		
570	3825.80				0.10	0.0101	3		
670	3826.13				0.15	0.0248	3		
770	3826.71				0.05	0.0030	3		
870	3827.05				0.03	0.0002	3		
970	3826.72				0.03	0.0006	3		
1070	3826.97				0.12	0.0141	3		
1170	3825.97				-0.23	0.0518	3		
1270	3825.38				-0.05	0.0015	3		
1371	3825.64				-0.12	0.0035	3		
1471	3824.73				0.12	0.0133	3		
ERRORS FOR 25 TH PERIOD ABS#									
371	3825.00				0.01	0.0000	3		0.16
471	3824.88				0.22	0.0468	2		
571	3825.12				0.01	0.0001	2		
671	3825.18				0.10	0.0062	2		
771	3825.02				-0.01	0.0002	2		
871	3825.23				0.10	0.0091	2		
971	3825.10				0.02	0.0003	2		
1071	3824.56				0.04	0.0013	2		
1171	3824.11				-0.34	0.0124	2		
1271	3823.48				-0.96	0.0124	2		
1372	3823.31				-0.06	0.0040	2		
1472	3823.11				-0.06	0.0035	2		
ERRORS FOR 26 TH PERIOD ABS#									
372	3822.47				-0.76	0.2540	2		0.31
472	3822.73				-0.50	0.0018	2		
572	3822.86				-0.09	0.0007	2		
672	3822.65				-0.08	0.0002	2		
772	3822.81				-0.19	0.0026	2		
872	3822.54				0.02	0.0004	2		
972	3822.91				-0.12	0.0004	2		
1072	3823.50				0.17	0.0020	2		
1172	3823.41				0.18	0.0010	2		
1272	3823.89				0.18	0.0097	2		
1373	3823.71				-0.12	0.0037	2		
1473	3824.99				-0.12	0.0150	2		
ERRORS FOR 27 TH PERIOD ABS#									
373	3824.49				-0.28	0.0018	2		0.18
473	3824.01				-0.04	0.1329	2		
573	3824.38				-0.14	0.0183	2		
673	3823.56				0.39	0.1530	2		
773	3824.19				0.02	0.0002	2		
873	3824.64				0.07	0.0047	2		
973	3824.68				0.03	0.0080	2		
1073	3824.53				0.03	0.0005	2		
1173	3824.58				0.18	0.0325	2		

	28 TH PERIOD	ABS		SQ(ABS)	RMS	0.17
1273	3824.03	3823.97	0.06	0.0034	3	
174	3823.82	3823.78	0.04	0.0017	3	
274	3823.64	3823.54	0.10	0.0094		
ERRORS FOR 28 TH PERIOD ABS						
374	3823.26	3823.03	1.15	0.3665	RMS	
474	3823.41	3823.69	0.23	0.0538	3	
574	3824.13	3824.34	-0.21	0.0784	3	
674	3824.55	3824.74	-0.19	0.0431	3	
774	3824.77	3824.92	-0.15	0.0359	3	
874	3824.93	3825.02	-0.09	0.0234	3	
974	3824.85	3825.13	-0.28	0.0784	3	
1074	3824.94	3824.78	0.16	0.0256	3	
1174	3824.94	3824.62	0.32	0.1024	3	
1274	3824.98	3824.97	0.01	0.0001	3	
1775	3825.01	3825.07	-0.06	0.0036	3	
2775	3825.01	3824.75	0.26	0.0676	3	

	29 TH PERIOD	ABS		SQ(ABS)	RMS	0.22
375	3825.25	3825.09	0.16	0.0270	3	
475	3825.41	3825.05	0.36	0.1296	3	
575	3825.70	3825.34	0.36	0.1296	3	
675	3825.70	3825.77	-0.07	0.0049	3	
775	3825.74	3825.57	0.17	0.0289	3	
875	3825.63	3825.74	-0.11	0.0121	3	
975	3825.06	3825.62	0.08	0.0064	3	
1075	3825.89	3825.24	0.65	0.4225	3	
1175	3824.73	3824.88	-0.15	0.0225	3	
1275	3824.73	3824.98	-0.25	0.0625	3	
1776	3824.59	3824.75	-0.16	0.0256	3	
2776	3824.59	3824.59	0.00	0.0000	3	

	30 TH PERIOD	ABS		SQ(ABS)	RMS	0.16
376	3824.85	3824.63	1.30	0.3053	RMS	
476	3825.42	3825.54	0.12	0.0144	3	
576	3825.81	3825.38	0.43	0.1849	3	
676	3825.58	3825.92	-0.34	0.1156	3	
776	3825.98	3825.84	0.14	0.0196	3	
876	3825.25	3824.60	0.65	0.4225	3	
976	3825.00	3825.57	-0.57	0.3249	3	
1076	3824.67	3825.45	-0.78	0.6084	3	
1176	3824.39	3824.43	0.04	0.0016	3	
1276	3824.15	3824.22	0.07	0.0049	3	
1777	3824.15	3823.94	0.21	0.0441	3	

	31 TH PERIOD	ABS		SQ(ABS)	RMS	0.27
377	3824.15	3823.94	1.02	0.8744	RMS	

SIMULATION ERRORS(TOTAL): ABSOLUTE ERROR= -4.08 SQUARE OF ABSOL.= 12.54 ROOT MEAN SQ. ERROR= 0.1

MODEL PREDICTIONS AND COMPUTATION OF ERRORS

DATE	PRED. DRAIN	OBS. DRAIN	ABS. DRAIN	ABS. ERROR	SO. OF ABS. ERR.	RMS	RMS
346	0.1405	0.1268	0.0137	0.0137	0.0002		
446	0.1835	0.2031	-0.0196	-0.0196	0.0004		
546	0.2216	0.2308	-0.0092	-0.0092	0.0001		
646	0.2336	0.2485	-0.0149	-0.0149	0.0002		
746	0.2658	0.2853	-0.0195	-0.0195	0.0004		
846	0.2632	0.3010	-0.0378	-0.0378	0.0014		
946	0.2231	0.2441	-0.0210	-0.0210	0.0004		
1046	0.1737	0.1613	0.0124	0.0124	0.0002		
1146	0.1459	0.1254	0.0205	0.0205	0.0004		
1246	0.1122	0.1122	0.0000	0.0000	0.0000		
1347	0.1185	0.0973	0.0212	0.0212	0.0005		
1447	0.1100	0.0773	0.0327	0.0327	0.0011		
ERRORS FOR 1 TH PERIOD ABS=							
347	0.1323	0.1103	0.0220	-0.00	0.0005	0.0056	0.02
447	0.1757	0.1827	-0.0070	0.0220	0.0000		
547	0.2054	0.1991	0.0063	-0.0070	0.0000		
647	0.2257	0.2097	0.0170	0.0063	0.0003		
747	0.2553	0.2498	0.0055	0.0170	0.0000		
847	0.2590	0.2636	-0.0046	0.0055	0.0000		
947	0.2201	0.2378	-0.0177	-0.0046	0.0000		
1047	0.1763	0.1372	0.0391	-0.0177	0.0015		
1147	0.1517	0.1059	0.0458	0.0391	0.0021		
1247	0.1340	0.0913	0.0427	0.0458	0.0018		
1348	0.1209	0.0822	0.0387	0.0427	0.0014		
1448	0.1111	0.0718	0.0393	0.0387	0.0015		
ERRORS FOR 2 TH PERIOD ABS=							
348	0.1193	0.0743	0.0450	0.25	0.0020	0.0094	0.03
448	0.1482	0.1482	0.0000	0.0450	0.0000		
548	0.1810	0.1810	0.0000	-0.0056	0.0000		
648	0.2000	0.1940	0.0060	-0.0056	0.0000		
748	0.2350	0.2534	-0.0184	-0.0060	0.0003		
848	0.2496	0.2770	-0.0274	-0.0184	0.0008		
948	0.2174	0.2424	-0.0250	-0.0274	0.0006		
1048	0.1773	0.1617	0.0156	-0.0250	0.0002		
1148	0.1519	0.1201	0.0318	0.0156	0.0010		
1249	0.1335	0.1132	0.0203	0.0318	0.0004		
1349	0.1201	0.1043	0.0158	0.0203	0.0002		
1449	0.1100	0.0821	0.0279	0.0158	0.0008		
ERRORS FOR 3 TH PERIOD ABS=							
349	0.1229	0.1180	0.0049	0.09	0.0000	0.0065	0.02
449	0.1685	0.1957	-0.0272	0.0049	0.0000		
549	0.1904	0.2188	-0.0284	-0.0272	0.0008		
649	0.2063	0.2280	-0.0217	-0.0284	0.0005		
749	0.2369	0.2755	-0.0386	-0.0217	0.0015		
849	0.2452	0.2839	-0.0387	-0.0386	0.0015		
949	0.2123	0.2226	-0.0103	-0.0387	0.0001		
1049	0.1755	0.1478	0.0277	-0.0103	0.0008		
1149	0.1536	0.1283	0.0253	0.0277	0.0006		

	ERRORS FOR	4 TH PERIOD	ABS	-0.01	SQ(ABS)	0.0099	RMS	0.03
1249	0.1378	0.1104	0.0274	0.0007				
150	0.1260	0.0945	0.0315	0.0010				
250	0.1172	0.0775	0.0397	0.0016				
	ERRORS FOR	4 TH PERIOD	ABS	-0.01	SQ(ABS) <td>0.0099</td> <td>RMS</td> <td>0.03</td>	0.0099	RMS	0.03
350	0.1303	0.1381	-0.0078	0.0001				
450	0.1688	0.1884	-0.0196	0.0004				
550	0.1982	0.2036	-0.0054	0.0000				
650	0.2119	0.2085	-0.0034	0.0000				
750	0.2272	0.2629	-0.0357	0.0013				
850	0.2000	0.2636	-0.0313	0.0010				
950	0.1627	0.2296	-0.0296	0.0009				
1050	0.1437	0.1706	0.0021	0.0007				
1150	0.1299	0.1171	0.0261	0.0007				
1251	0.1195	0.1017	0.0282	0.0008				
1351	0.1116	0.0922	0.0273	0.0007				
1451	0.0658	0.0658	0.0458	0.0021				
	ERRORS FOR	5 TH PERIOD	ABS	0.00	SQ(ABS) <td>0.0079</td> <td>RMS</td> <td>0.03</td>	0.0079	RMS	0.03
351	0.1288	0.0760	0.0528	0.0028				
451	0.1296	0.1249	0.0050	0.0000				
551	0.1252	0.1127	0.0125	0.0002				
651	0.1266	0.1150	0.0150	0.0002				
751	0.1318	0.1381	0.0045	0.0000				
851	0.1235	0.1488	-0.0045	0.0000				
951	0.1066	0.1171	0.0064	0.0000				
1051	0.0872	0.1171	0.0064	0.0000				
1151	0.0722	0.0557	0.0310	0.0010				
1252	0.0613	0.0531	0.0331	0.0004				
1352	0.0530	0.0461	0.0452	0.0002				
1452	0.0530	0.0360	0.0170	0.0003				
	ERRORS FOR	6 TH PERIOD	ABS	0.19	SQ(ABS) <td>0.0067</td> <td>RMS</td> <td>0.02</td>	0.0067	RMS	0.02
352	0.0429	0.0307	0.0129	0.0001				
452	0.0505	0.0970	-0.0167	0.0003				
552	0.0543	0.0837	-0.0285	0.0007				
652	0.0622	0.0893	-0.0294	0.0009				
752	0.0777	0.1082	-0.0271	0.0007				
852	0.0850	0.1084	-0.0305	0.0009				
952	0.0778	0.0689	-0.0334	0.0002				
1052	0.0728	0.0689	-0.0089	0.0001				
1152	0.0687	0.0522	0.0199	0.0004				
1253	0.0654	0.0522	0.0160	0.0003				
1353	0.0627	0.0471	0.0183	0.0003				
1453	0.0627	0.0306	0.0321	0.0010				
	ERRORS FOR	7 TH PERIOD	ABS	-0.04	SQ(ABS) <td>0.0059</td> <td>RMS</td> <td>0.02</td>	0.0059	RMS	0.02
353	0.0695	0.0284	-0.0411	0.0017				
453	0.0702	0.0798	-0.0096	0.0001				
553	0.0660	0.0774	-0.0114	0.0001				
653	0.0638	0.0766	-0.0128	0.0002				
753	0.0612	0.1000	-0.0388	0.0015				
853	0.0656	0.1165	-0.0509	0.0026				
953	0.0687	0.0933	-0.0246	0.0006				
1053	0.0622	0.0553	0.0069	0.0000				
1153	0.0582	0.0440	0.0142	0.0002				

ERRORS FOR	8 TH PERIOD	ABS#		SQ(ABS)#	RMS#	0.03
1253	0.0549	0.0381	0.0168	0.0003		
154	0.0521	0.0339	0.0182	0.0003		
254	0.0498	0.0239	0.0259	0.0007		
ERRORS FOR 8 TH PERIOD ABS#						
354	0.0426	0.0164	-0.0262	0.0007		
454	0.0269	0.0365	0.0096	0.0001		
554	0.0254	0.0394	-0.0140	0.0002		
654	0.0199	0.0323	-0.0124	0.0002		
754	0.0152	0.0307	-0.0155	0.0002		
854	0.0138	0.0279	-0.0141	0.0000		
954	0.0145	0.0184	-0.0039	0.0000		
1054	0.0155	0.0171	-0.0016	0.0000		
1154	0.0158	0.0131	0.0027	0.0000		
1254	0.0161	0.0117	0.0044	0.0000		
1354	0.0165	0.0110	0.0055	0.0000		
1454	0.0168	0.0073	0.0095	0.0001		
ERRORS FOR 9 TH PERIOD ABS#						
355	0.0126	0.0039	-0.0087	0.0001	0.0017	0.01
455	0.0064	0.0106	0.0042	0.0000		
555	0.0050	0.0093	-0.0043	0.0000		
655	0.0037	0.0029	-0.0008	0.0000		
755	0.0033	0.0089	-0.0056	0.0000		
855	0.0030	0.0093	-0.0065	0.0000		
955	0.0032	0.0111	-0.0079	0.0001		
1055	0.0037	0.0072	-0.0035	0.0000		
1155	0.0039	0.0036	0.0003	0.0000		
1255	0.0042	0.0037	0.0005	0.0000		
1355	0.0045	0.0032	0.0013	0.0000		
1455	0.0047	0.0029	0.0018	0.0000		
ERRORS FOR 10 TH PERIOD ABS#						
356	0.0039	0.0014	-0.0025	0.0000	0.0003	0.00
456	0.0022	0.0077	0.0052	0.0000		
556	0.0015	0.0058	-0.0036	0.0000		
656	0.0015	0.0037	-0.0022	0.0000		
756	0.0010	0.0050	-0.0049	0.0000		
856	0.0008	0.0035	-0.0027	0.0000		
956	0.0009	0.0029	-0.0020	0.0000		
1056	0.0010	0.0020	-0.0010	0.0000		
1156	0.0011	0.0016	-0.0005	0.0000		
1256	0.0012	0.0021	-0.0009	0.0000		
1356	0.0014	0.0023	-0.0009	0.0000		
1456	0.0015	0.0019	-0.0004	0.0000		
ERRORS FOR 11 TH PERIOD ABS#						
357	0.0012	0.0010	-0.0002	0.0000	0.0001	0.00
457	0.0007	0.0015	-0.0003	0.0000		
557	0.0006	0.0013	-0.0007	0.0000		
657	0.0005	0.0013	-0.0008	0.0000		
757	0.0006	0.0007	-0.0003	0.0000		
857	0.0010	0.0008	-0.0018	0.0001		
957	0.0016	0.0190	-0.0174	0.0003		
1057	0.0021	0.0144	-0.0123	0.0002		
1157	0.0024	0.0111	-0.0087	0.0001		

	12 TH PERIOD	ABS	ABS	SQ(ABS)	RMS	0.01
1257	0.0028	0.0082	-0.0054	0.0000	0.0006	
158	0.0032	0.0065	-0.0033	0.0000		
258	0.0037	0.0043	-0.0006	0.0000		
	ERRORS FOR	12 TH PERIOD	ABS	SQ(ABS)	RMS	0.01
358	0.0053	0.0105	-0.0052	0.0000		
458	0.0080	0.0295	-0.0215	0.0005		
558	0.0138	0.0394	-0.0256	0.0007		
658	0.0242	0.0580	-0.0338	0.0011		
758	0.0403	0.0790	-0.0381	0.0015		
858	0.0646	0.0953	-0.0407	0.0019		
958	0.0790	0.0746	-0.0156	0.0002		
1058	0.0634	0.0747	-0.0043	0.0000		
1158	0.0639	0.0563	0.0076	0.0001		
1258	0.0565	0.0459	0.0106	0.0001		
1358	0.0505	0.0388	0.0117	0.0001		
1458	0.0458	0.0352	0.0106	0.0001		
	ERRORS FOR	13 TH PERIOD	ABS	SQ(ABS)	RMS	0.02
359	0.0525	0.0560	-0.0035	0.0000	0.0053	
459	0.0871	0.0887	-0.0136	0.0002		
559	0.0807	0.1083	-0.0276	0.0008		
659	0.0941	0.1197	-0.0256	0.0007		
759	0.1150	0.1658	-0.0408	0.0026		
859	0.1260	0.1757	-0.0497	0.0025		
959	0.1114	0.1362	-0.0248	0.0006		
1059	0.0907	0.0898	0.0009	0.0000		
1159	0.0799	0.0730	0.0069	0.0000		
1259	0.0716	0.0671	0.0045	0.0000		
1359	0.0650	0.0581	0.0059	0.0000		
1459	0.0598	0.0428	0.0170	0.0003		
159	0.0598	0.0428	0.0170	0.0003		
	ERRORS FOR	14 TH PERIOD	ABS	SQ(ABS)	RMS	0.03
360	0.0703	0.0654	0.0049	0.0000	0.0077	
460	0.0943	0.1073	-0.0130	0.0002		
560	0.1087	0.1179	-0.0092	0.0001		
660	0.1214	0.1309	-0.0095	0.0001		
760	0.1421	0.1700	-0.0239	0.0008		
860	0.1589	0.2072	-0.0483	0.0023		
960	0.1460	0.1730	-0.0270	0.0007		
1060	0.1191	0.1089	0.0102	0.0001		
1160	0.1013	0.0793	0.0220	0.0005		
1260	0.0881	0.0647	0.0234	0.0005		
1360	0.0781	0.0589	0.0192	0.0004		
1460	0.0704	0.0504	0.0300	0.0004		
1560	0.0704	0.0504	0.0300	0.0004		
	ERRORS FOR	15 TH PERIOD	ABS	SQ(ABS)	RMS	0.02
361	0.0704	0.0565	-0.004	0.0002	0.0061	
461	0.0828	0.0888	0.0139	0.0003		
561	0.0920	0.0962	-0.0134	0.0002		
661	0.1051	0.1077	-0.0157	0.0002		
761	0.1145	0.1314	-0.0263	0.0007		
861	0.1091	0.1404	-0.0259	0.0007		
961	0.0925	0.1250	-0.0159	0.0003		
1061	0.0829	0.0875	-0.0050	0.0000		
1161	0.0829	0.0669	0.0160	0.0003		

1261	0.0754	0.0623	0.0131	0.0002	0.02
162	0.0695	0.0589	0.0106	0.0001	
262	0.0648	0.0423	0.0223	0.0003	
ERRORS FOR 16 TH PERIOD ABS=					
362	0.0716	0.0669	0.0047	0.0000	
462	0.0887	0.1065	-0.0178	0.0003	
562	0.1107	0.1155	-0.0048	0.0000	
662	0.1356	0.1378	-0.0022	0.0000	
762	0.1568	0.1722	-0.0054	0.0000	
862	0.1744	0.1820	-0.0076	0.0001	
962	0.1597	0.1685	-0.0088	0.0001	
1062	0.1371	0.1094	0.0277	0.0008	
1162	0.1171	0.0835	0.0336	0.0011	
1262	0.1025	0.0760	0.0265	0.0017	
1362	0.0815	0.0591	0.0274	0.0005	
1462	0.0631	0.0519	0.0312	0.0010	
ERRORS FOR 17 TH PERIOD ABS=					
363	0.0956	0.0739	0.0217	0.0005	0.02
463	0.1226	0.1029	0.0197	0.0004	
563	0.1190	0.0957	0.0233	0.0005	
663	0.1124	0.0949	0.0175	0.0003	
763	0.1102	0.1070	0.0032	0.0000	
863	0.1071	0.1205	-0.0134	0.0002	
963	0.0970	0.1048	-0.0078	0.0001	
1063	0.0854	0.0654	0.0200	0.0004	
1163	0.0779	0.0481	0.0294	0.0009	
1263	0.0719	0.0471	0.0248	0.0006	
1363	0.0671	0.0417	0.0254	0.0006	
1463	0.0632	0.0303	0.0329	0.0011	
ERRORS FOR 18 TH PERIOD ABS=					
364	0.0507	0.0283	0.0224	0.0005	0.02
464	0.0275	0.0114	-0.0039	0.0000	
564	0.0227	0.0192	0.0035	0.0000	
664	0.0163	0.0133	0.0030	0.0000	
764	0.0103	0.0150	-0.0047	0.0000	
864	0.0070	0.0120	-0.0050	0.0000	
964	0.0066	0.0137	-0.0071	0.0001	
1064	0.0074	0.0076	-0.0007	0.0000	
1164	0.0074	0.0053	0.0021	0.0000	
1264	0.0078	0.0049	0.0029	0.0000	
1364	0.0083	0.0047	0.0036	0.0000	
1464	0.0088	0.0034	0.0054	0.0000	
ERRORS FOR 19 TH PERIOD ABS=					
365	0.0071	0.0035	0.0036	0.0000	0.01
465	0.0048	0.0097	-0.0049	0.0000	
565	0.0038	0.0059	-0.0021	0.0000	
665	0.0061	0.0151	-0.0110	0.0001	
765	0.0061	0.0316	-0.0255	0.0007	
865	0.0121	0.0425	-0.0304	0.0009	
965	0.0176	0.0460	-0.0284	0.0008	
1065	0.0177	0.0266	-0.0089	0.0001	
1165	0.0177	0.0202	-0.0025	0.0000	

1265	0.0178	0.0159	0.0019	0.0000	0.0026	RMS=	0.01
166	0.0179	0.0174	0.0005	0.0000			
166	0.0160	0.0144	0.0036	0.0000			
ERRORS FOR 20 TH PERIOD ABS=							
366	0.0221	0.0193	0.0028	0.0000			
466	0.0274	0.0369	-0.0095	0.0001			
566	0.0382	0.0554	-0.0172	0.0003			
666	0.0515	0.0728	-0.0213	0.0005			
766	0.0619	0.0918	-0.0199	0.0004			
866	0.0861	0.1308	-0.0347	0.0012			
966	0.0989	0.1098	-0.0109	0.0001			
1066	0.0836	0.0662	-0.0174	0.0003			
1166	0.0718	0.0480	0.0238	0.0006			
1266	0.0628	0.0451	0.0177	0.0003			
1366	0.0558	0.0377	0.0181	0.0003			
1466	0.0502	0.0288	0.0214	0.0005			
ERRORS FOR 21 TH PERIOD ABS=							
367	0.0328	0.0392	0.0136	0.0002	0.0045	RMS=	0.02
467	0.0604	0.0508	0.0096	0.0001			
567	0.0515	0.0447	0.0068	0.0000			
667	0.0461	0.0472	-0.0011	0.0000			
767	0.0455	0.0573	-0.0118	0.0001			
867	0.0344	0.0631	-0.0087	0.0001			
967	0.0604	0.0679	-0.0075	0.0001			
1067	0.0552	0.0486	0.0066	0.0000			
1167	0.0503	0.0333	0.0172	0.0003			
1267	0.0467	0.0350	0.0113	0.0001			
1367	0.0435	0.0350	0.0113	0.0001			
1467	0.0408	0.0295	0.0113	0.0001			
ERRORS FOR 22 TH PERIOD ABS=							
368	0.0473	0.0375	0.0098	0.0001	0.0013	RMS=	0.01
468	0.0547	0.0558	-0.0011	0.0000			
568	0.0481	0.0584	-0.0103	0.0001			
668	0.0505	0.0603	-0.0098	0.0001			
768	0.0630	0.0807	-0.0177	0.0003			
868	0.0834	0.0994	-0.0160	0.0003			
968	0.0845	0.0986	-0.0177	0.0002			
1068	0.0694	0.0617	0.0141	0.0001			
1168	0.0604	0.0486	0.0118	0.0001			
1268	0.0574	0.0487	0.0047	0.0000			
1368	0.0478	0.0462	0.0016	0.0000			
1468	0.0432	0.0309	0.0123	0.0002			
ERRORS FOR 23 TH PERIOD ABS=							
369	0.0484	0.0451	-0.0032	0.0000	0.0014	RMS=	0.01
469	0.0708	0.0745	-0.0014	0.0000			
569	0.0821	0.0770	-0.0037	0.0000			
669	0.1182	0.1193	-0.0051	0.0000			
769	0.1534	0.1723	-0.0011	0.0000			
869	0.1449	0.1465	-0.0189	0.0004			
969	0.1183	0.1948	-0.0016	0.0006			
1069	0.1026	0.0767	0.0235	0.0007			
1169			0.0259				

24 TH PERIOD	ABS	0.10	SQ(ABS)	0.0031	RMS	0.02
1269	0.0909	0.0704	0.0205	0.0004		
170	0.0598	0.0820	0.0222	0.0005		
270	0.0750	0.0507	0.0243	0.0006		
ERRORS FOR 24 TH PERIOD						
370	0.0862	0.0744	0.0118	0.0001		
470	0.1087	0.1035	0.0052	0.0000		
570	0.1157	0.1019	0.0138	0.0002		
670	0.1356	0.1269	0.0087	0.0001		
770	0.1622	0.1650	0.0028	0.0000		
870	0.1788	0.1746	-0.0042	0.0000		
970	0.1610	0.1553	0.0057	0.0000		
1070	0.1303	0.0949	0.0354	0.0013		
1170	0.1081	0.0712	0.0369	0.0014		
1270	0.0920	0.0322	0.0322	0.0010		
171	0.0801	0.0505	0.0296	0.0009		
271	0.0709	0.0430	0.0279	0.0008		
ERRORS FOR 25 TH PERIOD						
371	0.0764	0.0577	0.021	0.0058		0.02
471	0.0816	0.0749	0.0187	0.0004		
571	0.0837	0.0661	0.0155	0.0002		
671	0.0851	0.0668	0.0169	0.0003		
771	0.0814	0.0847	0.0104	0.0000		
871	0.0728	0.0975	-0.0161	0.0003		
971	0.0611	0.0728	0.0102	0.0001		
1071	0.0481	0.0588	0.0023	0.0000		
1171	0.0386	0.0403	0.0078	0.0001		
1271	0.0314	0.0360	0.0027	0.0000		
172	0.0259	0.0302	-0.0046	0.0000		
272	0.0259	0.0302	-0.0043	0.0000		
ERRORS FOR 26 TH PERIOD						
372	0.0176	0.0461	0.03	0.0014		0.01
472	0.0218	0.0513	0.0285	0.0009		
572	0.0217	0.0338	-0.0021	0.0001		
672	0.0212	0.0323	-0.0021	0.0001		
772	0.0208	0.0372	-0.0064	0.0003		
872	0.0199	0.0371	-0.0172	0.0003		
972	0.0241	0.0351	-0.0110	0.0001		
1072	0.0321	0.0251	0.0070	0.0000		
1172	0.0424	0.0185	0.0187	0.0003		
173	0.0475	0.0188	0.0236	0.0006		
273	0.0524	0.0159	0.0316	0.0010		
373	0.0524	0.0119	0.0405	0.0016		
ERRORS FOR 27 TH PERIOD						
373	0.0577	0.0135	-0.00	0.0062		0.02
473	0.0555	0.0191	0.0442	0.0020		
573	0.0517	0.0121	0.0364	0.0013		
673	0.0439	0.0477	0.0194	0.0004		
773	0.0501	0.0735	-0.0338	0.0000		
873	0.0638	0.0735	-0.0234	0.0005		
973	0.0652	0.0839	-0.0201	0.0004		
1073	0.0601	0.0952	-0.0300	0.0009		
1173	0.0522	0.1005	-0.0404	0.0016		
		0.0527	-0.0005	0.0000		

1273	0.0459	0.0512	-0.0053	0.0000			
1274	0.0408	0.0436	-0.0028	0.0000			
1275	0.0367	0.0274	-0.0093	0.0001			
ERRORS FOR 28 TH PERIOD ABS= 0.02 RMS= 0.0073 SQ(ABS)= 0.0013							
374	0.0294	0.0336	-0.0042	0.0000			
474	0.0321	0.0725	-0.0404	0.0016			
574	0.0486	0.0783	-0.0297	0.0009			
674	0.0609	0.0873	-0.0264	0.0007			
774	0.0681	0.1050	-0.0369	0.0014			
874	0.0740	0.1189	-0.0449	0.0020			
974	0.0740	0.1168	-0.0428	0.0018			
1074	0.0712	0.1089	-0.0377	0.0014			
1174	0.0730	0.1045	-0.0085	0.0001			
1274	0.0745	0.0509	0.0236	0.0006			
1375	0.0759	0.0532	0.0227	0.0005			
1475	0.0770	0.0408	0.0362	0.0013			
ERRORS FOR 29 TH PERIOD ABS= 0.17 RMS= 0.0123 SQ(ABS)= 0.003							
375	0.0866	0.0577	0.0289	0.0008			
475	0.0847	0.0751	0.0096	0.0001			
575	0.0940	0.0790	0.0150	0.0002			
675	0.1060	0.0860	0.0200	0.0004			
775	0.1061	0.1022	0.0039	0.0000			
875	0.1078	0.1143	-0.0065	0.0000			
975	0.1028	0.1431	-0.0403	0.0016			
1075	0.0895	0.1418	-0.0523	0.0027			
1175	0.0800	0.1022	-0.0222	0.0005			
1275	0.0727	0.0631	0.0096	0.0001			
1376	0.0668	0.0709	-0.0041	0.0000			
1476	0.0621	0.0732	-0.0111	0.0001			
ERRORS FOR 30 TH PERIOD ABS= 0.05 RMS= 0.0067 SQ(ABS)= 0.002							
376	0.0711	0.0784	-0.0073	0.0001			
476	0.0808	0.0983	-0.0175	0.0003			
576	0.0934	0.1158	-0.0224	0.0005			
676	0.1114	0.1150	-0.0036	0.0000			
776	0.1007	0.1109	-0.0102	0.0001			
876	0.1079	0.1131	-0.0372	0.0014			
976	0.0866	0.1114	-0.0248	0.0006			
1076	0.0921	0.0822	0.0099	0.0001			
1176	0.0763	0.0630	0.0135	0.0002			
1276	0.0649	0.0572	0.0077	0.0001			
1377	0.0561	0.0509	0.0052	0.0000			
1477	0.0492	0.0436	0.0056	0.0000			
ERRORS FOR 31 TH PERIOD ABS= 0.08 RMS= 0.0034 SQ(ABS)= 0.002							

SIMULATION ERRORS(TOTAL): ABSOLUTE ERROR# 0.22 SQUARE OF ABSOL.# 0.15 ROOT MEAN SQ. ERROR# 0.02

MESILLA VALLEY, N. MEX.

CALC. RIVER LEAKAGE (AC-FT/MO)

	1967	1968	1969	1970	1971	1972	1973	1974	1975	1976
JAN	0.0	186.5	148.1	331.9	-418.9	495.5	46.7	303.2	1375.5	296.9
FEB	0.0	157.6	191.1	626.1	-548.3	235.6	47.5	200.8	1352.7	94.1
MAR	0.0	2331.7	6509.3	5471.3	-8384.4	2302.0	243.6	14678.5	6724.7	955.3
APR	0.0	1006.6	5343.1	3876.0	-5184.4	566.5	2287.6	9889.1	7828.4	1380.2
MAY	0.6	1225.5	4812.0	4245.8	-5617.8	5298.4	2371.8	8820.3	8424.6	1115.0
JUN	3319.7	12510.6	7798.0	6334.8	-7288.4	5214.8	23749.1	7027.8	8027.2	8295.7
JUL	3933.5	16379.8	6884.0	4349.3	-6448.8	305.8	37330.1	15676.0	7871.8	1095.5
AUG	2039.5	11574.4	8663.8	1301.1	-568.4	305.4	11107.8	33380.5	5232.5	3206.5
SEP	251.3	17.7	1204.0	307.9	-493.9	22.4	1107.6	35540.4	2662.7	26.7
OCT	355.3	80.3	430.9	154.5	-259.8	450.8	63.3	3580.6	3552.7	32.6
NOV	377.3		1660.9							
DEC										

PERCENTAGE OF YEARLY RIVER LEAKAGE INTO ALLUVIAL AQUIFER:

	10.0	5.1	16.9	12.2	-20.3	63.9	6.9	30.9	20.6	2.6
--	------	-----	------	------	-------	------	-----	------	------	-----

OBS. SURFACE DIVERSION (AC-FT/MO)

	1967	1968	1969	1970	1971	1972	1973	1974	1975	1976
JAN	42.0	140.0	0.0	2765.0	0.0	0.0	0.0	0.0	124.0	803.0
FEB	53564.0	38533.0	0.0	50867.0	0.0	0.0	0.0	0.0	1869.0	17997.0
MAR	20365.0	25543.0	34014.0	37779.0	22837.0	35763.0	28573.0	36116.0	32914.0	46384.0
APR	19464.0	19745.0	31427.0	38319.0	24843.0	14045.0	34098.0	41216.0	342217.0	461225.0
MAY	20373.0	41492.0	28564.0	52017.0	37280.0	61338.0	30615.0	42629.0	39952.0	51901.0
JUN	36916.0	47652.0	76173.0	52315.0	48677.0	7898.0	49954.0	60947.0	49920.0	537539.0
JUL	40136.0	47652.0	89025.0	70567.0	48677.0	2834.0	50579.0	49837.0	491914.0	63339.0
AUG	36241.0	55884.0	39025.0	39475.0	50774.0	2634.0	68790.0	44342.0	61515.0	568086.0
SEP	2520.0	33856.0	31047.0	4173.0	22229.0	6963.0	50858.0	32526.0	38279.0	39909.0
OCT	1310.0	2972.0	1400.0	1211.0	1120.0	40.0	4680.0	158.0	40220.0	3019.0
NOV										
DEC										

PERCENTAGE OF YEARLY RIVER FLOW THAT IS DIVERTED:

	50.8	55.7	58.9	58.2	57.0	47.8	55.4	53.6	55.2	59.0
--	------	------	------	------	------	------	------	------	------	------

MESQUILLA VALLEY, N. MEX.
BLANET-CRIDDLE

	CONSUMPTIVE USE (AC-FT/MO)											
1967	363.5	1968	1969	1970	1971	1972	1973	1974	1975	1976		
JAN	205.3	213.9	226.8	383.1	279.7	138.7	0.0	102.9	341.8	353.3		
FEB	119.2	150.6	99.4	149.3	160.4	160.4	7.4	114.0	97.9	44.0		
MAR	140.2	430.2	226.8	857.7	1304.2	1694.2	29.4	771.0	504.0	333.6		
APR	194.5	1350.3	161.3	157.4	3084.2	3084.2	47.3	1718.0	390.3	602.6		
MAY	182.6	332.8	208.9	335.4	1272.9	1308.4	43.3	1508.5	390.2	216.7		
JUN	282.9	384.8	328.6	583.8	1627.3	1538.3	26.3	1815.2	164.1	290.8		
JUL	282.9	251.9	201.6	345.8	1272.9	1308.4	37.3	1508.5	193.1	249.7		
AUG	14.5	670.2	1738.2	1980.7	4228.3	8874.4	34.3	3633.9	37.3	402.9		
SEPT	661.4	671.1	482.7	469.4	1922.9	6874.4	17.4	1132.7	193.1	74.7		
OCT	236.4	671.1	482.7	469.4	1922.9	6874.4	17.4	1132.7	193.1	74.7		
NOV	236.4	671.1	482.7	469.4	1922.9	6874.4	17.4	1132.7	193.1	74.7		
DEC	125.0	200.2	0.0	289.7	133.5	123.0	3.3	329.7	148.0	204.6		

CALC. GROUNDWATER PUMPAGE (AC-FT/MO)

	CONSUMPTIVE USE (AC-FT/MO)											
1967	703.6	1968	1969	1970	1971	1972	1973	1974	1975	1976		
JAN	410.0	427.8	453.6	766.2	589.4	477.4	0.0	205.6	495.7	0.0		
FEB	0.0	230.0	0.0	0.0	0.0	321.0	14.7	228.0	920.7	0.0		
MAR	186.4	1295.8	1498.6	1031.6	6520.3	4483.1	394.3	996.2	2999.7	0.0		
APR	181.2	5787.9	3596.6	2648.3	24208.3	24208.3	1181.5	2072.5	1788.7	0.0		
MAY	452.9	5413.5	3484.3	4848.3	6520.3	6520.3	1255.2	967.4	2050.2	0.0		
JUN	362.8	2418.9	0.0	3533.3	24208.3	24208.3	3780.1	2611.4	4004.8	0.0		
JUL	0.0	0.0	0.0	0.0	3884.0	3884.0	1249.8	2021.6	4924.3	0.0		
AUG	0.0	0.0	0.0	0.0	882.5	3189.2	0.0	0.0	0.0	0.0		
SEPT	0.0	1139.8	7084.9	11797.2	6864.9	9919.2	0.0	6410.5	1004.0	0.0		
OCT	4007.5	1007.6	4915.0	16279.4	5853.0	27531.0	17835.6	5896.8	5733.0	8167.4		
NOV	250.0	400.4	0.0	579.4	267.0	5247.0	648.0	158.0	297.4	409.2		
DEC	250.0	400.4	0.0	579.4	267.0	5247.0	648.0	158.0	297.4	409.2		

MESILLA VALLEY, N.MEX. HYDROSALINITY MODEL RESULTS FOR THE WATER YEAR 1967

MONTH	AG HEAD (PRD: FT)	DRAIN FLOW (PRO:AC-FT)	LEAS. OR LEAS. CHEM (OBS:MG/L)	AG. CHEM (PRD:MG/L)	EL PASO OR (PDI:AC-FT)	EL PASO OR (OBS:AC-FT)	EL PA CHEM (PRD:MG/L)	EL PA CHEM (OBS:MG/L)
JAN	3824.38	6058.82	0.00	0.00	0.00	0.00	0.00	0.00
FEB	3824.19	5447.88	0.00	0.00	0.00	0.00	0.00	0.00
MAR	3824.28	5737.75	0.00	0.00	0.00	0.00	0.00	0.00
APR	3824.53	6557.79	0.00	0.00	0.00	0.00	0.00	0.00
MAY	3824.24	5530.50	42150.00	1626.00	25004.92	24209.00	755.41	831.00
JUN	3824.04	5010.21	47930.00	1632.66	29807.89	25298.00	689.55	805.00
JUL	3824.01	4939.81	70059.00	1637.31	34773.15	31663.00	851.17	756.00
AUG	3824.34	5912.20	79419.00	1646.95	41261.74	36122.00	934.96	692.00
SEP	3824.53	5558.01	56609.00	1646.95	24886.53	27475.00	1034.97	822.00
OCT	3824.36	5297.94	5030.00	1662.97	8250.62	5094.00	1477.10	144.00
NOV	3824.20	5486.80	3890.00	1664.07	7781.46	2522.00	1479.15	1662.00
DEC	3824.06	5072.18	3770.00	1664.79	8464.69	2809.00	1486.93	1665.00
YEARLY AVERAGES:	3824.26	5696.99		1647.62	22653.88	19399.00	1083.65	1087.00

HYDRAULIC RESPONSE TIME = 2.9 MO.
 SOLUTE RESPONSE TIME = 36.4 YRS.

ERRORS: DIFF BETWEEN PREDICTED AND OBSERVED

MONTH	OPRED-OBS (AC-FT)	CHPRED-CMOBS (MG/L)	TONSOF SALT AT EL PASO (PRED)	TONS (OBS)	DIFF
JAN	0.0	0.0	0.0	0.0	0.00
FEB	0.0	0.0	0.0	0.0	0.00
MAR	0.0	0.0	0.0	0.0	0.00
APR	0.0	0.0	0.0	0.0	0.00
MAY	1795.9	-75.6	26727.1	27371.0	-643.91
JUN	4509.2	-115.5	27964.7	27707.7	257.33
JUL	3110.2	-195.2	40269.2	32567.7	7701.53
AUG	5139.7	243.0	52487.0	34008.7	18478.31
SEP	-2158.5	213.0	35043.3	30727.1	4316.14
OCT	3156.6	-12.8	16131.9	10000.0	6131.00
NOV	5259.5	-182.8	15659.8	5702.8	9957.01
DEC	5655.7	-198.1	17124.3	6439.7	10684.61
RMS ERRORS:	4122.6	160.2			9156.7

MESILLA VALLEY, N.MEX.; HYDROSALINITY MODEL RESULTS FOR THE WATER YEAR 1968

MONTH	AO HEAD (PRD; FT)	DRAIN FLOW (PRD; AC-FT)	LEAS OR (OB; AC-FT)	LEAS CHEM (OB; MG/L)	AO CHEM (PRD; MG/L)	EL PASO OR (PD; AC-FT)	EL PASO OR (OB; AC-FT)	EL PA CHEM (PRD; MG/L)	EL PA CHEM (OBS; MG/L)
JAN	3823.92	4725.48	1690.00	1118.00	1664.69	6329.02	2778.00	1526.18	1629.00
FEB	3823.82	4432.27	3220.00	1140.00	1664.66	7354.69	1878.00	1456.18	1669.00
MAR	3824.08	5135.12	8410.00	1622.00	1664.71	48380.25	39553.00	1768.43	847.00
APR	3824.35	5943.22	45220.00	721.00	1660.51	24613.48	27271.00	947.46	924.00
MAY	3824.10	5221.77	90560.00	662.00	1679.72	27951.15	23465.00	851.19	926.00
JUN	3824.21	5487.53	79640.00	628.00	1679.55	52045.06	41408.00	713.83	770.00
JUL	3824.62	6848.10	78930.00	514.00	1689.46	37199.49	46582.00	821.93	776.00
AUG	3825.16	9059.57	45090.00	649.00	1673.62	19838.80	40721.00	855.27	772.00
SEP	3825.20	9179.57	3320.00	747.00	1689.89	17862.90	23895.00	1123.10	972.00
OCT	3824.80	7540.30	2060.00	954.00	1691.21	7596.77	5000.00	1591.23	1540.00
NOV	3824.54	6564.43	2060.00	953.00	1691.21	7291.03	5227.00	1540.80	1503.00
DEC	3824.30	5801.36	1570.00						
YEARLY AVERAGES:		6328.19			1676.31	23115.72	22034.00	1153.92	1154.25

HYDRAULIC RESPONSE TIME = 2.9 MO.
 SOLUTE RESPONSE TIME = 21.9 YRS.

ERRORS: DIFF BETWEEN PREDICTED AND OBSERVED RIVER FLOW & IDS AT EL PASO

MONTH	OPRED-QOBS (AC-FT)	CMFRED-COBS (MG/L)	TONS OF SALT AT EL PASO (PRED)	TONS (OBS)	DIFF
JAN	3551.0	-102.8	13141.8	6157.0	6984.82
FEB	5476.7	-212.8	14571.1	4264.5	10306.65
MAR	9827.3	-218.6	50580.0	45380.5	15000.47
APR	-2657.5	-23.9	31741.5	32833.5	-2542.01
MAY	4486.2	-74.8	32369.8	29562.7	2807.12
JUN	10637.1	-56.2	50545.7	43379.8	7165.94
JUL	-9382.5	93.3	41599.2	46138.4	-4539.12
AUG	-9795.0	79.3	35986.5	42992.5	-7006.04
SEP	-4056.2	151.1	30314.3	31599.0	-1285.67
OCT	1232.9	111.2	17664.3	13891.4	3772.85
NOV	2596.8	324.0	16444.4	10659.9	5784.50
DEC	2064.0	37.8	15284.4	10688.7	4595.71
RMS ERRORS:	6283.6	101.5			5670.6

MESILLA VALLEY, N.MEX.; HYDROSALINITY MODEL RESULTS FOR THE WATER YEAR 1969

MONTH	AO HEAD (PRDI: FT)	DRAIN FLOW (PRDI: AC-FT)	LEAS. OR (OBI: AC-FT)	LEAS. CHEM (OBI: MG/L)	AO, CHEM (PRDI: MG/L)	EL PASO OR (PD: AC-FT)	EL PASO OR (OBI: AC-FT)	EL PA. CHEM (PRDI: MG/L)	EL PA. CHEM (OBI: MG/L)
JAN	3824.10	5192.08	875.00	897.00	1691.70	5919.96	4687.00	1594.11	1653.00
FEB	3823.92	4696.72	540.00	933.00	1691.63	5115.31	2571.00	1625.49	1831.00
MAR	3824.12	5256.64	92460.00	520.00	1691.59	37197.33	4230.00	685.58	750.00
APR	3824.73	7257.83	62990.00	505.00	1681.54	33477.69	3372.00	760.94	807.00
MAY	3824.84	7696.13	56990.00	517.00	1681.75	33310.04	3097.00	803.30	811.00
JUN	3825.14	8916.29	101800.00	500.00	1681.70	46819.34	5273.00	725.61	671.00
JUL	3825.25	123130.00	126690.00	515.00	1662.83	51592.83	6771.00	603.90	608.00
AUG	3826.61	16660.09	1426644.00	515.00	1662.85	50441.61	69224.00	894.12	621.00
SEP	3825.95	15737.81	12850.94	1213.00	1668.91	25371.61	36607.00	1288.91	894.00
OCT	3825.62	11148.77	3940.00	1080.00	1678.68	13288.78	10827.00	1645.14	1448.00
NOV	3825.36	9877.76	9810.00	1080.00	1681.63	18027.08	7180.00	1410.21	1594.00
DEC	3825.36	9877.76	9810.00	1080.00	1681.63	18027.08	7180.00	1410.21	1594.00
YEARLY AVERAGES:	3825.23	9843.91			1681.11	27701.10	30450.58	1154.77	1089.50

HYDRAULIC RESPONSE TIME = 2.9 MO.
SOLUTE RESPONSE TIME = 11.9 YRS.

ERRORS: DIFF BETWEEN PREDICTED AND OBSERVED
RIVER FLOW & IDS AT EL PASO

MONTH	QPRD-OBS (AC-FT)	CHPRD-CMOBS (MG/L)	TONS OF SALT AT EL PASO TONS(PRED)	TONS (OBS)	DIFF
JAN	1232.0	-58.9	12837.3	10541.0	2296.38
FEB	2574.3	-205.5	11379.1	6404.8	4974.34
MAR	-5112.7	-64.4	34693.4	43170.4	-8476.98
APR	-245.3	-46.1	34659.1	37026.0	-2367.41
MAY	363.0	-7.6	32219.5	34147.0	-1972.50
JUN	-3371.7	54.6	46221.5	48194.6	-1972.75
JUL	-16222.9	195.9	56429.5	56014.6	414.87
AUG	-18802.0	273.9	61361.0	58504.1	2856.97
SEP	-11235.4	394.9	44492.0	44526.1	-34.04
OCT	3014.1	197.1	31013.3	21359.5	9653.84
NOV	6465.8	36.0	31272.9	14637.9	16635.0
DEC	10847.1	14.2	34587.6	13637.1	20950.51
RMS ERRORS:	9029.6	174.2			8487.8

MESILLA VALLEY, N.MEX.; HYDROSALINITY MODEL RESULTS FOR THE WATER YEAR 1970

MONTH	AQ HEAD (PRDI FT)	DRAIN FLOW (PRDIAC-FT)	LEAS OR (OBTAC-FT)	LEAS, CHEM (OBTMG/L)	AQ, CHEM (PRDIAC-FT)	EL PASO OR (PDIAC-FT)	EL PASO OR (OBTAC-FT)	EL PA CHEM (PRDIAC/L)	EL PA CHEM (OBTMG/L)
JAN	3825.14	8906.45	2710.00	1118.00	1682.04	11284.56	5598.00	1563.17	1561.00
FEB	3824.96	8148.55	7880.00	1184.00	1681.93	12637.11	5572.00	1505.07	1339.00
MAR	3825.24	9365.57	95540.00	520.00	1683.69	48567.44	46386.00	1744.40	645.00
APR	3825.75	11803.17	69430.00	505.00	1673.09	39577.85	38232.00	853.36	743.00
MAY	3825.90	12572.85	72980.00	515.00	1674.02	42988.89	40803.00	853.98	750.00
JUN	3826.29	14725.95	92490.00	500.00	1694.32	50283.29	49775.00	851.25	737.00
JUL	3826.26	17618.3A	124040.00	456.00	1654.83	63008.51	67260.00	791.21	676.00
AUG	3827.05	19426.95	106080.00	449.00	1642.70	50590.64	54217.00	907.38	718.00
SEP	3826.75	17494.08	51570.00	743.00	1641.50	28101.79	31751.00	1302.22	925.00
OCT	3826.19	14157.13	6640.00	1088.00	1657.23	16321.99	7087.00	1581.73	1448.00
NOV	3825.74	11745.76	2500.00	1162.00	1659.79	12876.69	3933.00	1516.06	1530.00
DEC	3825.38	9998.40	2650.00	1129.00	1660.57	12323.85	3891.00	1511.39	1590.00
YEARLY AVERAGES:									
	3825.93	12996.86			1664.64	32377.23	29292.50	1175.10	1055.17

HYDRAULIC RESPONSE TIME = 2.9 MO.
 SOLUTE RESPONSE TIME = 11.4 YRS.

ERRORS: DIFF BETWEEN PREDICTED AND OBSERVED
 RIVER FLOW & TDS AT EL PASO

MONTH	OPRED-OBS (AC-FT)	CHPRED-CMOBS (MG/L)	TONSOF SALT AT EL PASO (PRED)	TONS (OBS)	DIFF
JAN	5685.6	2.2	2399.6	11891.29	1348.43
FEB	7065.1	166.1	25877.2	10150.1	15726.31
MAR	2181.4	199.4	49188.7	40706.1	8482.63
APR	1345.8	110.4	45951.0	38648.1	7302.91
MAY	1485.9	104.0	49947.6	41635.1	8311.84
JUN	1486.3	110.4	57503.8	46902.3	10601.53
JUL	4251.5	115.2	67827.5	61860.9	5966.64
AUG	3638.4	189.4	62355.9	52963.0	9492.92
SEP	3638.2	177.3	49799.4	39965.0	9834.36
OCT	9235.0	133.7	35125.2	13961.9	21163.30
NOV	8944.9	86.1	28312.7	8185.0	20127.71
DEC	8432.8	-48.6	25844.8	8417.0	17427.51
RMS ERRORS:	5675.6	155.6			13165.2

MESILLA VALLEY, N.MEX. HYDROSALINITY MODEL RESULTS FOR THE WATER YEAR 1971

MONTH	AO HEAD (PRDI: FT)	DRAIN FLOW (PRDI:AC-FT)	LEAS OR (OBI:AC-FT)	LEAS, CHEM (OBI:MG/L)	AO, CHEM (PRDI:MG/L)	EL PASO OR (PDI:AC-FT)	EL PASO OR (OBI:AC-FT)	EL PA CHEM (PRDI:MG/L)	EL PA CHEM (OBS:MG/L)
JAN	3825.09	8697.05	2060.00	1022.00	1660.45	11175.10	5008.00	1542.76	1560.00
FEB	3824.85	7703.12	2680.00	1029.00	1660.45	10926.98	3187.00	1503.58	1640.00
MAR	3825.10	8301.07	94150.00	544.00	1660.45	57998.57	43173.00	865.19	695.00
APR	3825.13	8721.07	50390.00	483.00	1659.47	39452.43	30835.00	908.13	743.00
MAY	3825.18	8862.34	53760.00	483.00	1654.18	42159.56	33571.00	886.55	743.00
JUN	3825.22	9089.36	73080.00	456.00	1646.56	52232.80	37097.00	823.05	687.00
JUL	3825.22	9245.97	8844.00	493.00	1642.52	50052.15	39612.00	863.56	699.00
AUG	3824.90	7911.48	72390.00	463.00	1642.51	35449.07	28088.00	906.64	854.00
SEP	3824.56	6637.38	23550.00	706.00	1648.82	11280.90	13143.00	1414.72	120.00
OCT	3824.11	5230.33	1320.00	272.00	1651.13	9560.51	4817.00	1300.59	120.00
NOV	3823.72	4195.36	1280.00	882.00	1651.31	6818.12	2849.00	1502.37	1480.00
DEC	3823.72	4195.36	1280.00	882.00	1651.31	5735.11	2776.00	1472.61	1470.00
YEARLY AVERAGES:		7786.51		1652.46		27736.77	20346.33	1166.73	1079.58

HYDRAULIC RESPONSE TIME = 2.9 MO.
 SOLUTE RESPONSE TIME = 16.5 YRS.

ERRORS: DIFF BETWEEN PREDICTED AND OBSERVED
 RIVER FLOW & TDS AT EL PASO

MONTH	OPRED-QOBS (AC-FT)	CHPRED-CMOBS (MG/L)	TONS OF SALT AT EL PASO (PRED)	TONS (OBS)	DIFF
JAN	6167.1	-17.2	23456.4	10629.2	12827.17
FEB	7740.0	-134.4	22382.9	7111.1	15271.72
MAR	14825.6	-170.2	68271.8	40823.4	27448.35
APR	18617.4	107.1	48531.0	33436.0	15094.98
MAY	8588.6	143.6	50852.6	33936.4	16916.20
JUN	15135.8	142.1	58916.8	34674.3	24242.44
JUL	10440.1	164.6	58806.8	37671.8	21134.99
AUG	17362.1	52.6	43727.1	32635.6	11091.49
SEP	-4743.5	290.6	21713.4	20027.4	1685.92
OCT	3969.1	27.4	16917.5	7930.0	8987.45
NOV	2959.1	29.6	13916.5	5736.8	8199.76
DEC			11545.2	5552.0	5993.22
RMS ERRORS:	8701.1	137.0			15836.1

MESILLA VALLEY, N.MEX.; HYDROBALINITY MODEL RESULTS FOR THE WATER YEAR 1972

MONTH	AO HEAD (PRD; FT)	DRAIN FLOW (PRD; AC-FI)	LEAS OR (OB; AC-FI)	LEAS, CHEM (OB; MG/L)	AO, CHEM (PRD; MG/L)	EL PASO OR (PDIAC-FI)	EL PASO OR (OB; AC-FI)	EL PA CHEM (PRD; MG/L)	EL PA CHEM (OB; MG/L)
JAN	3823.37	3414.49	776.00	904.00	1651.21	3695.01	2495.00	1594.58	1600.00
FEB	3823.05	2812.12	369.00	912.00	1651.09	2945.58	1740.00	1617.62	1690.00
MAR	3822.43	1907.65	77110.00	485.00	1650.58	16854.67	33845.00	1616.97	885.00
APR	3822.77	2370.11	34180.00	485.00	1658.79	9664.60	16312.00	773.34	807.00
MAY	3822.77	2361.47	18140.00	471.00	1659.79	6701.45	18885.00	883.94	961.00
JUN	3822.73	2306.14	16550.00	478.00	1655.06	5433.74	9885.00	981.80	1200.00
JUL	3822.70	2262.17	53150.00	441.00	1671.09	10923.74	7165.00	981.80	652.00
AUG	3822.63	2163.33	45470.00	478.00	1673.58	9078.53	24670.00	763.47	659.00
SEP	3822.93	2612.76	12140.00	757.00	1684.65	4484.17	21144.00	1437.90	1070.00
OCT	3823.41	3491.96	3790.00	824.00	1674.21	4847.52	15133.00	1437.90	1640.00
NOV	3823.66	4041.62	776.00	912.00	1674.99	4322.14	5937.00	1623.47	2060.00
DEC	3823.89	4600.84	706.00	706.00	1674.76	4856.05	809.00	1623.85	926.00
YEARLY AVERAGES:	3823.03	2862.06			1666.23	6996.69	11130.67	1159.52	1165.83

HYDRAULIC RESPONSE TIME = 2.9 MO.
SOLUTE RESPONSE TIME = 40.3 YRS.

ERRORS: DIFF BETWEEN PREDICTED AND OBSERVED
RIVER FLOW & TDS AT EL PASO

MONTH	OPRED-OBS (AC-FI)	CMRPRED-CMOBS (MG/L)	TONS(PRED)	TONS(OBS)	DIFF
JAN	1200.0	-5.4	8016.3	5431.3	2585.00
FEB	1205.6	-72.0	6482.8	4000.8	2481.93
MAR	-16290.3	-68.0	14148.0	4000.6	-17394.59
APR	-6663.4	-33.7	10151.8	31542.6	-17758.07
MAY	-2196.6	-71.2	8113.9	17909.9	-3520.13
JUN	-1731.3	-218.1	7228.3	11634.0	-4439.68
JUL	-13577.1	-20.4	10441.7	11698.9	-12785.05
AUG	-12065.5	104.5	9430.1	23226.7	-19527.55
SEP	-5955.8	227.5	7916.0	18957.7	-7282.39
OCT	285.5	-202.1	9483.5	115198.4	-1969.90
NOV	3405.1	-434.5	9558.5	112570.2	-1988.41
DEC	4047.1	-697.8	10728.6	11019.2	9709.33
RMS ERRORS:	7838.1	265.1			8475.2

MESILLA VALLEY, N. MEX.; HYDROSALINITY MODEL RESULTS FOR THE WATER YEAR 1973

MONTH	AQ HEAD (PRDI:FT)	DRAIN FLOW (PRDI:AC-FT)	LEAS OR LEAS (OBI:AC-FT)	LEAS. CHEM (OBI:MG/L)	AQ. CHEM (PRDI:MG/L)	EL PASO OR (PD:AC-FT)	EL PASO OR (OBI:AC-FT)	EL PASO OR (PRDI:MG/L)	EL PASO CHEM (OBI:MG/L)
JAN	3824.09	5154.48	680.00	618.00	1674.50	5787.78	465.00	1558.89	2040.00
FEB	3824.27	5688.74	692.00	824.00	1674.47	6333.21	344.00	1587.92	2090.00
MAR	3824.45	6267.06	63980.00	485.00	1674.44	39242.42	24938.00	674.96	552.00
APR	3824.37	6026.95	67400.00	507.00	1679.87	37641.87	30401.00	697.83	688.00
MAY	3824.24	5618.32	65380.00	552.00	1682.43	32995.77	31252.00	719.13	727.00
JUN	3823.95	4773.33	91760.00	522.00	1691.29	43708.23	42299.00	649.170	688.00
JUL	3824.19	5439.30	84180.00	456.00	1707.91	36803.47	53276.00	641.02	665.00
AUG	3824.64	6929.85	123380.00	485.00	1694.94	57770.78	58592.00	630.12	680.00
SEP	3824.68	7077.44	628770.00	529.00	1702.36	23759.30	38158.00	878.52	1080.00
OCT	3824.53	6529.50	62520.00	541.00	1710.79	7991.68	10008.00	1569.92	1478.00
NOV	3824.26	5665.25	812.00	118.00	1713.38	6514.62	5512.00	1635.76	1617.00
DEC	3824.03	4983.62	922.00	1191.00	1713.36	5842.30	4544.00	1636.58	1419.00
YEARLY AVERAGES:		5846.15			1693.31	25732.62	25149.08	1073.37	1143.67

HYDRAULIC RESPONSE TIME = 2.9 MO.
 SOLUTE RESPONSE TIME = 26.6 YRS.

ERRORS: DIFF BETWEEN PREDICTED AND OBSERVED

MONTH	OPRED-OBS (AC-FT)	CM-PRED-CMOBS (MG/L)	TONS OF SALT AT EL PASO (PRED)	TONS (OBS)	DIFF
JAN	5322.8	-481.1	12275.6	1290.6	10984.96
FEB	5969.2	-502.1	13682.5	978.2	12704.35
MAR	14304.4	123.0	36036.6	18728.9	17307.63
APR	6640.9	7.8	35168.8	28457.0	6711.79
MAY	6743.8	-7.8	37176.5	30911.8	6264.70
JUN	-590.8	-38.3	38635.3	41466.3	-2630.77
JUL	-16472.5	-24.0	32097.8	48202.1	-16104.30
AUG	-821.2	-49.9	49528.0	54207.6	-4678.99
SEP	-14398.7	-201.5	28398.7	56068.9	-27670.21
OCT	-22016.3	92.0	17070.1	20124.9	-3054.81
NOV	1002.6	18.8	14498.4	12426.9	2372.04
DEC	1298.3	217.6	13008.7	18772.9	4236.02
RMS ERRORS:	8390.7	223.6			12089.0

MESILLA VALLEY, N.MEX.; HYDROSALINITY MODEL RESULTS FOR THE WATER YEAR 1974

MONTH	AO HEAD (PRDI: FT)	DRAIN FLOW (PRDIAC-FT)	LEAS OR (OBIAC-FT)	LEAS CHEM (OBI:MG/L)	AO CHEM (PRDI:MG/L)	EL PASO OR (PD:AC-FT)	EL PASO OR (OBIAC-FT)	EL PA CHEM (PRDI:MG/L)	EL PA CHEM (OBI:MG/L)
JAN	3823.82	4435.21	980.00	1088.00	1713.34	5111.98	4270.00	1630.55	1574.00
FEB	3823.64	3987.38	647.00	1978.00	1713.22	4434.14	3712.00	1639.14	1640.00
MAR	3823.26	3194.88	101050.00	537.00	1713.13	35955.40	49334.00	636.95	1050.00
APR	3824.41	3489.94	168300.00	397.00	1730.61	22193.66	35722.00	606.71	840.00
MAY	3824.35	5282.08	74570.00	419.00	1716.54	27339.96	36454.00	669.69	1239.00
JUN	3824.35	6617.31	112780.00	419.00	1710.84	42412.25	56888.00	620.56	831.00
JUL	3824.77	7402.06	100500.00	441.00	1708.44	42389.02	47254.00	622.32	770.00
AUG	3824.93	8035.51	97590.00	412.00	1707.61	37901.80	44018.00	892.41	836.00
SEP	3824.85	8039.29	54520.00	412.00	1707.58	22227.84	25925.00	860.48	838.00
OCT	3824.90	7927.21	111600.00	436.00	1716.58	15633.64	7087.00	1057.22	1138.00
NOV	3824.94	8095.61	1230.00	1103.00	1713.57	9142.74	3932.00	1597.43	1592.00
DEC						8945.05	3891.00	1655.60	1505.00
YEARLY AVERAGES:		6186.56			1713.88	22890.65	26540.58	1027.42	1136.67

HYDRAULIC RESPONSE TIME = 2.9 MO.
 SOLUTE RESPONSE TIME = 15.3 YRS.

ERRORS: DIFF BETWEEN PREDICTED AND OBSERVED

MONTH	OPRED-OBS (AC-FT)	CHPRED-CMORS (MG/L)	TONS OF SALT AT EL PASO (PRED)	TONS (OBS)	DIFF
JAN	842.0	56.6	9144.2	9144.2	2196.40
FEB	722.1	-0.9	8282.6	8282.6	1606.12
MAR	-1378.6	-413.0	70477.1	70477.1	-39318.15
APR	-133628.3	-233.3	40825.0	40825.0	-22505.26
MAY	-9114.0	-569.3	61451.0	61451.0	-36540.60
JUN	-1475.8	-16.4	49302.5	49302.5	-13494.48
JUL	-4865.0	-156.7	52654.5	52654.5	-14456.84
AUG	-6116.2	-177.6	46114.5	46114.5	-10408.58
SEP	-2697.0	-24.5	29487.5	29487.5	12294.18
OCT	8546.8	-80.4	10972.5	10972.5	11514.82
NOV	5210.7	5.4	8516.7	8516.7	11353.93
DEC	5034.1	150.6	7967.3	7967.3	12181.62
RMS ERRORS:	9416.0	226.0			18950.9

MESILLA VALLEY, N.MEX.; HYDROSALINITY MODEL RESULTS FOR THE WATER YEAR 1975

MONTH	AO HEAD (PRD: FT)	DRAIN FLOW (PRD: AC-FT)	LEAS OR (OB: AC-FT)	LEAS CHEM (OB: MG/L)	AO CHEM (PRD: MG/L)	EL PASO OR (PD: AC-FT)	EL PASO OR (OB: AC-FT)	EL PA CHEM (PRD: MG/L)	EL PA CHEM (OBS: MG/L)
JAN	3824.98	8239.56	7200.00	412.00	1712.44	13540.04	9070.00	1203.97	1540.00
FEB	3825.01	8362.26	3580.00	463.00	1712.51	19720.72	6703.00	1537.89	1470.00
MAR	3825.23	9407.94	6550.00	412.00	1712.56	35319.62	32145.00	758.48	1106.00
APR	3825.21	9201.13	72260.00	426.00	1704.21	39405.74	38402.00	724.46	1836.00
MAY	3825.43	10205.60	79940.00	515.00	1702.01	42669.04	41462.00	798.91	676.00
JUN	3825.70	111510.96	92680.00	505.00	1690.28	45460.70	49137.00	812.59	868.00
JUL	3825.74	111524.09	96020.00	505.00	1690.28	38602.89	49144.00	812.59	868.00
AUG	3825.63	111709.08	104160.00	485.00	1698.87	42042.33	55385.00	824.51	1022.00
SEP	3825.63	111764.52	164160.00	551.00	1698.87	42042.33	55385.00	824.51	1022.00
OCT	3825.32	11721.46	5310.00	390.00	1707.77	31713.70	47040.00	952.86	1102.00
NOV	3825.09	8692.54	1380.00	390.00	1708.51	10745.70	13446.00	1582.17	1680.00
DEC	3824.89	7892.28	1710.00	412.00	1708.58	9249.56	8680.00	1570.48	1337.00
YEARLY AVERAGES:	3825.33	9802.62			1703.73	27348.21	30063.42	1094.97	1181.42

HYDRAULIC RESPONSE TIME = 2.9 MO
 SOLUTE RESPONSE TIME = 14.7 YRS.

ERRORS: DIFF BETWEEN PREDICTED AND OBSERVED RIVER FLOW & TDS AT EL PASO

MONTH	OPRED-OBS (AC-FT)	CHPRED-CMOBS (MG/L)	TONS OF SALT AT EL PASO (PRED)	TONS (OBS)	DIFF
JAN	4470.0	-336.0	22179.4	19003.6	3175.8
FEB	3017.7	67.9	20339.3	13406.0	6933.3
MAR	3174.6	-347.5	36447.8	48370.6	-11822.7
APR	1003.8	-111.5	38440.5	43679.0	-5238.4
MAY	1207.0	122.9	46379.2	38133.8	8245.4
JUN	4376.3	-55.4	50259.7	58855.1	-8595.4
JUL	-10541.1	-166.7	44918.6	58333.6	-2314.9
AUG	-13542.7	-55.5	47048.2	66399.5	-19351.3
SEP	-15337.0	-147.1	41199.4	70528.0	-29328.6
OCT	-2700.3	197.8	41331.3	30733.9	10597.4
NOV	591.8	233.5	20744.8	16456.9	4287.9
DEC	370.0	-144.0	19104.3	20079.7	-975.4
RMS ERRORS:	7059.8	184.0			13618.7

MESILLA VALLEY, N.MEX.; HYDROSALINITY MODEL RESULTS FOR THE WATER YEAR 1976

MONTH	AO HEAD (PRDI:FT)	DRAIN FLOW (PRDI:AC-FT)	LEAS OR LEAS, CHEM (OBS:MG/L)	AO, CHEM (PRDI:MG/L)	EL PASO OR (PRDI:AC-FT)	EL PASO OR (OBS:AC-FT)	EL PA CHEM (PRDI:MG/L)	EL PA CHEM (OBS:MG/L)
JAN	3824.73	7257.83	19980.00	1708.01	18138.67	15600.00	897.58	1190.00
FEB	3824.59	6746.27	21580.00	1709.65	10234.33	14198.00	1308.28	1288.00
MAR	3824.85	7728.04	82430.00	1715.31	42818.95	40200.00	653.78	1232.00
APR	3825.11	8781.97	85800.00	1706.12	47405.71	50279.00	663.96	630.00
MAY	3825.42	10148.36	102010.00	1698.05	60876.65	60697.00	650.55	777.00
JUN	3825.81	12100.65	95840.00	1695.86	53073.46	50099.00	727.11	1523.00
JUL	3825.58	87300.00	87300.00	1682.87	41077.60	47905.00	743.03	1822.00
AUG	3824.98	8241.92	109440.00	1709.72	48500.18	57382.00	659.92	875.00
SEP	3825.25	9403.00	50430.00	1721.90	20443.55	31511.00	1036.09	1551.00
OCT	3825.38	9999.95	4020.00	1706.10	10974.43	14340.00	1615.17	1242.00
NOV	3825.00	8308.38	1650.00	1707.31	9914.66	10475.00	1558.37	1382.00
DEC	3824.67	7051.05	1230.00	1707.28	8248.46	9999.00	1590.53	1358.00
YEARLY AVERAGES:								
	3825.11	8892.03		1705.68	30975.55	33557.08	1008.70	1155.83

HYDRAULIC RESPONSE TIME = 2.9 MO.
 SOLUTE RESPONSE TIME = 16.3 YRS.

ERRORS: DIFF BETWEEN PREDICTED AND OBSERVED
 RIVER FLOW & TDS AT EL PASO

MONTH	OPRED-OBS (AC-FT)	CMFRED-CMOBS (MG/L)	TONSOF SALT AT EL PASO TONS (PRED)	TONS (OBS)	DIFF
JAN	2538.7	-292.4	22150.9	25257.1	-3106.24
FEB	-3963.7	20.3	18216.9	24880.3	-6663.43
MAR	2618.9	-578.2	38087.3	67382.9	-29295.55
APR	-2873.3	14.0	42823.7	43026.3	-202.58
MAY	179.5	-125.4	53882.5	64165.4	-10282.94
JUN	2974.5	-195.9	53503.4	103810.6	-50307.21
JUL	-8827.4	-79.0	41526.5	53575.4	-12048.88
AUG	-8881.8	-215.1	43546.0	68311.9	-24765.90
SEP	-11067.5	-514.9	28818.3	66494.6	-37676.38
OCT	-3365.6	373.2	24116.4	24231.7	-115.27
NOV	-560.3	175.4	21021.4	19025.9	1325.50
DEC	-1750.5	232.5	17849.5	18474.3	-624.82
RMS ERRORS:	5073.2	366.6			22043.0

TOTAL RMS ERRORS:
 RMS ERROR IN RIVER FLOW = 7323.9
 RMS ERROR IN TDS = 213.4
 RMS ERROR IN TONS = 13769.0

Appendix F

Program to Estimate Porosity and River Leakage for Periods of

No Recharge

(Source: Pinzon, 1979)

```

00001    MAIN,    LN2K1S.FOR
00002    REAL    H, NP, NT, HO
00003    DIMENSION F(31,6), H(31,6), DH(31,6), G(31,6), X(31), XX(31), Y(31),
00004    1 XY(31), NM(31), CR(31), AP(31), PRED(31,6), DIFF(31,6), SDB(31,6),
00005    2 ENAS(31), RMS(31), BB(31), QDOBS(31,6)
00006    I=2
00007    I=3
00008    A=0.0755*9.05
00009    B=1.9*05
00010    HO=3825.0
00011    AREA=108626.0
00012    NRFC=31
00013    NT=0.0
00014    SX=0.0
00015    SXX=0.0
00016    SXY=0.0
00017    DO 1 J=1, NRFC
00018    READ(IR,5) NM(J)
00019    LIMIT=NM(J)
00020    WRITE(IW,6) J
00021    WRITE(IW,16) NM(J)
00022    NP(J)=NM(J)-2
00023    NT=NT+NP(J)
00024    DO 2 I=1, LIMIT
00025    READ(IR,15) MON, IYR, H(J,I), QDOBS(J,I)
00026    F(J,I)=6*ALOG(EXP(A*H(J,I)-HO))+I
00027    AREA=AREA+F(J,I)
00028    WRITE(IW,26) MON, IYR, H(J,I), QDOBS(J,I), F(J,I)
00029    F(J,I)=F(J,I)/AREA
00030    CONTINUE
00031    WRITE(IW,36)
00032    X(J)=0.0
00033    XX(J)=0.0
00034    XY(J)=0.0
00035    Y(J)=0.0
00036    LIMIT=NM(J)-1
00037    DO 3 I=2, LIMIT
00038    DH(J,I)=H(J,I+1)-H(J,I-1)
00039    G(J,I)=(F(J,I+1)+4*F(J,I)+F(J,I-1))/3.
00040    WRITE(IW,46) DH(J,I), G(J,I)
00041    X(J)=X(J)+DH(J,I)
00042    XX(J)=XX(J)+DM(J,I)**2
00043    XY(J)=XY(J)+G(J,I)
00044    CONTINUE
00045    WRITE(IW,56) X(J), XX(J), XY(J), Y(J)
00046    SY=SY+X(J)
00047    SXX=SXX+X(J)
00048    SXY=SXY+XY(J)
00049    SXX=XX+XX(J)
00050    SXY=SXY+XY(J)
00051    WRITE(IW,66) SX, SY, SXX, SXY
00052    PDR=0.218
00053    WRITE(IW,76) PDR
00054    DO 4 J=1, NRFC
00055    RB(J)=(Y(J)+PDR*H(J,I)+PDR*H(J,I))/NP(J)
00056    CR(J)=BB(J)/2.

```

```

00057 GRIV=AREA*QR(J)
00058 WRITE(IW,86)J,QR(J),QRIV
00059 EMS(J)=0
00060 DO I=1,2, LIMIT
00061 PRED(J,I)=POR*DH(J,I)+BB(J)
00062 DIFF(J,I)=G(J,I)-PRED(J,I)
00063 SQD(J,I)=DIFF(J,I)**2
00064 EMS(J)=EMS(J)+SQD(J,I)
00065 WRITE(IW,9)G(J,I),PRED(J,I),DIFF(J,I),SQD(J,I)
00066 RMS(J)=SQRT(EMS(J)/NP(J))
00067 *WRITE(IW,106)EMS(J),RMS(J)
00068 STOP
00069 *FORMAT(I2)
00070 *FORMAT(1H1,10X,13,'TH',3X,'RECESSION PERIOD'//)
00071 *FORMAT(2F10.4)
00072 *FORMAT(12F10.4)
00073 *FORMAT(2I2,2F10.2)
00074 *FORMAT(RX,W=U,3X,ORBS,MONTH OF RECESSION',//,3X,'MONTH',2X,'YEAR',3X,
00075 <'AVE. AU',W=U,3X,ORBS,MONTH OF RECESSION',//,3X,'MONTH',2X,'YEAR',3X,
00076 *FORMAT(4X,12.4X,19.2,4X,10.2,6X,16.1,9X,F10.4)
00077 *FORMAT(//,15X,'X=H(I+1),5X,'I=(DI/3)*(QD(I-1)+4*QD(I)+QD(I
00078 <+1))')
00079 *FORMAT(5X,F10.2,6X,F10.4)
00080 *FORMAT(//,40X,X(J)=F10.2,//,40X,'XX(J)=',F10.2,//,40X,'XY(J)=',
00081 *F10.2,//,40X,Y(J)=F10.4,//,40X,'SY=',F10.2,//,40X,'SXX=',F10.2,//,
00082 *FORMAT(40X,6X,SX=F10.2,//,40X,'SY=',F10.2,//,40X,'SXX=',F10.2,//,
00083 *FORMAT(5X,FINAL POROSITY IS ',3X,F10.5)
00084 *FORMAT(//,ACPT/NO,//)
00085 *FORMAT(//,ACPT/NO,//)
00086 *FORMAT(//,ACPT/NO,//)
00087 *FORMAT(//,ACPT/NO,//)
00088 *FORMAT(//,OBSERVED Y= ',F10.4,3X,'PREDICTED Y= ',F10.4,3X,
00089 *DIFFERENCE= ',F6.4,3X,'SUM OF SQ. OF DIFFS. = ',F6.4,3X,
00090 *FORMAT(//,SUM OF SQ. OF DIFFS. = ',F10.4,4X,'RMS= ',F10.4,4X,
00091 *FORMAT(//,SUM OF SQ. OF DIFFS. = ',F10.4,4X,'RMS= ',F10.4,4X,
00091 END

```

SUBPROGRAMS CALLED

SORT, EXP, ALOG.

SCALARS AND ARRAYS (** NO EXPLICIT DEFINITION = * NOT REFERENCED)

RMS	4	*AREA	40	*FCR	41	*SXX	42	*IW	43
DIFF	44	*NREC	336	H	767	*RH	667	*H	726
QR	727	*NPRED	7631	*SY	1323	*HO	770	*SXY	771
Y	1655	*SX	1654	*FMS	1657	*QRIV	1362	DH	1363
*J	2133	*S0002	2154	G	2155	*S0004	2151	*S0003	2152
*A	2116	*IYR	2255	F	2256	*S0000	2156	XY	2157
*NP	2646	*HO	3435	*IR	2647	*SQD	2315	XX	2607
*I	3143	*LIMIT	3435	*R	3436	NI	3141	*MON	3142

1TH RECESSION PERIOD				2TH RECESSION PERIOD			
5 MONTH OF RECESSION				5 MONTH OF RECESSION			
MONTH	YEAR	AVE. (PI)	OBS (AC-PT/MO)	MONTH	YEAR	AVE. (PI)	OBS (AC-PT/MO)
10	1946	3826.86	17520.0	10	1947	3826.84	14900.0
11	1946	3826.57	13620.0	11	1947	3826.38	11500.0
12	1946	3826.17	12190.0	1	1947	3826.18	9920.0
1	1947	3825.95	10570.0	2	1948	3825.86	9000.0
2	1947	3825.89	8400.0	2	1948	3825.79	7800.0

1TH RECESSION PERIOD				2TH RECESSION PERIOD			
5 MONTH OF RECESSION				5 MONTH OF RECESSION			
MONTH	YEAR	AVE. (PI)	OBS (AC-PT/MO)	MONTH	YEAR	AVE. (PI)	OBS (AC-PT/MO)
10	1946	3826.86	17520.0	10	1947	3826.84	14900.0
11	1946	3826.57	13620.0	11	1947	3826.38	11500.0
12	1946	3826.17	12190.0	1	1947	3826.18	9920.0
1	1947	3825.95	10570.0	2	1948	3825.86	9000.0
2	1947	3825.89	8400.0	2	1948	3825.79	7800.0

$X=H(I+1)-H(I-1)$ $Y=(DT/3)*(QD(I-1)+4*QD(I)+QD(I+1))$ $X=H(I+1)-H(I-1)$ $Y=(DT/3)*(QD(I-1)+4*QD(I)+QD(I+1))$
 -0.69 0.2967 -0.66 0.2892
 -0.57 0.2613 -0.52 0.2579
 -0.28 0.2391 -0.39 0.2318

$X(J)= -1.54$
 $X(J)= 0.80$
 $X(J)= 0.88$
 $Y(J)= 0.7991$

$X(J)= -1.57$
 $X(J)= 0.86$
 $X(J)= 0.81$
 $Y(J)= 0.7759$

$SX= -1.54$
 $SX= 0.80$
 $SXX= 0.88$
 $SXY= -0.42$

$SX= -3.11$
 $SX= 1.57$
 $SXX= 1.74$
 $SXY= -0.83$

3TH RECESSON PERIOD				4TH RECESSON PERIOD			
5 MONTH OF RECESSON				5 MONTH OF RECESSON			
MONTH	YEAR	AVE. (Ft)	OBS (AC-FI/MO)	MONTH	YEAR	AVE. (Ft)	OBS (AC-FI/MO)
10	1948	3826.95	17570.0	10	1949	3826.87	16060.0
11	1948	3826.58	13050.0	11	1949	3826.52	13940.0
11	1949	3826.25	12000.0	11	1950	3826.21	11990.0
1	1949	3825.89	11330.0	1	1950	3826.02	10270.0
2	1949	3825.86	8920.0	2	1950	3825.94	8419.0

CALC. DRAIN FLOW (AC-FI/MO)				CALC. DRAIN FLOW (AC-FI/MO)			
10			18803.1	10			18286.9
11			16468.5	11			16103.1
11			14504.3	11			14450.8
1			12537.9	1			13216.8
2			12337.0	2			12782.9

$X=H(I+1)-H(I-1)$ $Y=(DT/3)*(OD(I-1)+4*OD(I)+OD(I+1))$ $Y=(DT/3)*(OD(I-1)+4*OD(I)+OD(I+1))$
 -0.70 0.3046 0.2981
 -0.92 0.2670 0.2673
 -0.39 0.2361 0.2458

$X(J)= -1.78$ $X(J)= -1.43$
 $X(J)= -1.12$ $X(J)= -0.74$
 $X(J)= -0.46$ $X(J)= -0.40$
 $Y(J)= 0.8075$ $Y(J)= 0.8113$

$SX= -4.89$ $SX= -6.32$
 $SY= 21.866$ $SY= 3.19$
 $SXX= -1.32$ $SXX= 3.59$
 $SXY=$ $SXY= -1.72$

5TH RECESSON PERIOD

5	MONTH OF RECESSON	MONTH YEAR	AVE. AQ. W-L (Ft)	OBS (AC-FT/MO)	DRAIN FLOW (AC-FT/MO)	CALC (AC-FT/MO)	DRAIN FLOW (AC-FT/MO)	CALC (AC-FT/MO)
10	1950	3824.68	17447.0	17085.9	17085.9	17085.9	17085.9	17085.9
11	1950	3826.33	12766.0	14972.1	14972.1	14972.1	14972.1	14972.1
11	1950	3826.11	11046.0	13718.4	13718.4	13718.4	13718.4	13718.4
11	1951	3825.49	10020.0	12515.9	12515.9	12515.9	12515.9	12515.9
12	1951	3825.83	7150.0	12196.4	12196.4	12196.4	12196.4	12196.4

X=H(I+1)-H(I-1)
 -0.57
 -0.44
 -0.28

Y=(DT/3)*(Q(I-1)+4*Q(I)+Q(I+1))
 0.2783
 0.2527
 0.2331

X(J)= -1.29
 X(J)= 0.60
 X(J)= -0.34
 Y(J)= 0.7641

SX= 7.61
 SY= 3.66
 SXX= 4.19
 SXY= -2.05

6TH RECESSON PERIOD

5	MONTH OF RECESSON	MONTH YEAR	AVE. AQ. W-L (Ft)	OBS (AC-FT/MO)	DRAIN FLOW (AC-FT/MO)	CALC (AC-FT/MO)	DRAIN FLOW (AC-FT/MO)	CALC (AC-FT/MO)
10	1951	3825.82	7720.0	7720.0	7720.0	7720.0	7720.0	7720.0
11	1951	3825.74	6950.0	6950.0	6950.0	6950.0	6950.0	6950.0
11	1951	3824.76	5770.0	5770.0	5770.0	5770.0	5770.0	5770.0
11	1952	3824.69	5010.0	5010.0	5010.0	5010.0	5010.0	5010.0
12	1952	3824.59	3910.0	3910.0	3910.0	3910.0	3910.0	3910.0

X=H(I+1)-H(I-1)
 -1.06
 -0.37
 -0.17

Y=(DT/3)*(Q(I-1)+4*Q(I)+Q(I+1))
 0.2039
 0.1484
 0.1307

X(J)= -2.28
 X(J)= -1.25
 X(J)= -0.39
 Y(J)= 0.4830

SX= -9.89
 SY= 4.14
 SXX= 5.13
 SXY= -2.15

7TH RECESSION PERIOD				8TH RECESSION PERIOD			
5 MONTH OF RECESSION				5 MONTH OF RECESSION			
MONTH	YEAR	AVE. AQ. W-L (FT)	OBS. DRAIN FLOW (AC-FI/MO)	MONTH	YEAR	AVE. AQ. W-L (FT)	OBS. DRAIN FLOW (AC-FI/MO)
10	1952	3824.96	7480.0	10	1953	3824.46	6010.0
11	1953	3824.70	5720.0	11	1953	3824.40	4780.0
11	1953	3824.71	5720.0	11	1954	3824.77	4140.0
11	1953	3824.50	5720.0	12	1954	3824.53	3680.0
12	1953	3824.50	3320.0	12	1954	3823.95	2600.0

7TH RECESSION PERIOD				8TH RECESSION PERIOD			
CALC. DRAIN FLOW (AC-FI/MO)				CALC. DRAIN FLOW (AC-FI/MO)			
MONTH	YEAR	AVE. AQ. W-L (FT)	OBS. DRAIN FLOW (AC-FI/MO)	MONTH	YEAR	AVE. AQ. W-L (FT)	OBS. DRAIN FLOW (AC-FI/MO)
10	1952	3824.96	8156.8	10	1953	3824.46	6308.8
11	1953	3824.70	7620.8	11	1953	3824.40	6109.8
11	1953	3824.71	7120.8	11	1954	3824.77	5695.7
11	1953	3824.50	7120.8	12	1954	3824.53	5884.0
12	1953	3824.50	6433.7	12	1954	3823.95	4770.4

$X = H(I+1) - H(I-1)$ $Y = (DT/3) * (QR(I-1) + 4 * QR(I) + QR(I+1))$
 $X = H(I+1) - H(I-1)$ $Y = (DT/3) * (QR(I-1) + 4 * QR(I) + QR(I+1))$
 -0.26 0.1413 0.119
 -0.13 0.1334 0.1067
 -0.20 0.1300 0.1043

$X(U) = -0.59$ $X(U) = -0.59$
 $X(U) = 0.12$ $X(U) = 0.12$
 $X(U) = -0.08$ $X(U) = -0.08$
 $X(U) = 0.4047$ $X(U) = 0.3229$

$SX = -10.48$ $SX = -11.06$
 $SY = 4.64$ $SY = 3.17$
 $SXX = 6.57$ $SXX = 6.71$
 $SXY = -2.53$ $SXY = -2.58$

9TH RECESSION PERIOD				10TH RECESSION PERIOD					
5 MONTH OF RECESSION				5 MONTH OF RECESSION					
MONTH	YEAR	AVE. (ft)	OBS. DRAIN FLOW (AC-FI/MO)	CALC. DRAIN FLOW (AC-FI/MO)	MONTH	YEAR	AVE. (ft)	OBS. DRAIN FLOW (AC-FI/MO)	CALC. DRAIN FLOW (AC-FI/MO)
10	1954	3822.43	1860.0	1921.5	10	1955	3820.34	780.0	487.1
11	1954	3822.38	1420.0	1823.0	11	1955	3820.56	387.0	564.3
11	1954	3822.40	1270.0	1876.6	11	1955	3820.77	466.0	632.0
11	1955	3822.55	1200.0	2027.6	11	1955	3820.77	348.0	649.0
11	1955	3822.41	194.0	1888.5	12	1955	3820.65	310.0	603.2

$X=H(I+1)-H(I-1)$ $Y=(DT/3)*(QD(I-1)+4*QD(I)+QD(I+1))$
 $X=H(I+1)-H(I-1)$ $Y=(DT/3)*(QD(I-1)+4*QD(I)+QD(I+1))$
 -0.97 0.0344
 0.01 0.0350
 0.01 0.0369

$X(J)=0.15$
 $XY(J)=0.03$
 $Y(J)=0.1063$
 $SX= -10.91$
 $SX^2= 5.74$
 $SXY= -2.58$

$X(J)=0.570$
 $XY(J)=0.01$
 $Y(J)=0.0396$
 $SX= -10.38$
 $SX^2= 5.31$
 $SXY= -2.59$

11TH RECESSION PERIOD

5 MONTH OF RECESSION					
MONTH	YEAR	AVE. AQ (PI)	W=L	OBS. DRAIN FLOW (AC-PI/MO)	CALC. DRAIN FLOW (AC-PI/MO)
10	1956	3818.37		215.0	215.0
11	1956	3818.59		173.0	173.0
12	1956	3818.71		228.0	228.0
1	1957	3818.99		246.0	246.0
2	1957	3819.00		205.0	205.0

X=H(I+1)-H(I-1) Y=(DT/3)*(Q(I-1)+4*Q(I)+Q(I+1))

0.36 0.0027

0.40 0.0031

0.27 0.0038

X(J)= 1.03

X(J)= 0.36

X(J)= 0.00

Y(J)= 0.0093

SX= -9.35

SY= 5.13

SXX= 7.31

SXY= -2.56

12TH RECESSION PERIOD

5 MONTH OF RECESSION					
MONTH	YEAR	AVE. AQ (PI)	W=L	OBS. DRAIN FLOW (AC-PI/MO)	CALC. DRAIN FLOW (AC-PI/MO)
10	1957	3819.33		1564.0	1564.0
11	1957	3819.67		1201.0	1201.0
12	1957	3819.81		1880.0	1880.0
1	1958	3820.14		710.0	710.0
2	1958	3820.07		470.0	470.0

X=H(I+1)-H(I-1) Y=(DT/3)*(Q(I-1)+4*Q(I)+Q(I+1))

0.48 0.0056

0.50 0.0065

0.26 0.0076

X(J)= 1.24

X(J)= 1.53

X(J)= 0.53

Y(J)= 0.0197

SX= -8.11

SY= 5.34

SXX= 7.85

SXY= -2.57

13TH RECESSION PERIOD

5 MONTH OF RECESSION

MONTH	YEAR	AVE. (P)	AO, W-L	DRAIN FLOW {AC=PI/MO}	CALC {AC=PI/MO}	DRAIN FLOW {AC=PI/MO}	CALC {AC=PI/MO}	DRAIN FLOW {AC=PI/MO}	CALC {AC=PI/MO}
10	1958	3824.92		8110.0	8036.4	9760.0	9080.5	7837.8	
11	1958	3824.62		6119.0	6306.2	7930.0	7579.1	6550.6	
12	1958	3824.46		4050.0	5329.7	6310.0	6175.5		
1	1959	3824.27		4270.0		4650.0			
2	1959	3824.96		3820.0	4797.3	4650.0			

X=H(I+1)-H(I-1) Y=(DT/3)*(O(I-1)+4*O(I)+O(I+1))
 -0.47 0.1300
 -0.39 0.1164
 -0.50 0.1040

X(J)= -1.36
 X(J)= -0.62
 Y(J)= 0.3504

SX= 5.87
 SY= 5.89
 SXX= 8.79
 SXY= -2.79

14TH RECESSION PERIOD

5 MONTH OF RECESSION

MONTH	YEAR	AVE. (P)	AO, W-L	DRAIN FLOW {AC=PI/MO}	CALC {AC=PI/MO}	DRAIN FLOW {AC=PI/MO}	CALC {AC=PI/MO}
10	1959	3825.18		9760.0	9080.5	7837.8	
11	1959	3824.68		7930.0	7579.1	6550.6	
12	1959	3824.56		6310.0	6175.5		
1	1960	3824.42		4650.0			

X=H(I+1)-H(I-1) Y=(DT/3)*(O(I-1)+4*O(I)+O(I+1))
 -0.50 0.1458
 -0.32 0.1314
 -0.26 0.1223

X(J)= -1.08
 X(J)= 0.42
 Y(J)= 0.3995

SX= -10.55
 SY= 6.90
 SXX= 8.90
 SXY= -2.87

15TH RECESSION PERIOD

15TH RECESSION PERIOD		16TH RECESSION PERIOD		17TH RECESSION PERIOD					
MONTH	YEAR	AVE. AQ (F)	OBS. DRAIN FLOW (AC-FT/MO)	CALC. DRAIN FLOW (AC-FT/MO)	MONTH	YEAR	AVE. AQ (F)	OBS. DRAIN FLOW (AC-FT/MO)	CALC. DRAIN FLOW (AC-FT/MO)
10	1960	3825.93	11830.0	12729.2	10	1961	3425.26	9500.0	9433.3
11	1960	3825.28	8610.0	10640.3	11	1961	3825.00	7770.0	8319.8
1	1961	3825.01	7030.0	19522.8	12	1961	3824.87	6770.0	7738.5
2	1961	3824.80	5480.0	8360.9	1	1962	3824.71	6400.0	7189.4
				7527.6	2	1962	3824.58	4590.0	6755.9

X=H(I+1)-H(I-1)
 -0.55
 -0.51
 -0.28

Y=(DT/3)*(QD(I-1)+4*QD(I)+QD(I+1))
 0.1989
 0.1782
 0.1543

X(J)= -1.64
 X(J)= -0.91
 XY(J)= -0.28
 Y(J)= 0.5280

SX= -12.19
 SY= 6.61
 SXX= 9.84
 SXY= -3.11

Y=(DT/3)*(QD(I-1)+4*QD(I)+QD(I+1))
 0.1550
 0.1433
 0.1329

X=H(I+1)-H(I-1)
 -0.39
 -0.29
 -0.28

X(J)= -0.96
 X(J)= -0.14
 XY(J)= -0.14
 Y(J)= 0.4312

SX= -13.15
 SY= 7.05
 SXX= 10.12
 SXY= -3.31

17TH RECESSION PERIOD

17TH RECESSION PERIOD		18TH RECESSION PERIOD							
5 MONTH OF RECESSION		5 MONTH OF RECESSION							
MONTH	YEAR	AVE. AQ. W-L (FT)	OBS. DRAIN FLOW (AC-FI/MO)	CALC. DRAIN FLOW (AC-FI/MO)	MONTH	YEAR	AVE. AQ. W-L (FT)	OBS. DRAIN FLOW (AC-FI/MO)	CALC. DRAIN FLOW (AC-FI/MO)
10	1962	3626.29	1880.0	1719.4	10	1963	3825.13	7104.0	8864.7
11	1962	3825.67	8070.0	12405.8	11	1963	3825.89	5226.0	9279.8
12	1962	3825.62	8260.0	11128.4	12	1964	3825.74	5150.0	7837.8
1	1963	3825.41	7505.0	10118.4	1	1964	3825.50	3235.0	7300.9
2	1963	3825.07	5638.0	8610.3	2	1964	3825.50	3235.0	8443.7

$X = H(I+1) - H(I-1)$ $Y = (DT/3) * (QD(I-1) + 4 * QD(I) + QD(I+1))$ $Y = (DT/3) * (QD(I-1) + 4 * QD(I) + QD(I+1))$
 -0.67 0.2317 0.1529
 -0.86 0.2057 0.1440
 -0.85 0.1848 0.1334

$X(I) = -1.68$
 $X(I) = -0.36$
 $X(I) = -0.36$
 $Y(I) = 0.6222$

$X(I) = -0.25$
 $X(I) = -0.38$
 $Y(I) = 0.4303$

$SX = -14.83$
 $SY = 1.67$
 $SXX = 1.09$
 $SXY = -3.66$

$SX = -15.71$
 $SY = 1.10$
 $SXX = 1.76$
 $SXY = -3.78$

19TH RECESSION PERIOD 20TH RECESSION PERIOD

5 MONTH OF RECESSION 5 MONTH OF RECESSION

MONTH YEAR	AVE. AQ (PT)	W-L	OBS (AC-PI/MO)	DRAIN FLOW (AC-PI/MO)	CALC (AC-PI/MO)	DRAIN FLOW (AC-PI/MO)	OBS (AC-PI/MO)	W-L	AQ (PT)	MONTH YEAR	AVE. AQ (PT)	W-L	OBS (AC-PI/MO)	DRAIN FLOW (AC-PI/MO)	CALC (AC-PI/MO)	DRAIN FLOW (AC-PI/MO)
10 1964	321.10		874.0	807.8		2890.0	382.23		382.23	10 1965	382.23		2890.0	1673.8		
11 1964	321.28		570.0	945.2		2701.0	382.23		382.23	11 1965	382.23		2701.0	1815.1		
12 1964	321.48		270.0	1031.2		1983.0	382.23		382.23	1 1966	382.23		1983.0	1815.1		
1 1965	321.45		367.0	1011.0		1566.0	382.23		382.23	2 1966	382.23		1566.0	1750.1		

X=H(I+1)-H(I-1) Y=(DT/3)*(OP(I-1)+4*OP(I)+OP(I+1))

0.38 0.0173

0.15 0.0188

-0.03 0.0191

X(J)= 0.50

XY(J)= 0.01

Y(J)= 0.0552

SX= -15.21

SY= 1.122

SXY= -1.177

X(J)= 0.07

XY(J)= 0.05

Y(J)= 0.1011

SX= -15.14

SY= 1.259

SXY= -3.77

21TH RECESSON PERIOD

5 MONTH OF RECESSON

MONTH	YEAR	AVE. AQ. W-L (ft)	OBS. DRAIN FLOW (AC-FT/MO)	CALC. DRAIN FLOW (AC-FT/MO)
10	1966	3825.17	7188.0	9037.2
11	1966	3824.83	5213.0	7642.8
12	1966	3824.94	4864.0	6581.1
1	1967	3824.44	4096.0	6241.8
2	1967	3824.11	3132.0	5216.6

$X=H(I+1)-H(I-1)$
 $Y=(DT/3)*(QD(I-1)+4*QD(I)+QD(I+1))$
 -0.93
 -0.23
 -0.43
 0.1417
 0.1224
 0.1124

$X(J)= -1.45$
 $Y(J)= -0.73$
 $X(J)= -0.3779$

$SX= -16.59$
 $SY= 4.63$
 $SXX= 12.31$
 $SXY= -1.96$

22TH RECESSON PERIOD

5 MONTH OF RECESSON

MONTH	YEAR	AVE. AQ. W-L (ft)	OBS. DRAIN FLOW (AC-FT/MO)	CALC. DRAIN FLOW (AC-FT/MO)
10	1967	3824.17	5282.0	5392.5
11	1967	3823.66	3688.0	4797.3
12	1967	3823.86	3806.0	4533.2
1	1968	3823.87	3800.0	4559.1
2	1968	3823.57	3203.0	3834.7

$X=H(I+1)-H(I-1)$
 $Y=(DT/3)*(QD(I-1)+4*QD(I)+QD(I+1))$
 -0.31
 -0.09
 -0.29
 0.8893
 0.8844
 0.8816

$X(J)= -0.69$
 $Y(J)= 0.19$
 $X(J)= -0.06$
 $Y(J)= 0.2353$

$SX= -17.28$
 $SY= 8.89$
 $SXX= 12.50$
 $SXY= -4.02$

23TH RECESSION PERIOD					24TH RECESSION PERIOD						
5 MONTH OF RECESSION					5 MONTH OF RECESSION						
MONTH	YEAR	AVE. (Ff)	AQ W-L	OBS {AC-FI/HO}	DRAIN FLOW {AC-FI/HO}	CALC {AC-FI/HO}	OBS {AC-FI/HO}	AQ W-L	Ff	DRAIN FLOW {AC-FI/HO}	CALC {AC-FI/HO}
10	1968	3824.49		6704.0	5704.0	6409.7	1969	3825.74		10298.0	16778.6
11	1968	3824.45		5284.0	6273.1	6495.0	1969	3825.38		8236.0	9300.0
11	1968	3824.70		5698.0	5442.1	6495.0	1969	3825.23		7847.0	8736.8
11	1969	3824.00		5015.0	4806.4	4183.8	1970	3825.10		5504.0	7225.4
2	1969	3823.72		3356.0			1970	3824.72			

$X = H(I+1) - H(I-1)$
 $Y = (DT/3) * (OP(I-1) + 4 * OP(I) + OP(I+1))$
 $X = H(I+1) - H(I-1)$
 $Y = (DT/3) * (OP(I-1) + 4 * OP(I) + OP(I+1))$

-0.29
 -0.28
 -0.48
 0.1135
 0.1106
 0.0899

$X(J) = -1.22$
 $XX(J) = -0.52$
 $XY(J) = -0.12$
 $Y(J) = 0.3050$

$SX = -18.50$
 $SY = 73.01$
 $SXX = 13.01$
 $SXY = -4.04$

-0.51
 -0.28
 -0.51
 0.1876
 0.1716
 0.1580

$X(J) = -1.20$
 $XX(J) = -0.60$
 $XY(J) = -0.82$
 $Y(J) = 0.5166$

$SX = -19.80$
 $SY = 74.71$
 $SXX = 13.01$
 $SXY = -4.36$

25TH RECESSION PERIOD 26TH RECESSION PERIOD

5 MONTH OF RECESSION 5 MONTH OF RECESSION

MONTH	YEAR	AVE. (FT)	OBS (AC-FT/MO)	DRAIN FLOW (AC-FT/MO)	CALC (AC-FT/MO)	MONTH	YEAR	AVE. (FT)	OBS (AC-FT/MO)	DRAIN FLOW (AC-FT/MO)	CALC (AC-FT/MO)
10	1970	3826.07	10311.0	13492.2	13492.2	10	1971	3824.52	6383.0	6383.0	6312.1
11	1970	3825.97	7736.0	12946.7	12946.7	11	1971	3824.98	4379.0	4379.0	7079.1
1	1971	3825.43	6498.0	11076.8	11076.8	12	1971	3823.41	4489.0	4489.0	3288.1
2	1971	3824.73	5484.0	8950.6	8950.6	1	1972	3823.11	3911.0	3911.0	2915.0
			4671.0	7263.6	7263.6	2	1972	3823.11	3279.0	3279.0	

X=H(I+1)-H(I-1) Y=(DT/3)*(Q(I-1)+4*Q(I)+Q(I+1)) X=H(I+1)-H(I-1) Y=(DT/3)*(Q(I-1)+4*Q(I)+Q(I+1))

-0.46 0.2343 0.1187

-0.82 0.4222 0.1162

-0.88 0.1662 0.0710

X(J)= -2.16

SX(J)= 1.66

XY(J)= -0.42

Y(J)= 0.6036

X(J)= -2.55

SX(J)= -0.73

XY(J)= 0.3060

Y(J)= 0.3060

SX= -21.96

SY= 10.31

SXX= 15.27

SXY= -4.78

SX= -24.51

SY= 10.62

SXX= 16.06

SXY= -5.01

27TH RECESSON PERIOD

5 MONTH OF RECESSON

MONTH	YEAR	AVE. (FT)	AG, W-L	OBS (AC-FT/MO)	DRAIN FLOW (AC-FT/MO)	CALC (AC-FT/MO)	DRAIN FLOW (AC-FT/MO)
10	1972	3823.53		2730.0	3745.8		
11	1972	3823.53		2012.0	3658.7		
12	1972	3823.53		2047.0	4153.7		
1	1973	3823.53		1277.0	4878.9		
2	1973	3824.54		1288.0	6077.2		

X=H(I+1)-H(I-1) Y=(DT/3)*(QD(I-1)+4*QD(I)+QD(I+1))
 0.18 0.0692
 0.20 0.0773
 0.28 0.0913

X(J)= 1.36
 X(J)= 0.74
 Y(J)= 0.11
 Y(J)= 0.2377

SX= -23.15
 SY= 10.86
 SXX= 19.84
 SXY= -4.89

28TH RECESSON PERIOD

5 MONTH OF RECESSON

MONTH	YEAR	AVE. (FT)	AG, W-L	OBS (AC-FT/MO)	DRAIN FLOW (AC-FT/MO)	CALC (AC-FT/MO)	DRAIN FLOW (AC-FT/MO)
10	1973	3824.50		1093.0	3745.8		
11	1973	3824.50		5725.0	3658.7		
12	1973	3823.54		4732.0	4153.7		
1	1974	3823.54		2911.0	4878.9		
2	1974	3823.54		2911.0	6077.2		

X=H(I+1)-H(I-1) Y=(DT/3)*(QD(I-1)+4*QD(I)+QD(I+1))
 -0.53 0.0976
 -0.30 0.0883
 -0.43 0.0793

X(J)= -1.26
 X(J)= 0.56
 Y(J)= -0.831
 Y(J)= 0.2633

SX= -24.41
 SY= 10.74
 SXX= -5.01

29TH RECESSION PERIOD

5 MONTH OF RECESSION

MONTH	YEAR	AVE. (PI)	AD W-L	UBS (AC-PI/MO)	DRAIN FLOW (AC-PI/MO)	CALC (AC-PI/MO)	DRAIN FLOW (AC-PI/MO)	ORB	AD W-L	UBS (AC-PI/MO)	DRAIN FLOW (AC-PI/MO)	CALC (AC-PI/MO)	DRAIN FLOW (AC-PI/MO)
10	1974	3824.78		1704.0	1824.0	7451.5	15402.0	3825.24		1704.0	15402.0	9344.2	9344.2
11	1974	3824.62		574.0	1704.0	6724.4	12651.0	3824.88		574.0	12651.0	7877.8	7877.8
11	1975	3824.58		574.0	574.0	8610.3	7704.0	3824.68		574.0	7704.0	6910.1	6910.1
12	1975	3824.75		4436.0	4436.0	7338.3	7950.0	3824.59		4436.0	7950.0	6755.9	6755.9

$X=H(I+1)-H(I-1)$

-0.20
0.49
0.17

$Y=(DT/3)*(OD(I-1)+4*OD(I)+OD(I+1))$

0.1277
0.1300
0.1488

$X=H(I+1)-H(I-1)$

-0.76
-0.23
0.11

$Y=(DT/3)*(OD(I-1)+4*OD(I)+OD(I+1))$

0.1444
0.1237
0.1259

$X(J)=$
 $X(J)=$
 $X(J)=$
 $X(J)=$

0.427
0.27
0.06
0.4065

$X(J)=$
 $X(J)=$
 $X(J)=$
 $X(J)=$

-0.88
0.64
-0.12
0.3940

$SX=$
 $SXM=$
 $SXX=$
 $SXY=$

-23.99
11.43
10.63
-4.95

$SX=$
 $SXM=$
 $SXX=$
 $SXY=$

-24.87
11.8
11.28
-3.07

31TH RECESSON PERIOD

5 MONTH OF RECESSON

MONTH	YEAR	AVE. AQ. W-L (FT)	OBS DRAIN FLOW (AC-FT/MO)	CALC. DRAIN FLOW (AC-FT/MO)
10	1976	3825.27	8934.0	9477.9
11	1976	3824.45	6843.0	6275.1
12	1976	3824.43	6217.0	6208.3
1	1977	3824.22	5534.0	5542.3
2	1977	3823.94	4737.0	4743.5

$$X=H(I+1)-H(I-1) \quad Y=(DT/3)*(QD(I-1)+4*QD(I)+QD(I+1))$$

-0.84
-0.23
-0.49

X(J)= -1.56
XY(J)= -1.00
X²(J)= -0.18
Y(J)= 0.3393

SX= -26.43
SY= 12.56
SXX= 22.27
SXY= -5.25

FINAL POROSITY IS = 0.21800

1 TH RECESSON QR(J)= 0.076895 FT/MO = 8352.8 AC-FT/MO

BSERVED Y= 0.2967 PREDICTED Y= 0.0034 DIFFERENCE= .2933 SQUARE OF DIFF.= .0860
 BSERVED Y= 0.2613 PREDICTED Y= 0.0295 DIFFERENCE= .2318 SQUARE OF DIFF.= .0537
 BSERVED Y= 0.2391 PREDICTED Y= 0.0927 DIFFERENCE= .1463 SQUARE OF DIFF.= .0214

UM OF SQ. OF DIFFS.= 0.1612 RMS= 0.2318

2 TH RECESSION OR(J)= 0.072275 FT/MO = 7850.9 AC-FT/MO

BSERVED Y= 0.2862 PREDICTED Y= 0.0007 DIFFERENCE= .2855 SQUARE OF DIFF.= .0815
 BSERVED Y= 0.2579 PREDICTED Y= 0.0312 DIFFERENCE= .2268 SQUARE OF DIFF.= .0514
 BSERVED Y= 0.2318 PREDICTED Y= 0.0595 DIFFERENCE= .1722 SQUARE OF DIFF.= .0297
 UM OF SQ. OF DIFFS.= 0.1626 RMS= 0.2328

3 TH RECESSION OR(J)= 0.069902 FT/MO = 7593.2 AC-FT/MO

BSERVED Y= 0.3044 PREDICTED Y= -0.0128 DIFFERENCE= .3172 SQUARE OF DIFF.= .1006
 BSERVED Y= 0.2670 PREDICTED Y= -0.0106 DIFFERENCE= .2776 SQUARE OF DIFF.= .0771
 BSERVED Y= 0.2361 PREDICTED Y= 0.0548 DIFFERENCE= .1813 SQUARE OF DIFF.= .0329
 UM OF SQ. OF DIFFS.= 0.2105 RMS= 0.2649

4 TH RECESSION OR(J)= 0.083255 FT/MO = 9043.6 AC-FT/MO

BSERVED Y= 0.2981 PREDICTED Y= 0.0292 DIFFERENCE= .2689 SQUARE OF DIFF.= .0723
 BSERVED Y= 0.2673 PREDICTED Y= 0.0575 DIFFERENCE= .2098 SQUARE OF DIFF.= .0440
 BSERVED Y= 0.2459 PREDICTED Y= 0.1011 DIFFERENCE= .1447 SQUARE OF DIFF.= .0209
 UM OF SQ. OF DIFFS.= 0.1373 RMS= 0.2139

5 TH RECESSION OR(J)= 0.080485 FT/MO = 8742.8 AC-FT/MO

OBSERVED Y= 0.2783 PREDICTED Y= 0.0367 DIFFERENCE= .2416 SQUARE OF DIFF.= .0584
 OBSERVED Y= 0.2527 PREDICTED Y= 0.0651 DIFFERENCE= .1876 SQUARE OF DIFF.= .0352
 OBSERVED Y= 0.2331 PREDICTED Y= 0.0999 DIFFERENCE= .1332 SQUARE OF DIFF.= .0177

UM OF SQ. OF DIFFS.= 0.1113 RMS= 0.1926

6 TH RECESSION QR(J)= -0.002336 FT/MO = -253.8 AC-FT/MO

OBSERVED Y= 0.2039 PREDICTED Y= -0.2358 DIFFERENCE= .4397 SQUARE OF DIFF.= .1933
 OBSERVED Y= 0.1484 PREDICTED Y= -0.2336 DIFFERENCE= .3819 SQUARE OF DIFF.= .1459
 OBSERVED Y= 0.1307 PREDICTED Y= -0.0417 DIFFERENCE= .1724 SQUARE OF DIFF.= .0297

UM OF SQ. OF DIFFS.= 0.3689 RMS= 0.3507

7 TH RECESSION QR(J)= 0.046007 FT/MO = 4997.6 AC-FT/MO

OBSERVED Y= 0.1413 PREDICTED Y= 0.0353 DIFFERENCE= .1059 SQUARE OF DIFF.= .0112
 OBSERVED Y= 0.1334 PREDICTED Y= 0.0637 DIFFERENCE= .0698 SQUARE OF DIFF.= .0049
 OBSERVED Y= 0.1300 PREDICTED Y= 0.0484 DIFFERENCE= .0816 SQUARE OF DIFF.= .0067

UM OF SQ. OF DIFFS.= 0.0227 RMS= 0.0871

8 TH RECESSION QR(J)= 0.032742 FT/MO = 3556.6 AC-FT/MO

OBSERVED Y= 0.1118 PREDICTED Y= 0.0241 DIFFERENCE= .0878 SQUARE OF DIFF.= .0077

OBSERVED Y= 0.1067 PREDICTED Y= 0.0502 DIFFERENCE= .0565 SQUARE OF DIFF.= .0032
 OBSERVED Y= 0.1043 PREDICTED Y= -0.0043 DIFFERENCE= .1086 SQUARE OF DIFF.= .0118
 SUM OF SQ. OF DIFFS.= 0.0227 RMS= 0.0870

9 TH RECESSION QR(J)= 0.023165 FT/MO = 2516.3 AC-FT/MO

OBSERVED Y= 0.0344 PREDICTED Y= 0.0398 DIFFERENCE=-.0054 SQUARE OF DIFF.=0.0000
 OBSERVED Y= 0.0350 PREDICTED Y= 0.0834 DIFFERENCE=-.0483 SQUARE OF DIFF.= .0023
 OBSERVED Y= 0.0369 PREDICTED Y= 0.0485 DIFFERENCE=-.0116 SQUARE OF DIFF.= .0001
 SUM OF SQ. OF DIFFS.= 0.0025 RMS= 0.0289

10 TH RECESSION QR(J)= 0.024856 FT/MO = 2700.0 AC-FT/MO

OBSERVED Y= 0.0104 PREDICTED Y= 0.1347 DIFFERENCE=-.1244 SQUARE OF DIFF.= .0155
 OBSERVED Y= 0.0115 PREDICTED Y= 0.0955 DIFFERENCE=-.0840 SQUARE OF DIFF.= .0071
 OBSERVED Y= 0.0118 PREDICTED Y= 0.0345 DIFFERENCE=-.0227 SQUARE OF DIFF.= .0005
 SUM OF SQ. OF DIFFS.= 0.0230 RMS= 0.0876

11 TH RECESSION QR(J)= 0.038978 FT/MO = 4234.0 AC-FT/MO

OBSERVED Y= 0.0027 PREDICTED Y= 0.1564 DIFFERENCE=-.1537 SQUARE OF DIFF.= .0236
 OBSERVED Y= 0.0031 PREDICTED Y= 0.1652 DIFFERENCE=-.1621 SQUARE OF DIFF.= .0263

BSERVED Y= 0.0035 PREDICTED Y= 0.1368 DIFFERENCE=-.1333 SQUARE OF DIFF.= .0178
UM OF SQ. OF DIFFS.= 0.0677 RMS= 0.1502

12 TH RECESSION QR(J)= 0.048341 FT/MO = 5251.0 AC-FT/MO

BSERVED Y= 0.0056 PREDICTED Y= 0.2013 DIFFERENCE=-.1957 SQUARE OF DIFF.= .0383
BSERVED Y= 0.0065 PREDICTED Y= 0.2057 DIFFERENCE=-.1992 SQUARE OF DIFF.= .0397
BSERVED Y= 0.0076 PREDICTED Y= 0.1534 DIFFERENCE=-.1457 SQUARE OF DIFF.= .0212

UM OF SQ. OF DIFFS.= 0.0992 RMS= 0.1819

13 TH RECESSION QR(J)= 0.008990 FT/MO = 976.5 AC-FT/MO

BSERVED Y= 0.1300 PREDICTED Y= -0.0845 DIFFERENCE= .2145 SQUARE OF DIFF.= .0460
BSERVED Y= 0.1164 PREDICTED Y= -0.0670 DIFFERENCE= .1835 SQUARE OF DIFF.= .0337
BSERVED Y= 0.1040 PREDICTED Y= -0.0910 DIFFERENCE= .1950 SQUARE OF DIFF.= .0380

UM OF SQ. OF DIFFS.= 0.1177 RMS= 0.1981

14 TH RECESSION QR(J)= 0.027336 FT/MO = 2969.4 AC-FT/MO

BSERVED Y= 0.1458 PREDICTED Y= -0.0543 DIFFERENCE= .2001 SQUARE OF DIFF.= .0400
BSERVED Y= 0.1314 PREDICTED Y= -0.0151 DIFFERENCE= .1464 SQUARE OF DIFF.= .0214
BSERVED Y= 0.1223 PREDICTED Y= -0.0020 DIFFERENCE= .1243 SQUARE OF DIFF.= .0155

UM OF SQ. OF DIFFS. = 0.0769 RMS = 0.1602

15 TH RECESSION QR(J) = 0.028586 FT/MO = 3105.2 AC-FT/MO

BSERVED Y = 0.1989 PREDICTED Y = -0.0843 DIFFERENCE = .2834 SQUARE OF DIFF. = .0803

BSERVED Y = 0.1752 PREDICTED Y = -0.0540 DIFFERENCE = .2292 SQUARE OF DIFF. = .0525

BSERVED Y = 0.1549 PREDICTED Y = -0.0475 DIFFERENCE = .2024 SQUARE OF DIFF. = .0410

UM OF SQ. OF DIFFS. = 0.1738 RMS = 0.2407

16 TH RECESSION QR(J) = 0.036989 FT/MO = 4018.0 AC-FT/MO

BSERVED Y = 0.1550 PREDICTED Y = -0.0110 DIFFERENCE = .1660 SQUARE OF DIFF. = .0276

BSERVED Y = 0.1433 PREDICTED Y = 0.0108 DIFFERENCE = .1326 SQUARE OF DIFF. = .0176

BSERVED Y = 0.1329 PREDICTED Y = 0.0129 DIFFERENCE = .1200 SQUARE OF DIFF. = .0144

UM OF SQ. OF DIFFS. = 0.0595 RMS = 0.1409

17 TH RECESSION QR(J) = 0.042663 FT/MO = 4634.3 AC-FT/MO

BSERVED Y = 0.2317 PREDICTED Y = -0.0607 DIFFERENCE = .2924 SQUARE OF DIFF. = .0855

BSERVED Y = 0.2057 PREDICTED Y = -0.0150 DIFFERENCE = .2207 SQUARE OF DIFF. = .0487

BSERVED Y = 0.1848 PREDICTED Y = -0.0346 DIFFERENCE = .2193 SQUARE OF DIFF. = .0481

UM OF SQ. OF DIFFS. = 0.1823 RMS = 0.2465

18 TH RECESSION OR(J)= 0.039747 FT/MO = 4317.6 AC-FT/MO

BSERVED Y= 0.1529 PREDICTED Y= 0.0250 DIFFERENCE= .1279 SQUARE OF DIFF.= .0164
BSERVED Y= 0.1440 PREDICTED Y= 0.0250 DIFFERENCE= .1190 SQUARE OF DIFF.= .0142
BSERVED Y= 0.1334 PREDICTED Y= -0.0033 DIFFERENCE= .1368 SQUARE OF DIFF.= .0187
UM OF SQ. OF DIFFS.= 0.0492 RMS= 0.1281

19 TH RECESSION OR(J)= 0.027372 FT/MO = 2973.3 AC-FT/MO

BSERVED Y= 0.0173 PREDICTED Y= 0.1376 DIFFERENCE=-.1203 SQUARE OF DIFF.= .0145
BSERVED Y= 0.0188 PREDICTED Y= 0.0874 DIFFERENCE=-.0686 SQUARE OF DIFF.= .0047
BSERVED Y= 0.0191 PREDICTED Y= 0.0482 DIFFERENCE=-.0291 SQUARE OF DIFF.= .0008
UM OF SQ. OF DIFFS.= 0.0200 RMS= 0.0817

20 TH RECESSION OR(J)= 0.019388 FT/MO = 2106.1 AC-FT/MO

BSERVED Y= 0.0332 PREDICTED Y= 0.0802 DIFFERENCE=-.0469 SQUARE OF DIFF.= .0022
BSERVED Y= 0.0343 PREDICTED Y= 0.0388 DIFFERENCE=-.0044 SQUARE OF DIFF.= 0.0000
BSERVED Y= 0.0335 PREDICTED Y= 0.0126 DIFFERENCE= .0209 SQUARE OF DIFF.= .0004
UM OF SQ. OF DIFFS.= 0.0027 RMS= 0.0298

21 TH RECESSION OR(J)= 0.010307 FT/MO = 1119.6 AC-FT/MO
 OBSERVED Y= 0.1417 PREDICTED Y= -0.1167 DIFFERENCE= .2585 SQUARE OF DIFF.= .0668
 OBSERVED Y= 0.1234 PREDICTED Y= -0.0644 DIFFERENCE= .1878 SQUARE OF DIFF.= .0353
 OBSERVED Y= 0.1128 PREDICTED Y= -0.0731 DIFFERENCE= .1859 SQUARE OF DIFF.= .0346
 UM OF SQ. OF DIFFS.= 0.1366 RMS= 0.2134

22 TH RECESSION OR(J)= 0.017486 FT/MO = 1899.4 AC-FT/MO
 OBSERVED Y= 0.0893 PREDICTED Y= -0.0320 DIFFERENCE= .1220 SQUARE OF DIFF.= .0149
 OBSERVED Y= 0.0844 PREDICTED Y= 0.0154 DIFFERENCE= .0690 SQUARE OF DIFF.= .0048
 OBSERVED Y= 0.0816 PREDICTED Y= -0.0282 DIFFERENCE= .1099 SQUARE OF DIFF.= .0121
 UM OF SQ. OF DIFFS.= 0.0317 RMS= 0.1028

23 TH RECESSION OR(J)= 0.006507 FT/MO = 706.8 AC-FT/MO
 OBSERVED Y= 0.1135 PREDICTED Y= -0.0502 DIFFERENCE= .1637 SQUARE OF DIFF.= .0268
 OBSERVED Y= 0.1016 PREDICTED Y= -0.0851 DIFFERENCE= .1867 SQUARE OF DIFF.= .0349
 OBSERVED Y= 0.0899 PREDICTED Y= -0.0916 DIFFERENCE= .1815 SQUARE OF DIFF.= .0329
 UM OF SQ. OF DIFFS.= 0.0946 RMS= 0.1776

24 TH RECESSION OR(J)= 0.038862 FT/MO = 4221.4 AC-FT/MO

OBSERVED Y= 0.1870 PREDICTED Y= -0.0335 DIFFERENCE= .2205 SQUARE OF DIFF.= .0486
 OBSERVED Y= 0.1716 PREDICTED Y= 0.0167 DIFFERENCE= .1549 SQUARE OF DIFF.= .0240
 OBSERVED Y= 0.1580 PREDICTED Y= -0.0335 DIFFERENCE= .1914 SQUARE OF DIFF.= .0366
 UM OF SQ. OF DIFFS.= 0.1092 RMS= 0.1908

25 TH RECESSION OR(J)= 0.022126 FT/MO = 2403.4 AC-FT/MO

OBSERVED Y= 0.2343 PREDICTED Y= -0.0560 DIFFERENCE= .2903 SQUARE OF DIFF.= .0843
 OBSERVED Y= 0.2032 PREDICTED Y= -0.1345 DIFFERENCE= .3377 SQUARE OF DIFF.= .1140
 OBSERVED Y= 0.1662 PREDICTED Y= -0.1476 DIFFERENCE= .3137 SQUARE OF DIFF.= .0984
 UM OF SQ. OF DIFFS.= 0.2968 RMS= 0.3145

26 TH RECESSION OR(J)= -0.041651 FT/MO = -4524.4 AC-FT/MO

OBSERVED Y= 0.1187 PREDICTED Y= -0.0484 DIFFERENCE= .1672 SQUARE OF DIFF.= .0279
 OBSERVED Y= 0.1162 PREDICTED Y= -0.3318 DIFFERENCE= .4481 SQUARE OF DIFF.= .2008
 OBSERVED Y= 0.0710 PREDICTED Y= -0.4256 DIFFERENCE= .4966 SQUARE OF DIFF.= .2466
 UM OF SQ. OF DIFFS.= 0.4753 RMS= 0.3980

27 TH RECESSION OR(J)= 0.089035 FT/MO = 9671.5 AC-FT/MO

BSERVERD Y= 0.0692 PREDICTED Y= 0.2173 DIFFERENCE=-.1481 SQUARE OF DIFF.= .0219
 BSERVERD Y= 0.0773 PREDICTED Y= 0.2871 DIFFERENCE=-.2098 SQUARE OF DIFF.= .0440
 BSERVERD Y= 0.0913 PREDICTED Y= 0.3263 DIFFERENCE=-.2350 SQUARE OF DIFF.= .0552

UM OF SQ. OF DIFFS.= 0.1212 RMS= 0.2010

28 TH RECESSION QR(J)= -0.001558 FT/MO = -169.3 AC-FT/MO

BSERVERD Y= 0.0976 PREDICTED Y= -0.1187 DIFFERENCE= .2162 SQUARE OF DIFF.= .0467
 BSERVERD Y= 0.0883 PREDICTED Y= -0.0685 DIFFERENCE= .1568 SQUARE OF DIFF.= .0246
 BSERVERD Y= 0.0795 PREDICTED Y= -0.0969 DIFFERENCE= .1764 SQUARE OF DIFF.= .0311

UM OF SQ. OF DIFFS.= 0.1024 RMS= 0.1848

29 TH RECESSION QR(J)= 0.083014 FT/MO = 9017.5 AC-FT/MO

BSERVERD Y= 0.1277 PREDICTED Y= 0.1224 DIFFERENCE= .0053 SQUARE OF DIFF.=0.0000
 BSERVERD Y= 0.1300 PREDICTED Y= 0.2641 DIFFERENCE=-.1342 SQUARE OF DIFF.= .0180
 BSERVERD Y= 0.1488 PREDICTED Y= 0.2031 DIFFERENCE=-.0543 SQUARE OF DIFF.= .0029

UM OF SQ. OF DIFFS.= 0.0210 RMS= 0.0836

30 TH RECESSION QR(J)= 0.033693 FT/MO = 3659.9 AC-FT/MO

BSERVERD Y= 0.1444 PREDICTED Y= -0.0983 DIFFERENCE= .2427 SQUARE OF DIFF.= .0589

OBSERVED Y# 0.1237 PREDICTED Y# 0.0172 DIFFERENCE# .1065 SQUARE OF DIFF.# .0113
OBSERVED Y# 0.1259 PREDICTED Y# 0.0914 DIFFERENCE# .0345 SQUARE OF DIFF.# .0012
SUM OF SQ. OF DIFFS.# 0.0714 RMS# 0.1343

31 TH RECESSION OR(J)# -0.000135 FT/MO # -14.7 AC-FT/MO

OBSERVED Y# 0.1252 PREDICTED Y# -0.1834 DIFFERENCE# .3085 SQUARE OF DIFF.# .0952
OBSERVED Y# 0.1125 PREDICTED Y# -0.0504 DIFFERENCE# .1629 SQUARE OF DIFF.# .0265
OBSERVED Y# 0.1016 PREDICTED Y# -0.1071 DIFFERENCE# .2087 SQUARE OF DIFF.# .0436

SUM OF SQ. OF DIFFS.# 0.1653 RMS# 0.2347

Water Management Options in the Mesilla Valley Study

00100

LUMPED PARAMETER MODEL FOR THE MESILLA VALLEY, NEW MEXICO

DIMENSION COBS(10,12), IYR(10), MON(12),
1 GST(10,12), OPT(10,12), ETEFM(11,12), ORINT(10,12), GLRT(10,12)
COMMON / FITFUS(10,12), GLR(10,12), OP(10,12), CONUSE(10,12),
1 ORIN(10,12), CSUR(10,12), TC(10,12)
COMMON / GRAP6/CAQ(11,12), GROUT(10,12), COUT(10,12), H(11,12),
1 GD(10,12), GURS(10,12), TITLE(SU), ENET(11,12),
1 COMMON / LABEL/LRX(40), LB1(40), LB2(40), LB3(40), LB4(40), LB5(40),
1 LB6(40)

DATA MUN/5H JAN, 5H FEB, 5H MAR, 5H APR, 5H MAY, 5H JUN,
1 5H JUL, 5H AUG, 5H SEP, 5H OCT, 5H NOV, 5H DEC /
DATA IRI1, IRI2, IRIW/25, 26, 3 /
DATA USAVED, BETA1, BETA2, CLEAK/1, 0, 0, 50, 0, 75, 0, 45 /
DATA ALPHA, THO, POR, Z, AREA, HR/0, 0, 755, 3825, 0, 0, 218, 9, 05,
1 OPFN(UNIT=26, DEVICE='DISK', ACCESS='SEQIN', FILE='AMES.DAT'),
2000 OPEN(UNIT=25, DEVICE='DISK', ACCESS='SEQIN', FILE='OBSV.DAT'),
2100 FACTOR=(BETA2-BETA1)/(BETA1+BETA2)

DO 6 J=1,10
2200 READ(IRI,100) IYR(J)
2300 FORMAT(I4)

100

READ(IRI,101) (GST(J,I), I=1,12)
2500 READ(IRI,101) (ORINT(J,I), I=1,12)
2600 READ(IRI,101) (CSUR(J,I), I=1,12)
2700 READ(IRI,101) (OBS(J,I), I=1,12)
2800 READ(IRI,101) (CUBS(J,I), I=1,12)
2900 READ(IRI,101) (COBS(J,I), I=1,12)
3000 READ(IRI,101) (CONUSE(J,I), I=1,12)
3100 READ(IRI,101) (ETEM(J,I), I=1,12)
3200 READ(IRI,101) (GLRT(J,I), I=1,12)
3300 DO 6 I=1,12

OPT(J,I)=CMUSE(J,I)/BETA1-(1.0-CLEAK)*GST(J,I)
3400 IF(OPT(J,I).LT.0.0) OPT(J,I)=0.0

101

CONTINUE
3700 FORMAT(RF10,0,0, / 4F10,0)
3800 READ(IRI,104) (LRX(K), K=1,8)
3900 READ(IRI,104) (LR1(K), K=1,8)
4000 READ(IRI,104) (LR2(K), K=1,8)
4100 READ(IRI,104) (LR3(K), K=1,8)
4200 READ(IRI,104) (LR4(K), K=1,8)
4300 READ(IRI,104) (LR5(K), K=1,8)
4400 READ(IRI,104) (LR6(K), K=1,8)
4500 CALL INITIAL(1)
4600 DO 8 JR=1,5
4700 CAQ(1,1)=0.0
4800 CAQ(1,2)=0.0
4900 CAQ(1,3)=0.0
5000 CAQ(1,4)=0.0
5100 CAQ(1,5)=1626.0
5200 JGR=JR

104

READ(IRI,104) (TITLE(K), K=1,10)
5300 FORMAT(10A5)
5400 DO 7 J=1,10
5500 ORIN(J,I)=ORINT(J,I)
5600 QS(J,I)=QSI(J,I)
5700 ENET(J,I)=ETEM(J,I)
5800 GLR(J,I)=GLRI(J,I)
5900

06000


```

06100 7 QP(J,I)=QPT(J,I)
06200 GO TO (9,1,2,3,4),JGR
06300 1 DO 111 J=1,10
06400 DO 111 I=1,12
06500 ENET(J,I)=COMUSE(J,I)/AREA
06600 111 QS(J,I)=0.0
06700 GO TO 9
06800 2 DO 222 J=1,10
06900 DO 222 I=1,12
07000 IF(OLR(J,I).LT.0.0)GO TO 223
07100 DIFF=ORIN(J,I)-OS(J,I)-QP(J,I)-QLR(J,I)
07200 GO TO 224
07300 DIFF=ORIN(J,I)-OS(J,I)-QP(J,I)
07400 223 IF(DIFF.LT.0.0)QP(J,I)=QP(J,I)+DIFF
07500 224 QS(J,I)=QS(J,I)+QP(J,I)
07600 ENET(J,I)=(QS(J,I)-COMUSE(J,I))/AREA
07700 222 QP(J,I)=0.0
07800 GO TO 9
07900 4 DO 333 J=1,10
08000 DO 333 I=1,12
08100 QS(J,I)=(1.0-CLEAK)*QS(J,I)
08200 ENET(J,I)=(QS(J,I)-COMUSE(J,I))/AREA
08300 CONTINUE
08400 333 GO TO 9
08500 3 DO 444 J=1,10
08600 DO 444 I=1,12
08700 WATSAV=FACTOR*CONUSF(J,I)
08800 QS(J,I)=QS(J,I)-QSAVED+WATSAV
08900 IF(QS(J,I).LT.0.0)QS(J,I)=0.0
09000 444 ENET(J,I)=ENET(J,I)-QSAVED+WATSAV/AREA
09100 CONTINUE
09200 CALL WATBAL(JGR,ALPHA,POR,Z,HQ,AREA,HR)
09300 CALL WASHAL(WR,AREA,POR)
09400 CALL DELTA(JGR)
09500 CALL GRAFF(JGR)
09600 WRITE(IM,201)(TITLE(K),K=1,10),(IYR(K),K=1,10)
09700 FORMAT(1H,10(1X,10A5),5X,MESILLA VALLEY,N,MEX,-SIMULATION RESULTS',/)
09800 201 1 5X,AVE AQ W=L (FT),5X,10A5,12X,10(3X,14,3X)
09900 DO 10 I=1,12
10000 WRITE(IM,202)MON(I),(H(J,I),J=1,10)
10100 202 FORMAT(5X,A5,10F10.2)
10200 WRITE(IM,301)(A1(J),J=1,10)
10300 FORMAT(1X,10A1)
10400 301 WRITE(IM,203)(TITLE(K),K=1,10),(IYR(K),K=1,10)
10500 203 1 5X,AVE AQ W=L (FT),5X,10A5,12X,10(3X,14,3X)
10600 1 5X,AVE AQ W=L (FT),5X,10A5,12X,10(3X,14,3X)
10700 DO 11 I=1,12
10800 11 WRITE(IM,202)MON(I),(CAQ(J,I),J=1,10)
10900 WRITE(IM,301)(A3(J),J=1,10)
11000 204 1 5X,AVE RI=1,12
11100 WRITE(IM,204)(TITLE(K),K=1,10),(IYR(K),K=1,10)
11200 204 1 5X,AVE RI=1,12
11300 1 5X,AVE RI=1,12
11400 12 WRITE(IM,202)MON(I),(OROUT(J,I),J=1,10)
11500 WRITE(IM,301)(A4(J),J=1,10)
11600 12 WRITE(IM,205)(TITLE(K),K=1,10),(IYR(K),K=1,10)
11700 205 1 5X,AVE RI=1,12
11800 1 5X,AVE RI=1,12
11900 13 WRITE(IM,202)MON(I),(COMT(J,I),J=1,10)
12000 13 WRITE(IM,301)(A5(J),J=1,10)

```

```

12100 WRITE(IW,301)(A5(J),J=1,10)
12200 WRITE(IW,206)((TITLE(K),K=1,10),(IYR(K),K=1,10)
12300 DO 14 I=1,12
12400   14 WRITE(IW,202)MON(I),(CD(J,I),J=1,10)
12500   WRITE(IW,301)(A2(J),J=1,10)
12600   FORMAT(IH1,/,/,/,5X,MESILLA VALLEY,N,MEX.-SIMULATION RESULTS:’,/,/
12700   1 5X,DRAIN FLOW (AC-PT/MO)’,/,/,5X,10A5,/,/,12X,10(3X,14,3X))
12800   WRITE(IW,207)((TITLE(K),K=1,10),(IYR(K),K=1,10)
12900   1 5X,NET RECHARGE (FT)’,/,/,5X,10A5,/,/,12X,10(3X,14,3X))
13000   DO 15 I=1,12
13100     15 WRITE(IW,208)MON(I),(ENET(J,I),J=1,10)
13200     FORMAT(5X,A5(10F10.4)
13300   208 WRITE(IW,302)(A6(J),J=1,10)
13400   302 FORMAT(/,1X,AVERAGES:’,10F10.4)
13500   8 CONTINUE
13600   CALL RSTR(0)
13700   CALL RSTR(2)
13800   CLOSE(UNIT=26,DEVICE='DSK',ACCESS='SEQIN',FILE='AMES.DAT')
13900   CLOSE(UNIT=23,DEVICE='DSK',ACCESS='SEQIN',FILE='OBSV.DAT')
14000   STOP
14100   END
14200
14300
14400
14500   C
14600   C
14700   COMMON /FIT/US(10,12),CLR(10,12),GR(10,12),CONUSE(10,12),
14800   1 CORIN(10,12),CSUR(10,12),TC(10)
14900   1 GD(10,12),SUBS(10,12),GROUT(10,12),COUT(10,12),H(11,12),
15000   DIMENSION CRT(11),GR(10,12)
15100   DATA CRT/0,0103,0,0175,0,0065,0,0389,0,0221,-0,0416,0,0890
15200   1 -0,0016,0,0836,0,0334,-0,0061/
15300   DO 1 JK=1,10
15400     DO 1 IK=1,12
15500     IF(IK.GT.10)GO TO 2
15600     GR(JK,IK+2)=GR(JK+1)
15700     GO TO 1
15800     2 GR(JK,IK-10)=GR(JK)
15900     CONTINUE
16000     1 ENET(I,I)=ENET(10,12)
16100     TOL=0.01
16200     H(1,1)=3824.38
16300     DO 23 J=1,10
16400     DO 22 I=1,12
16500     GR(J,I)=1.0/Z*ALOG(EXP(ALPHA*Z*(H(J,I)-HO)))+1.0)
16600     IF(J.GT.1)AND(I.GT.1)GO TO 13
16700     HOLD=H(J,I)
16800     HREG=HOLD
16900     GO TO 18
17000     13 HREG=HNEW
17100     GO TO 18
17200     14 HOLD=HNEW
17300     GREG=1.0/Z*ALOG(EXP(ALPHA*Z*(HBEG-HO)))+1.0)
17400     GOLD=1.0/Z*ALOG(EXP(ALPHA*Z*(HOLD-HO)))+1.0)
17500     AVG=(GOLD+GREG)/2.0
17600     IF(I.EQ.12)GO TO 19
17700     AVREC=(ENET(J,I+1)+ENET(J,I))/2.0
17800     GO TO 20
17900     19 AVREC=(ENET(J+1,I)+ENET(J,I))/2.0
18000     20 HNEW=HBEG+(AVREC+OR(J,I)-AVG)/POR

```

```

18100 DIFF=ABS(HNEW-HOLD)
18200 IF(DIFF.GT.TOL)GO TO 14
18300 IF(I.LT.12)GO TO 21
18400 H(J,I,1)=HNEW
18500 GO TO 22
18600
18700
18800
18900
19000
19100
19200
19300
19400
19500
19600
19700
19800
19900
20000
20100
20200
20300
20400
20500
20600
20700
20800
20900
21000
21100
21200
21300
21400
21500
21600
21700
21800
21900
22000
22100
22200
22300
22400
22500
22600
22700
22800
22900
23000
23100
23200
23300
23400
23500
23600
23700
23800
23900
24000
DIFF=ABS(HNEW-HOLD)
IF(DIFF.GT.TOL)GO TO 14
IF(I.LT.12)GO TO 21
H(J,I,1)=HNEW
GO TO 22
H(J,I+1)=HNEW
22 CONTINUE
23 CONTINUE
DO 24 J=1,10
DO 24 I=1,12
24 OD(J,I)=AREA*OD(J,I)
RETURN
END
C
SUBROUTINE MASBAL(HR,AREA,POR)
C
COMMON /FIT/QS(10,12),CLR(10,12),OP(10,12),CONUSE(10,12),
COMMON /GRAPH/CAQ(10,12),TC(10)
COMMON /GORS(10,12),GROUT(10,12),COUT(10,12),H(11,12),
COMMON /AVE/A1(10),A2(10),A3(10),A4(10),A5(10),A6(10)
DIMENSION WL(10,12)
DO 3 J=1,10
A1(J)=0.0
A2(J)=0.0
A3(J)=0.0
A4(J)=0.0
A5(J)=0.0
A6(J)=0.0
TC(J)=0.0
FTOT=0.0
DO 4 I=1,12
WL(J,I)=H(J,I)-HR
TC(J)=TC(J)+WL(J,I)
IF(CLR(J,I).LT.0.0)GO TO 44
OLEAK=CLR(J,I)
GO TO 43
44 OLFAK=0.0
43 TOTREC=ENT(J,I)+OLEAK/AREA
FTOT=FTOT+TOTREC
IF(J.EQ.1.AND.I.LT.9)GO TO 4
AK1=TOTREC/(POR*WL(J,I))
AK2=(QS(J,I)+OLEAK)/(AREA*POR*WL(J,I))
IF(I.EQ.12)GO TO 5
CAQ(J,I+1)=CSUR(J,I)+2.*AK2/(AK1+2.)-CAQ(J,I)*(AK1-2.)/(AK1+2.)
GO TO 6
5 CAQ(J,I+1)=CSUR(J,I)+2.*AK2/(AK1+2.)-CAQ(J,I)*(AK1-2.)/(AK1+2.)
6 GROUT=QS(J,I)+CLR(J,I)
GROUT(J,I)=OD(J,I)+ORIN(J,I)-GROUT
IF(CLR(J,I).LT.0.0)GO TO 45
COUT(J,I)=((ORIN(J,I)-GROUT)*CSUR(J,I)+OD(J,I)*CAQ(J,I))/GROUT(J,I)
GO TO 4
45 COUT(J,I)=((ORIN(J,I)-QS(J,I))+CSUR(J,I)+CAQ(J,I))*OD(J,I)-
CLR(J,I))/GROUT(J,I)
4 CONTINUE
IF(FTOT.EQ.0.0)GO TO 75
TC(J)=POR*TC(J)/(12.*FTOT)
75 CONTINUE
DO 46 I=1,12
A1(J)=A1(J)+H(J,I)

```


30100	Y2(K)=0D(J,I)
30200	CONTINUE
30300	Y1MIN=YS1(JR)
30400	DY1=DELY1(JR)
30500	Y2MIN=YS2(JR)
30600	DY2=DELY2(JR)
30700	DO 111 KI=1,8
30800	LBY1(KI)=UB1(KI)
30900	LBY2(KI)=UB2(KI)
31000	GO TO 50
31100	DO 12 J=1,10
31200	IF(J,GT,1)IN=1
31300	DO 12 I=IN,12
31400	K=12*(J-1)+I
31500	AK=K
31600	X(K,4)=1967.0+(AK-0.5)/12.0
31700	Y1(K,4)=GROUT(J,I)
31800	Y2(K,4)=GROUT(J,I)
31900	CONTINUE
32000	Y1MIN=YS5(JR)
32100	DY1=DELY5(JR)
32200	Y2MIN=Y1MIN
32300	DY2=DY1
32400	DO 112 KI=1,8
32500	LBY1(KI)=UB3(KI)
32600	LBY2(KI)=UB3(KI)
32700	M=-1
32800	J2=J1
32900	GO TO 50
33000	DO 14 J=1,10
33100	IF(J,GT,1)IN=1
33200	DO 14 I=IN,12
33300	K=12*(J-1)+I
33400	AK=K
33500	X(K,4)=1967.0+(AK-0.5)/12.0
33600	Y1(K,4)=GROUT(J,I)
33700	Y2(K,4)=GROUT(J,I)
33800	CONTINUE
33900	Y1MIN=YS6(JR)
34000	DY1=DELY6(JR)
34100	Y2MIN=Y1MIN
34200	DY2=DY1
34300	DO 114 KI=1,8
34400	LBY1(KI)=UB4(KI)
34500	LBY2(KI)=UB4(KI)
34600	GO TO 50
34700	DO 13 J=1,10
34800	IF(J,GT,1)IN=1
34900	DO 13 I=IN,12
35000	K=12*(J-1)+I
35100	AK=K
35200	X(K,4)=1967.0+(AK-0.5)/12.0
35300	Y1(K,4)=GROUT(J,I)
35400	Y2(K,4)=ENET(J,I)
35500	CONTINUE
35600	Y1MIN=YS3(JR)
35700	DY1=DELY3(JR)
35800	Y2MIN=YS4(JR)
35900	DY2=DELY4(JR)
36000	M=1

36100
 36200
 36300
 36400
 36500
 36600
 36700
 36800
 36900
 37000
 37100
 37200
 37300
 37400
 37500
 37600
 37700
 37800
 37900
 38000
 38100
 38200
 38300
 38400
 38500
 38600
 38700
 38800
 38900
 39000
 39100
 39200
 39300
 39400
 39500
 39600
 39700
 39800
 39900
 40000
 40100
 40200
 40300
 40400
 40500
 40600
 40700
 40800
 40900
 41000
 41100
 41200
 41300
 41400
 41500
 41600
 41700
 41800
 41900
 42000

```

M=116
DO 113 KI=1,8
  LBY1(KI)=LB5(KI)
  LRY2(KI)=LB6(KI)
  113 XMIN=1967.0
  DX=1.0
  DO 444 I=1,N
    X(I)=(X(I)-XMIN)/DX
    Y(I)=(Y(I)-YMIN)/DY1
    Y2(I)=(Y2(I)-Y2MIN)/DY2
  444 CALL PLOT(0.35,1.0,-3)
  CALL AXIS(0.0,0.0, LBXR,=40.10,0.0, XMIN,DX,-1)
  CALL PLOT(0.7,3)
  CALL PLOT(10.1,2)
  CALL PLOT(0.6,3)
  CALL AXIS(0.0,0.0, LBY1,40.7,90, Y1MIN,DY1,-1)
  CALL SYMBOL(0.3,7,10,0.1, FILLF,0.50)
  CALL AXIS(10.0,0.0,3)
  CALL AXIS(10.0,0.0, LRY2,=40.7,90.0, Y2MIN,DY2,M)
  CALL PLOT(0.0,0.0,3)
  CALL LINE(X,V1,N,J1,1)
  CALL PLOT(0.0,3)
  CALL LINE(X,Y2,N,J2,1)
  CALL PLOT(0.0,1,-3)
  KOUNT=KOUNT+1
  IF(KOUNT.LI.5)GO TO 5
  RETURN
END
C
SUBROUTINE DELTA(JR)
C
COMMON /GRAPH/CAQ(11,12),GROUT(10,12),COUT(10,12),H(11,12),
1 GD(10,12),QRS(10,12),FILLE(50),ENER(11,12)
COMMON /AVE/A1(10),A2(10),A3(10),A4(10),A5(10),A6(10)
DIMENSION H(11,12),QD(10,12),CAQ(11,12),ENER(11,12)
1 GROUT(10,12),COUT(10,12),AA1(10),AA2(10),AA3(10),AA4(10),
2 AA5(10),AA6(10)
IF(JR.GT.1)GO TO 1
DO 2 J=1,10
  AA1(J)=A1(J)
  AA2(J)=A2(J)
  AA3(J)=A3(J)
  AA4(J)=A4(J)
  AA5(J)=A5(J)
  AA6(J)=A6(J)
DO 2 I=1,12
  HI(J,I)=H(J,I)
QD(I,J,I)=QD(J,I)
CAQ(I,J,I)=CAQ(J,I)
ENER(I,J,I)=ENER(J,I)
GROUT(I,J,I)=GROUT(J,I)
COUT(I,J,I)=COUT(J,I)
2 COUT(I,J,I)=COUT(J,I)
GO TO 3
DO 4 J=1,10
  1 A1(J)=A1(J)-AA1(J)
  A2(J)=A2(J)-AA2(J)
  A3(J)=A3(J)-AA3(J)
  A4(J)=A4(J)-AA4(J)
  A5(J)=A5(J)-AA5(J)
  A6(J)=A6(J)-AA6(J)

```

42100 DU 4 I=1, IZ
42200 H(J, I)=H(J, I)-HI(J, I)
42300 OD(J, I)=OD(J, I)-ODI(J, I)
42400 CAO(J, I)=CAO(J, I)-CAOI(J, I)
42500 ENET(J, I)=ENET(J, I)-ENETI(J, I)
42600 OROUT(J, I)=OROUT(J, I)-OROUTI(J, I)
42700 COUT(J, I)=COUT(J, I)-COUTI(J, I)
42800 3 RETURN
42900 END

MESILLA VALLEY, N.MEX.--SIMULATION RESULTS
AVE AQ W-L (FT)

BASE PERIOD SIMULATION

	1967	1968	1969	1970	1971	1972	1973	1974	1975	1976
JAN	3824.38	3823.87	3824.05	3825.12	3825.07	3823.37	3823.83	3823.76	3825.02	3824.75
FEB	3824.18	3823.76	3824.08	3824.94	3824.93	3823.05	3824.04	3823.58	3825.04	3824.61
MAR	3824.32	3824.03	3824.08	3825.23	3824.98	3822.42	3824.25	3823.57	3825.28	3824.87
APR	3824.52	3824.14	3824.59	3825.55	3825.09	3822.29	3824.16	3823.57	3825.37	3825.31
MAY	3824.88	3823.93	3824.72	3825.74	3825.12	3822.31	3824.06	3824.28	3825.36	3825.58
JUN	3823.88	3824.04	3825.04	3826.16	3825.17	3822.30	3824.79	3824.68	3825.80	3825.95
JUL	3824.21	3824.48	3825.87	3826.66	3825.81	3822.29	3824.04	3824.88	3825.78	3825.69
AUG	3824.47	3825.06	3826.55	3826.97	3825.12	3822.24	3824.51	3825.02	3825.80	3825.07
SEP	3824.42	3825.10	3826.41	3826.69	3824.89	3822.57	3824.57	3825.01	3825.80	3825.33
OCT	3824.12	3824.72	3825.91	3826.14	3824.55	3823.07	3824.43	3824.92	3825.37	3825.45
NOV	3823.98	3824.46	3825.36	3825.71	3824.11	3823.35	3824.18	3824.06	3825.33	3825.05
DEC	3823.98	3824.24	3825.33	3825.35	3823.72	3823.60	3823.95	3824.99	3824.92	3824.72
AVERAGES:	3824.16	3824.32	3825.17	3825.85	3824.82	3822.74	3824.15	3824.40	3825.40	3825.20

MESILLA VALLEY, N.MEX.--SIMULATION RESULTS:
AVE AQ CHEM. (MG/L)

BASE PERIOD SIMULATION

	1967	1968	1969	1970	1971	1972	1973	1974	1975	1976
JAN	0.00	1664.76	1691.84	1682.15	1660.50	1651.36	1674.65	1713.57	1713.78	1708.28
FEB	0.00	1664.73	1691.72	1682.04	1660.50	1651.14	1674.62	1713.46	1712.78	1709.92
MAR	0.00	1664.78	1691.72	1683.80	1660.50	1651.03	1674.60	1713.37	1713.03	1715.58
APR	0.00	1660.58	1685.67	1673.20	1659.53	1658.85	1680.04	1730.86	1704.48	1706.34
MAY	1626.00	1674.83	1681.87	1674.13	1654.24	1658.85	1682.60	1716.82	1702.29	1698.14
JUN	1632.68	1679.68	1684.83	1664.40	1646.62	1665.15	1681.49	1711.12	1690.57	1696.14
JUL	1637.34	1691.58	1676.34	1654.89	1646.62	1671.22	1708.15	1708.73	1690.30	1683.17
AUG	1646.36	1679.14	1662.94	1642.74	1642.56	1676.14	1708.15	1708.90	1697.15	1699.99
SEP	1647.00	1673.73	1664.19	1641.54	1648.87	1684.86	1695.58	1708.06	1698.15	1722.15
OCT	1664.00	1690.02	1679.01	1657.29	1651.18	1676.37	1711.03	1716.85	1708.04	1706.37
NOV	1664.14	1691.34	1681.77	1659.84	1651.36	1675.15	1713.62	1714.14	1708.76	1707.55
DEC	1664.86	1691.87	1682.74	1660.63	1651.36	1674.92	1713.59	1713.86	1708.55	1707.55
AVERAGES:	1647.66	1676.42	1681.22	1664.72	1652.52	1666.34	1693.51	1714.15	1704.01	1705.95

MESILLA VALLEY, N. MEX. - SIMULATION RESULTS
AVE RIVER OUTFLOW (AC-FT/MO)

BASE PERIOD SIMULATION

	1967	1968	1969	1970	1971	1972	1973	1974	1975	1976
JAN	0.00	6150.60	5761.31	11193.35	11072.55	3690.17	5029.99	4947.40	13702.51	18237.51
FEB	0.00	7202.57	5015.74	12565.18	10847.61	2941.77	5667.51	4297.85	19858.26	10314.33
MAR	0.00	48226.55	37067.74	48501.15	57927.88	16852.14	38607.36	35852.25	35443.63	42894.53
APR	0.00	23983.96	32972.06	38556.67	39410.46	9033.54	36389.14	22545.91	40118.37	48300.02
MAY	25468.15	27447.39	30856.98	42130.11	42124.46	6116.85	37449.29	27781.85	43291.78	61679.77
JUN	29370.13	21577.82	46393.32	49494.04	52207.09	4888.31	43230.86	42867.14	46001.80	53790.55
JUL	34384.06	36711.61	51152.39	62373.45	50027.18	10581.54	39338.14	42809.39	39035.86	41619.06
AUG	40862.10	30416.31	50044.61	50084.82	35428.85	8610.15	57319.96	38277.37	42334.04	46879.45
SEP	24507.10	19407.86	25072.98	27739.53	11265.08	3954.12	23363.36	15893.13	31988.11	20779.78
OCT	7941.93	19407.86	13645.39	16064.64	9548.72	4201.72	7669.30	15893.13	10951.19	10261.84
NOV	7524.84	7350.96	13103.04	12692.64	6809.55	3643.66	6262.09	93365.20	19866.73	10125.01
DEC	8253.21	7096.03	17909.21	12188.20	5728.70	4163.49	5640.31	9135.31	9373.77	8406.75
AVERAGES:	22289.35	22759.82	27416.28	31965.31	27699.54	6556.45	25260.61	23109.62	27668.77	31356.65

MESILLA VALLEY, N. MEX. - SIMULATION RESULTS
AVE RIVER CHEM (MG/L)

BASE PERIOD SIMULATION

	1967	1968	1969	1970	1971	1972	1973	1974	1975	1976
JAN	0.00	1522.22	1591.56	1562.29	1541.71	1594.55	1543.44	1628.00	1210.17	902.08
FEB	0.00	1451.82	1623.90	1504.13	1504.49	1617.62	1577.89	1637.00	1540.56	1311.58
MAR	0.00	765.58	682.16	743.14	864.23	616.82	658.54	623.88	1761.89	655.70
APR	0.00	929.17	746.78	831.69	903.36	713.06	780.24	624.31	741.93	683.31
MAY	737.06	836.10	790.43	837.84	885.93	816.35	780.24	886.39	811.97	664.24
JUN	842.28	705.09	716.83	782.44	828.61	905.61	639.43	632.17	822.99	740.09
JUL	928.01	810.53	888.06	899.97	863.19	644.55	628.58	672.44	864.59	755.34
AUG	1025.34	841.50	888.06	899.97	906.24	713.86	621.79	871.02	829.85	667.92
SEP	1428.37	1110.93	1284.50	1397.74	1414.43	1245.72	864.62	1068.03	961.35	1047.31
OCT	1472.97	1649.71	1649.32	1580.48	1300.20	1403.38	1504.52	1600.43	1584.77	1617.73
NOV	1482.41	1587.79	1619.32	1615.48	1502.23	1616.42	1632.83	1600.43	1572.93	1561.69
DEC	1482.41	1536.76	1408.47	1540.11	1479.46	1615.52	1634.03	1657.06	1520.76	1592.95
AVERAGES:	1074.01	1145.60	1149.43	1168.64	1166.17	1125.12	1062.56	1034.53	1101.98	1016.66

MESILLA VALLEY, N. MEX. - SIMULATION RESULTS:
DRAIN FLOW (AC-FT/MO)

BASE PERIOD SIMULATION

	1967	1968	1969	1970	1971	1972	1973	1974	1975	1976
JAN	6048.65	4547.07	5034.43	8815.225	8594.51	3409.60	4466.69	4270.63	8401.63	7356.69
FEB	5436.47	4280.15	4567.16	9076.629	7623.99	2808.38	5023.03	3851.73	8499.79	6826.47
MAR	5727.29	4981.41	5130.06	9299.299	8230.99	1905.12	5632.00	3091.73	9531.95	7803.63
APR	5866.49	5313.70	6752.19	10781.99	8679.52	1755.05	5374.22	3842.19	9913.74	9676.27
MAY	5053.75	4718.00	6243.07	11714.07	8827.24	1776.87	5071.84	5723.97	10828.34	10951.48
JUN	4573.47	5020.28	8490.27	13977.80	9058.62	1760.84	4345.96	7073.20	12052.06	12817.74
JUL	4550.41	6360.42	12395.67	13683.23	8220.03	1750.84	5033.97	7823.43	11956.66	11478.42
AUG	5512.56	8549.42	16263.57	18921.12	8824.76	1694.95	6479.03	8411.08	12009.79	18612.19
SEP	6178.87	7428.63	15439.17	17125.81	7895.66	2082.72	6681.50	8354.01	11439.63	9739.28
OCT	5683.26	7223.59	12640.22	13899.75	6625.59	2846.15	6207.12	7991.07	19926.97	10287.96
NOV	5233.17	6318.63	10959.02	11561.50	5221.68	3363.15	5412.72	8145.78	8850.46	8518.73
DEC	4860.70	5606.36	9759.89	9862.75	4188.94	3908.28	4781.63	8285.89	8016.10	7209.34
AVERAGES:	5393.36	5972.29	9559.08	12584.93	7749.27	2421.82	5374.14	6405.54	10123.18	9273.13

MESILLA VALLEY, N. MEX. - SIMULATION RESULTS:
NET RECHARGE (FT)

BASE PERIOD SIMULATION

	1967	1968	1969	1970	1971	1972	1973	1974	1975	1976
JAN	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000
FEB	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000
MAR	0.1238	0.1669	0.1664	0.2084	0.1685	-0.1458	0.0105	-0.0942	0.1039	0.1812
APR	-0.0328	-0.0342	0.0871	0.0714	0.1175	-0.0550	0.0566	0.1316	0.0447	0.1728
MAY	-0.0200	0.0208	0.0230	0.0719	0.1422	-0.0820	-0.0030	0.0741	0.1628	0.1349
JUN	-0.0494	0.1042	0.1821	0.1729	0.1293	-0.0697	-0.0262	0.0476	0.0862	0.2420
JUL	0.1580	0.1769	0.2926	0.2587	0.1384	-0.0798	0.2257	0.0998	0.0593	-0.1299
AUG	0.0066	0.2002	0.1896	0.1601	0.0708	-0.0887	0.0889	0.0373	0.1033	0.0452
SEP	0.0000	-0.0331	-0.0352	0.0050	0.0679	0.0000	0.0611	-0.0350	-0.0071	0.2360
OCT	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000
NOV	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000
DEC	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000	0.0000
AVERAGES:	0.0351	0.0501	0.0755	0.0899	0.0695	-0.0361	0.0345	0.0146	0.0461	0.0735

MESILLA VALLEY, N.MEX.-SIMULATION RESULTS
AVE AQ N=L (FT)

CASE(1): QS=0; CHANGES FROM BASE PERIOD

	1967	1968	1969	1970	1971	1972	1973	1974	1975	1976
JAN	0.00	-4.87	-9.98	-14.61	-19.94	-26.53	-29.00	-34.04	-37.28	-40.70
FEB	-0.01	-4.70	-9.78	-14.27	-19.61	-26.41	-28.80	-33.88	-36.96	-40.42
MAR	-0.51	-5.00	-10.18	-15.56	-20.93	-26.25	-28.81	-33.74	-37.06	-40.77
APR	-1.18	-5.48	-11.06	-16.21	-20.80	-26.33	-29.26	-34.30	-37.53	-41.72
MAY	-1.65	-5.02	-10.80	-15.21	-21.84	-26.76	-29.94	-35.50	-38.34	-42.72
JUN	-2.46	-7.41	-13.07	-17.84	-23.41	-27.43	-30.83	-36.93	-39.58	-45.27
JUL	-3.60	-9.16	-14.84	-19.85	-25.21	-28.35	-32.20	-38.13	-40.78	-45.60
AUG	-4.68	-10.39	-16.03	-21.30	-26.30	-28.79	-34.44	-38.58	-41.58	-46.35
SEP	-5.17	-10.69	-16.14	-21.01	-26.76	-29.04	-34.07	-38.44	-41.70	-46.35
OCT	-5.23	-10.56	-15.72	-21.01	-26.91	-29.37	-34.34	-38.12	-41.52	-46.77
NOV	-5.20	-10.42	-15.38	-20.70	-26.82	-29.21	-34.35	-37.91	-41.31	-46.47
DEC	-5.05	-10.19	-15.00	-20.32	-26.68	-29.18	-34.22	-37.62	-41.02	-46.17
AVERAGES:	-2.90	-7.91	-13.25	-18.09	-23.69	-27.80	-31.61	-36.43	-39.56	-43.93

MESILLA VALLEY, N.MEX.-SIMULATION RESULTS:
AVE AQ CHEM. (MG/L)

CASE(1): QS=0; CHANGES FROM BASE PERIOD

	1967	1968	1969	1970	1971	1972	1973	1974	1975	1976
JAN	0.00	52.68	178.01	324.37	572.68	914.10	1068.71	1428.45	1666.77	2124.60
FEB	0.00	59.52	178.31	324.84	572.68	914.25	1068.71	1428.21	1664.89	2123.42
MAR	0.00	59.98	178.59	322.91	572.77	914.25	1068.66	1428.21	1667.44	2119.88
APR	0.00	68.07	191.63	340.20	596.77	885.21	1085.18	1406.40	1695.42	2259.43
MAY	0.00	68.47	191.78	358.46	630.49	898.77	1171.27	1454.41	1700.23	2381.63
JUN	11.35	89.80	212.49	400.29	684.82	936.64	1171.14	1524.27	1928.77	2594.07
JUL	14.02	132.36	279.26	466.38	773.80	997.60	1250.77	1641.26	2050.27	2766.53
AUG	33.45	165.20	313.24	534.55	847.62	1052.22	1359.70	1671.41	2078.66	2866.53
SEP	61.24	178.19	323.59	556.09	875.02	1039.89	1359.70	1671.41	2078.66	2866.53
OCT	51.38	170.12	317.77	556.27	895.25	1061.46	1347.26	1652.83	2078.66	2937.98
NOV	57.75	177.32	321.85	566.37	906.46	1061.86	1347.26	1652.83	2078.66	2975.78
DEC	59.57	177.76	324.93	571.83	913.77	1069.12	1428.37	1667.19	2125.22	2987.08
AVERAGES:	41.10	117.24	254.31	443.54	736.81	978.77	1228.92	1553.95	1883.74	2526.85

MESILLA VALLEY, N. MEX. - SIMULATION RESULTS
 AVE RIVER OUTFLOW (AC-FT/MO)

CASE(1) Q8=0; CHANGES FROM BASE PERIOD

	1967	1968	1969	1970	1971	1972	1973	1974	1975	1976
JAN	0.00	-4349.91	-5027.58	-8814.66	-8594.50	-3409.60	-4666.69	-4270.63	-7877.63	1976.34
FEB	0.00	-3934.25	-4560.22	-5310.95	-7623.73	-2808.38	-5023.03	-3851.04	-6630.79	1446.34
MAR	0.00	33752.65	48890.04	41598.38	44606.03	33855.88	22941.00	50519.27	23382.95	1170.34
APR	0.00	20386.12	24679.54	26605.24	16163.49	12289.95	28723.78	37373.81	24303.26	38580.37
MAY	16285.03	15120.96	21324.06	26605.24	17252.77	4357.13	25543.17	36905.03	28223.66	36448.73
JUN	16792.53	36511.04	47684.36	38039.33	28155.38	6137.25	45608.04	53874.80	37867.94	40949.52
JUL	32829.31	41307.89	64079.89	55331.91	39456.97	26970.16	45498.97	42014.57	49957.34	44860.59
AUG	34905.88	47344.89	69762.04	51645.90	41449.24	28645.05	62310.97	45940.92	51454.21	59473.81
SEP	30295.88	25116.03	15608.34	22349.20	14833.30	4880.28	44116.50	24171.53	26839.37	29349.72
OCT	-2960.96	-4236.28	-8689.73	-9726.73	-5505.59	-4806.15	-1521.12	-7833.01	-5906.97	-7268.39
NOV	-3737.01	-5339.92	-9592.52	-10350.49	-5221.68	-3363.15	-5412.72	-8149.78	-8850.46	-8518.73
DEC	-4672.02	-5599.61	-9759.35	-9862.74	-4188.94	-3908.28	-4781.63	-8285.89	-8016.10	-7209.34
AVERAGES:	14967.26	16339.96	21199.91	18205.98	14231.90	8070.01	21131.78	21533.30	17062.16	23351.45

MESILLA VALLEY, N. MEX. - SIMULATION RESULTS
 AVE RIVER CHEM (MG/L)

CASE(1) Q8=0; CHANGES FROM BASE PERIOD

	1967	1968	1969	1970	1971	1972	1973	1974	1975	1976
JAN	0.00	-37.81	-685.48	-444.07	-315.47	-69.55	-925.44	-540.00	-198.17	1976.08
FEB	0.00	-274.99	-676.63	-220.05	-242.34	-705.62	-753.89	-659.00	-1077.56	-545.58
MAR	0.00	100.97	162.07	223.13	162.09	131.82	-173.54	-101.88	-349.89	-235.70
APR	0.00	-204.60	-241.67	-326.66	-245.87	-238.06	-173.24	-227.31	-315.93	-256.31
MAY	-170.85	-171.71	-273.35	-322.84	-230.67	-345.35	-153.43	-267.39	-296.97	-223.24
JUN	-114.72	-104.57	-216.81	-328.84	-202.74	-427.61	-117.43	-213.17	-307.99	-299.09
JUL	-108.72	-127.32	-281.42	-326.44	-222.19	-203.55	-172.58	-231.65	-364.59	-353.34
AUG	-205.71	-321.89	-608.48	-450.97	-324.84	-235.86	-136.79	-283.44	-344.85	-222.92
SEP	-554.78	-461.89	-908.48	-554.74	-468.84	-488.72	-335.62	-456.09	-410.35	-595.31
OCT	-402.34	-900.25	-431.61	-492.57	-750.79	-577.38	-223.22	-656.03	-1194.77	-935.73
NOV	-234.89	-630.74	-325.23	-453.47	-337.52	-706.40	-514.83	-762.03	-1182.93	-773.69
DEC	-242.91	-579.62	-328.41	-511.11	-313.51	-909.52	-443.03	-554.06	-1108.76	-689.95
AVERAGES:	-242.91	-356.38	-383.68	-396.19	-338.91	-470.70	-376.89	-413.03	-646.06	-492.50

MESILLA VALLEY, N. MEX. --SIMULATION RESULTS:
DRAIN FLOW (AC-FT/MO)

CASE(1): QS=0; CHANGES FROM BASE PERIOD

	1967	1968	1969	1970	1971	1972	1973	1974	1975	1976
JAN	0.00	-4349.91	-5027.58	-8814.66	-8594.50	-3409.60	-4466.69	-4270.63	-8401.63	-7356.66
FEB	-33.65	-4074.25	-4560.22	-8075.95	-7623.73	-2808.38	-5023.03	-3851.04	-8499.79	-6826.47
MAR	-144.18	-4780.35	-5123.86	-9298.62	-8230.97	-1905.12	-5632.00	-3043.73	-9511.95	-7803.63
APR	-2887.60	-5156.88	-6747.46	-10781.48	-8679.51	-1755.05	-3374.22	-3843.19	-9913.74	-9676.27
MAY	-3178.97	-4624.01	-7239.94	-11713.76	-8827.23	-1776.87	-5071.84	-5723.97	-10828.34	-10951.18
JUN	-3580.57	-4980.96	-8488.64	-13977.67	-9058.62	-1760.75	-4345.96	-7072.43	-12052.06	-12817.74
JUL	-4086.69	-6344.20	-12395.11	-16983.20	-9220.03	-1750.84	-5013.97	-7822.43	-11956.66	-11478.42
AUG	-5230.58	-8539.67	-16262.96	-17125.09	-8824.76	-1694.95	-6479.03	-8411.08	-12060.79	-9612.19
SEP	-5945.12	-8739.67	-15438.66	-17125.09	-7895.66	-2082.72	-6631.50	-8354.47	-11439.63	-9739.28
OCT	-5489.96	-7216.28	-13899.73	-13899.73	-6225.59	-2846.15	-5412.42	-7991.01	-9926.97	-10287.36
NOV	-5047.01	-6311.92	-10962.52	-11561.43	-5221.68	-3363.15	-5412.42	-8146.78	-8850.46	-10518.73
DEC	-4672.02	-5599.61	-9759.35	-9862.74	-4188.98	-3908.28	-4781.63	-8285.89	-8016.10	-7209.34
AVERAGES:	-3463.11	-5893.12	-9556.34	-12584.68	-7749.27	-2421.82	-5374.14	-6405.54	-10123.18	-9273.13

MESILLA VALLEY, N. MEX. --SIMULATION RESULTS:
NET RECHARGE (FT)

CASE(1): QS=0; CHANGES FROM BASE PERIOD

	1967	1968	1969	1970	1971	1972	1973	1974	1975	1976
JAN	-0.0033	-0.0020	-0.0023	-0.0035	-0.0026	-0.0022	0.0000	-0.0009	-0.0050	-0.0033
FEB	-0.0019	-0.0014	-0.0021	-0.0014	-0.0013	-0.0015	-0.0001	-0.0011	-0.0090	-0.0041
MAR	-0.2278	-0.2121	-0.2584	-0.2868	-0.2818	0.0253	-0.1009	-0.0050	-0.2010	-0.2855
APR	-0.1046	-0.0901	-0.2356	-0.2145	-0.2548	-0.0931	-0.1973	-0.3018	-0.1856	-0.2893
MAY	-0.1587	-0.2355	-0.2148	-0.3919	-0.3596	-0.1272	-0.1967	-0.3091	-0.3561	-0.3337
JUN	-0.2562	-0.4584	-0.4847	-0.5573	-0.5034	-0.1984	-0.2743	-0.4325	-0.3969	-0.5409
JUL	-0.3098	-0.4088	-0.4783	-0.6046	-0.4384	-0.2324	-0.4111	-0.2260	-0.3507	-0.5179
AUG	-0.2483	-0.2623	-0.3061	-0.2968	-0.2021	-0.0074	-0.2331	-0.1143	-0.2132	-0.1923
SEP	-0.0675	-0.0287	-0.0328	-0.0952	-0.1535	-0.1509	-0.1635	0.0299	-0.0585	-0.3042
OCT	-0.0218	-0.0600	-0.0426	-0.0649	-0.0424	-0.0128	-0.1635	-0.0299	-0.0585	-0.0452
NOV	-0.0012	-0.0071	-0.0262	-0.0318	-0.0269	-0.0255	-0.0361	-0.0071	-0.0264	-0.0132
DEC	-0.0012	-0.0018	0.0000	-0.0027	-0.0012	-0.0011	-0.0030	-0.0007	-0.0014	-0.0019
AVERAGES:	-0.1830	-0.1474	-0.1737	-0.2126	-0.1890	-0.0677	-0.1409	-0.1192	-0.1550	-0.1793

MESILLA VALLEY, N.MEX. - SIMULATION RESULTS
 AVE AQ W-L (FT)

CASE(2): OP=0: CHANGES FROM BASE PERIOD

	1967	1968	1969	1970	1971	1972	1973	1974	1975	1976
JAN	0.00	1.43	1.95	1.78	1.57	1.56	1.54	2.33	2.28	2.06
FEB	-0.01	1.25	1.66	1.48	1.28	1.39	1.54	1.75	1.86	2.35
MAR	-0.51	1.38	1.91	1.61	1.47	1.77	1.54	2.41	1.77	2.41
APR	-1.18	2.03	2.46	2.09	1.85	3.08	2.18	3.69	2.52	3.16
MAY	-0.84	2.57	2.90	2.48	2.00	2.35	2.05	3.57	3.06	3.18
JUN	0.04	3.17	3.52	2.78	2.20	2.18	4.18	4.24	3.98	4.27
JUL	1.04	3.77	4.23	3.32	2.71	2.84	4.76	5.56	4.63	5.27
AUG	1.78	3.95	4.23	3.88	3.30	2.99	4.86	5.58	4.76	4.93
SEP	2.28	4.09	4.56	4.05	3.26	2.48	5.16	4.91	4.12	3.80
OCT	2.38	3.66	3.86	3.50	2.72	2.48	4.55	4.78	3.09	3.89
NOV	1.95	2.85	2.92	2.59	2.12	2.11	3.48	3.78	2.38	2.29
DEC	1.67	2.33	2.25	1.98	1.79	1.79	2.78	2.89	2.38	2.29
AVERAGES:	0.72	2.71	3.00	2.63	2.19	2.32	3.28	3.93	3.05	3.18

MESILLA VALLEY, N.MEX. - SIMULATION RESULTS:
 AVE AQ CHEM. (MG/L)

CASE(2): OP=0: CHANGES FROM BASE PERIOD

	1967	1968	1969	1970	1971	1972	1973	1974	1975	1976
JAN	0.00	-79.54	-197.02	-328.41	-413.40	-453.66	-485.12	-589.97	-694.08	-745.11
FEB	0.00	-79.45	-196.96	-328.15	-413.07	-453.50	-485.11	-589.80	-693.77	-740.74
MAR	0.00	-79.49	-196.89	-330.13	-413.00	-453.59	-498.08	-589.67	-694.46	-759.63
APR	0.00	-92.82	-214.31	-339.79	-424.59	-480.89	-509.01	-623.50	-701.45	-758.74
MAY	0.00	-114.20	-227.50	-357.39	-426.21	-484.68	-524.51	-623.39	-701.45	-757.81
JUN	13.88	-120.71	-223.58	-357.27	-424.84	-474.09	-552.60	-644.20	-696.50	-763.26
JUL	-25.28	-137.88	-250.43	-391.24	-424.81	-478.76	-552.10	-659.70	-717.63	-758.09
AUG	-44.10	-151.36	-273.04	-374.61	-435.77	-498.93	-575.98	-677.26	-731.77	-786.23
SEP	-55.83	-172.35	-308.15	-396.47	-453.87	-488.70	-594.68	-697.23	-747.50	-806.63
OCT	-82.85	-199.38	-332.82	-416.96	-456.87	-487.32	-592.77	-696.09	-746.62	-805.84
NOV	-79.86	-196.58	-329.85	-414.92	-455.96	-487.32	-592.77	-696.09	-746.62	-805.84
DEC	-79.62	-197.08	-328.74	-413.41	-453.96	-485.27	-590.23	-694.29	-745.26	-805.18
AVERAGES:	-47.93	-135.07	-259.08	-367.87	-432.99	-476.79	-537.11	-642.95	-713.82	-776.72

MESILLA VALLEY, N. MEX. - SIMULATION RESULTS
 AVE RIVER OUTFLOW (AC-FT/MO)

		CASE(2)1 QP=01 CHANGES FROM BASE PERIOD											
		1967	1968	1969	1970	1971	1972	1973	1974	1975	1976		
JAN	0.00	4622.51	7527.76	5968.46	7423.66	5770.51	4364.97	5495.96	9076.01	11957.12	10575.47		
FEB	0.00	3937.34	5968.46	7935.53	8792.94	7426.44	3305.97	4874.32	9805.96	9066.56	10811.69		
MAR	0.00	5146.12	7325.53	7297.83	8792.94	7426.44	-1662.40	2420.04	9805.48	6954.13	12819.96		
APR	0.00	-4225.83	-2797.83	-2403.97	2403.97	-6128.53	921.52	-2158.44	7103.32	835.37	14819.80		
MAY	-22363.23	-11477.29	-7472.12	-10554.71	-16603.19	-23301.92	3176.24	-12457.81	-7433.18	-4607.53	23899.46		
JUN	-24680.88	-31068.92	-13114.57	-16418.25	-23301.92	-2087.99	2087.99	16368.60	14294.79	-13511.69	13825.57		
JUL	-26454.44	-2678.12	26582.67	-11119.95	-19418.69	-4229.96	3761.31	14527.90	37613.62	481.05	1708.50		
AUG	7552.55	25023.22	31377.88	30174.89	165018.92	6855.05	32562.83	29983.40	37776.75	32802.78	35466.76		
SEP	11025.20	25310.40	25989.23	22483.52	11873.23	6577.75	25768.73	25768.73	25666.13	33472.56	33703.72		
OCT	8913.07	2310.40	25989.23	22483.52	11873.23	6577.75	25768.73	25768.73	25666.13	33472.56	33703.72		
NOV	5931.76	13876.66	16343.32	15523.43	18097.53	6702.97	17395.84	17395.84	22173.32	17759.00	15480.43		
DEC	6169.51	10436.68	13345.03	10904.30	6167.30	5858.74	11986.54	11986.54	16772.80	12837.54	11515.11		
AVERAGES:	-4238.31	5092.78	12166.74	8147.86	1607.74	3143.27	9502.56	14946.79	11202.68	13278.87			

MESILLA VALLEY, N. MEX. - SIMULATION RESULTS
 AVE RIVER CHEM (MG/L)

		CASE(2)1 QP=01 CHANGES FROM BASE PERIOD											
		1967	1968	1969	1970	1971	1972	1973	1974	1975	1976		
JAN	0.00	12.02	24.87	-107.05	-227.39	-312.34	-397.14	-388.07	-505.59	-297.21	-167.82		
FEB	0.00	24.87	71.62	-129.10	-188.39	-288.47	-420.42	-410.24	-516.21	-534.41	-432.10		
MAR	0.00	71.62	393.81	369.74	40.10	122.34	185.42	32.45	65.26	-71.12	-37.31		
APR	0.00	393.81	722.53	369.74	148.67	142.34	464.90	123.10	273.77	5.42	-56.28		
MAY	888.94	653.88	663.94	383.34	342.10	342.10	358.83	323.60	407.17	38.41	-4.45		
JUN	841.21	653.88	663.94	383.34	342.10	342.10	358.83	323.60	407.17	38.41	-4.45		
JUL	769.78	567.38	637.60	477.65	354.98	393.16	277.84	500.39	448.56	171.09	123.08		
AUG	100.15	286.42	170.31	347.65	354.98	393.16	277.84	267.71	230.60	123.06	133.06		
SEP	165.81	191.73	-495.82	46.54	12.62	12.62	266.89	138.05	98.75	49.81	23.06		
OCT	151.78	159.06	-267.37	-144.64	-219.34	-105.89	-59.79	83.72	-70.28	-25.16	-23.41		
NOV	111.31	-83.66	-267.37	-370.59	-306.63	-105.89	-213.71	-447.87	-70.28	-624.23	-717.99		
DEC	23.47	-75.63	-125.92	-309.40	-308.92	-308.92	-426.24	-509.87	-635.27	-583.79	-690.56		
AVERAGES:	394.07	233.16	94.67	-11.31	-31.49	13.41	-66.58	-65.00	-213.18	-225.74			

MESILLA VALLEY, N.MEX.-SIMULATION RESULTS:
DRAIN FLOW (AC-FT/MO)

CASE(2): QP=0: CHANGES FROM BASE PERIOD

	1967	1968	1969	1970	1971	1972	1973	1974	1975	1976
JAN	0.00	5050.31	8025.36	9637.35	8230.29	4644.61	5495.95	9281.61	12752.52	10575.47
FEB	-33.65	4067.54	6117.05	7423.66	6059.91	3439.36	4888.69	6994.56	9987.21	10811.69
MAR	-2887.60	5146.12	7935.53	8790.94	7426.47	4857.85	6363.69	9805.48	9953.23	12819.96
APR	-1948.81	8732.72	12717.92	12722.72	10032.42	8200.11	9655.06	17070.52	12624.02	14419.80
MAY	-16.80	11252.10	16141.79	15038.24	11076.81	7516.22	14094.74	21784.70	15834.08	17152.83
JUN	3378.91	15488.61	21729.68	19097.98	12564.08	5215.55	214026.45	28381.39	20438.08	21552.83
JUL	116.80	21511.28	26587.67	24274.40	16304.19	6951.11	27983.40	34510.84	27559.86	28895.45
AUG	7552.55	25023.22	31377.88	29287.89	20304.92	8725.75	32562.83	37767.75	32802.78	35466.76
SEP	11025.20	26310.57	33691.55	30174.89	19307.92	9725.75	32562.83	37767.75	32802.78	33703.72
OCT	11171.74	21683.00	28594.40	24648.38	14303.23	7933.32	27230.91	32076.53	27469.40	25145.83
NOV	8226.43	14884.26	18477.34	16656.59	19417.53	6983.38	18245.27	22076.53	18775.27	17086.71
DEC	6419.51	10837.08	13345.03	11483.70	6434.30	6106.54	12635.14	16931.20	13134.94	11924.31
AVERAGES:	3467.24	14161.40	18625.52	17511.35	11789.52	6264.60	17467.92	22968.00	19567.03	19963.47

MESILLA VALLEY, N.MEX.-SIMULATION RESULTS:
NET RECHARGE (FT)

CASE(2): QP=0: CHANGES FROM BASE PERIOD

	1967	1968	1969	1970	1971	1972	1973	1974	1975	1976
JAN	-0.0033	0.0020	0.0023	0.0035	0.0026	0.0004	0.0000	0.0009	0.0072	0.0778
FEB	-0.0019	0.0020	0.0020	0.0241	0.0013	-0.0003	0.0001	0.0011	0.0167	0.1616
MAR	-0.2278	0.1427	0.2349	0.1845	0.2046	0.4142	0.1984	0.4886	0.1296	0.1415
APR	-0.1046	0.2644	0.1916	0.2283	0.1227	0.1032	0.2253	0.1694	0.2379	0.1353
MAY	0.2084	0.1555	0.2655	0.2048	0.1353	-0.0308	0.3296	0.2864	0.1922	0.2796
JUN	0.1596	0.3521	0.3532	0.2485	0.1694	-0.0971	0.5336	0.4581	0.3752	0.5196
JUL	0.3046	0.3525	0.2257	0.3869	0.3220	0.1133	0.1689	0.4159	0.4686	0.6310
AUG	0.2212	0.2522	0.4858	0.3529	0.0816	0.2792	0.4001	0.3860	0.3716	0.8345
SEP	0.2661	0.2350	0.2530	0.2682	0.0098	-0.0696	0.3046	0.3200	0.2939	0.0357
OCT	-0.0192	-0.0296	0.0030	-0.0063	-0.0048	-0.0229	-0.0183	-0.0306	-0.0100	-0.0085
NOV	0.0114	0.0011	0.0061	0.0027	-0.0148	0.0001	-0.0282	-0.0160	-0.0170	0.0015
DEC	0.0012	0.0018	0.0000	0.0027	0.0012	0.0011	0.0030	0.0007	0.0014	0.0019
AVERAGES:	0.0895	0.1408	0.1689	0.1570	0.1071	0.0576	0.1764	0.2118	0.1723	0.1826

MESILLA VALLEY, N.MEX.-SIMULATION RESULTS
 AVE AQ W-L (FT)

CASE(3): IMPR IRR EFF; CHANGES FROM BASE PERIOD

	1967	1968	1969	1970	1971	1972	1973	1974	1975	1976
JAN	0.00	-2.22	-3.22	-2.58	-2.19	-3.50	-6.92	-7.34	-7.87	-6.80
FEB	-0.01	-2.09	-3.05	-2.32	-1.96	-3.39	-5.92	-7.17	-7.53	-6.51
MAR	-0.16	-2.02	-3.02	-2.18	-1.91	-3.43	-5.84	-7.18	-7.32	-6.38
APR	-0.49	-2.12	-3.17	-2.44	-2.06	-3.83	-5.96	-7.41	-7.28	-6.35
MAY	-0.90	-2.48	-3.44	-2.91	-2.34	-4.31	-6.27	-7.80	-7.35	-6.41
JUN	-1.47	-3.18	-3.90	-3.47	-2.95	-4.97	-6.84	-8.47	-7.65	-6.68
JUL	-2.14	-3.86	-4.25	-3.58	-3.65	-5.78	-7.38	-9.08	-8.09	-7.04
AUG	-2.51	-4.03	-4.17	-3.58	-3.97	-6.31	-7.95	-9.19	-8.17	-7.25
SEP	-2.55	-3.89	-3.86	-3.31	-3.99	-6.45	-7.75	-9.01	-8.15	-7.20
OCT	-2.54	-3.77	-3.52	-3.03	-3.90	-6.47	-7.75	-8.76	-7.69	-6.96
NOV	-2.49	-3.61	-3.20	-2.76	-3.78	-6.40	-7.68	-8.51	-7.42	-6.65
DEC	-2.36	-3.40	-2.89	-2.47	-3.64	-6.29	-7.52	-8.21	-7.11	-6.35
AVERAGES:	-1.47	-3.06	-3.47	-2.77	-3.03	-5.10	-6.89	-8.18	-7.62	-6.71

MESILLA VALLEY, N.MEX.-SIMULATION RESULTS:
 AVE AQ CHEM. (MG/L)

CASE(3): IMPR IRR EFF; CHANGES FROM BASE PERIOD

	1967	1968	1969	1970	1971	1972	1973	1974	1975	1976
JAN	0.00	44.22	101.60	156.51	225.01	303.22	394.92	480.79	555.06	623.63
FEB	0.00	44.37	101.77	156.76	225.23	303.51	394.91	480.78	554.88	623.99
MAR	0.00	44.40	101.94	156.80	225.33	303.60	394.90	480.82	555.56	624.88
APR	0.00	45.83	105.40	158.33	229.50	310.09	401.29	489.56	556.57	625.00
MAY	0.00	51.92	112.88	165.61	236.02	319.19	409.87	493.03	563.99	626.58
JUN	8.91	62.41	123.69	175.68	247.07	337.82	425.04	506.40	588.01	637.42
JUL	22.38	80.54	139.01	193.31	268.93	362.40	450.42	532.86	569.01	651.19
AUG	33.67	91.39	145.53	209.78	285.68	385.64	456.06	532.86	608.62	684.76
SEP	36.77	93.29	149.86	215.38	293.46	392.24	464.54	546.68	610.69	697.23
OCT	39.71	97.58	154.35	222.33	297.07	392.64	471.34	550.78	615.59	689.06
NOV	43.29	101.16	155.66	225.02	300.69	392.55	476.69	552.37	620.83	693.88
DEC	44.14	101.45	156.69	224.83	303.11	394.35	480.49	555.14	623.65	695.37
AVERAGES:	28.61	71.55	129.03	188.03	261.42	349.74	435.04	517.96	585.21	656.16

MESILLA VALLEY, N. MEX. - SIMULATION RESULTS
 AVE RIVER OUTFLOW (AC-FT/MO)

CASE(3)1 IMPR_IRR EFF; CHANGES FROM BASE PERIOD

	1967	1968	1969	1970	1971	1972	1973	1974	1975	1976
JAN	0.00	-3389.79	-4358.53	-6770.77	-6068.08	-3055.24	-4384.03	-4236.58	-7984.24	-7023.94
FEB	0.00	-3000.85	-3193.70	-5824.52	-5114.28	-2550.87	-4914.21	-3817.27	-7778.62	-6426.10
MAR	0.00	-247.86	2320.99	-813.78	2808.03	7013.14	1053.36	4115.03	-2400.48	-107.76
APR	0.00	5168.73	499.59	2896.76	4035.69	9104.89	4928.55	7064.99	395.27	-1042.19
MAY	10885.26	11848.68	7562.82	7395.67	9300.88	4357.63	9478.54	11035.39	3280.20	3672.61
JUN	16143.63	21321.37	14253.77	14876.45	19708.71	6209.79	17461.80	20834.69	10560.53	9066.40
JUL	15534.03	11023.56	2186.87	11147.11	13598.18	20894.92	8451.09	8075.69	9224.70	8068.53
AUG	2226.14	-3278.15	-5964.63	-5589.11	1520.72	4221.67	4015.27	-2811.69	-4024.51	2127.62
SEP	-432.85	-3404.22	-8457.09	-7178.58	-987.72	2527.83	778.88	-5840.93	-6600.70	-4694.97
OCT	-1948.03	-3511.45	-7684.59	-6803.56	-4906.03	-2767.53	-1486.41	-7804.58	-5826.47	-7128.69
NOV	-2783.91	-5120.65	-7729.78	-7705.30	-4737.46	-3314.20	-537.68	-8115.08	-8759.80	-9387.85
DEC	-3723.63	-4928.14	-7826.81	-7314.81	-3780.23	-3845.49	-4747.24	-8242.23	-9727.96	-9081.02
AVERAGES:	4487.57	1873.02	-1216.26	-973.71	2114.83	3244.79	2104.99	854.83	-2320.91	-1579.78

MESILLA VALLEY, N. MEX. - SIMULATION RESULTS
 AVE RIVER CHEM (MG/L)

CASE(3)1 IMPR_IRR EFF; CHANGES FROM BASE PERIOD

	1967	1968	1969	1970	1971	1972	1973	1974	1975	1976
JAN	0.00	-156.49	-262.63	-111.14	-11.65	-104.15	-197.86	-487.61	-779.32	-527.96
FEB	0.00	-152.42	-174.41	-111.01	-19.23	21.44	-573.97	-573.33	-1015.39	-730.07
MAR	0.00	-171.65	-136.69	-145.40	-72.36	-120.20	-168.25	-100.79	-334.41	-229.41
APR	0.00	-158.30	-208.09	-220.64	-182.29	-216.84	-168.88	-225.56	-311.11	-248.56
MAY	-128.68	-146.53	-243.05	-235.26	-182.29	-331.02	-150.27	-265.56	-292.65	-216.83
JUN	-129.19	-194.29	-198.90	-263.12	-193.55	-318.87	-116.15	-212.34	-304.55	-291.96
JUL	-98.32	-168.65	-253.69	-268.22	-192.33	-201.72	-171.06	-230.86	-361.55	-347.29
AUG	-88.03	-291.82	-318.65	-342.57	-232.23	-232.85	-171.53	-230.26	-341.58	-219.55
SEP	-161.20	-400.94	-467.36	-369.81	-554.40	-483.10	-332.58	-457.26	-404.38	-581.95
OCT	-415.05	-713.88	-237.08	-243.33	-633.68	-543.88	-325.04	-649.57	-1164.39	-877.80
NOV	-278.97	-376.67	-136.89	-71.14	-231.84	-532.68	-471.18	-722.78	-1038.34	-652.12
DEC	-138.91	-320.87	-182.87	-63.34	-229.52	-640.22	-404.41	-497.06	-991.70	-544.77
AVERAGES:	-179.31	-254.37	-235.03	-203.75	-225.64	-317.01	-338.77	-392.06	-612.48	-455.69

MESILLA VALLEY, N.MEX.-SIMULATION RESULTS:
DRAIN FLOW (AC-FT/MO)

CASE(3): IMPR IRR EFF; CHANGES FROM BASE PERIOD

	1967	1968	1969	1970	1971	1972	1973	1974	1975	1976
JAN	0.00	3389.79	4358.53	6770.77	6068.08	3055.24	4384.03	4236.58	8345.44	1976.47
FEB	-22.48	-3103.25	-3393.70	-6792.22	-5118.28	-2500.87	-4914.21	-3817.27	-8428.15	-7259.63
MAR	-475.70	-3518.66	-4342.21	-7468.31	-5395.23	-1716.46	-5499.58	-3065.63	-9434.41	-7663.49
APR	-1392.17	-3838.67	-5757.27	-8480.47	-5905.88	-1617.91	-5259.58	-3813.67	-9807.39	-9483.93
MAY	-2058.21	-3700.52	-6329.44	-10433.32	-6443.32	-1676.37	-4985.05	-5688.41	-10713.34	-10729.12
JUN	-2688.11	-4330.63	-7659.36	-13820.68	-7380.83	-1697.21	-4297.05	-7042.71	-11940.87	-12580.93
JUL	-3330.04	-5775.78	-11258.20	-13808.75	-8128.82	-1714.52	-4973.97	-7800.18	-11874.03	-11322.67
AUG	-4316.79	-7778.82	-14409.63	-15384.38	-7993.81	-1670.59	-8328.29	-8388.90	-11983.05	-8523.85
SEP	-4841.19	-7875.69	-13378.56	-13442.31	-7183.66	-2055.11	-6636.72	-8328.90	-11356.23	-9630.10
OCT	-4469.03	-8491.45	-10770.39	-10916.56	-6028.03	-2807.53	-6166.41	-7962.58	-19846.47	-10147.69
NOV	-4093.97	-5634.71	-9129.78	-10916.56	-4737.46	-3314.20	-5376.68	-8115.08	-8768.60	-8387.84
DEC	-3723.63	-4928.14	-7826.81	-7314.81	-3780.23	-3845.49	-4747.24	-8242.23	-7927.96	-7081.02
AVERAGES:	-2617.53	-5030.51	-8259.49	-9687.46	-6179.64	-2305.97	-5306.12	-6375.13	-10035.50	-9127.48

MESILLA VALLEY, N.MEX.-SIMULATION RESULTS:
NET RECHARGE (ft)

CASE(3): IMPR IRR EFF; CHANGES FROM BASE PERIOD

	1967	1968	1969	1970	1971	1972	1973	1974	1975	1976
JAN	-0.0022	-0.0013	-0.0015	-0.0024	-0.0017	-0.0015	0.0000	-0.0006	-0.0033	-0.0022
FEB	-0.0013	-0.0019	-0.0014	-0.0009	-0.0009	-0.0010	-0.0000	-0.0007	-0.0060	-0.0027
MAR	-0.0693	-0.0301	-0.0613	-0.0523	-0.0755	-0.0804	-0.0603	-0.0661	-0.0648	-0.0696
APR	-0.0916	-0.0829	-0.0960	-0.0954	-0.0915	-0.0987	-0.0938	-0.1001	-0.0939	-0.0777
MAY	-0.1193	-0.1431	-0.1279	-0.1460	-0.1499	-0.1395	-0.1332	-0.1540	-0.1288	-0.1326
JUN	-0.1734	-0.2361	-0.2017	-0.2366	-0.2494	-0.1788	-0.2003	-0.2566	-0.2071	-0.1993
JUL	-0.0602	-0.1547	-0.1238	-0.2307	-0.2000	-0.2081	-0.1236	-0.1461	-0.1942	-0.1785
AUG	-0.0406	-0.0412	-0.0777	-0.0911	-0.0876	-0.0542	-0.0683	-0.0513	-0.0733	-0.0981
SEP	-0.0421	-0.0400	-0.0453	-0.0601	-0.0570	-0.0422	-0.0683	-0.0229	-0.0438	-0.0454
OCT	-0.0145	-0.0047	-0.0284	-0.0432	-0.0283	-0.0085	-0.0499	-0.0199	-0.0376	-0.0302
NOV	-0.0047	-0.0017	-0.0174	-0.0212	-0.0180	-0.0170	-0.0240	-0.0181	-0.0176	-0.0088
DEC	-0.0008	-0.0012	-0.0000	-0.0018	-0.0008	-0.0008	-0.0020	-0.0005	-0.0009	-0.0013
AVERAGES:	-0.0986	-0.0648	-0.0655	-0.0818	-0.0796	-0.0692	-0.0710	-0.0698	-0.0726	-0.0705

MESILLA VALLEY, N.MEX.-SIMULATION RESULTS
AVE AQ W-L (FT)

CASE(4): QL=0; CHANGES FROM BASE PERIOD

	1967	1968	1969	1970	1971	1972	1973	1974	1975	1976
JAN	0.00	-0.73	-1.08	-0.30	-0.82	-2.65	-2.47	-1.13	0.24	-0.01
FEB	0.01	-0.67	-0.99	-0.24	-0.71	-2.56	-2.31	-1.05	0.09	0.30
MAR	0.08	-0.65	-0.88	-0.23	-0.65	-2.01	-2.05	-0.39	0.09	0.03
APR	0.16	-0.54	-0.92	-0.30	-0.86	-1.53	-1.85	-0.02	-0.37	0.03
MAY	0.01	-0.70	-1.13	-0.71	-1.51	-1.75	-1.85	-0.38	-0.34	-0.24
JUN	-0.46	-1.48	-1.57	-1.62	-2.51	-2.28	-1.89	-0.91	-0.94	-0.91
JUL	-1.03	-2.27	-1.98	-2.44	-3.36	-2.79	-2.13	-0.92	-1.16	-1.00
AUG	-1.31	-2.38	-1.55	-2.35	-3.35	-2.60	-2.05	-0.40	-0.16	-0.20
SEP	-1.02	-1.76	-0.72	-1.54	-2.94	-2.48	-1.39	0.39	-0.15	-0.07
OCT	-0.77	-1.34	-0.36	-1.12	-2.94	-2.69	-1.16	0.65	0.08	-0.35
NOV	-0.84	-1.29	-0.38	-1.06	-2.88	-2.67	-1.23	0.42	-0.06	-0.38
DEC	-0.80	-1.18	-0.35	-0.95	-2.78	-2.61	-1.21	0.23	-0.11	-0.36
AVERAGES:	-0.50	-1.25	-0.99	-1.07	-2.12	-2.39	-1.80	-0.27	-0.25	-0.24

MESILLA VALLEY, N.MEX.-SIMULATION RESULTS:
AVE AQ CHEM. (MG/L)

CASE(4): QL=0; CHANGES FROM BASE PERIOD

	1967	1968	1969	1970	1971	1972	1973	1974	1975	1976
JAN	0.00	5.34	6.49	-23.53	-2.38	63.39	66.19	29.89	39.56	-57.26
FEB	0.00	5.26	6.75	-23.13	-2.10	63.62	66.19	29.89	-39.33	-62.51
MAR	0.00	5.59	6.98	-25.36	-1.95	63.78	66.19	30.11	-39.61	-74.36
APR	0.00	0.40	-2.53	-28.90	-8.13	34.44	57.24	-8.48	-39.17	-73.27
MAY	0.00	-8.83	1.57	-31.32	2.45	34.93	57.24	-2.32	-41.37	-71.83
JUN	3.69	-2.67	5.38	-15.34	23.65	45.56	55.44	3.28	-29.30	-69.83
JUL	17.57	23.84	19.68	9.15	53.91	63.74	55.44	10.19	-20.39	-47.08
AUG	13.86	34.32	18.23	25.40	68.71	70.58	55.44	11.54	-24.39	-68.92
SEP	19.57	23.42	17.64	25.08	56.13	50.06	38.83	1.54	-43.89	-92.85
OCT	-1.07	3.42	-7.64	10.08	55.91	61.71	20.65	-22.80	-63.64	-85.09
NOV	3.91	1.83	-26.25	-8.55	60.09	63.07	25.36	-45.89	-60.24	-82.50
DEC	5.22	6.28	-23.54	-2.66	63.24	66.08	25.36	-39.65	-57.42	-81.10
AVERAGES:	9.10	8.89	-1.66	-9.93	30.79	56.75	49.47	-4.67	-41.52	-72.22

MESILLA VALLEY, N. MEX. - SIMULATION RESULTS
 AVE RIVER OUTFLOW (AC-FT/MO)

CASE(4): Q1=01 CHANGES FROM BASE PERIOD

	1967	1968	1969	1970	1971	1972	1973	1974	1975	1976
JAN	0.00	-1589.02	-2358.92	-1233.43	-2972.35	-2782.84	-3504.42	-2102.51	1261.30	3925.17
FEB	0.00	-1345.55	-2035.64	330.74	-2406.40	-2268.55	-3804.06	-1815.44	1681.37	9089.34
MAR	0.00	15776.50	22258.09	21849.61	21423.77	14599.86	887.59	23473.22	15229.47	24117.31
APR	0.00	10095.50	11454.82	15584.13	8053.80	5210.82	11750.12	18598.29	15843.79	20883.25
MAY	8777.41	7310.05	9505.60	13947.28	6835.30	2547.58	10362.59	18086.55	18171.27	22210.03
JUN	8095.50	15727.25	20397.01	16015.04	9908.79	2385.16	19539.09	24921.17	18778.90	19858.52
JUL	14533.22	16738.30	26635.17	21109.04	14007.65	11031.54	19035.18	22912.82	22778.72	29987.54
AUG	15119.56	18832.76	30738.07	19677.89	15050.31	11061.54	26428.96	16350.67	24816.72	29916.28
SEP	13593.40	19849.17	10016.41	9551.73	3672.00	1461.02	19234.69	16350.67	24816.72	17266.46
OCT	884.88	2449.61	-64.85	-4588.23	-4989.76	-2327.70	-835.06	2963.25	2195.56	189.89
NOV	-1418.89	-2861.61	-1144.19	-4055.38	-4339.08	-2752.70	-248.59	1833.34	-267.20	-1484.45
DEC	-1804.93	-2769.45	-1531.13	-3737.08	-3460.01	-3154.05	-2458.06	1260.60	-438.26	-1236.45
AVERAGES:	7001.30	6942.87	10322.20	8790.11	5065.34	2810.94	8485.59	12156.30	11086.29	13603.34

MESILLA VALLEY, N. MEX. - SIMULATION RESULTS
 AVE RIVER CHEM (MG/L)

CASE(4): Q1=01 CHANGES FROM BASE PERIOD

	1967	1968	1969	1970	1971	1972	1973	1974	1975	1976
JAN	0.00	-46.19	14.42	32.74	-45.36	-130.47	-248.27	-40.47	12.99	-117.66
FEB	0.00	-51.83	-40.62	-56.34	-45.91	-61.92	-154.17	-31.32	-97.83	-330.22
MAR	0.00	-49.73	-100.93	-88.92	-119.29	-111.84	-129.86	-52.05	-102.24	-65.73
APR	0.00	-100.19	-133.99	-129.59	-148.29	-173.29	-127.75	-101.80	-196.41	-85.40
MAY	-55.80	-83.32	-160.87	-152.34	-172.53	-255.27	-111.94	-136.69	-129.66	-84.67
JUN	-70.23	-71.70	-211.53	-213.34	-186.89	-359.33	-90.42	-125.89	-169.39	-162.48
JUL	-74.09	-149.17	-211.53	-246.12	-205.89	-287.29	-140.32	-133.29	-234.49	-213.13
AUG	-72.78	-273.08	-253.38	-331.37	-501.30	-245.67	-107.33	-138.43	-198.24	-104.51
SEP	-132.54	-341.46	-287.57	-337.55	-580.82	-434.61	-249.38	-130.43	-167.12	-304.02
OCT	-162.86	-266.27	-82.05	-113.53	-516.69	-333.28	-248.40	-170.60	-210.70	-194.38
NOV	-102.86	-133.62	-45.75	-156.43	-233.96	-138.31	-45.39	-19.35	-57.64	-87.32
DEC	-48.60	-95.17	-37.47	-55.22	-234.61	-136.20	-39.88	-29.52	-57.90	-87.32
AVERAGES:	-89.89	-138.48	-130.91	-151.18	-233.51	-209.79	-141.08	-73.06	-126.30	-153.48

MESILLA VALLEY, N. MEX. - SIMULATION RESULTS:
DRAIN FLOW (AC-FT/MO)

CASE(4): QLE0: CHANGES FROM BASE PERIOD

	1967	1968	1969	1970	1971	1972	1973	1974	1975	1976
JAN	0.00	-1589.02	-2358.92	-1233.43	-2972.35	-2782.84	-3504.42	-2102.51	1025.50	-36.18
FEB	-32.29	-1408.55	-2035.64	-913.51	-2406.40	-2268.55	-3804.06	-1815.44	840.32	990.59
MAR	253.65	-1563.25	-2048.21	-1000.54	-2352.88	-1392.59	-3980.26	-651.73	415.88	1244.52
APR	522.40	-1398.91	-2687.33	-1416.42	-3125.55	-1109.43	-3593.98	51.09	98.18	1127.00
MAY	18.60	-1575.90	-3348.20	-3296.31	-4900.70	-1212.72	-3411.16	-1036.81	-1723.61	-1145.42
JUN	-1072.35	-2944.15	-4880.84	-7392.57	-6837.51	-1368.94	-4611.24	-2504.81	-4292.73	-4329.48
JUL	-2941.58	-4705.10	-7973.18	-11432.71	-7897.00	-1473.35	-3641.11	-3033.48	-5082.40	-4365.01
AUG	-2715.05	-6315.04	-7973.18	-8212.02	-7572.99	-1391.46	-3526.55	-1541.08	-3765.03	-822.42
SEP	-2018.88	-3790.65	-3954.74	-5466.08	-6556.05	-1672.33	-3651.41	-1713.97	-759.99	-326.54
OCT	-2008.39	-3299.01	-1774.19	-4610.33	-4339.08	-2345.70	-2981.59	2893.15	-386.34	-1548.45
NOV	-1804.93	-2769.45	-1531.13	-3737.08	-3460.01	-2752.70	-2781.59	1233.34	-267.20	-1484.45
DEC						-3154.05	-2458.06	1260.60	-438.26	-1236.45
AVERAGES:	-1156.49	-3062.02	-3518.11	-5065.69	-4826.19	-1910.39	-3442.08	-416.17	-1147.11	-1077.72

MESILLA VALLEY, N. MEX. - SIMULATION RESULTS:
NET RECHARGE (FT)

CASE(4): QLE0: CHANGES FROM BASE PERIOD

	1967	1968	1969	1970	1971	1972	1973	1974	1975	1976
JAN	-0.0034	-0.0020	-0.0023	-0.0035	-0.0026	-0.0022	0.0000	-0.0009	-0.0023	0.0413
FEB	-0.0019	-0.0007	-0.0021	-0.0126	-0.0013	-0.0015	-0.0001	-0.0011	0.0005	0.0871
MAR	-0.0034	-0.0170	-0.0151	-0.0293	-0.0140	0.2063	0.0437	0.2665	-0.0344	-0.0508
APR	-0.0015	0.0393	-0.0765	-0.0232	-0.1290	-0.0247	-0.0247	-0.0931	-0.0123	-0.0558
MAY	-0.0601	-0.1356	-0.0702	-0.1979	-0.2275	-0.0961	-0.0414	-0.0332	-0.1583	-0.0710
JUN	-0.1531	-0.2484	-0.2002	-0.2940	-0.3149	-0.1586	-0.0214	-0.1239	-0.1441	-0.2687
JUL	-0.1829	-0.1676	-0.0911	-0.2385	-0.1920	-0.0870	-0.1553	-0.0233	-0.0372	0.1474
AUG	-0.0451	0.0207	0.1295	0.0605	-0.0524	0.1407	0.1152	0.1609	0.1084	0.1524
SEP	-0.1160	0.1428	0.1244	0.1047	-0.0384	-0.1157	0.0940	0.1853	0.1353	-0.1062
OCT	-0.0504	-0.0449	-0.0226	-0.0437	-0.0368	-0.0126	-0.0512	-0.0291	-0.0360	-0.0300
NOV	-0.0151	-0.0022	-0.0191	-0.0257	-0.0269	-0.0255	-0.0361	-0.0271	-0.0264	-0.0132
DEC	-0.0012	-0.0018	0.0000	-0.0027	-0.0012	-0.0011	-0.0030	-0.0007	-0.0014	-0.0019
AVERAGES:	-0.0369	-0.0348	-0.0179	-0.0567	-0.0777	-0.0146	-0.0067	0.0222	-0.0174	-0.0141

This dissertation is accepted on behalf of the faculty of the

Institute by the following committee:

Lyman W. Heller
Adviser

David B. Stephens

Raz Khaleel

10/16/81
Date